# Dolphin social phenotypes vary in response to food availability but not climate

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## Abstract

Social behaviours can allow individuals to flexibly respond to environmental change, potentially buffering adverse effects. However, individuals may respond differently to the same environmental stimulus, complicating predictions for population-level response to environmental change. Here we show that bottlenose dolphins (*Tursiops truncatus*) alter their social behaviour at yearly and monthly scales in response to a proxy for food availability (salmon abundance) but do not respond to variation in a proxy for climate (the North Atlantic Oscillation index). There was also individual variation in plasticity for gregariousness and connectedness to distant parts of the social network, although these traits showed limited repeatability. In contrast, individuals showed consistent differences in clustering with their immediate social environment at the yearly scale but no individual variation in plasticity for this trait at either time scale. These results indicate that social behaviour in free-ranging cetaceans can be highly resource dependent with individuals increasing their connectedness over short timescales but possibly reducing their wider range of connection at longer timescales.  Some social traits showed more individual variation in plasticity or mean behaviour than others, highlighting how predictions for the responses of populations to environmental variation must consider the type of individual variation present in the population.

30 **Keywords**: dolphin, environmental change, individual variation, plasticity, social network, Tursiops

## Introduction

Animals engage in social interactions with conspecifics which are fundamental for determining

health, access to resources, and reproductive success [1]. Consequently, social interactions have a

strong influence on ecological processes such as population dynamics and evolutionary processes

such as the response to selection [2–4]. To maintain the best fit with their environment, animals may

adjust their social behaviour as conditions change [5], for instance being more gregarious when

resources are plentiful but less tolerant of conspecifics when resources are scarce [6]. Animals may

also change their behaviour through development and during senescence [7], and may non-

adaptively adjust their behaviour due to direct effects of the environment and other limitations [8].

Such plasticity is a hallmark of behavioural traits and gives behaviour an important role in how

animals interact with their environment.

When trying to understand how animals may respond plastically to changing environments, most

examine responses at the population level [e.g., 9], presuming that any individual variation in

response is absent or simply aggregates to give the population-level response. However, individuals

may show variation in plasticity, and so each will respond differently to change [10,11, sometimes

referred to as 'I x E' i.e., individual by environment interactions,12,13]. For example, European field

crickets (*Gryllus campestris*) become bolder and more active as they age, but individuals vary in the

extent of this, with some not increasing or even decreasing [14]. The degree of plasticity animals

show can be correlated with their mean behaviour (an "intercept-slope correlation") which

determines how the magnitude of among-individual differences varies across environments and

indicates the extent to which plasticity is a separate trait in its own right [11,15]. Individuality in

plasticity can influence biological processes such as population growth and adaptive change at a

range of scales [16, see: 17 and accompanying papers], giving fundamentally different results to

when population-level only effects are assumed [15,18]. For example, Seebacher and Little

demonstrated that mosquitofish (*Gambusia holbrooki*) differ in how their performance changes with

temperature, resulting in a switching of the rank of swimming speed of individuals between cool and

warm temperatures. This changed which individuals might be predated, with a lower fraction of the

population reaching the critical speed to avoid predation in cool temperatures [19]. Therefore,

variation in plasticity will alter both the strength of selection and which genotypes produce

phenotypes that are selected for, altering evolution trajectories. Additionally, the extent of among-

individual variation in plasticity gives an upper-limit to the heritability of plasticity, which indicates

how rapidly plasticity itself can evolve [20–22]. Understanding how plasticity as a trait in its own

right can evolve is key for understanding how animals will adapt to more variable climates [22,23].

Marine mammals are a key group to study individual variation in response to environmental change.

They are typically long-lived, increasing the relative importance of plasticity versus adaptive

evolution for coping with contemporary environmental change [24]. They are also exposed to a wide

range of changing conditions during their lifetimes including climate, food availability, and pollution,

and their populations are often of conservation concern. All of these factors increase the need for us

to understand how they respond to changes in their environment [25–32].

Here, we studied a population of bottlenose dolphins (*Tursiops truncatus*) in the North Sea for over

30 years, regularly recording their social associations. Previous work in this study population has

demonstrated that critical group sizes increase in years of higher salmon abundance [33], and we

extend this by examining multiple facets of individual-level social behaviour at both the monthly and

yearly temporal scales. We achieved this by using social network analysis to quantify three different

dimensions of individual social behaviour: an individual's gregariousness (strength), how tightly its

immediate social group interact together (clustering coefficient), and how well connected an

individual is to the entire population (closeness; each described in more detail below). We then used

random regression models to quantify individual social phenotypes and determine how these social

phenotypes depend on yearly and monthly variation in available proxies for climate (at a broad

scale) and food availability (at a local scale). Our analyses also indicated whether bottlenose dolphins

show individual changes in response to the environment, or if population-level change was more

prominent.

Specifically, we were interested in how social behaviour depends on current environmental

conditions. Social behaviours are often highly dependent on both resource availability and spatial

distribution [6] and current climatic conditions can impose energetic constraints on individuals

[34,35] and impact their ability to move around their environment [36]. Bottlenose dolphin group

sizes off the north west coast of Spain showed a nonlinear relationship with the North Atlantic

Oscillation (NAO) index [37], while abundances of Indo-Pacific bottlenose dolphins (*T. aduncus*) are

impacted by a combination of the El Nino Southern Oscillation and season [38]. In our study system,

previous work indicated that the NAO index at a two-year lag was associated with dolphin critical

group size, but this appears to be entirely mediated through food availability [33], something we are

testing for directly. As such, we did not consider lagged effects here. We summarised climate

through the NAO index (see Methods), where positive values in this region indicate warmer and

wetter periods which would make rougher sea conditions, potentially resulting in the dolphins

travelling shorter distances. We therefore expect higher NAO values to lead to higher clustering

coefficients and lower closeness, but not to affect strength, at monthly and yearly scales. We

summarised resource availability through salmon abundances (see Methods). We expect that higher

salmon abundances allow dolphins to form larger groups (as found previously) and travel shorter

distances to find sufficient food, leading to higher strengths, higher clustering coefficients, and lower

closeness at both temporal scales.

## Material and methods

### Study site & group data collection

This study used data from a bottlenose dolphin population on the east coast of Scotland (Fig. 1). The population of over 200 individuals [39] has been studied intensively as part of a long-term individual based study [40–42]. We use data from boat-based photo-identification surveys carried out annually between 1990 and 2021 which regularly recorded dolphin groups within the Moray Firth Special Area of Conservation (SAC; 92/43/EEC), a core part of the population's range which over 50% of the population use each year (40). All surveys were made from small (5-6m) boats with outboard engines, carefully and slowly manoeuvring the boat around each group to obtain high quality images of the left and right side of as many dorsal fins as possible. Surveys initially followed a fixed survey route until 2001 when, as a result of changing dolphin distribution within the SAC, flexible survey routes were introduced to maximise sightings probability [more details in: ,43]. Data were available from a total of 690 surveys (between 9 and 35 surveys each year; average of 22) with the majority carried out between May to September.

