Moss species, sections, and land use type drive moss-associated microbial taxonomy and functions in boreal forests

Sarah Ishak[1*], Maria Faticov[1], Jonathan Rondeau-Leclaire[1], Sébastien Roy[12] and Isabelle Laforest-Lapointe[123]

[1]Département de biologie, Université de Sherbrooke, Sherbrooke, Québec, Canada

[2]Centre Sève, Université de Sherbrooke, Sherbrooke, Québec, Canada

[3]Centre d'Étude de la Forêt, Université du Québec à Montréal, Montréal, Québec, Canada

*Corresponding author: s.isaac1905@protonmail.com

ORCID NUMBERS

Sarah Ishak	(0000-0003-3412-3642)
Maria Faticov	(0000-0001-8206-9332)
Jonathan Rondeau-Leclaire	(0000-0002-4433-3566)
Sébastien Roy	(0000-0001-7183-0691)
Isabelle Laforest-Lapointe	(0000-0003-3630-9475)

Word counts (current total 5918 without sub-headings or citations)				
Summary	Introduction	Methods	Results	Discussion
197	1272	1297	1440	1909
Number of figures/tables/supporting information				
Figures (main text)	Tables (main text)	Supp. figures	Supp. tables	
5	0	1	9	

Summary

• Moss microbial communities play important roles for ecosystem processes in boreal
forests. Host moss identity and leaf litter presence can affect microbial community
structures of mosses. However, the extent to which host-specific characteristics and
land use type affect taxonomic and functional profiles of microbial communities of
boreal mosses is still poorly understood.

• We used shotgun metagenomic sequencing to characterize microbial taxonomy and
metabolic KEGG ortholog (KO) profiles between green and brown sections of five
moss species across different natural sites, and two moss species between a natural site
and a mine site in Eeyou Istchee, Québec, Canada.

• Our results demonstrate that the abundance of nitrogen metabolic genes differ between
moss sections and that mosses from natural and mine environment associate with
different microbial taxa and KO profiles. Importantly, conditions at the mine site appear
to favor microbial taxa that can tolerate perturbated environments, including taxa that
can oxidize sulfur and participate in biofilm formation.

• Overall, our results highlight that moss section, moss species identity, and land-use type
are strong drivers of diversity and community structure of moss-associated microbial
taxa and metabolic genes. These findings could have major implications for boreal
forests facing climate change and anthropogenic pressures.

Key words: bacteria, fungi, boreal forests, KEGG profiles, metagenomics, microbial ecology,
mining, mosses, shotgun sequencing.

Introduction

Boreal forests are one of the largest biomes covering 29% of land surfaces (Kayes & Mallik, 2020) and storing at least 30% of the Earth's carbon (Walker *et al.*, 2019). This terrestrial ecosystem is habitat to a diversity of understory plants including lichens, and bryophytes (small non-vascular plants including mosses, liverworts, and hornworts). Ubiquitously distributed in boreal forests, mosses are crucial contributors to two key ecosystem processes: nitrogen fixation and carbon sequestration (Turetsky, 2003). These processes are also facilitated by symbiotic micro-organisms colonizing mosses (Holland-Moritz *et al.*, 2018; Ininbergs *et al.*, 2011). In boreal ecosystems, moss-microbe interactions contribute up to 50% of nitrogen inputs to the forest and play crucial roles in facilitating carbon degradation in peatlands (Lin *et al.*, 2014a; Rousk & Michelsen, 2017). Driven by the pressures of the Anthropocene, boreal ecosystems face a temperature increase of ~2-5°C before the end of the century which could alter moss-associated processes with major implications for global N and C cycles (Masson-Delmotte *et al.*, 2021). Considering the strong impact of climate change on boreal forests combined with the ever-increasing anthropogenic impacts on land-use (e.g., urbanization and mining), improving our understanding of the drivers and functions of moss-microbe interactions is imperative to identify and mitigate the risks of altering this precious habitat (Jonsson *et al.*, 2015; Royles & Griffiths, 2015).

In the last two decades, substantial molecular evidence have provided insights into the drivers of moss microbial communities (through the genetic material recovered from the microbiota, including dead, dormant, and active cells) and how these moss-microbe interactions can influence ecosystem functions (Stuart *et al.*, 2020a). The sequencing of moss microbial communities allowed for the identification of microbial taxonomy and consequently the inference of their ecological functions, e.g., nitrogen fixation and phosphorus solubilization (Lin *et al.*, 2014a). To date, cyanobacteria are the most commonly-studied microbial group associated with mosses due to their nitrogen-fixing activities (Rousk *et al.*, 2013). Yet, other non-cyanobacterial microbial families such as *Acidobacteraceae, Acetobacteraceae*, and *Bradyrhizobiaceae* have been much less studied, despite their ability to fix nitrogen (Holland-Moritz *et al.*, 2018; Vile *et al.*, 2014). Outside of carbon and nitrogen metabolism, moss-associated micro-organisms likewise show the capacity for (1) phosphorus acquisition (Lin *et al.*, 2014b); (2) antagonism to fungal and bacterial pathogens (Shcherbakov *et al.*, 2013); (3) biofilm formation (Bragina *et al.*, 2014); (4) quorum sensing (Vesty *et al.*, 2020); and

(5) organic matter decomposition (Turetsky *et al.*, 2012). These functions can be attributed not

only to bacteria, but fungi as well. In fact, several studies have demonstrated that moss-fungal

interactions can influence ecosystem functions, such as nitrogen fixation and utilization (Chen

et al., 2019; Davey *et al.*, 2012). Overall, moss-microbe associations are crucial in various

biogeochemical processes. Importantly, their interactions are not static. Thus, understanding

the determinants of moss-associated microbial community assembly and function could be key

to facilitate predictions of their roles in boreal ecosystems in the Anthropocene.

Moss microbial communities can be affected by biotic and abiotic factors in their local

ecosystem (Jean *et al.*, 2020; Klarenberg *et al.*, 2021). Moss-associated cyanobacteria can be

affected by light exposure and moisture levels (Kox *et al.*, 2020; Liebner *et al.*, 2011). McLaren

et al. (2017) demonstrated that the expansion of temperate shrubs into boreal ecosystems in

response to higher temperatures results in a higher density of labile shrub litter. This, in turn,

alters the assembly of moss bacterial communities. Moss phylogeny is a major driver of moss

bacterial community structure (Holland-Moritz *et al.*, 2018). Moss diversity can reach up to as

many as 130 species in Canadian boreal forests (Macdonald & Fenniak, 2007). Yet a vast

number of studies on moss-microbe interactions have focused on dominant boreal mosses such

as the feathermoss *Pleurozium schreberi* (Elzen *et al.*, 2020) and peat mosses of the genus

Sphagnum sp. (Kostka *et al.*, 2016). Certainly, other mosses such as the acrocarpous mosses

Dicranum undulatum and *Polytrichum commune,* while less abundant, could have an impact

on ecosystem functions through their interactions with microbes (Rousk *et al.*, 2015). Finally,

physiological differences between green and brown sections can affect moss microbial

communities. Moss tissue can be separated along their vertical gradient into distinct sections:

an above-ground living "green" section, and an underlying, decaying "brown" section (Lindo

& Gonzalez, 2010). These sections can show distinctions in both bacterial and fungal

taxonomic diversity (Arróniz-Crespo *et al.*, 2022; Chen *et al.*, 2019). The green section is

thought to be photosynthetically active and is often in focus when studying bacterial

communities associated with mosses. The green section is also associated with nitrogen fixing

cyanobacteria (Renaudin *et al.*, 2022a). Meanwhile, microbial communities associated with

brown sections are much less studied, even though they are hypothesized to be an important

source of nutrients for certain bacterial taxa (Shen *et al.*, 2018). To gain a comprehensive

perspective of moss-microbe interactions, it appears necessary to characterize how moss

sections differ in microbial community assembly and functions.

Up until recently, moss microbial functions have been mostly inferred by looking at
microbial taxonomy and associating it with measures of nitrogen fixation (Stuart *et al.*, 2020b;
Warshan *et al.*, 2017). Recently, metagenomic analyses have enabled functional profiling of
microbial communities beyond the inference of functions from taxonomy (Quince *et al.*, 2017).
Instead, this molecular approach focuses on total genomic material present in a sample and
allows for the characterization of microbial genes, which in return provides information on
microbial metabolic profiles. This improved understanding of metabolic profiles across
different moss species could help to predict functional changes occurring as climate change
and human activities impact boreal ecosystems. Much remains to be done to provide a global
portrait of moss-microbe interactions and functions beyond a few dominant moss species and
the sole microbial kingdom of bacteria. Presently, most moss-microbiome studies have been
confined to amplicon-based metagenomic analysis of specific microbial kingdoms, such as
bacteria or fungi, and do not necessarily explore the functions of the moss microbiome. To gain
a comprehensive perspective of moss-microbe interactions, it appears necessary to characterize
moss microbial community assembly and functions, especially between different moss
sections.

