Differential contribution of nitrifying prokaryotes

to nitrification in groundwater

Markus Krüger[1], Narendrakumar Chaudhari[1,2], Bo Thamdrup[3], Will Overholt[1], Laura Bristow[3],

Martin Taubert[1], Kirsten Küsel[1,2], Nico Jehmlich[4], Martin von Bergen[2,4,5], Martina Herrmann[1,2*]

[1]Aquatic Geomicrobiology, Institute of Biodiversity, Friedrich Schiller University, Jena, Germany.

[2]German Center for Integrative Biodiversity Research (iDiv) Halle-Jena-Leipzig, Leipzig, Germany.

[3]Nordcee, Department of Biology, University of Southern Denmark, Odense, Denmark. [4]Department of

Molecular Systems Biology, Helmholtz Centre for Environmental Research – UFZ, Leipzig, Germany.

[5]Faculty of Biosciences, Pharmacy and Psychology, Institute of Biochemistry, University of Leipzig,

Germany.

* Corresponding author: martina.herrmann@uni-jena.de

Abstract

The ecophysiology of complete ammonia-oxidizing *Nitrospira* (CMX) and their widespread occurrence in groundwater suggests that CMX bacteria have a competitive advantage over ammonia-oxidizing bacteria (AOB) and archaea (AOA) in these environments. However, the specific contribution of their activity to nitrification processes has remained unclear. We aimed to disentangle the contribution of CMX, AOA and AOB to nitrification and to identify the environmental drivers of their niche differentiation at different levels of ammonium and oxygen in oligotrophic carbonate rock aquifers. CMX accounted for up to 95% of the ammonia oxidizer communities. Nitrification rates were positively correlated to CMX clade A associated phylotypes and AOB affiliated with *Nitrosomonas ureae*. Surprisingly, short-term incubations amended with the nitrification inhibitors allylthiourea and chlorate suggested that AOB contributed more than 90% to overall ammonia oxidation, while metaproteomics analysis confirmed an active role of CMX in both ammonia and nitrite oxidation. Ecophysiological niche differentiation of CMX clades A and B, AOB and AOA was linked to their requirements for ammonium, oxygen tolerance, and metabolic versatility. Our results demonstrate that despite numerical predominance of CMX, the first step of nitrification in oligotrophic groundwater is primarily governed by AOB. Higher growth yields at lower ammonia turnover rates and energy derived from nitrite oxidation most likely enable CMX to maintain consistently high populations.

Introduction

Nitrification is a globally relevant process of the nitrogen cycle mediated by a polyphyletic

group of microorganisms [1]. The discovery of complete ammonia oxidizers (comammox

bacteria, CMX) challenged the paradigm of a strict division of labor in the process of

nitrification [2]. While ammonia oxidation is carried out by ammonia-oxidizing bacteria (AOB)

and archaea (AOA) [3] and nitrite oxidation by nitrite-oxidizing bacteria (NOB) [2], CMX of the

genus *Nitrospira* can perform complete oxidation of ammonia (NH_3) to nitrate (NO_3^-) in one

cell [4, 5]. CMX have been frequently reported from engineered environments such as drinking

water or wastewater treatment facilities [6–8] and from natural environments like soils [9, 10],

surface freshwater [11, 12] and groundwater [13, 14]. However, our knowledge about the role

and relevance of CMX for nitrification activity in natural environments is still scarce.

In most environments, CMX will compete with canonical nitrifiers for either ammonium (NH_4^+)

or nitrite (NO_2^-) or both, raising the question of the actual contribution of CMX to nitrification

processes and of the factors that control their co-occurrence with AOB, AOA, and NOB.

Recent investigations consider affinity to NH_4^+, growth rates, growth yields, and metabolic

versatility as key factors that determine the niche differentiation and competition between

CMX, AOB, and AOA [2, 15]. Since CMX exhibit higher affinity to NH_4^+ and oxygen than

terrestrial AOB [15, 16] along with higher growth yields compared to AOB and AOA [2], CMX

might have a competitive advantage over AOB in environments that are limited in NH_4^+ such

as oligotrophic groundwater. Particular lineages of AOB and AOA exhibit adaptation to NH_4^+

limited conditions [17–20] and have previously been reported from groundwater environments

[21–23]. As such, oligotrophic groundwater represents an excellent microbial habitat to study

the co-occurrence of CMX, AOB and AOA and their contribution to nitrification activity.

Based on the phylogeny of the ammonia mono-oxygenase (Amo), the key enzyme of ammonia

oxidation, CMX *Nitrospira* are divided into two clades, A and B [2, 4]. Metagenomic studies

uncovering the phylogeny and metabolic features of CMX proposed an ecophysiological niche

differentiation between both clades in natural habitats [4, 21, 24–26]. Those findings are linked

to genomic variability in nitrogen uptake and alternative energy metabolism [25, 26], and

several studies suggest a more prominent role of clade A in nitrification. So far, CMX

enrichments and isolates comprise only clade A representatives [27]. ^{13}C stable isotope

probing confirmed growth of CMX clade A with NH_4^+ during incubations of wetland coast

sediments [28], and CMX clade A were highly abundant in salt marsh sediments showing

strong potential for nitrification [29]. However, the ecophysiological niche differentiation

between the clades has rarely been studied in the context of their co-occurrence and

competition with canonical nitrifiers.

Previous studies from karstic carbonate rock aquifers of the Hainich Critical Zone Exploratory

in central Germany pointed to strong links between in situ nitrification and

chemolithoautotrophic carbon fixation [30, 31]. In addition, genome resolved metagenomics

revealed an important role of CMX *Nitrospira* in the nitrogen and carbon cycle in this

oligotrophic habitat [21]. The aquifers are characterized by a high spatial heterogeneity of NH_4^+

and NO_3^- concentrations across an oxygen gradient, linked to differences in groundwater flow

regime and distance to groundwater recharge areas [28, 29, 30]. These settings are ideal for

the investigation of the ecophysiological niche differentiation of CMX clade A and B, AOB and

AOA.

We combined ^{15}N-based nitrification rate measurements with selective nitrification inhibitors

and genome-resolved metagenomics and metaproteomics to assess the contributions of CMX

Nitrospira, AOB and AOA to overall groundwater nitrification in these aquifers and to follow

the composition of the groundwater nitrifier populations across the hydrochemical gradients.

We hypothesized that (i) CMX act as key ammonia oxidizers in these oligotrophic groundwater

habitats and account for a large fraction of overall nitrification activity. We further hypothesized

that (ii) distribution patterns of ammonia oxidizers across gradients of NH_4^+ and oxygen

availability are linked to their ecophysiological niche differentiation, indicated by differences in

encoded metabolic potentials.