**Figure 1**. Images depicting a photo-identification survey (top left), a newborn calf with mother (middle left), unique markings used to identify individuals (bottom left), and location of encounters

with bottlenose dolphins within the Moray Firth Special Area of Conservation from 1990 to 2021

(right).

During surveys, when we located a bottlenose dolphin group (one or more individuals in close

proximity within 100m, hereafter an "encounter") we collected photo-identification data following a

standardised protocol [43]. We identified individuals from high-quality photographs based on unique

markings matched against a photo catalogue of previously identified individuals from the area

[40,41,44]. On average in a group 84% are successfully photographed, a rate of identification well

above the level at which social network metrics in incomplete networks are reliable [45]. All

individual identifications from photographs were confirmed by at least two experienced researchers.

For individuals first sighted as calves we could determine their year of birth and so their age [46], but

for individuals first sighted as juveniles or adults their exact age is unknown. Sex was determined

using genital photographs or if an adult was seen in repeat associations with a known calf [46].

Social network construction

Individuals sighted during the same encounter were assumed to be in the same group and therefore

associating [known as the 'Gambit of the Group' 47]. Aggregating many of these records of groups

allows one to infer which individuals are frequently associated and which individuals infrequently or

never associate. We removed observations of individuals younger than three years old (n = 2668

observations of 242 individuals), as these individuals are not likely to be independent of the mother

and so their social associations most likely represent her preferences. We then converted the

records of encounters into group by individual matrices (indicating which individuals were seen

together in each encounter) and then into weighted, undirected social networks using the R package

*asnipe* [48]. Edge weights were set as the simple ratio index, where the number of times two

individuals are seen together is divided by the total number of times they are seen, both together

and apart [49]. This measure ranges from 0 (individuals never seen together) to 1 (individuals always

seen together). We did this separately for each year, creating yearly social networks to assess how

social phenotypes vary at this temporal scale in response to environmental conditions. To assess

how social phenotypes vary at the monthly scale in response to environmental conditions we then

reconstructed social networks per month and removed any months with fewer than 10 encounters

[excluding 110 out of 215 months; as networks constructed using fewer than 10 observations can be

biased; ,50]. Histograms of the frequency of the number of encounters per year and per month are

shown in Fig. S1.

For each individual present in the network, for each year and again for each month we calculated

three network metrics. First was "strength", the sum of all an individual's associations, which as our

associations are based on observations of co-occurrence in groups is analogous to typical size of

groups an individual is in. Second was "weighted clustering coefficient", the rate at which an

individual associates with other individuals who also associate with each other. This metric

represents how tightly individuals interact in their immediate social environment (possibly

analogous to "alliances" between three or more individuals, see also: [51,52]), at the expense of

interacting with a wider range of individuals. Finally, we quantified "closeness", the inverse of the

mean of the path lengths between that individual and each other individual in the network,

corrected for network size to allow comparisons among networks which vary in the number of

individuals. Closeness represents the dolphin's connectedness to the wider population, and would

be high if an individual linked two communities or moved between different areas each containing

more sedentary individuals.

We removed an individual's scores for a given year if they had fewer than five observations in that

year (removing 811 observations and leaving 874), as the social network position of those individuals

would be highly uncertain. They would however still contribute to the social environments and

therefore social network measures of individuals in that year who had five or more observations. We

repeated this at the month level, removing individuals' monthly scores when they had fewer than

five observations that month (removing 3423 observations and leaving 320). Histograms of the

frequency of the number of encounters per individual per year and per month are shown in Fig. S2.

The social network measures were not strongly correlated; Pearson correlations between

individuals' strength and clustering coefficient were 0.141 (yearly) and 0.346 (monthly), for strength

and closeness they were -0.004 (yearly) and -0.245 (monthly), and for clustering coefficient and

closeness they were 0.129 (yearly) and 0.170 (monthly).

Environmental data

We used the NAO index in the same time period the grouping observations were made as a measure

of climate. We used monthly and yearly measures of the NAO index between 1990 and 2021

downloaded from https://www.cpc.ncep.noaa.gov/products/precip/CWlink/pna/nao.shtml (Fig. S3).

This index indicates the atmospheric pressure difference between the low pressure zone over

Iceland and the high pressure zone over the Azores [53,54]. This index has frequently been linked to

the ecology of animal populations [55,56], for example influencing the foraging behaviour of Cory's

shearwaters [*Calonectris borealis*; ,57]. Climatic effects on cetaceans are typically thought to occur

via changes in prey species [31,58]; for instance Lusseau *et al*. found the NAO at a two-year lag

influenced critical group size in our study population through the lagged variable's effect on food

availability [33]. However, it is also possible that cetaceans respond directly to climate, sometimes at

even faster rates than their prey species [37,38,59].

For our index of food availability we followed Lusseau *et al*. [33], using data on monthly catches by

fishing rods (as opposed to nets) from the wild of both one season and multiple season adult Atlantic

salmon (*Salmo salar*) from the Alness, Beauly, Canon, Ness, and Nairn rivers. These feed into the sea

where observations of dolphin groups take place and hence are expected to be a good proxy for

salmon availability in that area. Further, catches on rods are positively correlated among rivers and

among months and with automatic counter data, indicating they are a good proxy for actual

abundance [60,61]. Atlantic salmon are an important food source for this dolphin population [62,63],

with dolphins forming larger groups when salmon are more abundant [33]. We downloaded monthly

data from https://marine.gov.scot/data/marine-scotland-salmon-and-sea-trout-catches-salmon-

district-shinyapp (using "rod data" and summing retained and released fish for both MSW and 1SW)

and summed monthly catches within a calendar year for yearly measures of fish abundance (Fig. S4).

Data analysis

All analyses were performed in R [ver 4.3.1; ,64] using linear mixed-effect models in *glmmTMB* [65].

Using regression-based models as opposed to randomisation-based tests has been recommended

for analysing questions about node-level social network traits as it improves the ability to make

inference while accounting equally well as node permutations for common types of data non-

independence [66]. We fitted 12 models, with all combinations across the three social traits, the two

environmental variables, and the yearly and monthly timescales. Clustering coefficient cannot be

calculated when an individual only associates with one other individual, and so the datasets for

these models were slightly smaller than the dataset for strength and closeness (see below). We

included individuals with unknown birth and death dates to maximise our sample size, and so

neither age nor lifespan could be included as predictor variables. In all models we included the fixed

effect of sex [as males and females can differ in social behaviour; ,67,68; ranging behaviour, and

survival; ,69], and either the NAO index or the count of caught salmon for that month or year. We

did not include both the NAO index and salmon abundance in the same model as we encountered

estimation problems with two random slopes. Including sex meant we excluded individuals of

unknown sex (140 observations of 66 individuals for the yearly networks, 30 observations of 21

individuals for the monthly networks), but the interactions between individuals of known and

unknown sex were still used to build the networks and so associations with individuals of unknown

sex still influenced the social network traits of males and females. We mean centred and scaled to

unit variance the environmental variable [70], and included the interaction between it and sex, to

see if male and female social behaviour responded differently to environmental variation. Random

effects were the random intercept for individual ID, the random slope of individual ID with the

environmental variable fitted as a fixed effect in the model (again mean centred and with a standard

deviation of 1), and the correlation between these two terms. We also included a temporal

autocorrelation term (ar1) among years to account for unmodelled environmental variation that

changes slowly across years, which could influence social behaviour to make adjacent years more

similar than non-adjacent years. Similarly, in the models for monthly variation we included a random

effect of month alongside the yearly temporal autocorrelation term. We used a Gaussian error

structure for all models. For strength and closeness, we used log link functions as the distributions

were right skewed. For clustering coefficient, which is bounded between 0 and 1, we used a logit link

function, which is preferable to an arcsine transformation when handling response variables

bounded in this way [71]. We used the default optimising algorithm for all models except for both

the yearly models for clustering coefficient and closeness, the model for strength in response to

monthly variation in the NAO index, and the models for clustering coefficient and closeness in

response to monthly variation in salmon abundance, where we used the "BFGS" optimiser, as

otherwise the models did not converge [65].