Alongside climate change, boreal ecosystems are also frequently perturbed by human
activities, such as mining. Mining can result in one or a combination of consequences – from
biologically toxic concentrations of metals in soils to moisture loss and pH imbalances (Vareda
et al., 2019; Wong, 2003). Mosses have been found on mine sites (Monaci *et al.*, 2021) and
were shown to have positive impacts on degraded soils (Cheng *et al.*, 2021). They possess
characteristics including increased cation exchange capacity and weak hydro-phobic barriers
that enable their survival in mines (Antreich *et al.*, 2016). Mosses can also increase evaporation
rates during rainfall events, lowering the soil moisture content (Merlin *et al.*, 2019) and
mitigating desertification (Cao *et al.*, 2020; Pombubpa *et al.*, 2020). However, it remains
unclear whether the taxonomy and functions of moss-associated microbial communities are
distinct in this highly perturbed and unique environment. In vascular plants naturally growing
in mine sites, community structures of bacteria and fungi were found to be driven by plant
species and environmental factors, namely pH and humidity (Gagnon *et al.*, 2020a,b;
Klarenberg *et al.*, 2023). In moss microbiomes, pathways related to oxidative stress, such as
biofilm formation have been previously detected, which can contribute to resilience of moss
microbiomes in mine sites (Bragina *et al.*, 2014). Interestingly, a study by Goth et al. (2019)

demonstrated that the ability of moss microbes to fix nitrogen was not inhibited despite heavy metal pollution. Understanding the mechanisms of the moss microbiome in mine sites will be crucial in determining the capacity for mosses to facilitate reclamation of mine sites.

In this study, we quantified the relative influence of multiple drivers of moss microbial community taxonomical and functional profiles in boreal forests. Specifically, our objectives were to (1) provide a comprehensive portrait of moss-associated microbial communities beyond bacteria using shotgun metagenomic sequencing; (2) explore whether microbial taxonomy and functions differ among boreal moss sections as well as moss species; and (3) determine the taxonomical and functional profiles of moss microbial communities in natural and mine sites among multiple moss species. To this end, we investigated moss bacterial and fungal microbial communities in natural stands of boreal forest and at the Whabouchi mine site of the Eeyou-Istchee region of northern Québec, Canada. Supported by previous findings, we hypothesized that (i) location type (natural vs mine) will explain most of the variation in moss microbial community assembly and functions followed by section type (green vs brown) and moss species identity. We further hypothesize that (ii) green sections will harbor a higher relative abundance of cyanobacteria and nitrogen-fixing genes than the brown sections; and (iii) that there will be more functional genes related to biofilm formation in moss-associated microbial communities from the mine site in tolerating the stresses of the mine environment, compared to natural sites.

Materials and Methods

Study area

We conducted this study within the Eeyou-Istchee (James Bay) region of northern Québec (N51.7040721°, W76.0521982°) and the Whabouchi mine site (N51.68694°, W75.83597°; Fig. **1a**) (Nemaska Lithium Inc. *et al.*, 2014). All sampling locations were part of Québec's boreal forest biome which is dominated by Jack pine (*Pinus banksiana*) and black spruce (*Picea mariana*) (Moussaoui *et al.*, 2020). The annual average temperature of this region in 2021 was approximately 8.9°C, in June, and 11.3°C in September (Environment and Climate Change Canada Historical Climate Data, 2021). The soil of these boreal forests is mainly acidic and nutrient poor (Collin *et al.*, 2018). This project focused on five moss species: *Polytrichum juniperinum*, *Polytrichum commune*, *Polytrichum piliferum*, *Dicranum undulatum*, and *Tortella tortuosa* (Fig. **1b**).

 Study design

Sampling occurred over two periods in June and September of 2021. June samples were used

to compare the taxonomical and functional diversity of moss microbes from naturalized

environments and the Whabouchi mine site. September samples were used to characterize

taxonomical and functional profiles of moss microbes between green and brown moss sections.

Across the two periods, mosses were sampled from six different locations within the

Eeyou-Istchee (James Bay) region of northern Québec (Fig. **1a**). For the September sampling,

we randomly selected three microsites 5 to 10 meters apart (Fig. **1a**). In each of the microsites,

we further established 3-by-3-inch moss patches (Fig. **1a**). In September, a total of 69

individuals were sampled across *P. juniperinum*, *P. commune*, *P. piliferum*, and *D. undulatum*

from five sites in Eeyou Istchee (Table S1). Individual moss samples were further separated

into "green" and "brown" sections (determined visually, Fig. **1c**). Meanwhile, there was no

microsite sampling for June. Specifically, a total of 10 individuals of *P. juniperinum* and

T. tortuosa were sampled from one natural and one mine site in June (Table **S2**). Only green

parts of the moss were analyzed from these samples. During both sampling periods, mosses

were first visually identified to species level in the field and then confirmed with microscopy

in the laboratory. Moss samples were placed in 50 mL Falcon tubes and immediately stored at

-20°C until back to the laboratory. In the laboratory, samples were then stored at -80°C.

A breakdown of the sampling procedure can be viewed in Tables **S1-S2**.

DNA extraction and shotgun metagenomic sequencing

DNA extraction of moss samples were performed using modified protocols by Holland-Moritz,

et al. (2021) and Klarenberg, *et al.* (2021). In brief, 0.25g of each moss sample were initially

weighed out and divided into two sterile 2 mL bead beating tubes with three sterilised 2.3 mm

diameter stainless-steel beads (BioSpec Products, Bartlesville, Oklahoma) without extraction

buffer. Tubes were then frozen in liquid nitrogen for approximately 30 seconds and moss tissue

was broken down using the FisherbrandTM Bead Mill 24 Homogenizer for 30 seconds at 1.75

m/s. DNA was subsequently extracted using the protocol of QIAGEN DNEasy PowerSoil

DNA Extraction Kit with minor changes (see *Supplementary Materials* for details). All samples

were sent to Génome Québec (Montréal, Canada) for library preparation and metagenomic

sequencing. Ten metagenomic libraries were generated from green part of the mosses from

June samples, and 131 metagenomic libraries were generated from green and brown sections

from September samples. Seven samples from September were not included in the analysis due

to insufficient read quality. Thus, a total of 141 samples were sequenced on a single lane of the

Illumina NovaSeq6000 with 2×150bp chemistry at Génome Québec.

Assignment of taxonomy and functions

On average, library read numbers ranged from 10.4×10^6 to 22.5×10^6. Reads were quality-

filtered using KneadData v0.10.0 paired-end run with default parameters. To remove host

DNA, reads mapping to *Homo sapiens* and *Physcomitrella patens* (the closest available full

moss genome) were excluded using Bowtie 2.3.5 (Langmead *et al.*, 2019). Filtered reads were

input to Kraken 2.1.2 for taxonomical annotation using default parameters. Then, Bracken was

used for Bayesian re-estimation of abundance to obtain species tables which were used for

statistical analysis (Lu *et al.*, 2017). Next, a taxonomic profile report was generated from

KrakenTools, which was used to generate a custom subset of annotated genomes from the

ChocoPhlAn 201901b database. Annotated genomes together with the Uniref90 proteins

database were used to quantify gene families and metabolic pathways using the HUMAnN 3.0

pipeline (Beghini *et al.*, 2021). Finally, gene annotation step generated KEGG orthologs (KOs)

– genes that are homologous across different species, and are identified with K numbers

(Kanehisa *et al.*, 2016). KOs were used to explore the metabolic profiles of microbial

communities between green and brown moss sections, and among mosses from natural sites

and the Whabouchi mine site. To classify KOs into biologically meaningful categories, i.e.,

genes that are relevant for biogeochemical processes and stress responses in the moss

microbiome, we compiled a list from relevant pathways in the KEGG database. The final list

included pathways responsible for the nitrogen metabolism, photosynthesis, biofilm formation,

heat and cold adaptation, and oxidative stress.

Statistical analyses

All statistical analysis and visualization were performed in R v4.2.1. Each of four datasets

(microbial taxonomical and functional data from June and September samples) were handled

separately to answer slightly different questions (see Table **S3** for details). For univariate

responses, we ran linear models based on Shannon diversity using functions *lmer, lm*, and *glm*

from R base package. Prior to calculating Shannon diversity of microbial taxa, reads were

rarefied to 90% of the lowest sample sum of reads using *rarefy_even_depth* in the *phyloseq*

package to retain all samples while accounting for uneven sequencing depth (McMurdie &

Holmes, 2013; Wicaksono *et al.*, 2021). To calculate Shannon diversity of microbial KOs, reads were Hellinger transformed to correct for a double-zero effect (Franzetti *et al.*, 2020). We used Shapiro-Wilk from the base R package and Breusch-Pagan tests from the *bptest* package to determine the normality and homoskedasticity of Shannon values (Zeileis & Hothorn, 2002). Where variables were not normal, we transformed them using the Ordered Quantile Normalisation (ORQ) using the package *bestNormalize* (Peterson & Cavanaugh, 2020; Peterson, 2021). We then used the *Anova* function in the *car* package to assess significance of the models, and *emmeans* to estimate specific contrasts in our models (with Tukey adjustment). For the taxonomical and functional community assembly analyses, we first addressed uneven sequencing depth and heteroskedasticity, by transforming raw taxa counts using variance stabilizing transformations (VST) in the package *DESeq2* (Love *et al.*, 2014). To obtain community assembly matrices, for VST transformed data, we used Jaccard dissimilarity index calculated from presence-absence data (Tian *et al.*, 2019); while for Hellinger transformed data, we used the Euclidean dissimilarity index (Pakgohar *et al.*, 2021). Finally, *adonis2* (with the argument *by = "terms"*) in the vegan package was used to run PERMANOVAs (Oksanen *et al.*, 2022), every time making sure that the order of parameters didn't affect the effect size or significance.