Materials and methods

Study site characteristics and sampling of groundwater

Groundwater was obtained from karstic limestone aquifers at the Hainich Critical Zone

Exploratory (CZE) in Thuringia (Germany) within the framework of the Collaborative Research

Center (CRC) AquaDiva [34]. Further details about location and geological setting are reported

elsewhere [33, 34]. Sampling wells H14, H32, H42, H43, H52 and H53 at varying depths

provide access to oxygen-deficient (less than 15 µM DO) groundwater from an upper

mudstone-dominated low-flow aquifer (HTU) and wells H13, H31, H41 and H51 (mean values

154 to 672 µM DO) provide access to oxic groundwater from a lower fast-flow limestone-

dominated karstified aquifer (HTL) [32, 34] (Figure 1A). The main recharge zone for

groundwater is located at the hilltop of the transect near wells H13 and H14. Groundwater

collected from regular monthly sampling between 2018 to 2021 was obtained using

submersible motor pumps (Grundfos, Denmark). Hydrochemical parameters (pH, T, TOC,

DOC) were determined as described in [32, 33]. NH_4^+, NO_2^- and NO_3^- concentrations were

measured with colorimetric methods [35, 36] from groundwater filtered through 0.2 µm-pore

size sterile polyvinylidine fluoride filters. Groundwater for molecular analysis was collected in

sterile 10 L fluorinated polyethylene containers and filtered through 0.2 µm polycarbonate

filters (Nuclepore, Whatman). Additional groundwater samples for metaproteomic analysis

were collected from wells H14, H41, H43 and H52 in January 2019 and filtered through 0.2

µm pore size filters (Merck Millipore, Germany). All filters were immediately frozen on dry ice

and stored at -80 °C until extraction.

^{15}N-isotope labelling and rate measurements

Groundwater was collected in sterile 0.5 L glass bottles and filled from the bottom using sterile

tubing. Bottles were overfilled with three volumes exchanges before they were capped with

silicone septa without headspace to minimize alterations of in situ oxygen concentrations. At

each sampling campaign, groundwater was collected in triplicate bottles for each well along

with one control and kept at 4°C until further processing within 2 hours. Prior to the ^{15}N

labelling, 10 ml of each sampling bottle was removed for analysis of inorganic nitrogen

concentrations and pH, and replaced with N_2. Bottles containing anoxic groundwater were

amended with 1 % sterile oxygen (4 ml 100% oxygen) to stimulate nitrification under oxygen-

deficient conditions. Groundwater of control bottles was sterile filtered through 0.2 µm filters

(Supor, Pall Corporation, USA). Sterile filtered ^{15}N-labeled ammonium sulfate solution

(([15]NH_4)$_2SO_4$) (98 at%, Cambridge Isotope Laboratories, Tewksbury) as the substrate source

for nitrifiers was added to all 0.5 L bottles to a final concentration of 50 µM, at a natural

background of 0.5 to 55 µM NH_4^+. Nitrification rate measurements were conducted according

to Overholt *et al.* [21] and detailed descriptions are provided in Supplementary Methods.

Incubation experiments in the presence of nitrification inhibitors

The nitrification activity of CMX and NOB, and CMX and AOB was selectively inhibited by

either 1 mM potassium chlorate ($KClO_3$) or 100 µM allylthiourea (ATU), respectively [7, 37].

Nitrification rate measurements with ^{15}N tracer were carried out over five days as described

above. Groundwater from well H41 was collected in February 2020 and amended with 50 µM

NH_4^+ as ([15]NH_4)$_2SO_4$ and one of the nitrification inhibitors along with untreated controls.

Groundwater for mesocosm experiments was collected in March and May 2021 and was filled

in sterile 2 L glass bottles without headspace to assess potential growth of ammonia oxidizers

with NH_4^+ in the presence of nitrification inhibitors at the same concentrations as before.

Groundwater from oxic well H41 (10 µM NH_4^+ in situ) was adjusted to 10 and 50 µM NH_4^+, and

groundwater from hypoxic well H53 (36 µM NH_4^+ in situ) was used without additional

supplement. All samples were incubated for 15 to 25 days at 15°C without agitation in the

dark. Subsamples for nitrogen compound analysis were taken at the onset and every two to

three days until the end of the incubation to monitor changes of the nitrogen chemistry

(Supplementary Methods). Incubation of well H41 with an initially 10 µM NH_4^+ was adjusted

again to 15 µM NH_4^+ after 15 days. After incubation, the groundwater was filtered on 0.2 µm

polycarbonate filters and stored at -80°C for later extraction of nucleic acids.

Nucleic acid extraction, cDNA-synthesis and quantitative PCR

Genomic DNA and total RNA of microbial biomass from the 0.2 µm filters were extracted using the DNeasy PowerSoil kit (Qiagen, Germany) and the RNeasy PowerWater kit (Qiagen, Germany), respectively, according to the manufacturer's specifications. RNA was further processed with the TURBO DNA-free kit (Thermo Fischer Scientific, Germany) to remove remnant DNA in the samples before cDNA synthesis [38]. Reverse transcription was performed using the GoScript Reverse Transcription System (Promega, USA). Gene and transcript abundances of *amoA* from CMX *Nitrospira*, AOB, AOA and of *nxrB* from *Nitrospira*-like NOB were determined using a CFX96 qPCR cycler (Bio-Rad, USA). Primer pairs for quantification were comamoAF/ comamoAR for CMX *amoA* [11], amoA-1F/ amoA-2R for AOB *amoA* [39], Arch-amoAF/ Arch-amoAR for AOA *amoA* [40] and nxrB169f/ nxrB638r for *Nitrospira nxrB* [41]. More details are provided in the Supplementary Methods.

Amplicon sequencing of *amoA* genes and sequence analysis

Amplicon libraries of CMX, AOB and AOA *amoA* genes were created using the NEBNext® Ultra™ DNA Library Prep Kit for Illumina® (New England Biolabs, MA) according to the manufacturer's guideline. The same primers as for qPCR were used for CMX and AOB *amoA*, while primers Arch-amoA-104F/ Arch-amoA-616Rmod [42, 43] were chosen for sequencing of AOA *amoA* genes due to shorter fragment length (Supplementary Methods). Sequencing was conducted on an Illumina Miseq using v3 chemistry. After trimming primer sequences, raw reads were first quality filtered with DADA2 1.22.0 [44] and further processed with Mothur v.1.46.1 [45] (Supplemental Methods). OTUs were assigned at 95% sequence identity for CMX, 95% for AOB [46] and 96% for AOA [42].

Inferring metabolic potential from metagenome assembled genomes of CMX *Nitrospira* and other nitrifiers

To assess metabolic capabilities of the two CMX clades, genomic features of metagenome assembled genomes (MAGs) were investigated. Groundwater sampling for metagenomic sequencing and generation of MAGs were described in detail in [21, 47]. Here, Hainich

groundwater MAGs, which were taxonomically classified as *Nitrospira* on genus level and 243

Nitrospira reference genomes downloaded from NCBI (July 2021) were aligned and screened

for both *amoA* and *nxrB* genes with HMMER in Anvi'o v7.1 [21] using custom AmoA and NxrB

databases containing sequences from the FunGene repository [48].

After sorting good quality MAGs (> 90 % completeness, except *N. inopinata* with 42.25 %)

and dereplication, 35 CMX *Nitrospira* genomes were kept for final analysis. The affiliation of

those genomes to CMX clade A and clade B was achieved by alignment of protein sequences

of bacterial single copy core genes detected with default 'Bacteria_71' hmm profile using

MUSCLE [49] in Anvi'o and subsequent construction of a maximum likelihood phylogenetic

tree using FastTree with 1000 bootstrap iterations. The tree was then processed with iTOL for

visualization [50]. The genomic repertoire of representative groundwater nitrifier MAGs was

inferred by KEGG orthology (KO) previously generated using kofamscan [21] (Supplementary

Methods). Relative transcriptional activity of each investigated MAG was extracted from

Overholt *et al.* [21], who determined the transcriptomic coverages for each open reading frame

from each MAG from previously generated metatranscriptomic libraries of the groundwater

[30].