We report the coefficients and standard errors for each fixed effect, along with p-values, to give an

idea of the magnitude and uncertainty of each effect. We used the p-values from the Anova function

of the *car* package [72], using a Chi-squared test with type III sum of squares. We describe these p-

values in terms of "clarity" rather than "significance"; see Dushoff *et al*. [73] for a discussion on this.

To test whether individuals clearly differed in their response to the environmental variable, we first

tested whether there was a correlation between an individual's plasticity and its mean behaviour by

re-fitting each model (12 in total) with the correlation between random intercepts and slopes

suppressed to zero and conducted a likelihood ratio test between the full model and this reduced

model with a single degree of freedom. If there was a clear difference between the models, we

concluded that the correlation between intercepts and slopes was non-zero. If there was no clear

difference between the models, we then tested the importance of the random slopes by comparing

the model with an intercept-slope correlation of zero to a model without the random slopes (but still

with the random intercepts) using a likelihood ratio test with a mix of zero and one degree of

freedom [as is appropriate for testing the clarity of single variance components, 74]. If there was a

clear difference between the models, we concluded that individuals differ in their response to the

environmental variable. When there is variation in plasticity, the magnitude of among-individual

differences varies across environments. We therefore calculated the marginal repeatability for each

trait following Schielzeth and Nakagawa [75]. We did this for all trait-environmental variable-time

scale combinations (12 in total), even when there was no evidence for random slopes, to aid

comparison among traits. Data and R code are available online [76].

Results

Change with environmental variables at a yearly scale

For the analysis of how dolphin social phenotypes change at the yearly scale in response to

environmental variables, our dataset included 129 unique individuals. For strength and closeness

there were 874 measures, and for clustering coefficient there were 873 measures. All traits had a

mean of 6.78 (sd = 5.39) measures per individual.

Dolphins' strength and clustering coefficient were not affected by the NAO index (Fig. 2a and b) or

salmon abundance (Fig. 2d and e) in both sexes, and the sexes did not differ in mean strength or

clustering coefficient (Table 1, full model results in Tables S1-4, Supplementary Materials). For both

strength and clustering coefficient with the NAO index there were no intercept-slope correlations

and the random slopes were not statistically clear (Table 2). Individuals did differ in how their

strengths changed with salmon abundance, with a clear negative intercept-slope correlation (Table

2). There was no intercept-slope correlation and no random slopes for the change of clustering

coefficient with salmon abundance (Table 2). The marginal repeatabilities for strength were low:

0.017 in the NAO index model and 0.018 in the salmon abundance model, suggesting the trait has

limited repeatability. The marginal repeatabilities for clustering coefficient were higher: 0.172 and

0.178 for the NAO index and salmon abundance models respectively.

**Figure 2.** Plots of each of the three social network traits and yearly variation in the North Atlantic

Oscillation (NAO) index (a. strength, b. clustering coefficient, c. closeness) and salmon abundances

(d. strength, e. clustering coefficient, f. closeness). For each individual dolphin we have predicted its

network trait on the observed scale based on the model results, using the "predict" function in R

with the individual's sex, the range of NAO values or salmon counts that individual was exposed to,

and picking a random year that individual experienced.

**Table 1**. Main effects ("M") and interactions with sex ("I", also shaded in grey) for the two

environmental variables effects on the three social network traits, at each of the monthly and yearly

scales. Effects clearly different from zero (p < 0.05) are highlighted in bold, while effects with p

values between 0.07 and 0.05 are shown in italics.

| | | | Strength | Clustering coefficient | Closeness |
|---|---|---|---|---|---|
| NAO index | Yearly | M | $\beta = -0.035 \pm 0.066, \chi^2 = 0.280, p = 0.597$ | $\beta = -0.102 \pm 0.081, \chi^2 = 1.595, p = 0.207$ | $\beta = -0.115 \pm 0.102, \chi^2 = 1.275, p = 0.259$ |
| | | I | $\beta = -0.002 \pm 0.021, \chi^2 = 0.005, p = 0.942$ | $\beta = 0.031 \pm 0.035, \chi^2 = 0.793, p = 0.373$ | $\beta = 0.006 \pm 0.008, \chi^2 = 0.458, p = 0.499$ |
| | Monthly | M | $\beta = 0.053 \pm 0.034, \chi^2 = 2.437, p = 0.119$ | *$\beta = -0.102 \pm 0.056, \chi^2 = 3.304, p = 0.069$* | $\beta = 0.023 \pm 0.033, \chi^2 = 0.464, p = 0.496$ |
| | | I | $\beta = 0.039 \pm 0.037, \chi^2 = 1.105, p = 0.293$ | $\beta = 0.057 \pm 0.069, \chi^2 = 0.678, p = 0.410$ | $\beta = 0.031 \pm 0.037, \chi^2 = 0.713, p = 0.398$ |
| Salmon | Yearly | M | $\beta = 0.108 \pm 0.069, \chi^2 = 2.422, p = 0.120$ | $\beta = -0.027 \pm 0.091, \chi^2 = 1.595, p = 0.207$ | **$\beta = -0.258 \pm 0.094, \chi^2 = 7.489, p = 0.006$** |
| | | I | $\beta = -0.009 \pm 0.026, \chi^2 = 0.126, p = 0.723$ | $\beta = -0.070 \pm 0.038, \chi^2 = 0.794, p = 0.373$ | $\beta = -0.002 \pm 0.008, \chi^2 = 0.091, p = 0.763$ |
| | Monthly | M | **$\beta = 0.137 \pm 0.038, \chi^2 = 12.755, p < 0.001$** | *$\beta = 0.116 \pm 0.062, \chi^2 = 3.569, p = 0.059$* | **$\beta = 0.087 \pm 0.040, \chi^2 = 4.714, p = 0.030$** |
| | | I | $\beta = -0.011 \pm 0.045, \chi^2 = 0.065, p = 0.799$ | $\beta = -0.105 \pm 0.074, \chi^2 = 2.011, p = 0.156$ | $\beta = 0.009 \pm 0.030, \chi^2 = 0.081, p = 0.777$ |

Closeness did not vary with NAO index in either sex (Fig. 2c), but it did decrease with increasing

salmon numbers in both sexes (Fig. 2f ; Table 1; full results in Supplementary Material Tables S5 &

S6). Dolphins showed a negative mean-plasticity relationship for the NAO index, with individuals

with lower than average mean closeness increasing their closeness with increasing NAO indices and

individuals with higher than average closeness decreasing their closeness. This also indicates there

was individual variation in plasticity (Table 2). In contrast, there was a positive intercept-slope

correlation for closeness in response to salmon abundance (Table 2), indicating that individuals with

lower means also decreased the most. The marginal repeatabilities were low for closeness: 0.003 in

the NAO index model and 0.002 in the salmon abundance model.