To examine the relationship between microbial taxonomic and KO diversity and moss characteristics, we independently modelled Shannon diversity of microbial taxa (n=131) and microbial functions (n=131) as a function of the fixed effects '*moss section*', '*moss species identity*', and their interaction in September samples. We included "*microsite*" as a random effect to account for variation among sites. To explore the relationship between the taxonomic community assembly, the KEGG functional gene data, and moss characteristics of September samples, we independently modelled multivariate taxonomic (n=131) or KO (n=131) community assembly as a function of fixed effects '*moss section*', '*moss species identity*', and their interaction. We restricted permutations within blocks (i.e., '*microsite*') using the function *how* in the package *permute* (Simpson *et al.*, 2022). June samples (n=10) were used to examine whether mosses growing in natural sites and the Whabouchi mine site differed in their microbial taxonomic and functional diversity. We modelled Shannon diversity of microbial taxa (n=10) and KOs (n=10), as a function of fixed effects '*location*', '*moss species identity*', and their interaction. Multivariate analysis was then used to examine the relationship between multivariate taxonomic community assembly (n=10), KOs potential profiles (n=10) and

'location', 'moss species identity', and their interaction. Finally, to investigate which microbial
taxa and KOs differed in relative abundance between moss sections, the natural and mine
locations, as well as across moss species, we conducted differential abundance analysis with
the *ANCOM-BC* package without structural zeroes (Lin & Peddada, 2020). See Table S3 for
details on the data used, statistical model structure, and transformations.

Results
In September, samples collected from green and brown sections contained several relatively
abundant taxa. In green sections, the phylum *Proteobacteria* (including the classes *Alpha-,*
Beta-, Gamma-, and *Delta-*) was the most prevalent, making up to 59.5% of the relative
abundance of all taxa. Next in order were *Actinobacteria* (27.1%), *Cyanobacteria* (4.7%),
Acidobacteria (4.6%), *Firmicutes* (1.8%), and *Bacteroidetes* (1.7%). For brown sections, these
were also *Proteobacteria (Alpha-, Beta-, Gamma-* and *Delta-*) (52.5%), *Actinobacteria*
(27.1%), *Cyanobacteria* (5.5%), *Acidobacteria* (4.7%), *Bacteroidetes* (2.8%), *Firmicutes*
(2.2%), and *Planctomycetes* (1.4%). *Ascomycota* was the most prevalent fungal taxa (a relative
abundance of 1.3% in both green and brown sections) and *Evosea,* an amoeba, was found to be
most relatively abundant eukaryote in green sections (2.6%) and brown sections (1.4%).

When comparing respectively samples from the mine and natural sites in June, the most
prevalent phyla included *Proteobacteria* (63.2% vs 65.2%), *Actinobacteria* (19.4% vs 15.6%),
Bacteroidetes (4.8% vs 1.1%), *Cyanobacteria* (4.4% vs 11.1%), and *Acidobacteria* (3.3% vs
2.8%). Relative abundances of microeukaryotes and fungi were less than 1% overall. Archaea
was the least abundant kingdom making up only 0.21% and 0.13% of the relative abundance
in September and June samples, respectively.

<u>Effect of moss sections on microbial community assembly and diversity</u>
Overall, moss microbiome taxonomical and functional profiles in natural environments were
distinct between green and brown sections. For community composition, section type was the
strongest driver of microbial community assembly ($R^2 = 21.0\%$, p = 0.001; Fig. **2a**; Table **S4a**),
followed by moss species ($R^2 = 17.3\%$, p = 0.001; Table **S4a**). The effect of moss sections on
microbial community assembly also differed among moss species ($R^2 = 4.79\%$, p = 0.001;
Table **S4a**). Altogether this model accounted for 43.1% of the variation in taxonomic structure
of microbial communities.

Regarding alpha-diversity, we observed lower Shannon diversity of microbial taxa in green sections compared to brown sections for all moss species (Fig. **3a**; Table **S5a**). Shannon diversity of microbial taxa also differed across moss species (Fig. **3a**; Table **S5a**). Overall, our model explained 78.3% of variation in Shannon diversity. We did not find a significant interactive effect between section and moss species on microbial taxonomic diversity (Table **S5a**). Furthermore, pairwise comparisons revealed species-specific differences in Shannon diversity of moss microbial taxa between sections. *D. undulatum* had a significantly higher diversity of microbial taxa in brown section compared to three *Polytrichum* species (Fig. **S1a**; Table **S6a**). However, in brown sections of *Polytrichum* species taxonomical microbial diversity did not change significantly (Fig. **S1a**; Table **S6a**). Similarly, the green sections of *P. juniperinum* and *P. piliferum*, as well as of *D. undulatum* and *P. commune*, did not significantly differ in diversity of microbial taxa (Fig. **S1b**; Table **S6b**). Our results also showed that *D. undulatum* had lower microbial Shannon diversity in green sections than both *P. juniperinum* and *P. piliferum* (Fig. **S1b**; Table **S6b**). Overall, all four species differed in microbial diversity between green and brown sections (Fig. **3a**; Table **S7a**).

The differential abundance analysis identified several bacterial and fungal taxa that changed in relative abundance between green and brown sections. The bacterial families *Verrucomicrobia* (log$_2$-fold = 0.38), *Planctomycetes* (log$_2$-fold = 0.37), and the fungal families *Agariomycetes* (log$_2$-fold = 0.51), and *Leotiomycetes* (log$_2$-fold = 0.26) were relatively more abundant in brown compared to green sections (all *P*-values <0.001; **Fig. 4a,b**; Table **S8a**). In green compared to brown sections, *Proteobacteria* (log$_2$-fold = 0.41), *Acidobacteria* (log$_2$-fold = 0.27), and *Saccharomycetes* (log$_2$-fold = 0.24) were relatively more abundant (all *P*-values <0.001; Fig. **4a,b**; Table **S8a**).

Effect of moss sections on community assembly and diversity of KO profiles

For community composition, moss section type was a strong primary driver ($R^2 = 11.6\%$, p = 0.001; Fig. **2c**; Table **S4b**), while moss species identity was a strong secondary driver ($R^2 = 9.59\%$, p = 0.001; Fig. **2c**; Table **S4b**) of KO profiles. Interaction between moss section type and moss species identity accounted for 4.94% of the total variation (Table **S4b**). Altogether this model explained 26.1% of the variation in moss microbial KO profiles. Similarly, the Shannon diversity of microbial KOs was lower in green sections compared to brown sections across all moss species except for *D. undulatum* (Fig. **3c**; Table **S7b**). The

univariate model for Shannon diversity of KOs explained 75.5% of variation in September samples. Further exploration of pairwise differences of KO diversity between moss sections and moss species revealed that only *D. undulatum* showed strong differences in alpha-diversity in brown section (Fig. **S1c**; Table **S6b**). In contrast, within green section, the greatest differences in alpha-diversity were observed between *D. undulatum* and *P. piliferum* as well as between *D. undulatum* and *P. juniperinum* (Fig. **S1d**; Table **S6b**).

Using ANCOM-BC, we found differentially more abundant moss-associated nitrogen fixing genes in green sections (Fig. **5a,b**; Table **S8b**). Among them are KOs involved in nitrate/nitrite transportation, K02575 ($\log_2$-fold $= 0.06$, *P*-value <0.001) nitrogen fixating protein *nifU*, K04488 ($\log_2$-fold $= 0.05$, *P*-value <0.001; Fig. **5a**, Table **S8b**). Interestingly, KOs that were differentially abundant in moss brown sections were part of gene families involved in nitrogen assimilation rather than nitrogen fixation (Fig. **5a**; Table **S8b**).This list includes KOs involved in amino acid production, for example K00264 ($\log_2$-fold $= 0.03$, *P*-value <0.001), K01915 ($\log_2$-fold $= 0.03$, *P*-value $= 0.001$) and K00266 ($\log_2$-fold $= 0.07$, *P*-value <0.001), and ammonia production, namely K01725 ($\log_2$-fold $= 0.04$, *P*-value $= 0.001$; Fig. **5a**; Table **S8b**). In green sections, several KOs that are part of gene families that make-up chloroplasts and facilitate photosynthesis were shown to increase (Fig. **5b**; Table **S8b**). However, only a single KO of the *atpD* gene involved in photosynthesis was differentially abundant in the brown section (K02112; $\log_2$-fold $= 0.05$, *P*-value <0.001; Fig. **5b**; Table **S8b**).