Metaproteomic analysis

Proteins were extracted from filters using a phenol-chloroform based protocol as previously

described [51]. Proteins were subjected to SDS polyacrylamide gel electrophoresis,

subsequently in-gel tryptic cleavage and LC-MS/MS analysis was performed as previously

described [52] (Supplemental Methods). Translated amino acid sequences from all genes of

the MAGs obtained from the metagenomic datasets [21, 47] were used as a reference

database for protein identification. Taxonomic and functional annotations of identified peptides

were transferred from the metagenomic dataset [21, 47]. Mass spectrometry proteomics data

have been deposited into the ProteomeXchange Consortium via the PRIDE [30] partner

repository.

Statistical analysis

All statistical analyses were conducted using R 4.12 (R Core Team, 2020) and packages FSA

[54], ComplexHeatmap [55] and vegan [56]. Analysis details are explained in the

Supplementary Methods. To analyze communities of CMX, AOB and AOA together, the

absolute *amoA* read abundances of OTUs > 2% of all reads for each ammonia oxidizer

community were determined as followed: The percentage of each OTU was calculated from

the relative abundance obtained by *amoA* amplicon sequencing and was multiplied by the

total *amoA* counts from corresponding samples measured with qPCR.

Results

Nitrification activity and distribution of nitrifiers across the different

groundwater wells

Nitrification rates in groundwater varied over two orders of magnitude ranging from 0.54 – 136 nmol N L^{-1} d^{-1} across the wells (Figure 1B). We observed the highest rates at oxic well H41 with 93 ± 24 nmol N L^{-1} d^{-1} and high activity at hypoxic well H52 with 58 ± 4.0 nmol N L^{-1} d^{-1}. Rates were positively correlated with NH_4^+ concentrations in the groundwater ($R^2 = 0.67$, $p <$ 0.05), which were on average below 40 µM and significantly higher in the hypoxic wells (Figure 1C). NO_3^- concentrations ranged from 1.7 to 38 µM in hypoxic and from 80 to 384 µM in oxic groundwater and showed no correlation with the rates across the transect.

CMX accounted for 9 to 95 % of the total groundwater ammonia oxidizer community. Proportions of CMX and AOB increased substantially at more distant wells (H41, H51, H52, H53) to maxima of 95% and 34%, and were positively correlated to NH_4^+ ($R^2 = 0.48$, $p < 0.01$; $R^2 = 0.44$, $p < 0.01$), respectively (Figure 1D, Figure 2D). In contrast, AOA dominated the ammonia-oxidizing community by up to 90% in wells with lower NH_4^+ concentrations located closer to the main recharge area (H13 and H14). Oxic well H41 and anoxic well H53 harbored the highest absolute *amoA* gene and transcript abundances of CMX and AOB (Figure 1D, Figure S 1A). Nitrification rates were strongly correlated with AOB *amoA* gene abundances ($R = 0.87$, $p < 0.001$), and AOB also had the highest *amoA* transcript/ gene ratios across all wells (Figure S 1B).

Across all wells, abundances of CMX *amoA* genes and *Nitrospira nxrB* genes encoding nitrite oxidoreductase were positively correlated to each other ($R = 0.56$, $p < 0.001$). Assuming one *amoA* and 1.5 *nxrB* copies per genome, CMX accounted for 18 to 44 % and 48 to 70 % of the *Nitrospira* population at oxic well H41 and hypoxic well H52, respectively (Figure S 1C), which exhibited the highest observed nitrification rates (Figure 1C). Furthermore, CMX and *Nitrosomonadaceae* showed the highest abundance on peptide level at wells H41 and H52

(Figure S 2). Out of the *Nitrosomonadaceae* affiliated peptides, 88.3% were assigned to

Nitrosomonas.

The ammonia oxidizer community composition showed formation of clusters depending on

local groundwater chemistry and the location in the transect (Figure 2A). The CMX community

was dominated by three OTUs affiliated to clade A and six OTUs affiliated to clade B, of which

clade A OTUs were closely related to *Ca.* Nitrospira nitrificans (Figure 2C). These clade A

OTUs increased in abundance at wells H41, H51, H52 and H53 and accounted for up to 50%

of the CMX community (Figure 2B). CMX clade A affiliated OTUs were also positively

correlated to the nitrification rates (R^2 = 0.72 – 0.88, p < 0.05), while only two low abundant

clade B affiliated OTUs correlated with the rates (Figure 2D). The AOB community was mainly

dominated by AOB OTU1 affiliated to *Nitrosomonas ureae*, which was the only AOB phylotype

showing a strong correlation with nitrification rates (R^2 = 0.88, p < 0.01). The most dominant

AOA OTU1 related to *Nitrosoarchaeum koreense* accounted for an average of 70% of all AOA

across the groundwater sites (Figure S 3). Clustering based on the correlation with

hydrochemical parameters, suggest similar ecological preferences of (i) AOA and some clade

B affiliated CMX OTUs, and (ii) *Nitrosomonas ureae* affiliated AOB OTU1 and CMX clade A

affiliated OTUs (Figure 2D).

Metabolic capabilities of CMX *Nitrospira* clades and AOB

We analyzed nine CMX affiliated metagenome assembled genomes (MAGs) from the Hainich

groundwater [21], which were phylogenomically affiliated to two CMX clade A MAGs and

seven clade B MAGs (Figure 3A). Both clade A MAGs (H51-bin291-1 and H41-bin216-1) were

closely related to *Ca.* Nitrospira nitrificans on the genome level and showed 99% amino acid

sequence identity to the AmoA of CMX OTU5 (Figure 2C). AmoA sequences of Clade B

representative MAGs showed closest relation to AmoA of CMX OTU2, OTU3, OTU8 and

OTU9.

Genes encoding essential components for ammonia oxidation, nitrite oxidation and urea

degradation were shared in MAGs representing both CMX clades (Figure 3B) and most of

these genes were highly transcribed in the groundwater. Metaproteomic analysis revealed the

presence of peptides related to AmoA, AmoB, AmoC, Hao and NxrA of *Nitrospira*,

demonstrating that nitrification-related proteins of CMX *Nitrospira* populations were formed in

situ. We found different genomic repertoires between CMX clade A and clade B affiliated

Nitrospira MAGs containing genes for sulfur and alternative energy metabolism such as

oxidation of formate and hydrogen, nitrogen transporter types and cellular regulation of

reactive oxygen species. CMX clade B MAGs exclusively harbored genes involved in

conversion of bicarbonate to carbon dioxide, oxidation of thiosulfate via the *sox* pathway,

dissimilatory sulfate reduction, as well as genes encoding two superoxide dismutases (SOD1/

SOD2) and a formate transporter (*focA*). Transporters for cellular import of urea as well as for

NO_2^-/NO_3^- were present in representative genomes from both CMX clade A and clade B.

Interestingly, we found that clade B MAGs encoded Amt-type NH_4^+ transporters, while clade

A MAGs employed Rh-type NH_4^+ transporters. MAGs from both clades harbored genes for

ATP citrate lyase (*acl*), which were transcriptionally active, indicating an active role in carbon

fixation.