**Table 2**. Intercept slope correlations ("ISC") and their statistical tests, and whether random slopes

("RS") were present or absent and their statistical tests (likelihood ratio tests in all cases, an asterix is

given for the test of the random slopes if it was not performed as the intercept-slope correlation was

first found to be clear). Clear positive correlations are noted in bold text and highlighted with blue

fill, and clear negative correlations with bold text and orange fill, while clear random slopes are

highlighted in bold text and the same colour as the associated correlation.

|  |  |  | Strength | Clustering coefficient | Closeness |
|---|---|---|---|---|---|
| NAO index | Yearly | ISC | -0.377 ($\chi_1^2$ = 1597, p = 0.206) | 0.996 ($\chi_1^2$ = 1.274, p = 0. 259) | **-0.957 ($\chi_1^2$ = 5.269, p = 0.022)** |
|  |  | RS | Absent ($\chi_{0,1}^2$ = 1.688, p = 0.097) | Absent ($\chi_{0,1}^2$ = 0.023, p = 0.500) | **Present \*** |
|  | Monthly | ISC | -0.976 ($\chi_1^2$ = 0.868, p = 0.352) | 0.075 ($\chi_1^2$ = 0.000, p = 0.985) | -0.486 ($\chi_1^2$ = 0.447, p = 0.504) |
|  |  | RS | Absent ($\chi_{0,1}^2$ = 0.001, p = 0.486) | Absent ($\chi_{0,1}^2$ = 0.030, p = 0.431) | **Present ($\chi_{0,1}^2$ = 3.448, p = 0.032)** |
| Salmon | Yearly | ISC | **-0.603 ($\chi_1^2$ = 7.195, p = 0.007)** | 0.998 ($\chi_1^2$ = 2.660, p = 0.103) | **0.982 ($\chi_1^2$ = 3.967, p = 0.046)** |
|  |  | RS | **Present \*** | Absent ($\chi_{0,1}^2$ = 0.000, p = 0.500) | **Present \*** |
|  | Monthly | ISC | **-0.769 ($\chi_1^2$ = 6.191, p = 0.013)** | -0.967 (NA, model did not converge) | -0.999 ($\chi_1^2$ = 2.005, p = 0.157) |
|  |  | RS | **Present \*** | Absent ($\chi_2^2$ = 0.035, p = 0.983) | Absent ($\chi_{0,1}^2$ = 0.000, p = 0.500 |

For all social traits there was substantial among-year variation and social traits in consecutive years

were positively correlated (year to year correlations: strength; $r_{NAO}$ = 0.452, $r_{Salmon}$ = 0.566; clustering

coefficient; $r_{NAO}$ = 0.590, $r_{Salmon}$ = 0.657; and closeness; $r_{NAO}$ =0.283, $r_{Salmon}$ = 0.472), showing that, as

expected, adjacent years were more similar than non-adjacent years.

In summary, dolphins had lower closeness scores in years of high salmon abundance, but there were

no trait-environment associations at the yearly scale for the other two social traits or for the effect

of NAO. For closeness, individuals showed variation in plasticity that was related to their mean

behaviour for both NAO and salmon abundance, but for clustering coefficient individuals showed no

variation in individual plasticity. Individual dolphins showed variation in their plasticity of strength in

response to salmon abundance only, with a great spread of values at high than low values. Clustering

coefficient showed consistent differences between individuals across environments, but strength

and closeness did not.

Change with environmental variables at a monthly scale

We analysed how dolphin social phenotypes change in response to environmental variables at the

monthly scale with a dataset of 88 unique individuals and 320 measures for all traits. Traits had a

mean of 3.64 measures (sd = 2.64) each.

Strength increased with monthly salmon abundance for both sexes (Fig. 3d, Table 1), and there was

individual variation in plasticity and a negative intercept-slope correlation, with individuals with

lower means increasing more than those with higher means (Table 2). However, strength did not

respond to the monthly NAO index (Fig. 3a) and there was no individual variation in plasticity due to

the NAO (Table 1, full model results in Tables S7-12 in the supplementary materials).  As for the

yearly models, the marginal repeatability of strength was low (0.003 in the NAO index model, 0.007

in the salmon abundance model).

Clustering coefficient showed no clear relationships with either the monthly NAO index (Fig. 3b), or

the monthly salmon abundance (Fig. 3e and Table 1), and the sexes did not differ in their mean

clustering coefficient or how it varied. Individuals showed no mean-plasticity relationship and no

individual variation in response to either variable (Table 2, note that the model for clustering

coefficient and monthly salmon would not converge with random intercepts and slopes but no

correlation between them, so we compared the full model with the model without random slopes

using two degrees of freedom). Clustering coefficient was slightly repeatable, with marginal

repeatabilites of 0.038 and 0.020 in the NAO index and salmon abundance models respectively.

**Figure 3**. Plots of each of the three social network traits and monthly variation in the North Atlantic

Oscillation (NAO) index (a. strength, b. clustering coefficient, c. closeness) and salmon abundances

(d. strength, e. clustering coefficient, f. closeness). For each individual dolphin we have predicted its

network trait on the observed scale based on the model results, using the "predict" function in R

with the individual's sex, the range of NAO values or salmon counts that individual was exposed to,

picking a random year that individual experienced, and the month to June (an arbitrary choice that

was approximately in the middle of the calendar year).

Both sexes increased their closeness with higher monthly salmon abundance (Fig. 3f), but individuals

did not vary in how their closeness changed in response to salmon abundance (Table 2). There was

no response to the NAO index (Fig. 3c and Table 1), with no mean-plasticity relationship. However,

individuals did vary around the population-level stability in their responses to the NAO index (Table

2). There was limited repeatability of closeness at the monthly scale, with a marginal repeatability of

0.063 in the NAO index model and 0.033 in the salmon abundance model.

There was among-month and among-year variation, but in the monthly models social traits in

consecutive years were either negatively or not correlated (year to year correlations: strength: $r_{NAO} =$

-0.016, $r_{Salmon} = 0.024$; clustering coefficient: $r_{NAO} = -0.307$, $r_{Salmon} = -0.399$; closeness: $r_{NAO} = -0.144$,

$r_{Salmon} = -0.139$).

In summary, months with higher salmon numbers led to higher strength and higher closeness.

Meanwhile, the NAO index did not clearly affect any network trait. Individuals differed in how their

strength changed in response to salmon abundance, and they also showed variation in responses of

closeness to the NAO index, but they did not differ in how their clustering coefficient changed for

either variable. Clustering coefficient and closeness were slightly repeatable across environments

but strength was not.