Effect of location on community assembly and diversity of microbial taxa

In June samples, variation in moss microbial community assembly was driven primarily by location ($R^2 = 33.6\%$, p $= 0.001$; Fig. **2b**; Table **S4c**), then by moss species ($R^2 = 23.0\%$, p $= 0.004$; Fig. **2b**; Table **S4c**) and finally by their interaction ($R^2 = 14.5\%$, p $= 0.025$; Table **S4c**). Overall, this model accounted for 71.1% of the variation in moss microbial community assembly. Furthermore, we found that moss microbial Shannon diversity (taxonomic level) were lower in mine sites compared to natural sites between two moss species (Fig. **3b**; Table **S5c**). This model explained 79.6% of variation in June samples. Shannon diversity of microbial taxa associated with *T. tortuosa* and with *P. juniperinum* was lower in the mine site compared to natural sites (Fig. **3b**; Table **S6b**).

Based on ANCOM-BC analyses, we noted differential abundant bacteria from the classes *Cytophagia* (log$_2$-fold = 2.38), *Actinomycetia* (log$_2$-fold = 0.74), and *Alpha-proteobacteria* (log$_2$-fold = 0.91) in moss samples from different sites in the mine (all *P*-values <0.001; Fig. 4c; Table **S9a**). Conversely, several cyanobacterial families were found to be differentially abundant in natural sites compared to the mine site including *Scytonemataceae* (log$_2$-fold = 4.38), *Nostocaceae* (log$_2$-fold = 3.67), *Calotrichaceae* (log$_2$-fold = 3.40), and *Aphanizomenonaceae* (log$_2$-fold = 2.80; all *P*-values <0.001; Fig. **4d**; Table **S9a**).

Effect of location on community assembly and diversity of KO profiles

In June samples, location was a primary driver of KO profiles (R^2 = 20.3%, p = 0.001; Fig. **2d**; Table **S4d**), followed by moss species (R^2 = 19.1%, p = 0.005; Fig. **2d**; Table **S4d**) accounting in total for 33.4% of variation in KO profiles. However, ANOVA results, based on the univariate model of Shannon diversity, showed no significant differences in the KO diversity between sections (Fig. **3c**; Table **S5d**), mine and natural sites nor between moss species (Fig. **3d**; Table **S5d**).

ANCOM-BC showed that there were differential increases of various microbial nitrogen-fixing genes from mosses in natural sites compared to samples from the mine site (Fig. **5c-e**; Table **S9b**). These KOs specifically come from a family of genes that encode for components of the nitrogenase enzyme: *nifH* (K02588, log$_2$-fold = 0.21, *P*-value <0.001; Fig. **5c**; Table **S9b**), *nifB* (K02585, log$_2$-fold = 0.13, *P*-value <0.001; Fig. **5c**; Table **S9b**), *nifD* (K02586, log$_2$-fold = 0.10, *P*-value <0.001; Fig. **5c**; Table **S9b**) and *nifW* (K02595, log$_2$-fold = 0.12, *P*-value <0.001; Fig. **5c**; Table **S9b**), as well as differential increases of KOs involved in catalysing the oxidation of organic sulfur, such as taurine dioxygenase (K03119, log$_2$-fold = 0.11, *P*-value = 0.001; Fig. **5d;** Table **S9b**), a sulfur-containing organic compound. There were also differentially more genes associated with sulfur oxidation (K20035 and K00387; log$_2$-fold = 0.11, *P*-value <0.001 and log$_2$-fold = 0.15, *P*-value <0.001, respectively; Fig. **5d**; Table **S9b**) of the moss microbial community in the mine site. Stress adaptation KOs that were differentially abundant in the mine site include genes that encode heat shock proteins (K04761, log$_2$-fold = 0.074; *P*-value <0.001; Fig. **5e**; Table **S9b**), as well as those involved in biofilm formation (K04761 and K01991; log$_2$-fold = 0.12, *P*-value <0.001 and log$_2$-fold = 0.15, *P*-value = 0.001, respectively Fig. **5e**; Table **S9b**).

Discussion

By using deep shotgun sequencing over multiple moss species and across contrasting land-use types, this study highlights the importance of moss sections, location, and moss species identity as key ecological drivers of microbial community assembly across multiple kingdoms. In addition, our work goes beyond taxonomy by providing novel evidence that these drivers also impact the potential metabolic profiles and diversity of moss microbial communities. Our results show that specific functional pathways preferentially occur between green and brown sections from the same moss species. In addition, we observed pathways and taxa that were more tolerant of the anthropogenic perturbations at the Whabouchi mine site. Most importantly, given the importance of moss-microbe interactions for boreal forests, these findings can improve our assessment of the impact of anthropogenic activities and climate change on microbe-mediated ecosystem functions of mosses.

Moss microbial community assembly varies strongly between moss sections

Our results show that the physiological and chemical differences between moss sections drive moss microbial community assembly and diversity for both microbial taxonomy and KO profiles. Interestingly, the alpha-diversity of microbial taxa in green sections was consistently lower than in brown sections across all moss species. This contrast could potentially be explained either by (i) a phenomenon of niche dominance of a few taxa in green compartments or by (ii) the proximity of brown sections to soil which is extremely rich in microorganisms. As such, brown sections could behave as an interface in which both soil and moss-associated micro-organisms can thrive. This area is also mostly shaded from UV by the overlying green sections, and is generally more humid, due to the boundary layer effect (i.e., the physical phenomenon in which wind velocity is reduced), minimizing moisture loss (Halpern *et al.*, 2014). In general, green sections lose moisture before brown sections, which leads to higher humidity in brown compared to green sections (Lindo & Gonzalez, 2010). To sum up, our results suggest that brown sections represent a more conducive environment for the proliferation of various micro-organisms whereas green sections are generally dominated by fewer taxa that tolerate harsher conditions.

Regarding individual taxa, the *Proteobacteria, Acidobacteria*, and *Saccharomycetes* were found to be differentially abundant in green sections. These groups are closely associated with nitrogen fixation (Rahimlou *et al.*, 2021; Reji & Zhang, 2022). For example,

Saccharomycetes is a soil-dwelling taxon and was previously found to be differentially abundant in mosses growing in soil supplied with extra nitrogen, indicating an opportunistic pathway in obtaining nitrogen (Allison *et al.*, 2008). In line with previous literature, our results suggest that there is an association with nitrogen metabolism and the microbial taxa found in green sections. On the other hand, bacterial families of *Verrucomicrobia* and *Planctomycetes* and the fungal families of *Agariomycetes* and *Leotiomycetes*, which were differentially abundant in brown sections, are often found in boreal forest soils (Hofer, 2016; Ivanova *et al.*, 2016; Sun *et al.*, 2013). These phyla are less associated with nitrogen fixation. Instead, *Planctomycetes* and *Agariomycete*s were shown to efficiently degrade complex polysaccharides found in soil (Dedysh & Ivanova, 2019; Elliott *et al.*, 2015; Wang *et al.*, 2015). Likewise, *Verrucomicrobia* were found in soil organic matter and carbon rich substrates (Nam *et al.*, 2021). Furthermore, many studies on *Verrucomicrobia* (e.g., Methylacidiphilaceae) have demonstrated its ability to consume methane and hydrogen oxidation to obtain energy (Carere *et al.*, 2017). Taken together, these results suggest that moss-associated micro-organisms in brown sections may have broader metabolic activity related to carbon cycling, rather than nitrogen metabolism.

Finally, in contrast with our hypothesis, we detected no significant differences in differential abundances of cyanobacterial taxa between sections. However, cyanobacteria activity may not have necessarily been constant throughout the whole moss gametophyte. Rather, this finding could be a caveat of metagenomic analysis on DNA. Cyanobacterial DNA may linger in brown sections, though the actual cells could be dormant or even dead, not actively fixing nitrogen as shown by in Renaudin *et al.* (2022b), whose work was based on cDNA and *nifh* sequencing. Further quantitative analyses may reveal true differences in activity between different moss sections. Nevertheless, our data suggests that microbial taxa associated with mosses diverge between green and brown sections, a conserved phenomenon repeated across various boreal moss species.