Compared to the CMX and AOA affiliated MAGs, transcription of *amo* genes was more than

100 times higher for AOB affiliated MAG H51-bin202-1 (Figure S 4). The AmoA sequence of

this MAG had 93% sequence identity to the AmoA of the dominant, *N. ureae* affiliated AOB-

OTU1 obtained from amplicon sequencing in this study (Figure S 5), suggesting that both

approaches probably identified the same population of AOB. Interestingly, this MAG employed

Rh-type NH_4^+ transporters as representatives of CMX clade A, while AOA affiliated MAGs had

Amt-type NH_4^+ transporters as representatives of CMX clade B (Figure S 4).

Contribution of ammonia-oxidizing groups to nitrification activity

Short-term incubations in the presence of the nitrification inhibitor chlorate, which is supposed

to target only nitrite oxidizers including CMX, did not affect the nitrification rates (74.5 ± 9.4

nmol N L^{-1} d^{-1}) compared to the controls without inhibitor (75.9 ± 3.7 nmol N L^{-1} d^{-1}) (Figure

4A). In the presence of allylthiourea, rates were not significantly different from zero.

Allylthiourea is supposed to inhibit AOB as well as CMX but affects AOA much less at the 100

μM concentrations used in this study.

Long-term incubations (15-28 days) were performed to further investigate the potential growth

and shifts in community composition of CMX, AOB, AOA and NOB under the inhibitor

treatments, using oxic groundwater from well H41 (condition 1: oxic/ 10 μM NH_4^+; condition 2:

oxic/ 50 μM NH_4^+) and hypoxic groundwater from H53 (condition 3: hypoxic/ without

supplement). Nitrification was always completely inhibited in the presence of allylthiourea,

regardless of groundwater source or initial NH_4^+ concentrations (Figure S 6). NH_4^+ was only

oxidized in incubations without inhibitor and in the chlorate treatment, while NO_2^- oxidation

was only measured in incubations without inhibitor. Condition 3 without inhibitor showed

complete consumption of NH_4^+ after 13 days, while it took 19 and 25 days under condition 1

and 2, respectively.

In incubations without inhibitor, CMX increased two to three-fold compared to initial

groundwater under condition 1 and 2 while we did not detect growth of CMX under condition

3 (Figure 4B). CMX *amoA* gene abundance also increased in the allylthiourea treatment under

condition 1 but decreased in the chlorate treatments. NOB targeted using *nxrB* genes had

similar growth patterns across condition and treatment types as CMX, and CMX accounted

for an estimated 14 to 63 % of the NOB population. AOB *amoA* abundance increased by one

to two orders of magnitude in all untreated and chlorate treatments of the different conditions,

while they decreased in the presence of allylthiourea. AOA *amoA* gene abundance showed

only minor variations during all incubation settings. CMX OTU1 affiliated to CMX clade A

(Figure 4C) and AOB-OTU1 related to *Nitrosomonas ureae* (Figure 4D) increased in

abundance after incubations without inhibitor. Notably, these two OTUs were already

dominant in the ammonia oxidizer communities of the original groundwater.

Discussion

We present evidence that despite numerical predominance of CMX bacteria, they contribute only a small fraction to ammonium oxidation in the studied oligotrophic groundwater. Instead, ammonium oxidation in this system seemed to be largely driven by AOB. Ranging from 0.54 up to 136 nmol N L^{-1} d^{-1}, the measured nitrification rates matched those from oligotrophic lakes and open marine environments including oxygen minimum zones [57–60], but were lower than in rivers, and estuaries [61, 62] (Table S1). The observed nitrification rates did not reflect the spatial heterogeneity of groundwater NO_3^- concentrations, suggesting that surface-derived inputs mask the effects of in situ NO_3 formation on groundwater NO_3^- loads in these carbonate-rock aquifers.

The oligotrophic conditions in the groundwater appeared to select for ammonia oxidizers with the highest reported NH_4^+ affinities of their respective group. As the three groups differ in their NH_4^+ affinity, with 0.65 to 1.1 µM for *Ca.* Nitrospira inopinata [15], 0.3 to 4 µM for the *Nitrosomonas oligotropha* lineage [15, 21], and < 0.01 µM for AOA closely related to *Nitrosoarchaeum koreensis* [15], fluctuating resource availability at overall limiting conditions may have enabled their coexistence. Previous metagenomics analysis work at this study site identified CMX as key nitrifiers and demonstrated that measured groundwater carbon fixation rates could fully be accounted for by carbon fixation associated to nitrification, assuming nitrification and carbon fixation kinetics of CMX *Nitrospira* [21]. Similarly, in our study, CMX accounted for up to 80% of the ammonia oxidizer community at the groundwater well showing the highest nitrification activity, supporting our original assumption that CMX were the most competitive ammonia-oxidizing group. Our inhibitor-based approach strongly suggested that ammonium oxidation appeared to be largely driven by AOB, even though CMX outnumbered AOB by a factor of 10 in the samples used for rate measurements.

Treatment with 1 mM chlorate in nitrification assays was expected to inhibit canonical nitrite oxidizers as well as CMX, since reduction of chlorate to toxic chlorite from the activity of nitrite oxido-reductase would eventually also inhibit ammonium oxidation in the same cell [7].

Surprisingly, ammonium oxidation was not affected in the chlorate treatment, but was

completely inhibited in the treatment with allylthiourea, which specifically targets the ammonia

monooxygenase and acts as a reversible inhibitor preventing the oxidation of NH_4^+ [63]. At a

concentration of below 160 µM, it selectively inhibits the activity of AOB and CMX but not of

AOA [37]. Consequently, this inhibitor approach strongly suggested that AOB were primarily

responsible for the highest observed ammonium oxidation activity at well H41. A strong bias

of our results by potential residual activity of CMX in the presence of chlorate appears unlikely.

Although *Nitrospira* is capable of detoxifying hazardous chlorite by chlorite dismutases [64],

the chlorate concentrations used here obviously resulted in growth inhibition and cell death,

which is in line with previous observations in wetland sediment mesocosms [28]. Similarly, we

can rule out that our results were caused by excessive growth of AOB, since nitrification rates

were linear across these five-days incubations. A negligible contribution of AOA to ammonium

oxidation was indicated by no detectable activity in the presence of allylthiourea.

Despite the presumably minor contribution of CMX to ammonium oxidation rates, we obtained

strong evidence of their active role in overall nitrification. CMX expressed genes and formed

peptides involved in both ammonia and nitrite oxidation in situ, which was confirmed by growth

of CMX in the presence of NH_4^+ in long-term incubations. However, it remains unclear if the

major energy source to sustain cell metabolism originated from complete ammonium oxidation

or rather from nitrite oxidation. In fact, a recent subsurface study detected transcriptional

activity in association with nitrite oxidation but not ammonium oxidation for a CMX-related

MAG [65], while CMX *Nitrospira* in water from rapid sand filters did not prefer to oxidize

external nitrite alone [7]. As we found NO_2^- transporters in representative groundwater MAGs

from both CMX clades, the cells may also be capable of taking up the NO_2^- formed by AOB,

supporting their primary role as nitrite oxidizers along with *Nitrosomonas ureae* and

Nitrosospira sp. affiliated AOB as key ammonia oxidizers.