Discussion

We explored whether bottlenose dolphin social behaviour responded to environmental variation.

Social behaviour responded to variation in food availability, with a measure of connectedness to the

wider network (closeness) decreasing at higher salmon abundances at a yearly scale, and both

overall gregariousness (strength) and closeness increasing at higher salmon abundances at a

monthly scale. Clustering of the local social environment, clustering coefficient, also showed an

increase with monthly salmon abundance but this trend was statistically unclear. In contrast, social

behaviours showed no population-level responses to climatic change at either scale. In addition, we

found that individuals showed consistent differences in mean clustering coefficient, especially at the

yearly scale, but not in individual plasticities, while strength and closeness showed some variation in

individual plasticity but limited consistent differences in mean behaviour. This plasticity was often,

but not always, negatively associated with mean behaviour, causing among individual differences to

typically be greater for low values of salmon abundance and the NAO index.

Months of higher salmon abundance led to increase in all three social behaviours, although this was

not statistically clear for clustering coefficient (see Fig. 4 for a comparison of networks between

months with low vs. high salmon abundance). This indicates that dolphins were increasing all kinds

of social associations in response to increased immediate food availability. Similar results have been

found in spotted hyenas (*Crocuta crocuta*) where seasons with high prey availability have denser

social networks [77], and in Grant's gazelles (*Nanger granti*), where increased rainfall (and so food

availability) led to higher closeness scores [78]. It is presumed that an increase in social interactions

at higher food availability is facilitated by a reduction in the intensity of resource competition

[reviewed in: ,6,see also: ,79]. Therefore, in our study system higher rates of social interaction may

be beneficial for non-foraging reasons, such as mating. Both mating and calving seasons

approximately align with the months of highest salmon abundance in the summer (Fig. S4),

supporting this suggestion. Additionally, dolphins could move into areas where food availability is

especially high [42,80], causing more individuals to be seen together and therefore inferred social

networks to be denser, even if actual rates of social interaction are not changing, or only changing as

a byproduct. Finally, it is possible that months with fewer salmon also differ in an unidentified

variable which causes dolphins to group less. However, it is unlikely that this unidentified factor is

predation threat that changes month to month, as predators are absent in this area [81]. Therefore,

a change in social behaviour and/or movement related to the seasonal availability of salmon,

perhaps influenced by reproductive behaviour, seems the most likely.

Interestingly, the change in closeness at the yearly scale was in opposite direction to that at the

monthly scale, showing a decrease with increased salmon abundance, as well as being a

considerably stronger effect (standardised effect size of -0.258 compared to 0.087). A decrease in

yearly closeness suggests dolphins are more poorly connected to distant parts of the network when

salmon are more abundant. This decrease might be indicative of a reduced need to travel long

distances to find food, generating a network with a smaller diameter and so lower individual

closeness [see also 82 who demonstrated that more patchy food availability increases network

connectedness]. This effect might not be apparent at the monthly scale as the increase in strength in

months of high salmon abundance could also lead to higher closeness scores. Whether this is the

case or not, the fact that changes in different timescales can be in opposite directions is intriguing

and should be kept in mind when attempting to generalise results from one timescale to another.

**Figure 4**. Plots of dolphin social networks in a month of low salmon abundance (May 1990) and a

month of high salmon abundance (August 2007). Circles are individual dolphins and grey lines

indicate associations i.e., those seen in the same group at least once in that month.

Despite the changes in response to salmon abundance, we did not see any responses to climatic

variation (the NAO index) on a yearly or monthly scale. The NAO index varied considerably at both

scales, ranging from -1.15 to 1.08 at the yearly scale and -3.18 to 2.12 at the monthly scale (Fig. S3),

hence a lack of necessary variability seems unlikely. Lusseau *et al*. [33] also observed no variation in

group size in our study population in response to contemporary variation in the NAO index (they did

see an effect at a two-year lag, likely mediated by food availability). The robustness or inflexibility of

social behaviour in response to variation in climatic conditions might indicate that the variation in

the NAO index is inconsequential and so they have no need to respond to it. Additionally or

alternatively, dolphins may change other phenotypes, such as foraging behaviour or metabolism, to

cope with this stressor [31], leaving social behaviour unchanged. Finally, local conditions might be

more relevant to dolphin behaviour, as opposed to the regional conditions summarised by the NAO

index. For example, movements of bottlenose dolphins depend on tidal currents and fronts [83], and

changes to these might be important for their social behaviour.

Alongside the plasticity at the population level in strength and closeness, these traits also showed

individual variation in plasticity (although strength only showed this for salmon abundance, and

closeness only at the yearly level). Therefore, even if the population as a whole showed no overall

change for some trait-environmental variable combinations, some individuals might still show an

increase in their overall gregariousness and/or connectedness to wider parts of the network, while

others a decrease. Variation in individual plasticity leads to environment-dependent repeatability

(and possibly heritability), can dampen population responses to environmental variability, and

enhance population persistence [18]. For example, due to the negative intercept-slope correlation,

individual strength shows the most among-individual variation at low salmon abundances (the

approximate marginal repeatability of strength at the monthly scale two standard deviations below

the mean salmon abundance was 0.124, compared to 0.02 at the mean). If strength is linked to

foraging strategy, for instance if individuals with more social connections have access to more

information about prey availability [82], a wider range of social phenotypes could increase the

possibility that at least some individuals are successful despite low food availability. Determining the

genetic basis and importance of early life conditions for the development of these different

responses [12] and how this variation impacts population dynamics [84] is key. Models projecting

how these traits in dolphin populations will change in the future should account for both among-

individual variation in mean behaviour and in behavioural plasticity. Additionally, Kebke *et al*. [31]

suggest that cetacean ranging and foraging behaviour may well be under selection for increased

plasticity as environments change, and so estimating selection on both means and plasticities of

behaviours is a logical and important next step.

In contrast, clustering coefficient showed no individual variation in plasticity at any timescale-

environmental variable combination. Clustering coefficient at the yearly scale was the only trait with

more than slight repeatability, indicating some individual consistency [see also 85 who found

consistency over lifetime in the social behaviours of Indo-Pacific bottlenose dolphins, *T. aduncus*].

Therefore, plasticity might be more limited with individuals keeping the same pattern of local

connections across environmental conditions. As clustering coefficient depends on the frequency of

connections among triads, an individual's trait value cannot change without impacting the trait value

of others. This interdependence may then constrain the degree of plasticity possible at the individual

level. There is no evidence for male alliances in this population [86], and so determining what these

clusters of individuals represent and why they might be so stable would be useful.

In conclusion, we found that, at the population level, individual dolphin social behaviour is more

responsive to variation in food availability than climatic variation, with this being particularly

apparent at the monthly scale. We observed that individuals increased overall gregariousness and

connectedness to wider parts of the network in months of higher salmon abundance. In contrast,

dolphins decreased their connectedness to wider parts of the network in years of high salmon

availability. Traits that tended to show higher repeatability tended to show limited individual

variation in plasticity, although there was considerable variation in this trend. As such, whether

individual heterogeneity in both mean and plasticity in behaviour needs to be accounted for when

predicting species responses to environmental change might have to be considered on a case-by-

case basis, and individual plasticities as well as means may be targets of selection and hence

evolvable.

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## Statement of authorship

DF conceived the idea for the manuscript, BC provided data, DF ran analyses with guidance from BC, DF wrote first draft of the manuscript and both authors contributed to revisions.

## Data accessibility statement

Data and R code used for the analysis are available at [76].

## References

1. Snyder-Mackler N *et al.* 2020 Social determinants of health and survival in humans and other animals. *Science* **368**. (doi:10.1126/science.aax9553)

2. Moore AJ, Brodie EDI, Wolf JB. 1997 Interacting phenotypes and the evolutionary process: I. Direct and indirect genetic effects of social interactions. *Evolution* **51**, 1352–1362.

3.  Wolf JB, Brodie III ED, Moore AJ. 1999 Interacting Phenotypes and the Evolutionary Process. II.
Selection Resulting from Social Interactions. *The American Naturalist* **153**, 254–266.
(doi:10.1086/303168)

4.  McGlothlin JW, Moore AJ, Wolf JB, Brodie ED. 2010 Interacting phenotypes and the evolutionary
process. III. Social evolution. *Evolution; international journal of organic evolution* **64**, 2558–74.
(doi:10.1111/j.1558-5646.2010.01012.x)

5.  Komdeur J, Ma L. 2021 Keeping up with environmental change: The importance of sociality.
*Ethology* **127**, 790–807. (doi:10.1111/ETH.13200)

6.  Fisher DN, Kilgour RJ, Siracusa ER, Foote JR, Hobson EA, Montiglio PO, Saltz JB, Wey TW, Wice
EW. 2021 Anticipated effects of abiotic environmental change on intraspecific social
interactions. *Biological Reviews* **96**, 2661–2693. (doi:10.1111/brv.12772)

7.  White DJ, Gersick AS, Snyder-Mackler N. 2012 Social networks and the development of social
skills in cowbirds. *Philosophical Transactions of the Royal Society of London B: Biological Sciences*
**367**.

8.  Ghalambor CK, McKay JK, Carroll SP, Reznick DN. 2007 Adaptive versus non-adaptive phenotypic
plasticity and the potential for contemporary adaptation in new environments. *Functional*
*Ecology* **21**, 394–407. (doi:10.1111/j.1365-2435.2007.01283.x)

9.  Abrahms B, Rafiq K, Jordan NR, McNutt JW. 2022 Long-term, climate-driven phenological shift in
a tropical large carnivore. *Proceedings of the National Academy of Sciences* **119**.
(doi:10.1073/PNAS.2121667119)

10. Nussey DH, Wilson AJ, Brommer JE. 2007 The evolutionary ecology of individual phenotypic
plasticity in wild populations. *Journal of Evolutionary Biology* **20**, 831–844. (doi:10.1111/j.1420-
9101.2007.01300.x)

11. Dingemanse NJ, Kazem AJN, Réale D, Wright J. 2010 Behavioural reaction norms: animal
personality meets individual plasticity. *Trends in Ecology & Evolution* **25**, 81–89.
(doi:10.1016/j.tree.2009.07.013)

12. Stamps JA. 2016 Individual differences in behavioural plasticities. *Biological Reviews* **91**, 534–
567. (doi:10.1111/brv.12186)

13. Stamps J, Biro P. 2016 Personality and individual differences in plasticity. *Current Opinion in*
*Behavioral Sciences*

14. Fisher DN, David M, Tregenza T, Rodriguez-Munoz R. 2015 Dynamics of among-individual
behavioral variation over adult lifespan in a wild insect. *Behavioral Ecology* **26**, 975–985.
(doi:10.1093/beheco/arv048)

15. Brommer JE. 2013 Variation in plasticity of personality traits implies that the ranking of
personality measures changes between environmental contexts: calculating the cross-
environmental correlation. *Behavioral Ecology and Sociobiology* **67**, 1709–1718.
(doi:10.1007/s00265-013-1603-9)

16. Vindenes Y, Langangen Ø. 2015 Individual heterogeneity in life histories and eco-evolutionary
dynamics. *Ecology Letters* **18**, 417–432. (doi:10.1111/ele.12421)

17. Hamel S, Gaillard J-M, Yoccoz NG. 2018 Introduction to: Individual heterogeneity - the causes
and consequences of a fundamental biological process. *Oikos* **127**, 643–647.
(doi:10.1111/oik.05222)

18. Dingemanse NJ, Wolf M. 2013 Between-individual differences in behavioural plasticity within
populations: causes and consequences. *Animal Behaviour* **85**, 1031–1039.
(doi:10.1016/j.anbehav.2012.12.032)

19. Seebacher F, Little AG. 2021 Plasticity of Performance Curves in Ectotherms: Individual Variation
Modulates Population Responses to Environmental Change. *Frontiers in Physiology* **12**.

20. Chevin L-M. 2013 Genetic constraints on adaptation to a changing environment. *Evolution* **67**,
708–721. (doi:10.1111/j.1558-5646.2012.01809.x)

21. Araya-Ajoy YG, Dingemanse NJ. 2017 Repeatability, heritability, and age-dependence of
seasonal plasticity in aggressiveness in a wild passerine bird. *Journal of Animal Ecology* **86**, 227–
238. (doi:10.1111/1365-2656.12621)

22. Nussey DH, Postma E, Gienapp P, Visser ME. 2005 Selection on Heritable Phenotypic Plasticity in
a Wild Bird Population. *Science* **310**, 304–306.

23. Brommer JE. 2013 Phenotypic plasticity of labile traits in the wild. *Current Zoology* **59**, 485–505.

24. Hoffmann AA, Sgrò CM. 2011 Climate change and evolutionary adaptation. *Nature* **470**, 479–
485. (doi:10.1038/nature09670)

25. Simmonds MP, Isaac SJ. 2007 The impacts of climate change on marine mammals: early signs of
significant problems. *Oryx* **41**, 19–26. (doi:10.1017/S0030605307001524)

26. Tyack PL. 2008 Implications for marine mammals of large-scale changes in the marine acoustic
environment. *Journal of Mammalogy* **89**, 549–558. (doi:10.1644/07-mamm-s-307r.1)

27. MacLeod CD. 2009 Global climate change, range changes and potential implications for the
conservation of marine cetaceans: a review and synthesis. *Endangered Species Research* **7**, 125–
136. (doi:10.3354/esr00197)

28. Evans PG, Bjørge A. 2013 Impacts of climate change on marine mammals. *Marine Climate
Change Impacts Partnership (MCCIP) Science Review* **2013**, 134–148.

29. Avila IC, Kaschner K, Dormann CF. 2018 Current global risks to marine mammals: Taking stock of
the threats. *Biological Conservation* **221**, 44–58. (doi:10.1016/j.biocon.2018.02.021)

30. van Weelden C, Towers JR, Bosker T. 2021 Impacts of climate change on cetacean distribution,
habitat and migration. *Climate Change Ecology* **1**, 100009. (doi:10.1016/j.ecochg.2021.100009)

31. Kebke A, Samarra F, Derous D. 2022 Climate change and cetacean health: impacts and future
directions. *Philosophical Transactions of the Royal Society B: Biological Sciences* **377**.

32. Kershaw JL, Hall AJ. 2019 Mercury in cetaceans: Exposure, bioaccumulation and toxicity. *Science
of The Total Environment* **694**, 133683. (doi:10.1016/j.scitotenv.2019.133683)

33. Lusseau D, Williams R, Wilson B, Grellier K, Barton TR, Hammond PS, Thompson PM. 2004
Parallel influence of climate on the behaviour of Pacific killer whales and Atlantic bottlenose
dolphins. *Ecology Letters* **7**, 1068–1076. (doi:10.1111/j.1461-0248.2004.00669.x)

34. Pörtner HO, Gutt J. 2016 Impacts of Climate Variability and Change on (Marine) Animals:
Physiological Underpinnings and Evolutionary Consequences. *Integrative and Comparative*
*Biology* **56**, 31–44. (doi:10.1093/icb/icw019)

35. Mitchell D, Snelling EP, Hetem RS, Maloney SK, Strauss WM, Fuller A. 2018 Revisiting concepts of
thermal physiology: Predicting responses of mammals to climate change. *Journal of Animal*
*Ecology* **87**, 956–973. (doi:10.1111/1365-2656.12818)

36. Domenici P, Claireaux G, McKenzie D j. 2007 Environmental constraints upon locomotion and
predator–prey interactions in aquatic organisms: an introduction. *Philosophical Transactions of*
*the Royal Society B: Biological Sciences* **362**, 1929–1936. (doi:10.1098/rstb.2007.2078)

37. Methion S, Giralt Paradell O, Padín XA, Corrège T, Díaz López B. 2023 Group size varies with
climate and oceanographic conditions in bottlenose dolphins. *Mar Biol* **170**, 7.
(doi:10.1007/s00227-022-04154-4)

38. Sprogis KR, Christiansen F, Wandres M, Bejder L. 2018 El Niño Southern Oscillation influences
the abundance and movements of a marine top predator in coastal waters. *Global Change*
*Biology* **24**, 1085–1096. (doi:10.1111/gcb.13892)

39. Arso Civil M, Quick NJ, Mews S, Hague E, Cheney BJ, Thompson PM, Hammond PS. 2021
Improving understanding of bottlenose dolhin movements along the east coast of Scotland.

40. Wilson B, Hammond PS, Thompson PM. 1999 Estimating Size and Assessing Trends in a Coastal
Bottlenose Dolphin Population. *Ecological Applications* **9**, 288–300. (doi:10.1890/1051-
0761(1999)009[0288:ESAATI]2.0.CO;2)

41. Cheney B *et al.* 2013 Integrating multiple data sources to assess the distribution and abundance
of bottlenose dolphins Tursiops truncatus in Scottish waters. *Mammal Review* **43**, 71–88.
(doi:10.1111/j.1365-2907.2011.00208.x)

42. Arso Civil M, Quick NJ, Cheney B, Pirotta E, Thompson PM, Hammond PS. 2019 Changing
distribution of the east coast of Scotland bottlenose dolphin population and the challenges of
area-based management. *Aquatic Conservation: Marine and Freshwater Ecosystems* **29**, 178–
196. (doi:10.1002/aqc.3102)

43. Cheney B *et al.* 2014 Long-term trends in the use of a protected area by small cetaceans in
relation to changes in population status. *Global Ecology and Conservation* **2**, 118–128.
(doi:10.1016/j.gecco.2014.08.010)

44. Würsig B, Jefferson TA. 1990 Methods of photo-identification for small cetaceans. , 43–52.

45. Silk MJ, Jackson AL, Croft DP, Colhoun K, Bearhop S. 2015 The consequences of unidentifiable
individuals for the analysis of an animal social network. *Animal Behaviour* **104**, 1–11.
(doi:10.1016/j.anbehav.2015.03.005)

46. Cheney BJ, Dale J, Thompson PM, Quick NJ, Barbara Cheney CJ. 2022 Spy in the sky: a method to
identify pregnant small cetaceans. *Remote Sensing in Ecology and Conservation*
(doi:10.1002/RSE2.258)

47. Whitehead H, Dufault S. 1999 Techniques for Analyzing Vertebrate Social Structure Using
Identified Individuals: Review and Recommendations. *Advances in the Study of Behavior* **28**.

48. Farine DR. 2013 Animal social network inference and permutations for ecologists in R using
asnipe. *Methods in Ecology and Evolution* **4**, 1187–1194. (doi:10.1111/2041-210X.12121)

49. Cairns SJ, Schwager SJ. 1987 A comparison of association indices. *Animal Behaviour* **35**, 1454–
1469. (doi:10.1016/S0003-3472(87)80018-0)

50. Franks DW, Ruxton GD, James R. 2010 Sampling animal association networks with the gambit of
the group. *Behavioral Ecology and Sociobiology* **64**, 493–503. (doi:10.1007/s00265-009-0865-8)

51. Connor RC, Smolker RA, Richards AF. 1992 Two levels of alliance formation among male
bottlenose dolphins (Tursiops sp.). *Proceedings of the National Academy of Sciences of the
United States of America* **89**, 987–90.

52. Connor RC, Smolker R, Bejder L. 2006 Synchrony, social behaviour and alliance affiliation in
Indian Ocean bottlenose dolphins, Tursiops aduncus. *Animal Behaviour* **72**, 1371–1378.
(doi:10.1016/j.anbehav.2006.03.014)

53. Hurrell JW. 1995 Decadal Trends in the North Atlantic Oscillation: Regional Temperatures and
Precipitation. *Science* **269**, 676–679. (doi:10.1126/science.269.5224.676)

54. Stenseth NChr, Ottersen G, Hurrell JW, Mysterud A, Lima M, Chan K, Yoccoz NG, Ådlandsvik B.
2003 Review article. Studying climate effects on ecology through the use of climate indices: the
North Atlantic Oscillation, El Niño Southern Oscillation and beyond. *Proceedings of the Royal
Society of London. Series B: Biological Sciences* **270**, 2087–2096. (doi:10.1098/rspb.2003.2415)

55. Gordo O, Barriocanal C, Robson D. 2011 Ecological Impacts of the North Atlantic Oscillation
(NAO) in Mediterranean Ecosystems. In *Hydrological, Socioeconomic and Ecological Impacts of
the North Atlantic Oscillation in the Mediterranean Region* (eds SM Vicente-Serrano, RM Trigo),
pp. 153–170. Dordrecht: Springer Netherlands. (doi:10.1007/978-94-007-1372-7_11)

56. Wan X, Holyoak M, Yan C, Le Maho Y, Dirzo R, Krebs CJ, Stenseth NChr, Zhang Z. 2022 Broad-
scale climate variation drives the dynamics of animal populations: a global multi-taxa analysis.
*Biological Reviews* **97**, 2174–2194. (doi:10.1111/brv.12888)

57. Pereira JM, Paiva VH, Ceia FR, Ramos JA. 2020 Facing extremes: Cory's shearwaters adjust their
foraging behaviour differently in response to contrasting phases of North Atlantic Oscillation.
*Reg Environ Change* **20**, 77. (doi:10.1007/s10113-020-01662-1)

58. Víkingsson GA *et al.* 2015 Distribution, abundance, and feeding ecology of baleen whales in
Icelandic waters: have recent environmental changes had an effect? *Frontiers in Ecology and
Evolution* **3**.

59. Thorne LH, Nye JA. 2021 Trait-mediated shifts and climate velocity decouple an endothermic
marine predator and its ectothermic prey. *Sci Rep* **11**, 18507. (doi:10.1038/s41598-021-97318-z)

60. Youngson AF, MacLean JC, Fryer RJ. 2002 Rod catch trends for early-running MSW salmon in
Scottish rivers (1952–1997): divergence among stock components. *ICES Journal of Marine
Science* **59**, 836–849. (doi:10.1006/jmsc.2002.1195)

61. Thorley JL, Eatherley DMR, Stephen AB, Simpson I, MacLean JC, Youngson AF. 2005 Congruence
between automatic fish counter data and rod catches of Atlantic salmon (Salmo salar) in Scottish
rivers. *ICES Journal of Marine Science* **62**, 808–817. (doi:10.1016/j.icesjms.2005.01.016)

62. Janik VM. 2000 Food–related bray calls in wild bottlenose dolphins (Tursiops truncatus).
*Proceedings of the Royal Society of London. Series B: Biological Sciences* **267**, 923–927.
(doi:10.1098/rspb.2000.1091)

63. Santos MB, Pierce GJ, Reid RJ, Patterson I a. P, Ross HM, Mente E. 2001 Stomach contents of
bottlenose dolphins (Tursiops truncatus) in Scottish waters. *Journal of the Marine Biological
Association of the United Kingdom* **81**, 873–878. (doi:10.1017/S0025315401004714)

64. R Development Core Team, Team RDC. 2016 R: A language and environment for statistical
computing. *R Foundation for Statistical Computing*. **1**, 409. (doi:10.1007/978-3-540-74686-7)

65. Brooks ME, Kristensen K, van Benthem KJ, Magnusson A, Berg CW, Nielsen A, Skaug HJ,
Maechler M, Bolker BM. 2017 glmmTMB Balances speed and flexibility among packages for
modeling. *The R Journal* **9**, 378–400.

66. Hart JDA, Weiss MN, Brent LJN, Franks DW. 2021 Common Permutation Methods in Animal
Social Network Analysis Do Not Control for Non-independence. *bioRxiv*

67. Shane SH, Wells RS, Würsig B. 1986 Ecology, Behavior and Social Organization of the Bottlenose
Dolphin: A Review. *Marine Mammal Science* **2**, 34–63. (doi:10.1111/j.1748-
7692.1986.tb00026.x)

68. Mann J, Stanton MA, Patterson EM, Bienenstock EJ, Singh LO. 2012 Social networks reveal
cultural behaviour in tool-using dolphins. *Nat Commun* **3**, 980. (doi:10.1038/ncomms1983)

69. Arso Civil M, Cheney B, Quick NJ, Islas-Villanueva V, Graves JA, Janik VM, Thompson P,
Hammond PS. 2019 Variations in age- and sex-specific survival rates help explain population
trend in a discrete marine mammal population. *Ecology and Evolution* **9**, 533–544.

70. Schielzeth H. 2010 Simple means to improve the interpretability of regression coefficients.
*Methods in Ecology and Evolution* **1**, 103–113. (doi:10.1111/j.2041-210X.2010.00012.x)

71. Warton D, Hui F. 2011 The arcsine is asinine: the analysis of proportions in ecology. *Ecology*

72. Fox J, Weisberg S. 2019 *An R Companion to Applied Regression*. Thousand Oaks (CA): Sage.

73. Dushoff J, Kain MP, Bolker BM. 2019 I can see clearly now: Reinterpreting statistical significance.
*Methods in Ecology and Evolution* **10**, 756–759. (doi:10.1111/2041-210X.13159)

74. Self SG, Liang KY. 1987 Asymptotic properties of maximum likelihood estimators and likelihood
ratio tests under nonstandard conditions. *Journal of the American Statistical Association* **82**,
605–610. (doi:10.1080/01621459.1987.10478472)

75. Schielzeth H, Nakagawa S. 2022 Conditional repeatability and the variance explained by reaction
norm variation in random slope models. *Methods in Ecology and Evolution* **13**, 1214–1223.
(doi:10.1111/2041-210X.13856)

76. Fisher DN, Cheney BJ. 2023 Dataset and R code for 'Dolphin social phenotypes vary in response
to food availability but not climate', available at https://figshare.com/s/e22d4f0b1d89915aa3f2.

77. Holekamp KE, Smith JE, Strelioff CC, Van Horn RC, Watts HE. 2012 Society, demography and
genetic structure in the spotted hyena. *Molecular Ecology* **21**, 613–632. (doi:10.1111/j.1365-
294X.2011.05240.x)

78. Williams AE, Worsley-Tonks KEL, Ezenwa VO. 2017 Drivers and consequences of variation in
individual social connectivity. *Animal Behaviour* **133**, 1–9.

79. Lin YH, Chan SF, Rubenstein DR, Liu M, Shen SF. 2019 Resolving the paradox of environmental
quality and sociality: The ecological causes and consequences of cooperative breeding in two
lineages of birds. *American Naturalist* **194**, 207–216. (doi:10.1086/704090)

80. Hastie GD, Wilson B, Wilson LJ, Parsons KM, Thompson PM. 2004 Functional mechanisms
underlying cetacean distribution patterns: hotspots for bottlenose dolphins are linked to
foraging. *Marine Biology* **144**, 397–403. (doi:10.1007/s00227-003-1195-4)

81. Cheney BJ, Thompson PM, Cordes LS. 2019 Increasing trends in fecundity and calf survival of
bottlenose dolphins in a marine protected area. *Scientific Reports* **9**, 1767. (doi:10.1038/s41598-
018-38278-9)

82. Tanner CJ, Jackson AL. 2012 Social structure emerges via the interaction between local ecology
and individual behaviour. *The Journal of animal ecology* **81**, 260–7. (doi:10.1111/j.1365-
2656.2011.01879.x)

83. Bailey H, Thompson P. 2010 Effect of oceanographic features on fine-scale foraging movements
of bottlenose dolphins. *Marine Ecology Progress Series* **418**, 223–233. (doi:10.3354/meps08789)

84. Vindenes Y, Engen S, Sæther B. 2008 Individual Heterogeneity in Vital Parameters and
Demographic Stochasticity. *The American Naturalist* **171**, 455–467. (doi:10.1086/528965)

85. Evans T, Krzyszczyk E, Frère C, Mann J. 2021 Lifetime stability of social traits in bottlenose
dolphins. *Commun Biol* **4**, 1–8. (doi:10.1038/s42003-021-02292-x)

86. Wilson DRB. 1995 The ecology of bottlenose dolphins in the Moray Firth, Scotland : a population
at the northern extreme of the species' range. Ph.D., University of Aberdeen. See
https://ethos.bl.uk/OrderDetails.do?uin=uk.bl.ethos.261914.