In terms of functional profiles, our results show similar patterns of diversity of KOs to that of taxonomic microbial diversity, where alpha-diversity of KOs from green sections were found to be lower than those of brown sections across all taxa except in *D. undulatum*. This may be related to the capacity of microbes associated with green sections to fix nitrogen. KOs specifically related to nitrogen fixation were indeed found to be differentially abundant in

green sections compared to brown sections, except in *D. undulatum*. Because nitrogen fixation is such a metabolically costly process (Rastetter *et al.*, 2001), the capacity for other microbial functions to occur might have been reduced, thereby reducing overall microbial-associated KO diversity in moss green sections. Differentially abundant KOs in brown sections, on the other hand, were part of gene families that were more often related to nitrogen assimilation than fixation (Carrell *et al.*, 2022; Newsholme *et al.*, 2003). Hence, nitrogen metabolic processes may diverge between the two sections, where nitrogen fixation is more prominent in green sections, and brown sections may be utilizing nitrogen and incorporating it into microbial cellular functions. It remains unclear why this is occurring, although it may be a consequence of photosynthesis mediated nitrogen fixation. Correspondingly, various photosynthetic KOs were differentially abundant in green compared to brown sections. Yet, our results confirm the hypothesis that nitrogen fixation is associated with moss green sections. This work provides unprecedented demonstration that (1) microbial taxa distribution is strongly driven by moss sections and (2) although microbial nitrogen metabolism can occur throughout the entire moss gametophyte, nitrogen-fixation is mostly occurring in green sections.

Moss microbial community assembly differs between mine and natural sites

The comparison of moss microbial communities in mine and natural environments revealed that location type respectively explains 33.6% and 20.3% of taxonomy and functions, suggesting a significant shift in both microbial community assembly and potential metabolic profile. In addition, the mine site samples displayed lower microbial alpha-diversity compared to natural sites. These differences in beta- and alpha-diversity could be explained by the difficult conditions of mine tailings. Environmental factors such as shade, UV (Cardona-Correa *et al.*, 2015), wind exposure, and moisture (Whiteley & Gonzalez, 2016) can influence the survival of various micro-organisms. Our results showed that several cyanobacterial taxa were differentially abundant in natural sites. For example, the *Nostocaceae* were found to have a drastic change in relative abundance – from 74.6% in natural sites to 14.9% in mine site. Most importantly, this sharp change in relative abundance of *Nostocaceae* may have implications for nitrogen fixation, as several studies have previously shown that cyanobacteria were found to be the main nitrogen fixers in mosses (Stewart *et al.*, 2011). Conversely, differential abundance of bacterial taxa such as *Alphaproteobacteria* and *Cytophagia* were observed from the mine site moss samples. *Alphaproteobacteria* is a known contributor to nitrogen fixation and has been previously found in the rhizosphere of vegetation growing in mine sites (Bragina *et al.*,

2012; Gagnon *et al.*, 2020c). Our results also showed that there was a great differential increase

of *Cytophagia* at the mine site, a taxa generally known to grow in stressful environments

including mine tailings (Gagnon *et al.*, 2020c; Ren *et al.*, 2021). Future work will confirm if

conditions at mine sites are unfavorable for cyanobacteria so that other bacterial taxa may be

filling the nitrogen-fixing niche. Taken together, our results demonstrate that between the

natural sites of Nemaska and the Whabouchi mine site, there is a shift in moss microbial

community structure, where bacteria that can survive and thrive in the mine environment are

favored over cyanobacteria.

Although there was no difference in alpha-diversity of microbial KO between mosses

from natural sites and the mine site, our results suggest that differential abundances of certain

KOs can be related to moss microbial community structure. Several KOs in the *nif* family were

differentially abundant in natural sites compared to the mine site. This may be explained by the

higher relative abundance of cyanobacteria in natural sites. Our study identified the presence

of K04761 and K01991, which are involved in the formation of biofilms, among differentially

abundant KOs. In addition, K11940, a heat-shock protein was found to be the most

differentially abundant KO from mine site mosses. This KO was previously identified to be

associated with a *nif* gene island in *Bradyrhizobium*, a genus of *Alphaproteobacteria* (Tao *et*

al., 2021). Although mine tailings at the Whabouchi mine site were found to have low sulfur

content, various KOs associated with sulfur metabolism, such as K00387 is a sulfite oxidase,

were differentially abundant (Chen *et al.*, 2016; Gao *et al.*, 2020). The KEGG ortholog of

dmdC, K20035, which is another enzyme involved in sulfur metabolism found to be

differentially abundant in the mine site, plays an important role in breaking down

dimethylsulfoniopropionate (DMSP), generating a source of organic sulfur. The production of

DMSP by marine micro-organisms is highly correlated with salinity (Rinnan *et al.*, 2014).

Conversely, the increased abundance of taurine dioxygenase can be due to the presence of this

gene in the genome of *Nostocaceae*, which was relatively abundant in natural sites (Warshan

et al., 2018). In summary, the differences in taxonomical structure and functional capacity of

the moss microbiome between mine and natural sites could reflect contrasting conditions in

mine sites. Future studies will confirm the role of biotic and abiotic conditions in mine sites in

driving moss-microbe interactions. Overall, our results suggest that natural sites are more

conducive for microbe-associated nitrogen fixation processes and enable the proliferation of

various micro-organisms.

Moss species was consistently found to drive moss microbial taxonomic community structure and functional profiles (KOs). Additionally, moss section and moss species identity interactively explained microbial taxonomic community assembly. Through pairwise dissimilarities analysis, the diversity of moss microbial taxa and KOs of *D. undulatum* were different to that of all *Polytrichum* species in brown sections. Interestingly, there were also no differences in microbial diversity of moss taxa and KOs among *Polytrichum* species. These results point to the important role of moss phylogeny in structuring microbial communities in boreal mosses (Holland-Moritz *et al.*, 2021). For example, moss physiological characteristics such as leaf morphology and phenolic compounds may explain the differences in microbial community assembly of taxa and functional profiles among moss species (Jassey *et al.*, 2013; Renaudin *et al.*, 2022a).

Furthermore, the interaction between moss section and moss species identity may explain the different taxonomical and functional microbial assembly patterns in green sections compared to brown sections among different moss species. For example, in green sections, microbial taxonomic diversity differed between *P. commune* and the two other *Polytrichum* species: *P. piliferum* and *P. juniperinum*. This result may be explained by the fact that *P. piliferum* and *P. juniperinum* are phylogenetically closer to each other than to *P. commune* (Bell & Hyvönen, 2010). Nevertheless, our findings recapitulate previous literature in which moss species was shown to be an important driver of microbial taxonomical and functional community assembly (Kostka *et al.*, 2016; Holland-Moritz *et al.*, 2018; Jean *et al.*, 2020).

Conclusion

Overall, our findings provide a comprehensive portrait of the drivers of moss microbial communities, both taxonomically and functionally, for multiple species and different land use types. Importantly, our results show that mosses from natural sites and the Whabouchi mine site show significant differences in microbial communities, which can have major implications for boreal ecosystem functions facing accelerating climate and anthropogenic pressures. In line with previous studies, our results show that abundances of nitrogen metabolic genes differ between green sections and brown sections of mosses. In the mine site, contrasting conditions could favor microbial taxa that can tolerate the harsh environments, and functions such as sulfur oxidation and stress tolerance processes. Future studies using true abundance measurements of

nitrogen fixation genes and *in situ* nitrogen fixation activity are needed to confirm the correlation with differences in taxonomic and functional profiles. As global change continues to affect ecosystems worldwide but especially boreal and northern ecosystems, it is crucial to understand moss microbe functions and adaptability beyond focusing entirely on bacteria and taxonomical profiles. In this view, our work suggests novel avenues of research to leverage the key actors of nitrogen fixation and carbon sequestration by protecting these pristine environments and by leveraging mosses to restore perturbed sites.

Acknowledgements

First and foremost, the authors would like to thank the Cree Community of Nemaska and Nemaska Lithium for their permission to sample on their land. We would also like to acknowledge Joey Chamard and members of Sébastien Roy Lab (Louis Garneau, Julie Beaudin, Olivier Alix-Paré, and Samuel Malo) for the much-appreciated help with sampling and dissections, Ema Lussier for extensive support with laboratory work, Jean-Francois Lucier for assistance with metagenomic analysis and Mélanie Desrochers for map generation. We also thank Julie Beaudin, Peter Moffett, and Pascale B. Beauregard for critical reading of the manuscript. This work was supported by funding from (1) the Fonds de recherche du Québec – Nature et technologies (FRQNT) Programme de recherche en partenariat sur le développement durable du secteur minier (Grant 285683 awarded to lead PI Pr. Roy and co-PI Pr. Laforest-Lapointe); and (2) Pr. Laforest-Lapointe's Canada Research Chair T2 on Applied Microbial Ecology.

Competing interests

Authors declare no conflict of interests.

Author contributions

SI, ILL, and SR designed the experiment. SI conducted the sample collection, processing, and molecular work. SI, JRL, and ILL conducted the bioinformatic analyses. SI analyzed the data with support from ILL, MF, and JRL. SI wrote the first draft and ILL, MF, JRL, and SR contributed to the final manuscript.

Data availability

Scripts for statistical analyses, along with taxonomy tables, functional abundance tables, and metadata used for analysis will be available on FigShare. Raw sequence data is being deposited on NCBI.