AOB with faster NH_4^+ oxidation rates than CMX [3, 15] did not outcompete CMX for NH_4^+ but

simply processed a much larger fraction of the available NH_4^+, thereby driving the total

oxidation rates and providing the resulting NO_2^- to both CMX and canonical NOB. The question

remains unclear, why CMX were apparently favored as nitrite oxidizers over canonical NOB,

although CMX accounted for more than 50% of the *Nitrospira* population. Co-occurrence of

CMX and canonical *Nitrospira* appears to be common across a broad range of terrestrial and

aquatic habitats [27], and their respective roles in nitrification might be highly variable.

Physiological studies of *Ca.* Nitrospira inopinata demonstrated a much lower affinity to external

NO_2^- compared to canonical *Nitrospira* [15], a trait which is, however, not shared among all

CMX *Nitrospira* [66].

CMX also dominated the ammonia oxidizer communities in groundwater-fed rapid gravity sand

filters (RGSF) [7, 67] and in coastal wetlands [28] at low NH_4^+ concentrations (71 μM or 1.1

μmol g^{-1}, respectively). In contrast to our findings, CMX *Nitrospira* were identified as the driver

of the two steps of nitrification in the RGSF [7] and to make a similar contribution to nitrification

as AOB in wetland soils (27% of total nitrification) [28]. These contrasting results demonstrate

that high abundance of CMX in an environment do not necessarily imply that they also control

the complete nitrification process. The competitive advantage of CMX has been associated

with a biofilm-associated lifestyle such as in RGSF. However, division of labor in nitrification

may remain the preferred strategy in the planktonic communities of groundwater environments

at low NH_4^+ concentrations.

Distribution patterns of ammonia oxidizers across the groundwater wells suggested different

niche preferences for AOB and CMX clade A compared to AOA and some clade B affiliated

CMX. Both AOA and CMX clade B dominated in groundwater closer to the major recharge

area. AOA can be the most abundant ammonia oxidizers in soils [68] and several studies

reported CMX clade B in soil habitats [9, 69]. Consequently, transfer from soils to shallow

groundwater could result in the observed high fractions of AOA and CMX clade B in the

ammonia oxidizer communities of the respective wells [70]. In turn, deeper and more distant

groundwater favored both AOB and clade A affiliated CMX under slightly increased availability

of NH_4^+ and lower DO concentrations.

The observed ecophysiological niche differentiation of CMX clade A and clade B across

different groundwater conditions was not reflected by differences in central nitrogen

metabolism of representative MAGs [21, 30] and metaproteomic data from the groundwater.

Genes encoding enzymes of ammonia and nitrite oxidation were highly transcribed and

respective peptides represented both CMX clades. However, differences in the encoded NH_4^+

transporters could be linked to different ecological preferences. Rh-type NH_4^+ transporters in

CMX clade A convey a better adaptation to higher NH_4^+ concentrations than Amt-type

transporters. As Rh-type transporters have higher uptake capacity but lower NH_4^+ affinity than

Amt-type transporters [25] and are also encoded in AOB [71], CMX clade A might have a

competitive advantage against clade B in the groundwater when more NH_4^+ is available. Along

with NH_4^+ uptake efficiency, also oxygen tolerance appeared to drive the distribution patterns

of the two clades, as CMX clade B but not clade A encode two types of superoxide dismutases

(SOD1, SOD2) [25, 26]. Finally, given the broader metabolic versatility of clade B including

the potential for formate and thiosulfate utilization, CMX clade B representatives might be less

strongly tied to energy generation by nitrification compared to the clade A representatives.

In conclusion, we provide strong evidence for a major contribution of AOB to ammonia

oxidation activity in oligotrophic groundwater, despite numerical predominance of CMX.

Conditions of low NH_4^+ availability and varying oxygen concentrations appeared to generally

support coexistence of CMX, AOB and AOA with representatives closely related to those with

the highest NH_4^+ affinities in their respective groups. Our results suggest that high substrate

affinity, high growth yields, and a potentially major role in nitrite oxidation enable CMX

Nitrospira to maintain consistently high populations in the groundwater. We further propose

niche differentiation of CMX clade A and B with a more important role in nitrification for clade

A and a higher metabolic versatility, including the use of alternative electron donors, for clade

B.

References

1. Kuypers MMM, Marchant HK, Kartal B. The microbial nitrogen-cycling network. *Nat Rev Microbiol* . 2018. Nature Publishing Group. , **16**: 263–276

2. Koch H, van Kessel MAHJ, Lücker S. Complete nitrification: insights into the ecophysiology of comammox Nitrospira. *Appl Microbiol Biotechnol* . 2018. Springer Berlin Heidelberg. , 1–13

3. Lehtovirta-Morley LE. Ammonia oxidation: ecology, physiology, biochemistry and why they must all come together. *FEMS Microbiol Lett* 2018; **365**.

4. Daims H, Lebedeva E V., Pjevac P, Han P, Herbold C, Albertsen M, et al. Complete nitrification by Nitrospira bacteria. *Nature* 2015; **528**: 504–509.

5. Van Kessel MAHJ, Speth DR, Albertsen M, Nielsen PH, Op Den Camp HJM, Kartal B, et al. Complete nitrification by a single microorganism. *Nature* 2015; **528**: 555–559.

6. Fowler SJ, Palomo A, Dechesne A, Mines PD, Smets BF. Comammox *Nitrospira* are abundant ammonia oxidizers in diverse groundwater-fed rapid sand filter communities. *Environ Microbiol* 2018; **20**: 1002–1015.

7. Gülay A, Fowler SJ, Tatari K, Thamdrup B, Albrechtsen H-J, Al-Soud WA, et al. DNA- and RNA-SIP Reveal Nitrospira spp. as Key Drivers of Nitrification in Groundwater-Fed Biofilters. *MBio* 2019; **10**.

8. Spasov E, Tsuji JM, Hug LA, Doxey AC, Sauder LA, Parker WJ, et al. High functional diversity among Nitrospira populations that dominate rotating biological contactor microbial communities in a municipal wastewater treatment plant. *ISME J* 2020; **14**: 1857–1872.

9. Li C, Hu HW, Chen QL, Chen D, He JZ. Niche differentiation of clade A comammox Nitrospira and canonical ammonia oxidizers in selected forest soils. *Soil Biol Biochem* 2020; **149**: 107925.

10. Osburn ED, Barrett JE. Abundance and functional importance of complete ammonia-oxidizing bacteria (comammox) versus canonical nitrifiers in temperate forest soils. *Soil Biol Biochem* 2020; **145**: 2015–2018.

11. Zhao Z, Huang G, He S, Zhou N, Wang M, Dang C, et al. Abundance and community composition of comammox bacteria in different ecosystems by a universal primer set. *Sci Total Environ* 2019; **691**: 146–155.

12. Shi Y, Jiang Y, Wang S, Wang X, Zhu G. Biogeographic distribution of comammox bacteria in diverse terrestrial habitats. *Sci Total Environ* 2020; **717**: 137257.

13. Liu S, Wang H, Chen L, Wang J, Zheng M, Liu S, et al. Comammox Nitrospira within the Yangtze River continuum: community, biogeography, and ecological drivers. *ISME J* 2020; 1–17.

14. Yu C, Hou L, Zheng Y, Liu M, Yin G, Gao J, et al. Evidence for complete nitrification in enrichment culture of tidal sediments and diversity analysis of clade a comammox Nitrospira in natural environments. *Appl Microbiol Biotechnol* 2018; **102**: 9363–9377.