Fig.1. Sampling design: For September sampling, within each location **(a)**, three microsites were randomly selected (yellow circles, inlaid). Three replicates were then sampled within each microsite (green circles). Top right, inlaid in **(a)**: map of Québec showing sampling sites within the continuous boreal forest section in Québec. **(b)** Images of each moss species that were sampled (displayed with permission from Michael Lüth). **(c)** Moss samples separated into green and brown sections.

Fig. 2. Two-dimensional principal coordinate analysis (PCoA) of community similarities for **(a, b)** microbial taxonomy based on Jaccard indices of variance-stabilizing transformed taxonomical abundances and **(c, d)** functional gene abundances based on Euclidean distances calculated from Hellinger transformed data. Panels **a** and **c** show samples collected in September, while panels **b** and **d** show samples collected in June. Shapes represent moss species, while colours correspond to different variable types. Each point is one sample. Ellipses indicate a 95% confidence interval around samples from each moss section or location.

Fig. 3. Differences in alpha-diversity (Shannon diversity) of moss microbial **(a, b)** taxa and **(c, d)** functional genes. Colors indicate different variable types: green and brown for section types (n=131) and teal and grey for natural vs. mine sites (n=10), respectively. Boxplots are faceted according to moss species identity. Letters indicate Tukey post-hoc pairwise comparisons of mean differences.

Fig. 4. Microbial taxonomic profiles between moss sections **(a, b)** and locations **(c, d)** highlighting (stars) differentially abundant taxa levels. Bar chart colors represent different microbial taxa. Star colors indicate taxa that are relatively abundant in brown (brown) or green (green) sections **(a, b)** as well as mine (grey) or natural (teal) sites **(c, d)**.

Fig. 5. Differential abundance (Log$_2$ fold change) of microbial functional genes (KOs) part of various metabolic pathways identified by ANCOM-BC between **(a, b)** moss sections from September sampling and **(c-e)** locations from June sampling. Colours represent variable type according to which KOs were differentially abundant. Circle size shows differential abundance weight at each variable type.

REFERENCES

Allison SD, Czimczik CI, Treseder KK. **2008**. Microbial activity and soil respiration under nitrogen addition in Alaskan boreal forest. *Global Change Biology* **14**: 1156–1168.

Antreich S, Sassmann S, Lang I. **2016**. Limited accumulation of copper in heavy metal adapted mosses. *Plant Physiology and Biochemistry* **101**: 141–148.

Arróniz-Crespo M, Bougoure J, Murphy DV, Cutler NA, Souza-Egipsy V, Chaput DL, Jones DL, Ostle N, Wade SC, Clode PL, *et al.* 2022. Revealing the transfer pathways of cyanobacterial-fixed N into the boreal forest through the feather-moss microbiome. *Frontiers in Plant Science* **13**: 1036258.

Beghini F, McIver LJ, Blanco-Míguez A, Dubois L, Asnicar F, Maharjan S, Mailyan A, Manghi P, Scholz M, Thomas AM, *et al.* 2021. Integrating taxonomic, functional, and strain-level profiling of diverse microbial communities with bioBakery 3 (P Turnbaugh, E Franco, and CT Brown, Eds.). *eLife* **10**: e65088.

Bell NE, Hyvönen J. **2010**. Phylogeny of the moss class Polytrichopsida (Bryophyta): Generic-level structure and incongruent gene trees. *Molecular Phylogenetics and Evolution* **55**: 381–398.

Bragina A, Maier S, Berg C, Müller H, Chobot V, Hadacek F, Berg G. **2012**. Similar diversity of Alphaproteobacteria and nitrogenase gene amplicons on two related sphagnum mosses. *Frontiers in Microbiology* **2**: 275.

Bragina A, Oberauner-Wappis L, Zachow C, Halwachs B, Thallinger GG, Müller H, Berg G. **2014**. The Sphagnum microbiome supports bog ecosystem functioning under extreme conditions. *Molecular Ecology* **23**: 4498–4510.

Cao W, Xiong Y, Zhao D, Tan H, Qu J. **2020**. Bryophytes and the symbiotic microorganisms, the pioneers of vegetation restoration in karst rocky desertification areas in southwestern China. *Applied Microbiology and Biotechnology* **104**: 873–891.

Cardona-Correa C, Graham JM, Graham LE. **2015**. Anatomical Effects of Temperature and UV-A + UV-B Treatments and Temperature-UV Interactions in the Peatmoss Sphagnum compactum. *International Journal of Plant Sciences* **176**: 159–169.

Carere CR, Hards K, Houghton KM, Power JF, McDonald B, Collet C, Gapes DJ, Sparling R, Boyd ES, Cook GM, *et al.* 2017. Mixotrophy drives niche expansion of verrucomicrobial methanotrophs. *The ISME Journal* **11**: 2599–2610.

Carrell AA, Veličković D, Lawrence TJ, Bowen BP, Louie KB, Carper DL, Chu RK, Mitchell HD, Orr G, Markillie LM, *et al.* 2022. Novel metabolic interactions and environmental conditions mediate the boreal peatmoss-cyanobacteria mutualism. *The ISME Journal* **16**: 1074–1085.

Chen L, Huang L, Méndez-García C, Kuang J, Hua Z, Liu J, Shu W. **2016**. Microbial communities, processes and functions in acid mine drainage ecosystems. *Current Opinion in Biotechnology* **38**: 150–158.

Chen K-H, Liao H-L, Bellenger J-P, Lutzoni F. **2019**. Differential gene expression associated with fungal trophic shifts along the senescence gradient of the moss Dicranum scoparium. *Environmental Microbiology* **21**: 2273–2289.

Cheng C, Gao M, Zhang Y, Long M, Wu Y, Li X. **2021**. Effects of disturbance to moss biocrusts on soil nutrients, enzyme activities, and microbial communities in degraded karst landscapes in southwest China. *Soil Biology and Biochemistry* **152**: 108065.

Collin A, Messier C, Kembel SW, Bélanger N. **2018**. Can sugar maple establish into the boreal forest? Insights from seedlings under various canopies in southern Quebec. *Ecosphere* **9**: e02022.

Davey ML, Heegaard E, Halvorsen R, Ohlson M, Kauserud H. **2012**. Seasonal trends in the biomass and structure of bryophyte-associated fungal communities explored by 454 pyrosequencing. *New Phytologist* **195**: 844–856.

Dedysh SN, Ivanova AA. **2019**. Planctomycetes in boreal and subarctic wetlands: diversity patterns and potential ecological functions. *FEMS Microbiology Ecology* **95**: fiy227.

Elliott DR, Caporn SJM, Nwaishi F, Nilsson RH, Sen R. **2015**. Bacterial and Fungal Communities in a Degraded Ombrotrophic Peatland Undergoing Natural and Managed Re-Vegetation. *PLOS ONE* **10**: e0124726.

Elzen E van den, Bengtsson F, Fritz C, Rydin H, Lamers LPM. **2020**. Variation in symbiotic N2 fixation rates among Sphagnum mosses. *PLOS ONE* **15**: e0228383.

Environment and Climate Change Canada Historical Climate Data. **2021**. *Monthly Climate Summaries*.

Franzetti A, Pittino F, Gandolfi I, Azzoni RS, Diolaiuti G, Smiraglia C, Pelfini M, Compostella C, Turchetti B, Buzzini P, et al. **2020**. Early ecological succession patterns of bacterial, fungal and plant communities along a chronosequence in a recently deglaciated area of the Italian Alps. *FEMS Microbiology Ecology* **96**: fiaa165.

Gagnon V, Rodrigue-Morin M, Migneault M, Tardif A, Garneau L, Lalonde S, Shipley B, Greer CW, Bellenger J-P, Roy S. **2020a**. Survival, growth and element translocation by 4 plant species growing on acidogenic gold mine tailings in Québec. *Ecological Engineering* **151**: 105855.

Gagnon V, Rodrigue-Morin M, Tremblay J, Wasserscheid J, Champagne J, Bellenger J-P, Greer CW, Roy S. **2020b**. Life in mine tailings: microbial population structure across the bulk soil, rhizosphere, and roots of boreal species colonizing mine tailings in northwestern Québec. *Annals of Microbiology* **70**: 41.

Gagnon V, Rodrigue-Morin M, Tremblay J, Wasserscheid J, Champagne J, Bellenger J-P, Greer CW, Roy S. **2020c**. Vegetation drives the structure of active microbial communities on an acidogenic mine tailings deposit. *PeerJ* **8**: e10109.

Gao S-M, Schippers A, Chen N, Yuan Y, Zhang M-M, Li Q, Liao B, Shu W-S, Huang L-N. **2020**. Depth-related variability in viral communities in highly stratified sulfidic mine tailings. *Microbiome* **8**: 89.

Goth A, Michelsen A, Rousk K. **2019**. Railroad derived nitrogen and heavy metal pollution
does not affect nitrogen fixation associated with mosses and lichens at a tundra site in
Northern Sweden. *Environmental Pollution* **247**: 857–865.

Halpern CB, Dovčiak M, Urgenson LS, Evans SA. **2014**. Substrates mediate responses of
forest bryophytes to a gradient in overstory retention. *Canadian Journal of Forest Research*
44: 855–866.

Hofer U. **2016**. A small soil bacterium dominates. *Nature Reviews Microbiology* **14**: 729–
729.

Holland-Moritz H, Stuart J, Lewis LR, Miller S, Mack MC, McDaniel SF, Fierer N.
2018. Novel bacterial lineages associated with boreal moss species. *Environmental*
Microbiology **20**: 2625–2638.

Holland-Moritz H, Stuart JEM, Lewis LR, Miller SN, Mack MC, Ponciano JM,
McDaniel SF, Fierer N. **2021**. The bacterial communities of Alaskan mosses and their
contributions to N2-fixation. *Microbiome* **9**: 1–14.

Ininbergs K, Bay G, Rasmussen U, Wardle DA, Nilsson M-C. **2011**. Composition and
diversity of nifH genes of nitrogen-fixing cyanobacteria associated with boreal forest feather
mosses. *New Phytologist* **192**: 507–517.

Ivanova AA, Kulichevskaya IS, Merkel AY, Toshchakov SV, Dedysh SN. **2016**. High
Diversity of Planctomycetes in Soils of Two Lichen-Dominated Sub-Arctic Ecosystems of
Northwestern Siberia. *Frontiers in Microbiology* **7**: 2065.

Jassey VE, Chiapusio G, Binet P, Buttler A, Laggoun-Défarge F, Delarue F, Bernard N,
Mitchell EA, Toussaint M-L, Francez A-J, *et al*. 2013. Above- and belowground linkages
in Sphagnum peatland: climate warming affects plant-microbial interactions. *Global Change*
Biology **19**: 811–823.

Jean M, Holland-Moritz H, Melvin AM, Johnstone JF, Mack MC. **2020**. Experimental
assessment of tree canopy and leaf litter controls on the microbiome and nitrogen fixation
rates of two boreal mosses. *New Phytologist* **227**: 1335–1349.

Jonsson M, Kardol P, Gundale MJ, Bansal S, Nilsson M-C, Metcalfe DB, Wardle DA.
2015. Direct and Indirect Drivers of Moss Community Structure, Function, and Associated
Microfauna Across a Successional Gradient. *Ecosystems* **18**: 154–169.

Kanehisa M, Sato Y, Kawashima M, Furumichi M, Tanabe M. **2016**. KEGG as a
reference resource for gene and protein annotation. *Nucleic Acids Research* **44**: D457–D462.

Kayes I, Mallik A. **2020**. Boreal Forests: Distributions, Biodiversity, and Management. In:
Leal Filho W, Azul AM, Brandli L, Lange Salvia A, Wall T, eds. Life on Land. Cham:
Springer International Publishing, 1–12.

Klarenberg IJ, Keuschnig C, Colmenares AJR, Warshan D, Jungblut AD, Jónsdóttir
IS, Vilhelmsson O. **2021**. Long-term warming effects on the microbiome and nifH gene
abundance of a common moss species in sub-Arctic tundra. *New Phytologist* **234**: 2044–
2056.

Klarenberg IJ, Keuschnig C, Salazar A, Benning LG, Vilhelmsson O. **2023**. Moss and underlying soil bacterial community structures are linked to moss functional traits. *Ecosphere* **14**: e4447.

Kostka JE, Weston DJ, Glass JB, Lilleskov EA, Shaw AJ, Turetsky MR. **2016**. The Sphagnum microbiome: new insights from an ancient plant lineage. *New Phytologist* **211**: 57–64.

Kox MAR, van den Elzen E, Lamers LPM, Jetten MSM, van Kessel MAHJ. **2020**. Microbial nitrogen fixation and methane oxidation are strongly enhanced by light in Sphagnum mosses. *AMB Express* **10**: 1–11.

Langmead B, Wilks C, Antonescu V, Charles R. **2019**. Scaling read aligners to hundreds of threads on general-purpose processors. *Bioinformatics* **35**: 421–432.

Liebner S, Zeyer J, Wagner D, Schubert C, Pfeiffer E-M, Knoblauch C. **2011**. Methane oxidation associated with submerged brown mosses reduces methane emissions from Siberian polygonal tundra. *Journal of Ecology* **99**: 914–922.

Lin H, Peddada SD. **2020**. Analysis of compositions of microbiomes with bias correction. *Nature Communications* **11**: 3514.

Lin X, Tfaily MM, Green SJ, Steinweg JM, Chanton P, Imvittaya A, Chanton JP, Cooper W, Schadt C, Kostka JE. **2014a**. Microbial Metabolic Potential for Carbon Degradation and Nutrient (Nitrogen and Phosphorus) Acquisition in an Ombrotrophic Peatland. *Applied and Environmental Microbiology* **80**: 3531–3540.

Lin X, Tfaily MM, Steinweg JM, Chanton P, Esson K, Yang ZK, Chanton JP, Cooper W, Schadt CW, Kostka JE. **2014b**. Microbial Community Stratification Linked to Utilization of Carbohydrates and Phosphorus Limitation in a Boreal Peatland at Marcell Experimental Forest, Minnesota, USA. *Applied and Environmental Microbiology* **80**: 3518–3530.

Lindo Z, Gonzalez A. **2010**. The bryosphere: an integral and influential component of the Earth's biosphere. *Ecosystems* **13**: 612–627.

Love MI, Huber W, Anders S. **2014**. Moderated estimation of fold change and dispersion for RNA-seq data with DESeq2. *Genome Biology* **15**: 550.

Lu J, Breitwieser FP, Thielen P, Salzberg SL. **2017**. Bracken: estimating species abundance in metagenomics data. *PeerJ Computer Science* **3**: e104.

Macdonald SE, Fenniak TE. **2007**. Understory plant communities of boreal mixedwood forests in western Canada: Natural patterns and response to variable-retention harvesting. *Forest Ecology and Management* **242**: 34–48.

Masson-Delmotte V, Zhai P, Pirani A, Connors SL, Pean C, Chen Y, Goldfarb L, Gomis MI, Matthews JR, Berger S, *et al.* **2021**. *Climate Change 2021: The Physical Science Basis. Contribution of Working Group I to the Sixth Assessment Report of the Intergovernmental Panel on Climate Change*. Switzerland: Intergovernmental Panel on Climate Change.

McLaren JR, Buckeridge KM, van de Weg MJ, Shaver GR, Schimel JP, Gough L. **2017**. Shrub encroachment in Arctic tundra: Betula nana effects on above- and belowground litter decomposition. *Ecology* **98**: 1361–1376.

McMurdie PJ, Holmes S. **2013**. phyloseq: An R Package for Reproducible Interactive Analysis and Graphics of Microbiome Census Data. *PLOS ONE* **8**: e61217.

Merlin M, Leishman F, Errington RC, Pinno BD, Landhäusser SM. **2019**. Exploring drivers and dynamics of early boreal forest recovery of heavily disturbed mine sites: a case study from a reconstructed landscape. *New Forests* **50**: 217–239.

Monaci F, Ancora S, Bianchi N, Bonini I, Paoli L, Loppi S. **2021**. Combined use of native and transplanted moss for post-mining characterization of metal(loid) river contamination. *Science of The Total Environment* **750**: 141669.

Moussaoui L, Leduc A, Girona MM, Bélisle AC, Lafleur B, Fenton NJ, Bergeron Y. **2020**. Success Factors for Experimental Partial Harvesting in Unmanaged Boreal Forest: 10-Year Stand Yield Results. *Forests* **11**: 1199.

Nam S, Alday JG, Kim M, Kim H, Kim Y, Park T, Lim HS, Lee BY, Lee YK, Jung JY. **2021**. The relationships of present vegetation, bacteria, and soil properties with soil organic matter characteristics in moist acidic tundra in Alaska. *Science of The Total Environment* **772**: 145386.

Nemaska Lithium Inc., Cree Nation of Nemaska, Grand Council of the Crees (Eeyou Istchee), Cree Nation Government. **2014**. Chinuchi agreement.

Newsholme P, Lima MMR, Procopio J, Pithon-Curi TC, Doi SQ, Bazotte RB, Curi R. **2003**. Glutamine and glutamate as vital metabolites. *Brazilian Journal of Medical and Biological Research* **36**: 153–163.

Oksanen J, Simpson GL, Blanchet FG, Kindt R, Legendre P, Minchin PR, O'Hara RB, Solymos P, Stevens MHH, Szoecs E, *et al.* 2022. vegan: Community Ecology Package.

Pakgohar N, Eshaghi Rad J, Gholami G, Alijanpour A, Roberts D. **2021**. Comparison of two non-hierarchal clustering performance in vegetation community datasets. *Iranian Journal of Forest and Poplar Research* **29**: 315–325.

Peterson RA. **2021**. Finding Optimal Normalizing Transformations via bestNormalize. *The R Journal* **13**: 310–329.