15. Kits DK, Sedlacek CJ, Lebedeva E V., Han P, Bulaev A, Pjevac P, et al. Kinetic analysis of a complete nitrifier reveals an oligotrophic lifestyle. *Nature* 2017; **549**: 269–272.

16. Roots P, Wang Y, Rosenthal AF, Griffin JS, Sabba F, Petrovich M, et al. Comammox Nitrospira are the dominant ammonia oxidizers in a mainstream low dissolved oxygen nitrification reactor. *Water Res* 2019; **157**: 396–405.

17. Martens-Habbena W, Berube PM, Urakawa H, De La Torre JR, Stahl DA. Ammonia oxidation kinetics determine niche separation of nitrifying Archaea and Bacteria. *Nature* 2009; **461**: 976–979.

18. Prosser JI, Nicol GW. Archaeal and bacterial ammonia-oxidisers in soil: the quest for niche specialisation and differentiation. *Trends Microbiol* . 2012. Elsevier Current Trends. , **20**: 523–531

19. Herber J, Klotz F, Frommeyer B, Weis S, Straile D, Kolar A, et al. A single Thaumarchaeon drives nitrification in deep oligotrophic Lake Constance. *Environ Microbiol* 2020; **22**: 212–228.

20. Sedlacek CJ, McGowan B, Suwa Y, Sayavedra-Soto L, Laanbroek HJ, Stein LY, et al. A Physiological and Genomic Comparison of Nitrosomonas Cluster 6a and 7 Ammonia-Oxidizing Bacteria. *Microb Ecol* 2019; **78**: 985–994.

21. Overholt WA, Trumbore S, Xu X, Bornemann TL V., Probst AJ, Krüger M, et al. Carbon fixation rates in groundwater similar to those in oligotrophic marine systems. *Nat Geosci* 2022; 1–7.

22. Opitz S, Küsel K, Spott O, Totsche KU, Herrmann M. Oxygen availability and distance to surface environments determine community composition and abundance of ammonia-oxidizing prokaroytes in two superimposed pristine limestone aquifers in the Hainich region, Germany. *FEMS Microbiol Ecol* 2014; **90**: 39–53.

23. Van Der Wielen PWJJ, Voost S, Van Der Kooij D. Ammonia-oxidizing bacteria and archaea in groundwater treatment and drinking water distribution systems. *Appl Environ Microbiol* 2009; **75**: 4687–4695.

24. Palomo A, Dechesne A, Cordero OX, Smets BF. Evolutionary Ecology of Natural Comammox Nitrospira Populations. *mSystems* 2022.

25. Palomo A, Pedersen AG, Fowler SJ, Dechesne A, Sicheritz-Pontén T, Smets BF. Comparative genomics sheds light on niche differentiation and the evolutionary history of comammox Nitrospira. *ISME J* 2018; **12**: 1779–1793.

26. Poghosyan L, Koch H, Lavy A, Frank J, van Kessel MAHJ, Jetten MSM, et al. Metagenomic recovery of two distinct comammox Nitrospira from the terrestrial subsurface. *Environ Microbiol* 2019; 1462-2920.14691.

27. Palomo A, Dechesne A, Pedersen AG, Smets BF. Genomic profiling of Nitrospira species reveals ecological success of comammox Nitrospira. *Microbiome* 2022; **10**: 1–14.

28. Sun D, Tang X, Li J, Liu M, Hou L, Yin G, et al. Chlorate as a comammox Nitrospira specific inhibitor reveals nitrification and N2O production activity in coastal wetland. *Soil Biol Biochem* 2022; **173**: 108782.

29. Wang DQ, Zhou CH, Nie M, Gu JD, Quan ZX. Abundance and niche specificity of different types of complete ammonia oxidizers (comammox) in salt marshes covered by different plants. *Sci Total Environ* 2021; **768**: 144993.

30. Wegner CE, Gaspar M, Geesink P, Herrmann M, Marz M, Küsel K. Biogeochemical regimes in shallow aquifers reflect the metabolic coupling of the elements nitrogen, sulfur, and carbon. *Appl Environ Microbiol* 2019; **85**: e02346-18.

31. Herrmann M, Rusznyák A, Akob DM, Schulze I, Opitz S, Totsche KU, et al. Large Fractions of CO 2 -Fixing Microorganisms in Pristine Limestone Aquifers Appear To Be Involved in the Oxidation of Reduced Sulfur and Nitrogen Compounds. *Appl Environ Microbiol* 2015; **81**: 2384–2394.

32.   Lehmann R, Totsche KU. Multi-directional flow dynamics shape groundwater quality in
sloping bedrock strata. *J Hydrol* 2020; **580**: 124291.

33.   Kohlhepp B, Lehmann R, Seeber P, Küsel K, Trumbore SE, Totsche KU. Aquifer
configuration and geostructural links control the groundwater quality in thin-bedded
carbonate-siliciclastic alternations of the Hainich CZE, central Germany. *Hydrol Earth*
Syst Sci 2017; **21**: 6091–6116.

34.   Küsel K, Totsche KU, Trumbore SE, Lehmann R, Steinhäuser C, Herrmann M. How
deep can surface signals be traced in the Critical Zone? Merging biodiversity with
biogeochemistry research in a central German Muschelkalk landscape. *Front Earth Sci*
2016; **4**: 32.

35.   Scheiner D. A modified version of the sodium salicylate method for analysis of
wastewater nitrates. *Water Res* 1974; **8**: 835–840.

36.   Seelmeyer G. Deutsche Einheitsverfahren zur Wasseruntersuchung. Herausgegeben
im Auftrage der Fachgruppe Wasserchemie in der Gesellschaft Deutscher Chemiker E.
V. Bearbeitet von L. W. Haase, H. Stooff, G. Gad, W. Wesly u.a. Verlag Chemie,
Weinheim/Bergstraße 1954, 180. *Mater Corros* 1954; **5**: 357–357.

37.   Shen T, Stieglmeier M, Dai J, Urich T, Schleper C. Responses of the terrestrial
ammonia-oxidizing archaeon Ca. Nitrososphaera viennensis and the ammonia-
oxidizing bacterium Nitrosospira multiformis to nitrification inhibitors. *FEMS Microbiol*
Lett 2013; **344**: 121–129.

38.   Herrmann M, Hädrich A, Küsel K. Predominance of thaumarchaeal ammonia oxidizer
abundance and transcriptional activity in an acidic fen. *Environ Microbiol* 2012; **14**:
3013–3025.

39.   Rotthauwe JH, Witzel KP, Liesack W. The ammonia monooxygenase structural gene
amoA as a functional marker: molecular fine-scale analysis of natural ammonia-
oxidizing populations. *Appl Environ Microbiol* 1997; **63**: 4704–12.

40.   Francis CA, Roberts KJ, Beman JM, Santoro AE, Oakley BB. Ubiquity and diversity of
ammonia-oxidizing archaea in water columns and sediments of the ocean. *Proc Natl*
Acad Sci 2005; **102**: 14683–14688.

41.   Pester M, Maixner F, Berry D, Rattei T, Koch H, Lücker S, et al. *NxrB* encoding the beta
subunit of nitrite oxidoreductase as functional and phylogenetic marker for nitrite-
oxidizing N itrospira. *Environ Microbiol* 2014; **16**: 3055–3071.

42.   Alves RJE, Minh BQ, Urich T, Von Haeseler A, Schleper C. Unifying the global
phylogeny and environmental distribution of ammonia-oxidising archaea based on
amoA genes. *Nat Commun* 2018; **9**: 1517.

43.   Tourna M, Freitag TE, Nicol GW, Prosser JI. Growth, activity and temperature
responses of ammonia-oxidizing archaea and bacteria in soil microcosms. *Environ*
Microbiol 2008; **10**: 1357–1364.

44.   Callahan BJ, McMurdie PJ, Rosen MJ, Han AW, Johnson AJA, Holmes SP. DADA2:
High-resolution sample inference from Illumina amplicon data. *Nat Methods* 2016; **13**:
581–583.

45.   Schloss PD, Westcott SL, Ryabin T, Hall JR, Hartmann M, Hollister EB, et al.
Introducing mothur: open-source, platform-independent, community-supported
software for describing and comparing microbial communities. *Appl Environ Microbiol*
2009; **75**: 7537–7541.

46.   Purkhold U, Wagner M, Timmermann G, Pommerening-Röser A, Koops HP. 16S rRNA

and *amoA*-based phylogeny of 12 novel betaproteobacterial ammonia-oxidizing
isolates: Extension of the dataset and proposal of a new lineage within the
nitrosomonads. *Int J Syst Evol Microbiol* 2003; **53**: 1485–1494.

47. Chaudhari NM, Overholt WA, Figueroa-Gonzalez PA, Taubert M, Bornemann TL V.,
Probst AJ, et al. The economical lifestyle of CPR bacteria in groundwater allows little
preference for environmental drivers. *Environ Microbiome* 2021; **16**: 1–18.

48. Fish JA, Chai B, Wang Q, Sun Y, Brown CT, Tiedje JM, et al. FunGene: The functional
gene pipeline and repository. *Front Microbiol* 2013; **4**: 291.

49. Edgar RC. MUSCLE: Multiple sequence alignment with high accuracy and high
throughput. *Nucleic Acids Res* 2004; **32**: 1792–1797.

50. Letunic I, Bork P. Interactive tree of life (iTOL) v5: An online tool for phylogenetic tree
display and annotation. *Nucleic Acids Res* 2021; **49**: W293–W296.

51. Taubert M, Overholt WA, Heinze BM, Matanfack GA, Houhou R, Jehmlich N, et al.
Bolstering fitness via CO2 fixation and organic carbon uptake: mixotrophs in modern
groundwater. *ISME J* 2021; 1–10.

52. Taubert M, Stöckel S, Geesink P, Girnus S, Jehmlich N, von Bergen M, et al. Tracking
active groundwater microbes with D2O labelling to understand their ecosystem
function. *Environ Microbiol* 2018; **20**: 369–384.

53. Team RC. A Language and Environment for Statistical Computing. *R Found Stat
Comput* . 2020. R Foundation for Statistical Computing, Vienna, Austria. , **2**:
https://www.R--project.org

54. Ogle D, Wheeler P, Dinno A. FSA: Fisheries Stock Analysis. R package version 0.8.31.
2020.

55. Gu Z, Eils R, Schlesner M. Complex heatmaps reveal patterns and correlations in
multidimensional genomic data. *Bioinformatics* 2016; **32**: 2847–2849.

56. Oksanen J, Kindt R, Legendre P, O'Hara B, Simpson GL, Solymos PM, et al. The vegan
package. *Community Ecol Packag* 2008; 190.

57. Small GE, Bullerjahn GS, Sterner RW, Beall BFN, Brovold S, Finlay JC, et al. Rates
and controls of nitrification in a large oligotrophic lake. *Limnol Oceanogr* 2013; **58**: 276–
286.

58. Newell SE, Babbin AR, Jayakumar A, Ward BB. Ammonia oxidation rates and
nitrification in the Arabian Sea. *Global Biogeochem Cycles* 2011; **25**: 1–10.

59. Beman JM, Chow CE, King AL, Feng Y, Fuhrman JA, Andersson A, et al. Global
declines in oceanic nitrification rates as a consequence of ocean acidification. *Proc Natl
Acad Sci U S A* 2011; **108**: 208–213.

60. Bristow LA, Dalsgaard T, Tiano L, Mills DB, Bertagnolli AD, Wright JJ, et al. Ammonium
and nitrite oxidation at nanomolar oxygen concentrations in oxygen minimum zone
waters. *Proc Natl Acad Sci* 2016; **113**: 10601–10606.

61. Gribsholt B, Boschker HTS, Struyf E, Andersson M, Tramper A, De Brabandere L, et
al. Nitrogen processing in a tidal freshwater marsh: A whole-ecosystem 15N labeling
study. *Limnol Oceanogr* 2005; **50**: 1945–1959.

62. Miranda J, Balachandran KK, Ramesh R, Wafar M. Nitrification in Kochi backwaters.
Estuar Coast Shelf Sci 2008; **78**: 291–300.

63. Bédard C, Knowles R. Physiology, biochemistry, and specific inhibitors of CH4, NH4+,

588   and CO oxidation by methanotrophs and nitrifiers. *Microbiol Rev* 1989; **53**: 68–84.

589   64.   Kostan J, Sjöblom B, Maixner F, Mlynek G, Furtmüller PG, Obinger C, et al. Structural
590   and functional characterisation of the chlorite dismutase from the nitrite-oxidizing
591   bacterium ' Candidatus Nitrospira defluvii': Identification of a catalytically important
592   amino acid residue. *J Struct Biol* 2010; **172**: 331–342.

593   65.   Mosley OE, Gios E, Close M, Weaver L, Daughney C, Handley KM. Nitrogen cycling
594   and microbial cooperation in the terrestrial subsurface. *ISME J* 2022; 1–13.

595   66.   Sakoula D, Koch H, Frank J, Jetten MS, van Kessel MA, Lücker S. Enrichment and
596   physiological characterization of a novel comammox Nitrospira indicates ammonium
597   inhibition of complete nitrification. *ISME J* 2020.

598   67.   Tatari K, Gülay A, Thamdrup B, Albrechtsen HJ, Smets BF. Challenges in using
599   allylthiourea and chlorate as specific nitrification inhibitors. *Chemosphere* 2017; **182**:
600   301–305.

601   68.   Cardarelli EL, Bargar JR, Francis CA. Diverse Thaumarchaeota Dominate Subsurface
602   Ammonia-oxidizing Communities in Semi-arid Floodplains in the Western United
603   States. *Microb Ecol* 2020; **80**: 778–792.

604   69.   Zhu G, Wang X, Wang S, Yu L, Armanbek G, Yu J, et al. Towards a more labor-saving
605   way in microbial ammonium oxidation: A review on complete ammonia oxidization
606   (comammox). *Sci Total Environ* . 2022. Elsevier. , **829**: 154590

607   70.   Krüger M, Potthast K, Michalzik B, Tischer A, Küsel K, Deckner FFK, et al. Drought and
608   rewetting events enhance nitrate leaching and seepage-mediated translocation of
609   microbes from beech forest soils. *Soil Biol Biochem* 2021; **154**: 108153.

610   71.   Weidinger K, Neuhäuser B, Gilch S, Ludewig U, Meyer O, Schmidt I. Functional and
611   physiological evidence for a Rhesus-type ammonia transporter in Nitrosomonas
612   europaea. *FEMS Microbiol Lett* 2007; **273**: 260–267.

Data availability

Sequence data generated in this study was submitted to the European Nucleotide Sequence

Archive with project number PRJEB57321 under accessions ERS13665998 to ERS13665900.

The mass spectrometry proteomics data have been deposited to the ProteomeXchange

Consortium via the PRIDE [1] partner repository with the dataset identifier PXD039573.

Raw data of nitrification rates, gene abundance, hydrochemical data and OTU sequences will

be released upon publication of the manuscript.

Acknowledgements

This study was part of the Collaborative Research Centre 1076 AquaDiva (CRC AquaDiva) of the Friedrich Schiller University Jena, funded by the Deutsche Forschungsgemeinschaft under project number 218627073. Climate chambers to conduct experiments under controlled temperature conditions and the infrastructure for MiSeq Illumina sequencing were financially supported by the Thüringer Ministerium für Wirtschaft, Wissenschaft und Digitale Gesellschaft (TMWWDG; project B 715-09075 and project 2016 FGI 0024 "BIODIV"). Martin Taubert gratefully acknowledges funding by the Deutsche Forschungsgemeinschaft (DFG, German Research Foundation) under Germany´s Excellence Strategy – EXC 2051 – Project-ID 390713860.

We thank Heiko Minkmar, Falko Gutmann, Jens Wurlitzer, Stefan Krümmling, Swantantar Kumar, Patricia Geesink, Lena Carstens and Linda Gorniak for support of field and laboratory work and Gundula Rudolph for ion chromatography measurements. We acknowledge Kai Uwe Totsche and Robert Lehmann for field site management and provision of physicochemical data from groundwater.

Author contributions

MH and MK designed the study. MK conducted all molecular work, incubation experiments

and data analysis. BT and LB performed IRMS measurements and analysis. NC and WO

performed metagenomic data analysis. MT, NJ and MvB conducted metaproteomic data

analysis. MK and MH wrote the manuscript with contributions of all authors.

Declaration of competing interest

The authors declare that there was no financial or personal conflict of interest in the

preparation of this study.

Figures

Figure 1

Figure 1: **A)** Schematic cross-section of the Hainich groundwater monitoring transect showing two
superimposed aquifer assemblages accessible by 10 wells with various depths. **B)** Overview the
concentrations of dissolved oxygen, ammonium, and nitrate from oxic (blue) and hypoxic/ anoxic
(green) groundwater wells (n = 24 per well). Single dots represent one measurement, the crossbar
indicates the mean of all values, and the error bars show the variation from the mean. Outliers are
beyond the error bars. Lowercase letter code indicates significant differences based on Dunn's multiple
comparison test ($p < 0.05$). **C)** Nitrification rates assessed by ^{15}N based labeling of groundwater from
five days incubation close to in situ conditions (n ≥ 3 per well; n.d. = not detected; n.a. = not available).
D) Gene abundances of *amoA* per L groundwater from different groups of ammonia-oxidizing
prokaryotes in 10 groundwater wells (n ≥ 3).

Figure 2

Figure 2: **A)** Redundancy analysis (RDA) demonstrates association between distribution of ammonia-
oxidizing community (relative *amoA* read abundances of CMX, AOB and AOA OTUs which account for
< 2% sequence reads of each community were normalized to *amoA* gene abundances to generate
absolute amoA read abundances) and significant hydrochemical parameters fitting ($p < 0.05$) with
nitrifier community distribution across eight groundwater wells (n ≥ 3). Dotted ellipses show the
confidence of clusters assuming a multivariant normal distribution. **B)** Community composition of CMX
Nitrospira showing dominant OTUs (< 2% of sequence reads) affiliated to clade A and clade B across
eight groundwater wells (n ≥ 3 per well). Each bar displays the mean absolute *amoA* read abundance
of OTUs. **C)** Maximum likelihood phylogenetic tree of deduced AmoA amino acid sequences of CMX
Nitrospira depicting evolutionary relationship between reference genomes, and sequences of OTUs
and MAGs from Hainich groundwater. Tree was constructed using JTT substitution model with gamma
distribution and 1000 bootstrap iterations. Bootstrap support values greater than 75% are denoted with
black dots. An AmoA sequence from *Nitrosomonas oligotropha* served as outgroup indicated by the
arrow. **D)** Pairwise correlations between total amoA gene abundances of CMX, AOB and AOA, absolute
amoA read abundances of OTUs and groundwater hydrochemical parameters (TIC = total inorganic
carbon, TOC = total organic carbon, DOC = dissolved organic carbon, ORP = redox potential, EC =
electrical conductivity) using Spearman's rank correlation. Hierarchical clustering is based on Euclidean
distance. Color code of correlation coefficient displays strength of association using following
significance levels: * $p < 0.05$, ** $p < 0.01$, *** $p < 0.001$.

Figure 3

Figure 3: **A)** Maximum likelihood tree of bacterial core protein sequences of CMX *Nitrospira* from nine
MAGs originating from Hainich groundwater depicting the evolutionary relationship to reference
genomes. Tree was constructed using JTT substitution model with gamma distribution and 1000
bootstrap iterations. Bootstrap support values greater than 75% are shown as black dots. AmoA
sequences from *Nitrosomonas oligotropha* and *Nitrosomonas europaea* served as an outgroup
indicated by the arrow. **B)** Metabolic capacities of nine groundwater CMX *Nitrospira* MAGs inferring
genomic potential for nitrogen and sulfur energy metabolism, and corresponding transporter types as
well as protective systems against radical oxygen species, and genes involved in carbon fixation (rTCA
= reductive tricarboxylic acid cycle). Sizes of circles display the transcript coverage of each gene
mapped from groundwater metatranscriptome. If more than one gene was present, maximum
transcriptional coverage is shown. Filled circles indicate peptides from metaproteomic data matching
for displayed genes. rTCA = reverse tricarboxylic acid cycle.

Figure 4

Figure 4: Groundwater incubations from wells H41 and H53 supplemented with different NH_4^+
concentrations and nitrification inhibitors across 5 days (short-term) and 25 to 35 days (long-term). **A)**
Nitrification rates from short-term incubation of oxic groundwater supplemented with 50 μM NH_4^+ and
selective inhibition with chlorate and allylthiourea (n.d. = not detected). Dots depict single rates of each
replicate. Crossbar shows the average of three replicates and error bars show the variation between
rates. **B)** Fold-change of *amoA* genes from ammonia oxidizers and *nxrB* genes from *Nitrospira*-like
nitrite oxidizers after long-term incubation of groundwater. Single points represent one replicate and
crossbar shows the average. Error bars display variance between conditions (n = 3). Condition 1: oxic
groundwater supplemented with 10 μM NH_4^+. Condition 2: oxic groundwater supplemented with 50 μM
NH_4^+. Condition 3: hypoxic groundwater without additional NH_4^+. **C)** Composition of CMX *Nitrospira*
clade A and clade B affiliated OTUs of groundwater and after long-term incubation (n = 3). Bars show
the mean relative *amoA* read abundance of OTUs representing > 2%. **D)** Composition of bacterial
ammonia-oxidizing community and taxonomic affiliation of OTUs.