Peterson RA, Cavanaugh JE. **2020**. Ordered quantile normalization: a semiparametric transformation built for the cross-validation era. *Journal of Applied Statistics* **47**: 2312–2327.

Pombubpa N, Pietrasiak N, De Ley P, Stajich JE. **2020**. Insights into dryland biocrust microbiome: geography, soil depth and crust type affect biocrust microbial communities and networks in Mojave Desert, USA. *FEMS Microbiology Ecology* **96**: fiaa125.

Quince C, Walker AW, Simpson JT, Loman NJ, Segata N. **2017**. Shotgun metagenomics, from sampling to analysis. *Nature Biotechnology* **35**: 833–844.

Rahimlou S, Bahram M, Tedersoo L. **2021**. Phylogenomics reveals the evolution of root nodulating alpha- and beta-Proteobacteria (rhizobia). *Microbiological Research* **250**: 126788.

Rastetter EB, Vitousek PM, Field C, Shaver GR, Herbert D, gren GI. **2001**. Resource Optimization and Symbiotic Nitrogen Fixation. *Ecosystems* **4**: 369–388.

Reji L, Zhang X. **2022**. Genome-Resolved Metagenomics Informs the Functional Ecology of Uncultured Acidobacteria in Redox Oscillated Sphagnum Peat. *mSystems* **7**: e00055-22.

Ren J, Liu F, Luo Y, Zhu J, Luo X, Liu R. **2021**. The Pioneering Role of Bryophytes in Ecological Restoration of Manganese Waste Residue Areas, Southwestern China. *Journal of Chemistry* **2021**: e9969253.

Renaudin M, Blasi C, Bradley RL, Bellenger J-P. **2022a**. New insights into the drivers of moss-associated nitrogen fixation and cyanobacterial biomass in the eastern Canadian boreal forest. *Journal of Ecology* **110**: 1403–1418.

Renaudin M, Laforest-Lapointe I, Bellenger J-P. **2022b**. Unraveling global and diazotrophic bacteriomes of boreal forest floor feather mosses and their environmental drivers at the ecosystem and at the plant scale in North America. *Science of The Total Environment* **837**: 155761.

Rinnan R, Steinke M, MCGENITY T, Loreto F. **2014**. Plant volatiles in extreme terrestrial and marine environments. *Plant, Cell & Environment* **37**: 1776–1789.

Rousk K, Jones D, DeLuca T. **2013**. Moss-cyanobacteria associations as biogenic sources of nitrogen in boreal forest ecosystems. *Frontiers in Microbiology* **4**: 1–10.

Rousk K, Michelsen A. **2017**. Ecosystem nitrogen fixation throughout the snow-free period in subarctic tundra: effects of willow and birch litter addition and warming. *Global Change Biology* **23**: 1552–1563.

Rousk K, Sorensen PL, Lett S, Michelsen A. **2015**. Across-Habitat Comparison of Diazotroph Activity in the Subarctic. *Microbial Ecology* **69**: 778–787.

Royles J, Griffiths H. **2015**. Invited review: climate change impacts in polar regions: lessons from Antarctic moss bank archives. *Global Change Biology* **21**: 1041–1057.

Shcherbakov AV, Bragina AV, Kuzmina EYu, Berg C, Muntyan AN, Makarova NM, Malfanova NV, Cardinale M, Berg G, Chebotar VK, *et al*. 2013. Endophytic bacteria of Sphagnum mosses as promising objects of agricultural microbiology. *Microbiology* **82**: 306–315.

Shen JC, Zhang ZH, Liu R, Wang ZH. **2018**. Ecological restoration of eroded karst utilizing pioneer moss and vascular plant species with selection based on vegetation diversity and underlying soil chemistry. *International Journal of Phytoremediation* **20**: 1369–1379.

Simpson GL, R Core Team, Bates DM, Oksanen J. **2022**. permute: Functions for Generating Restricted Permutations of Data.

Stewart KJ, Coxson D, Grogan P. **2011**. Nitrogen Inputs by Associative Cyanobacteria across a Low Arctic Tundra Landscape. *Arctic, Antarctic, and Alpine Research* **43**: 267–278.

Stuart JEM, Holland-Moritz H, Lewis LR, Jean M, Miller SN, McDaniel SF, Fierer N, Ponciano JM, Mack MC. 2020a. Host identity as a driver of moss-associated N2 fixation rates in Alaska. *Ecosystems* **24**: 530–547.

Stuart RK, Pederson ERA, Weyman PD, Weber PK, Rassmussen U, Dupont CL. 2020b. Bidirectional C and N transfer and a potential role for sulfur in an epiphytic diazotrophic mutualism. *The ISME Journal* **14**: 3068–3078.

Sun H, Terhonen E, Koskinen Mora K, Paulin L, Kasanen R, Asiegbu F. 2013. Bacterial diversity and community structure along different peat soils in boreal forest. *Applied Soil Ecology* **74**: 37–45.

Tao J, Wang S, Liao T, Luo H. 2021. Evolutionary origin and ecological implication of a unique nif island in free-living Bradyrhizobium lineages. *The ISME Journal* **15**: 3195–3206.

Tian W, Wang H, Xiang X, Wang R, Xu Y. 2019. Structural Variations of Bacterial Community Driven by Sphagnum Microhabitat Differentiation in a Subalpine Peatland. *Frontiers in Microbiology* **10**: 1–11.

Turetsky MR. 2003. The role of bryophytes in carbon and nitrogen cycling. *The Bryologist* **106**: 395–409.

Turetsky MR, Bond-Lamberty B, Euskirchen E, Talbot J, Frolking S, McGuire AD, Tuittila E-S. 2012. The resilience and functional role of moss in boreal and arctic ecosystems. *New Phytologist* **196**: 49–67.

Vareda JP, Valente AJM, Durães L. 2019. Assessment of heavy metal pollution from anthropogenic activities and remediation strategies: A review. *Journal of Environmental Management* **246**: 101–118.

Vesty EF, Whitbread AL, Needs S, Tanko W, Jones K, Halliday N, Ghaderiardakani F, Liu X, Cámara M, Coates JC. 2020. Cross-kingdom signalling regulates spore germination in the moss Physcomitrella patens. *Scientific Reports* **10**: 2614.

Vile MA, Kelman Wieder R, Živković T, Scott KD, Vitt DH, Hartsock JA, Iosue CL, Quinn JC, Petix M, Fillingim HM, *et al.* 2014. N2-fixation by methanotrophs sustains carbon and nitrogen accumulation in pristine peatlands. *Biogeochemistry* **121**: 317–328.

Walker XJ, Baltzer JL, Cumming SG, Day NJ, Ebert C, Goetz S, Johnstone JF, Potter S, Rogers BM, Schuur EAG, *et al.* 2019. Increasing wildfires threaten historic carbon sink of boreal forest soils. *Nature* **572**: 520–523.

Wang X, Sharp CE, Jones GM, Grasby SE, Brady AL, Dunfield PF. 2015. Stable-Isotope Probing Identifies Uncultured Planctomycetes as Primary Degraders of a Complex Heteropolysaccharide in Soil. *Applied and Environmental Microbiology* **81**: 4607–4615.

Warshan D, Espinoza JL, Stuart RK, Richter RA, Kim S-Y, Shapiro N, Woyke T, C Kyrpides N, Barry K, Singan V, *et al.* 2017. Feathermoss and epiphytic Nostoc cooperate differently: expanding the spectrum of plant–cyanobacteria symbiosis. *The ISME Journal* **11**: 2821–2833.

Warshan D, Liaimer A, Pederson E, Kim S-Y, Shapiro N, Woyke T, Altermark B,
Pawlowski K, Weyman PD, Dupont CL, *et al.* **2018**. Genomic Changes Associated with the
Evolutionary Transitions of Nostoc to a Plant Symbiont. *Molecular Biology and Evolution*
35: 1160–1175.

Whiteley JA, Gonzalez A. **2016**. Biotic nitrogen fixation in the bryosphere is inhibited more
by drought than warming. *Oecologia* **181**: 1243–1258.

Wicaksono WA, Cernava T, Berg C, Berg G. **2021**. Bog ecosystems as a playground for
plant–microbe coevolution: bryophytes and vascular plants harbour functionally adapted
bacteria. *Microbiome* **9**: 170.

Wong MH. **2003**. Ecological restoration of mine degraded soils, with emphasis on metal
contaminated soils. *Chemosphere* **50**: 775–780.

Zeileis A, Hothorn T. **2002**. Diagnostic Checking in Regression Relationships. *R News* **2**: 7–
10.

 Supporting Information

886 **Article title**: Moss species, sections, and land use type drive moss-associated microbial

887 taxonomy and functions in boreal forests

889 **Authors**: Sarah Ishak, Maria Faticov, Jonathan Rondeau-Leclaire, Sébastien Roy and Isabelle

890 Laforest-Lapointe

891 The following supporting documents are available for this paper: