Bird communities in a selectively logged tropical montane forest are dominated by small, low-elevation species

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Abstract

Tropical montane forests are critical centres of terrestrial biodiversity and endemism, and face a range of threats, including selective logging and climate change. However, few studies have explored the joint influence of these threats, particularly in tropical mountains, despite how crucial it is to understand the cumulative impact of forest loss and climate change on biodiversity. We used mist-netting and bird ringing data from a long-term community and population monitoring program to examine how the composition of the mid-elevation Eastern Himalayan understorey bird community changed over a ten-year period in mid-elevation primary and logged forest. Because logged forest is warmer than primary forest, we hypothesised that the bird community would shift towards lower elevation species, with a faster transition in logged forest, and that the primary forest community would have larger species than the logged forest community. Our study shows that the bird community in the logged forest shifted more rapidly towards lower-elevation species than in the primary forest. Moreover, we found that smaller species were better able to colonise warmer logged forest, whereas larger species reached higher densities in primary forest. These trends are likely to be driven by climate change causing upslope range shifts and logged forest selecting for species that are likely to tolerate the consequently altered abiotic (higher temperatures) and biotic (lower resources) environment. These findings have significant implications for understanding the impacts of forest loss and climate change on biodiversity, particularly in tropical montane forests.

Keywords: Bird community, Selective logging, Climate change, Body mass, Elevational range shifts, Montane forests

Introduction

Tropical montane forests are the richest centres of terrestrial biodiversity and endemism globally (Rahbek et al., 2019). These forests face multiple threats (Bradshaw et al. 2009), of which climate change and habitat degradation/conversion are major drivers of biodiversity declines (Asner et al., 2009; Edwards and Laurance, 2013; Constantini, Edwards, & Simons, 2016; Burivalova et al., 2014; Dawson et al., 2011; Corlett, 2012).

While the separate impacts of habitat loss and climate change on biodiversity have been studied extensively (Bellard et al., 2012; Matthews et al., 2014; Fahrig, 2003), few studies have examined the influence of both these threats together (Manytka-Pringle et al., 2012; Guo et al., 2018; Srinivasan and Wilcove, 2021), especially in the tropics, and particularly in tropical mountains. With climate change and forest loss accelerating dramatically (Mitchard, 2018; Arneth et al., 2020), understanding their cumulative impacts on biodiversity is essential. This is especially true because forest degradation (e.g., from selective logging) alters forest structure and creates novel habitats that tend to be hotter and drier than primary forest (Srinivasan and Wilcove, 2021), complicating climate impacts. Further, tropical species are thermal specialists and poor dispersers and tend to have narrow elevational ranges, and how climate- *and* land-use-driven changes in the abiotic environment might impact such species remains poorly known (Janzen, 1967; McCain, 2009; Sheard et al., 2020).

Tropical montane species are shifting their ranges to higher elevations rapidly, with strong evidence that such upslope shifts are a result of rising temperatures globally (Forero-Medina et al., 2011; Campos-Cerqueira et al., 2017; Freeman et al., 2018; Neate-Clegg et al., 2018, Freeman et al. 2021). As tropical species shift their ranges upslope in response to climate change, populations might either encounter primary forest or degraded forest (e.g., selectively logged forest) at higher elevations. With populations of the same species likely shifting their ranges into primary and logged forest at higher elevations, it is crucial to understand how such range shifts might lead to differences in community assembly in undisturbed versus disturbed forest. Although selectively logged forest can retain more biodiversity than other forms of forest loss/degradation such as outright conversion to agriculture (Putz et al., 2012; Gibson et al., 2011), logging does lead to the altered composition of ecological communities (Constantini, Edwards, & Simons, 2016; Burivalova et al., 2014).

Using mist-netting and bird ringing data from a long-term community and population monitoring program, we asked how the composition of the mid-elevation Eastern Himalayan understorey bird community changed over a ten-year period in primary and in logged forest. At any given elevation, the primary forest bird community should become progressively dominated by historically “low-elevation” species as they move upwards to replace “higher-elevation” species that are moving further upslope. Logged forest being warmer than primary forest (Srinivasan and

Wilcove, 2021) should be thermally more similar to lower-elevation primary forest than to primary forest at the same elevation. We, therefore, expected that upslope range shifts should lead to faster species turnover from “higher-elevation” to “lower-elevation” in logged than in primary forest. Further, we expected that smaller species, being more thermally tolerant would be better able to colonise warmer logged forest, whereas larger species would reach higher densities in primary forest.

We predicted that:

1. The bird community would gradually shift towards lower elevation species and the transition would occur faster in the logged forest.
2. The bird community in primary forest would have larger species as compared to the logged forest community.

Methods

Field Methods

We carried out this study in mid-elevation (2000m ASL) montane broadleaved forest (Champion and Seth, 1968) in Eaglenest Wildlife Sanctuary (27°06′0″N 92°24′0″E, EWLS henceforth), Arunachal Pradesh, Eastern Himalaya, India. In this habitat, we established three plots each in primary and logged forest (total of 9ha each in primary and logged forest; Fig. 1). We sampled each of these plots for three consecutive days from 0500 hrs to 1200 hrs in the early breeding season (April-May) from 2011 to 2021 using 24-28 mist nets (12m length, 2.4m height, 16mm mesh, 4-shelf). In 2020, a nationwide COVID-19-related lockdown precluded any bird sampling. We placed the mist nets in a grid-like pattern within the plot with ~40m distance between neighbouring nets (Fig. 1). We then checked each net every 20-30 minutes and any birds captured were brought back to the ringing station at the centre of the sampling plot. We identified and weighed each bird caught with a jeweller’s scale, ringed it with a uniquely numbered aluminium ring and then the bird them immediately.

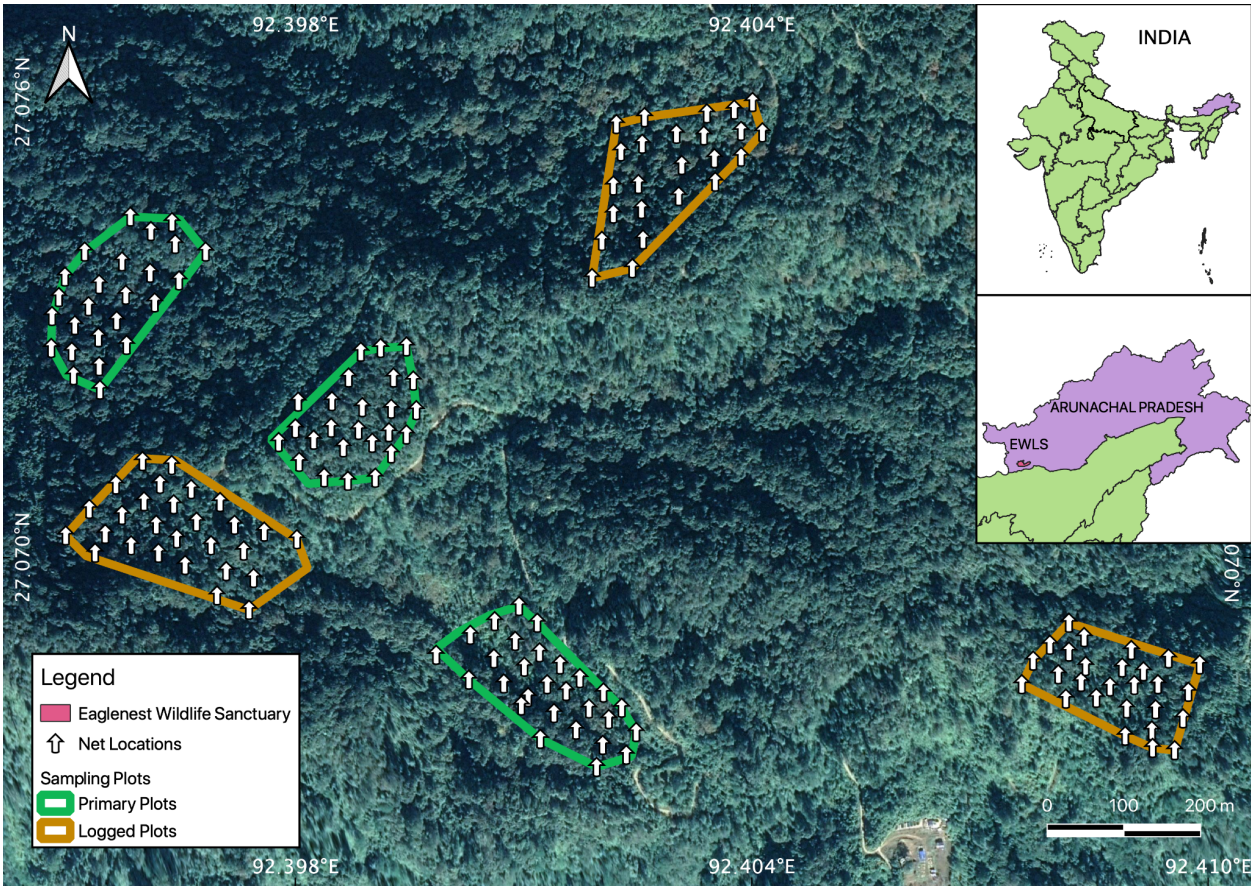

Figure 1: Map of the study area in Eaglenest Wildlife Sanctuary, Arunachal Pradesh, India. Primary forest and logged forest plots are outlined in green and brown, respectively. Arrows show the locations of mist nets.

Data Curation

We excluded high-elevation breeders which are passage migrants to higher-elevation breeding sites and do not breed in our study area (Table S1), aquatic species, carnivores, canopy dwellers, and non-invertivores (which included frugivores, granivores and nectarivores). We finally analysed the understorey insectivorous bird community of 61 species (Table S1). We pooled the raw counts of each species in each habitat in each year pooled from for the first two days of sampling effort in each site (i.e., six days each in logged and primary forest in each year). We did this because in some years in some plots, rain prevented the full three-day sampling protocol and sampling was limited to two consecutive days.

Analytical Methods:

Detection corrected abundance data

In primary and logged forest separately, we created capture histories of each individual bird. In the two habitats separately, we then ran Cormack-Jolly-Seber (CJS) open population mark-recapture models (Cormack 1964; Jolly 1965; Seber 1965) in the R package *RMark* (Laake 2013) in program R (R Core Team 2023), which simultaneously estimate apparent survival and recapture probability using a maximum likelihood approach (Seber 1965). We ran four CJS models with different formulations for apparent survival (Φ) and recapture probability (p) which were combinations of

(a) Φ ~ species identity alone, the model assumes that different species would have different apparent survival, but that apparent survival would not change between habitats,
(b) Φ ~ species * habitat, where apparent survival was allowed to fully vary between both species and habitats,
(c) p ~ habitat * capture class, where recapture probability was not time-varying but varied between habitats (primary and logged) and capture class (divided into small or large species), and
(d) p ~ habitat * capture class * time, where recapture probability was dependent on habitat and capture class, but was also allowed to vary with time.

Each of these models was run with a size cutoff to differentiate "small" and "large" species that differed maximally in capture probability. These size cutoffs were the 0.2 to 0.8 quantiles in body mass of the 61 species. We used the second-order Akaike's Information Criterion (AICc; Burnham and Anderson 2002) to select the best value for the size cutoff and the best model formulation to estimate apparent survival and recapture probability. We then divided year-wise raw counts of species by recapture probabilities to estimate detection-corrected abundances.

Differences in community composition

We first created a year-by-species matrix in primary forest and logged forest separately with detection-corrected abundance data (Table S2) transformed using the Hellinger transformation (Bocard et al., 2011) which reduces the weightage of rare species in contributing to compositional differences between communities (Legendre and Gallagher, 2001). We then performed a non-metric multidimensional scaling (NMDS) ordination analysis to visualise the bird community in logged and primary forests each year using the Bray-Curtis dissimilarity index on the transformed data in the R package *vegan* (Okansen et al., 2022; R Core Team 2023). To test whether the bird community composition differed significantly between primary and logged

forest, we used permutational multivariate analysis of variance (PERMANOVA), again using the Bray-Curtis dissimilarity index.

To understand the relative contributions of different species to community-level differences between primary and logged forest, we used similarity percentage routines (SIMPER, Clarke and Warwick 2001). From the SIMPER routine, we then selected the top species that cumulatively contributed to 50% of dissimilarity in bird community composition between primary and logged forest (Table S3). To visualise the importance of these species to the difference in community composition across habitats, we plotted the location of these species in NMDS space.

To understand potential differences between the morphology (i.e., body mass) and distribution (i.e., elevational ranges) of species that contributed in large part to community-level differences between primary and logged forest, we used:
(a) A generalized linear model (GLMs) with habitat type (logged/primary) as the categorical predictor variable and body mass as the response variable.
(b) An ANOVA to compare the elevation range midpoints of species in primary and logged forest.

We extracted the breeding elevation range data of each from Spierenburg (2005) and it was supplemented by data from Srinivasan et al. (2018). We used this data to calculate the elevation range midpoint.

Results:

We recorded 9,036 captures of 6,189 individuals from 130 species in our mist nets across all years and across all sampling plots. Upon curating the data, we finally analysed 6,926 captures of 4,801 individuals of 61 understorey insectivorous species. The best-supported body mass size cutoff to estimate different recapture probabilities for “small” and “large” species was 10.85g. The best model (with a delta AICc = 0; next best model delta AICc = 7.41) was one in which apparent survival was species-specific and recapture probabilities (Table S4) were time-varying and differed for “small” and “large” species in primary and in logged forest. Detection-corrected raw counts are in Table S2.

Community composition in primary and logged forest

NMDS axis 1 separated bird communities in primary and logged habitats (stress = 0.20, Fig. 2). Primary and logged forest communities were compositionally significantly different from each other (PERMANOVA ($R^2 = 0.14$, $F_{1,18} = 2.98$, $p < 0.01$). The first 20 species (Table S3) explained 50% of the variation in community composition between primary and logged forest and were

used in subsequent analyses. We did not use *Schoeniparus cinereus* and *Tesia castaneocoronata* out of these since they did not fall in either primary or logged forest clusters.

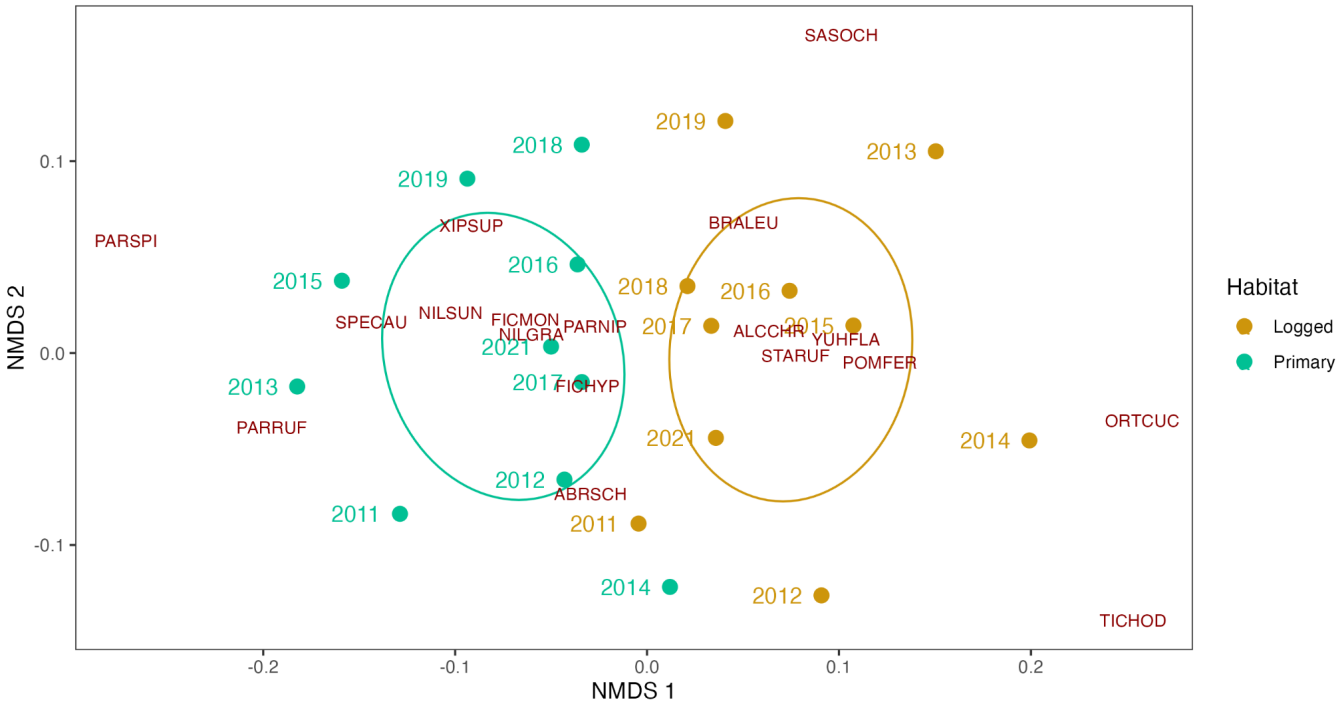

Figure 2: Non-metric multidimensional scaling (NMDS) showing dissimilarity between primary and logged forest communities over a ten-year period (see Table S3 for details of species codes and contribution to the SIMPER routine).

Body mass trends in primary and logged forest

Habitat was significantly related to body mass (Poisson ANOVA, β_{primary} = 0.26 [0.02, 0.50], McFadden's Pseudo R^2 = 0.04. On average, species typical to primary forest were 1.3g heavier than species typical to logged forest.

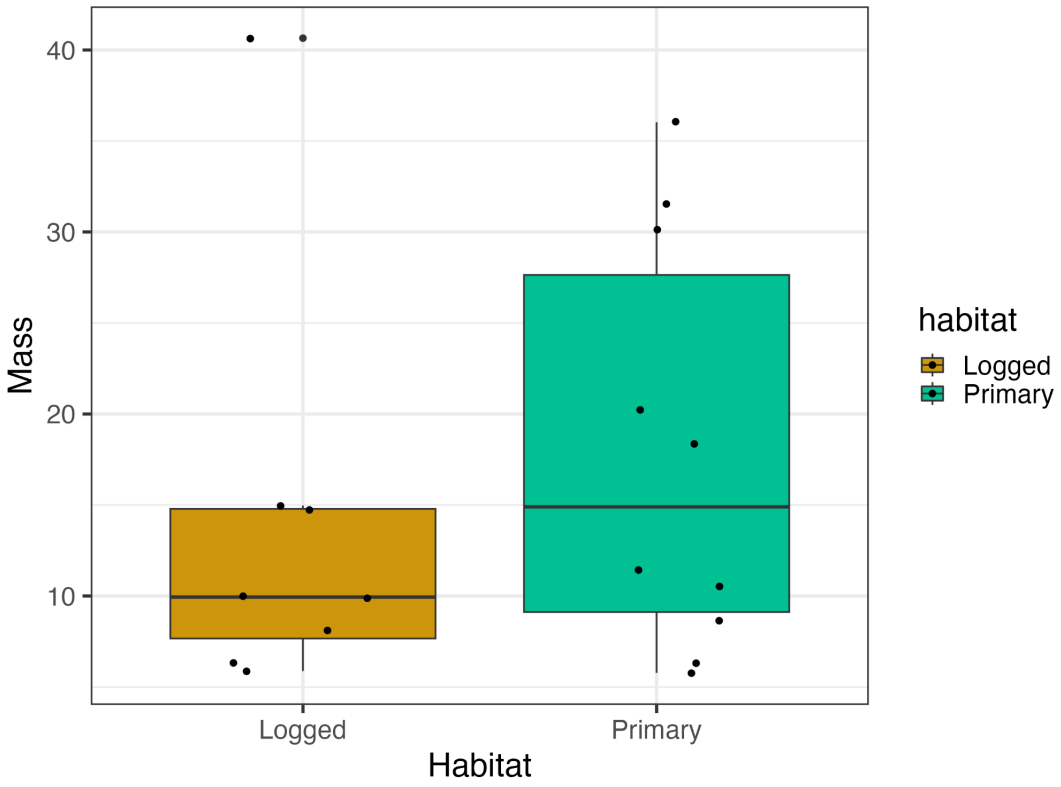

Figure 3: Mass of species that contributed to 50% of dissimilarity in community composition between primary and logged forests. Species typical to primary forest are, on average, larger than those typical to logged forest. The box shows the middle 50% of the data, with the vertical line inside the box indicating the median value. The whiskers of the boxplot extend from the box to show the minimum and maximum values of the data. Each dot represents individual data points and is plotted with a jitter to avoid overlap.

Trends of Breeding elevational range midpoints in primary and logged forest

Elevation range midpoints of bird species typical to primary forest was, on average 172.5m [-185.3, 530.3; R^2 = 0.06] higher than species representative of logged forest. While the mean difference in the mid-points of species' elevational ranges was as expected, the effect of habitat was weak (i.e., 95% CIs overlapping zero).

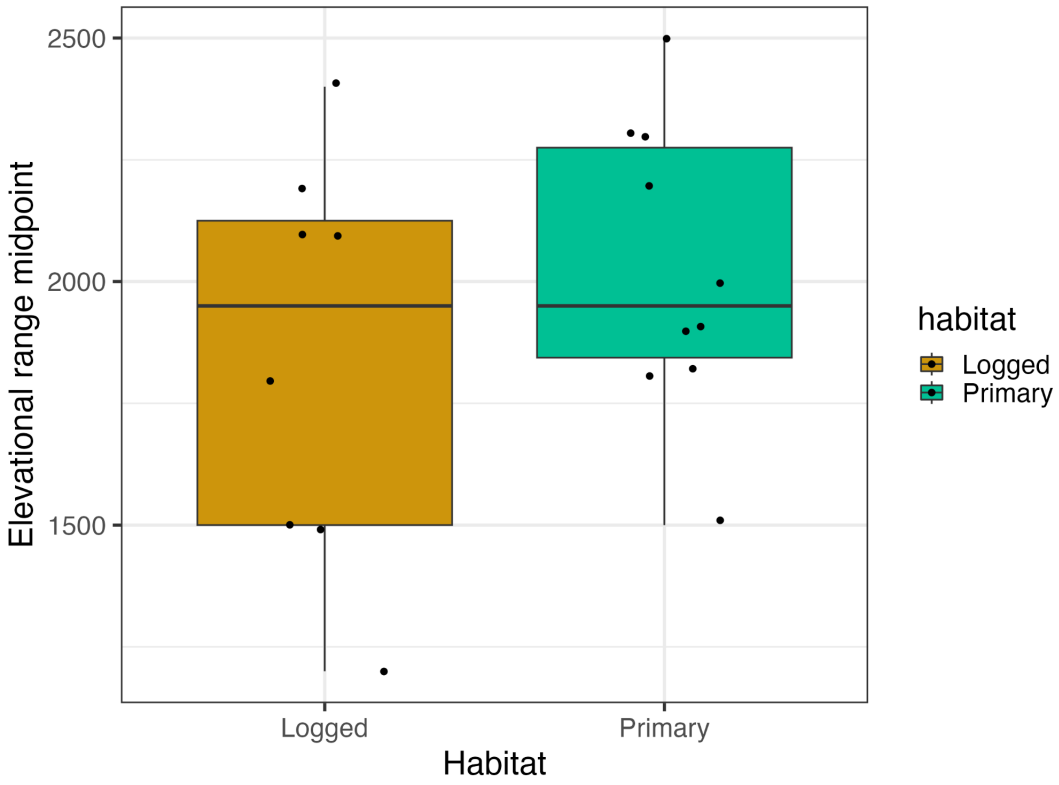

Figure 4: Mean breeding elevation range of the species that contributed to 50% dissimilarity in each habitat. The box shows the middle 50% of the data, with the vertical line inside the box indicating the median value. The whiskers of the boxplot extend from the box to show the minimum and maximum values of the data. Each dot represents individual data points and is plotted with a jitter to avoid overlap.

The mean low elevation range limits of species in the primary forest were 175m [-188.9, 538.9; R^2 = 0.06] higher than in logged forest. Similarly, the mean high elevation range limit of the species was higher in primary forest as well by 170 m [-270.3, 610.3; R^2 = 0.04], supporting the overall trend (although weak) of species representative of primary forest being "higher elevation" species than those representative of logged forest in mid-elevation montane temperate forest (See Figs. S1 and S2).

Discussion:

We investigated how climate change and selective logging might combine to impact tropical understorey bird communities in primary and selectively logged forest using a 10-year dataset from the Eastern Himalayas. We find that the bird community in primary forest was, on average, composed of larger species than in logged forest (Fig. 3). We also find weak support that the community in logged forest has a lower mean, minimum and maximum summer elevation range

than the primary forest community, which indicates that the logged forest community might resemble lower elevation communities (Fig. 4). However, this effect was not as strong as the effect for body mass.

Logging can lead to the loss of large-bodied, old-growth-dependent species and decrease overall biodiversity (Cleary et al., 2007). Understorey insectivores, which are often highly specialized, are negatively influenced by logging and show steep abundance declines (Chaves et al. 2017). Further, the body size of birds is an important factor affecting their response to logging. Birds with larger body sizes were found to be less common in logged forests, whereas smaller-bodied species were found to be more resilient to logging (Burivalova et al., 2015). We find a similar pattern, with species typical of primary forest being significantly larger than species typical of logged forest (Fig. 2). Smaller-bodied bird species were more abundant in logged forest, while larger-bodied species declined (Velho et al., 2012, Srinivasan 2013). Such patterns might arise from changes in the abiotic environment with logging leading to changes in arthropod prey composition and abundance, in turn resulting in reduced resource availability for birds (Aggarwal et al., 2022). Logged forest is warmer than old-growth forest because of changes in forest structure and microclimates (Bharadwaj et al., 2022). On average, primary forest was found to be 2.3°C cooler and 14.6% more humid than logged forest (Aggarwal et al., 2022), and have lower densities of foliage-dwelling arthropods (Aggarwal et al., 2022). Because large species have higher absolute energy requirements, a reduction in resource availability is likely to disproportionately reduce the abundance of large species (Srinivasan 2013).

The altered abiotic environment in logged forest can not only reduce the abundances of large species through indirect impacts on prey availability but can also directly influence the composition of bird communities in terms of body mass and elevational range. Climate warming is believed to have driven a decrease in the body size of warm-blooded animals and recent work in the Neotropics has found mean body mass and the mass: wing length ratio of birds decreasing over the past four decades (Jirinec et al., 2021). Smaller individuals and species might do better in warmer habitats for several reasons. Smaller individuals tend to have a higher surface area to volume ratio, which means they lose heat more quickly and can adapt to warmer temperatures more easily (Sheridan and Bickford, 2011; Speakman et al., 2003). Further, smaller organisms tend to have higher metabolic rates, which enables them to grow and reproduce faster in warmer environments (Angilletta, 2009).

Across the world's tropical mountains, bird populations are shifting to higher elevations in response to rising temperatures (Forero-Medina et al., 2011; Freeman et al., 2018; Neate-Clegg et al., 2018; Freeman and Class Freeman, 2014), including in the Eastern Himalaya (Srinivasan & Wilcove, 2021; Girish and Srinivasan, 2022). At our study site, as populations shift their elevational ranges upslope, they might encounter primary or logged forest. Our results show that logged forest bird communities are dominated by species that have a mean elevation range

midpoint that is lower than that of primary forest species, although this effect was weak (Fig. 4). Logged forest, which is consistently warmer than primary forest at our site (with a mean midday difference of 2.3°C) might be more suitable abiotically for lower-elevation, comparatively warm-adapted, species compared with primary forest.

These indirect impacts of climate change on bird species highlight the complex interplay between biotic and abiotic factors in shaping ecological communities. The weak shift in species composition of understory birds in logged forest towards communities resembling those at lower elevations suggests that while climate change will cause upslope range shifts, habitat degradation might then select for species that are adapted to the altered abiotic conditions typical of such habitats. At our site, bird populations at the cold-edge elevational range limit have increased survival rates over time, while those at the warm-edge elevational range limit suffer declines (Srinivasan and Wilcove, 2021). This suggests that species are adapting to climate change by tracking favourable conditions and shifting their ranges upslope. However, individuals of the same species are unable to maintain demographic vital rates in higher-elevation logged forests in response to climate change in the long term (Srinivasan and Wilcove, 2021). This could be due to a decrease in resource availability in logged forest which can severely impact bird demographics (Grames et al., 2023). Therefore, protecting large tracts of well-preserved primary forests across elevational gradients is essential for enabling tropical montane species to survive the dual effects of climate change and land-use change.

This study aimed to investigate the effects of climate change and selective logging on the insectivorous bird communities in tropical understorey in primary and selectively logged forests in the Eastern Himalayas of India. The results indicate that smaller species, more typical of low elevations, tended to dominate the bird communities in logged forests. This shift in body size and range may be due to various factors, such as changes in food resources, habitat, and microclimate resulting from logging and warming temperatures. This emphasizes the importance of considering long-term changes in bird communities when developing conservation strategies for bird species and their habitats. In light of our findings, continuous monitoring and management of bird habitats are crucial to prevent declines in bird diversity and abundance.

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Supplementary materials:

Supp. Table S1: Final list of bird species used for analyses after excluding high-elevation breeders which do not breed in our study area, aquatic species, carnivores, canopy dwellers, and non-invertivores (which included frugivores, granivores and nectarivores).

	Scientific Name	Species Code	English Name	Mass (g)	Forest Layer	Diet	Migrant	Flocking	Elevation - Summer Low	Elevation - Summer High	Elevation - Summer Mean
1	*Abroscopus schisticeps*	ABRSCH	Black-faced Warbler	5.78	Midstorey	Insectivore	No	Yes	1400	2400	1900
2	*Actinodura egertoni*	ACTEGE	Rusty-fronted Barwing	32.77	Midstorey	Insectivore	No	Yes	1000	2000	1500
3	*Actinodura nipalensis*	ACTNIP	Hoary-throated Barwing	46.46	Midstorey	Insectivore	No	Yes	1800	3200	2500
4	*Actinodura waldeni*	ACTWAL	Streak-throated Barwing	47.5	Midstorey	Insectivore	No	Yes	2400	3300	2850
5	*Aegithalos concinnus*	AEGCON	Black-throated Bushtit	6.14	Midstorey	Insectivore	No	Yes	1200	2600	1900
6	*Alcippe castaneceps*	ALCCAS	Rufous-winged Fulvetta	9.7	Midstorey	Insectivore	No	Yes	1800	3200	2500
7	*Alcippe cinerea*	ALCCIN	Yellow-throated Fulvetta	9.34	Understorey	Insectivore	No	Yes	1400	2400	1900
8	*Anthipes monileger*	FICMON	White-gorgeted Flycatcher	11.41	Midstorey	Insectivore	Yes	No	1600	2000	1800
9	*Blythipicus pyrrhotis*	BLYPYR	Bay Woodpecker	164.49	Midstorey	Insectivore	Yes	Yes	800	2200	1500
10	*Brachypteryx leucophris*	BRALEU	Lesser Shortwing	14.73	Terrestrial	Insectivore	Yes	No	1400	2800	2100

11	*Brachypteryx montana*	BRAMON	White-browed Shortwing	18.57	Terrestrial	Insectivore	Yes	No	2200	3400	2800
12	*Certhia discolor*	CERDIS	Brown-throated Treecreeper	10.86	Midstorey	Insectivore	Yes	Yes	160	2600	1380
13	*Cettia fortipes*	CETFOR	Brown-flanked Bush Warbler	10.23	Understorey	Insectivore	Yes	No	1400	2800	2100
14	*Cissa chinensis*	CISCHI	Common Green Magpie	127	Midstorey	Omnivore	No	No	400	2000	1200
15	*Dendrocitta frontalis*	DENFRO	Collared Treepie	80.99	Midstorey	Omnivore	No	No	600	2000	1300
16	*Ficedula hyperythra*	FICHYP	Snowy-browed Flycatcher	8.65	Understorey	Insectivore	Yes	No	1800	2800	2300
17	*Ficedula westermanni*	FICWES	Little Pied Flycatcher	7.8	Midstorey	Insectivore	Yes	No	1000	2400	1700
18	*Garrulax caerulatus*	GARCAE	Grey-sided Laughingthrush	80.61	Understorey	Omnivore	No	Yes	1600	2400	2000
19	*Harpactes erythrocephalus*	HARERY	Red-headed Trogon	100.81	Midstorey	Omnivore	Yes	Yes	400	2000	1200
20	*Hydrornis nipalensis*	PITNIP	Blue-naped Pitta	124	Terrestrial	Insectivore	Yes	No	1400	1800	1600
21	*Leiothrix argentauris*	LEIARG	Silver-eared Leiothrix	21.89	Understorey	Omnivore	No	No	600	1800	1200
22	*Leiothrix lutea*	LEILUT	Red-billed Leiothrix	22.36	Understorey	Omnivore	No	No	1800	3200	2500
23	*Liocichla bugunorum*	LIOBUG	Bugun Liocichla	39	Midstorey	Omnivore	No	Yes	2000	2600	2300
24	*Lioparus chrysotis*	ALCCHR	Golden-breasted Fulvetta	8.12	Understorey	Insectivore	No	Yes	2000	2800	2400
25	*Machlolophus spilonotus*	PARSPI	Yellow-cheeked Tit	18.39	Midstorey	Insectivore	Yes	Yes	1200	2450	1825

26	*Muscicapa ferruginea*	MUSFER	Ferruginous Flycatcher	10.21	Midstorey	Insectivore	Yes	No	1400	2800	2100
27	*Muscicapella hodgsoni*	MUSHOD	Pygmy Flycatcher	5.76	Midstorey	Insectivore	Yes	No	1400	2600	2000
28	*Myiomela leucura*	MYILEU	White-tailed Robin	25.72	Terrestrial	Insectivore	Yes	No	1200	2200	1700
29	*Niltava grandis*	NILGRA	Large Niltava	36.03	Midstorey	Insectivore	Yes	No	1600	2200	1900
30	*Niltava macgrigoriae*	NILMAC	Small Niltava	11.31	Understorey	Insectivore	Yes	No	800	2000	1400
31	*Niltava sundara*	NILSUN	Rufous-bellied Niltava	20.2	Understorey	Insectivore	Yes	No	1600	3000	2300
32	*Parus monticolus*	PARMON	Green-backed Tit	18.65	Midstorey	Insectivore	Yes	Yes	1200	3000	2100
33	*Phyllergates cuculatus*	ORTCUC	Mountain Tailorbird	6.31	Understorey	Insectivore	Yes	Yes	1400	2200	1800
34	*Phylloscopus chloronotus*	PHYCHL	Lemon-rumped Warbler	5.01	Midstorey	Insectivore	Yes	Yes	2400	3600	3000
35	*Phylloscopus reguloides*	PHYREG	Blyth's Leaf Warbler	7.05	Midstorey	Insectivore	Yes	Yes	1600	3200	2400
36	*Pnoepyga pusilla*	PNOPUS	Pygmy Wren-babbler	12.25	Understorey	Insectivore	No	Yes	1000	2600	1800
37	*Pomatorhinus ferruginosus*	POMFER	Coral-billed Scimitar Babbler	40.65	Understorey	Insectivore	No	Yes	1000	2000	1500
38	*Pomatorhinus ruficollis*	POMRUF	Streak-breasted Scimitar Babbler	29.76	Midstorey	Insectivore	Yes	Yes	1600	2600	2100
39	*Pomatorhinus superciliaris*	XIPSUP	Slender-billed Scimitar Babbler	30.12	Understorey	Omnivore	Yes	No	1600	3400	2500

40	*Psarisomus dalhousiae*	PSADAL	Long-tailed Broadbill	55.06	Midstorey	Insectivore	Yes	Yes	400	2000	1200
41	*Psittiparus ruficeps*	PARRUF	White-breasted Parrotbill	31.54	Midstorey	Insectivore	No	Yes	1000	2000	1500
42	*Rhipidura albicollis*	RHIALB	White-throated Fantail	12.02	Midstorey	Insectivore	Yes	Yes	1200	2400	1800
43	*Rimator malacoptilus*	RIMMAL	Long-billed Wren-Babbler	19.69	Understorey	Insectivore	Yes	Yes	900	2700	1800
44	*Sasia ochracea*	SASOCH	White-browed Piculet	10.03	Midstorey	Insectivore	Yes	Yes	400	2000	1200
45	*Seicercus affinis*	SEIAFF	White-spectacled Warbler	7.58	Midstorey	Insectivore	Yes	Yes	1000	2800	1900
46	*Seicercus poliogenys*	SEIPOL	Grey-cheeked Warbler	6.57	Midstorey	Insectivore	Yes	Yes	1400	2400	1900
47	*Spelaeornis caudatus*	SPECAU	Rufous-throated Wren-Babbler	10.53	Terrestrial	Insectivore	No	Yes	1600	2400	2000
48	*Spelaeornis formosus*	ELAFOR	Spotted Wren-Babbler	10.79	Terrestrial	Insectivore	No	No	800	2200	1500
49	*Sphenocichla humei*	SPHHUM	Sikkim Wedge-billed Babbler	34.1	Understorey	Insectivore	No	Yes	1400	2000	1700
50	*Stachyridopsis chrysaea*	STACHR	Golden Babbler	7.85	Understorey	Insectivore	No	Yes	800	2000	1400
51	*Stachyridopsis ruficeps*	STARUF	Rufous-capped Babbler	9.85	Understorey	Insectivore	Yes	Yes	1200	1800	1500
52	*Stachyris nigriceps*	STANIG	Grey-throated Babbler	14.17	Understorey	Insectivore	No	Yes	400	1800	1100
53	*Suthora nipalensis*	PARNIP	Black-throated Parrotbill	6.31	Understorey	Insectivore	No	Yes	1600	2800	2200

54	*Tesia castaneocoronata*	TESCAS	Chestnut-headed Tesia	8.31	Terrestrial	Insectivore	Yes	No	2200	3400	2800
55	*Tesia cyaniventer*	TESCYA	Grey-bellied Tesia	9.89	Terrestrial	Insectivore	Yes	No	1400	2600	2000
56	*Tesia olivea*	TESOLI	Slaty-bellied Tesia	8.79	Terrestrial	Insectivore	Yes	Yes	1400	2000	1700
57	*Tickellia hodgsoni*	TICHOD	Broad-billed Warbler	5.88	Understorey	Insectivore	Yes	Yes	1600	2600	2100
58	*Trochalopteron erythrocephalum*	GARERY	Chestnut-crowned Laughingthrush	71.63	Understorey	Omnivore	No	Yes	1600	3400	2500
59	*Trochalopteron squamatum*	GARSQU	Blue-winged Laughingthrush	81.48	Understorey	Omnivore	No	Yes	1000	2000	1500
60	*Trochalopteron subunicolor*	GARSUB	Scaly Laughingthrush	66	Understorey	Omnivore	No	Yes	2000	3400	2700
61	*Yuhina flavicollis*	YUHFLA	Whiskered Yuhina	14.97	Midstorey	Omnivore	Yes	No	1600	2800	2200

Supp. Table S2: Detention corrected raw counts of species in primary and logged forest. The first ten columns are from primary, followed by logged.

	Species Code	**2011.P**	**2012.P**	**2013.P**	**2014.P**	**2015.P**	**2016.P**	**2017.P**	**2018.P**	**2019.P**	**2021.P**	**2011.L**	**2012.L**	**2013.L**	**2014.L**	**2015.L**	**2016.L**	**2017.L**	**2018.L**	**2019.L**	**2021.L**
1	ABRSCH	3	5	4	10	3	2	5	0	1	5	3	2	0	2	1	6	5	3	3	3
2	ACTEGE	4	9	13	4	8	4	6	2	4	6	1	7	5	6	6	13	6	14	5	10
3	ACTNIP	0	0	0	0	0	0	0	0	0	1	0	0	0	0	0	0	0	1	0	0
4	ACTWAL	1	0	2	0	0	1	0	0	0	0	0	2	0	0	1	1	0	0	0	0

5	AEGCON	0	1	0	0	1	0	3	0	0	0	0	0	0	0	0	0	0	0	0	2
6	ALCCAS	16	32	14	17	16	20	17	11	7	29	41	23	10	12	19	22	21	27	17	48
7	ALCCHR	11	10	10	4	8	16	13	6	6	29	22	23	11	14	21	27	24	16	16	48
8	ALCCIN	16	50	28	10	24	32	16	16	6	42	50	30	26	7	25	18	26	23	7	70
9	BLYPYR	1	0	0	0	0	0	0	0	1	0	0	0	0	0	0	0	0	1	0	0
10	BRALEU	4	1	0	2	4	1	5	2	4	3	3	2	7	1	2	5	7	3	5	5
11	BRAMON	0	0	0	0	0	0	0	1	0	2	0	1	0	0	0	0	0	0	0	1
12	CERDIS	0	1	2	0	2	2	0	1	0	1	0	0	0	0	0	0	0	0	0	0
13	CETFOR	0	1	0	1	1	4	0	1	1	3	1	0	1	2	2	2	1	4	2	2
14	CISCHI	0	0	1	0	0	0	0	0	0	0	0	1	0	0	0	0	0	0	0	0
15	DENFRO	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	1
16	ELAFOR	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	1
17	FICHYP	18	13	8	13	11	14	10	7	6	18	19	13	4	2	15	10	10	16	4	12
18	FICMON	2	3	5	3	5	6	6	1	2	7	1	0	0	2	6	2	4	3	3	9
19	FICWES	0	0	0	0	0	0	0	0	0	0	1	0	0	0	0	0	0	0	0	0
20	GARCAE	3	2	1	1	3	3	5	3	1	11	4	6	1	0	2	4	3	4	1	2
21	GARERY	3	7	6	2	2	5	2	2	6	8	4	4	3	4	2	6	2	8	4	11
22	GARSQU	0	0	1	0	0	0	0	0	0	1	0	0	0	0	4	0	2	3	0	0
23	GARSUB	1	0	0	0	0	0	0	0	0	0	0	0	0	1	0	0	0	0	0	0
24	HARERY	0	2	0	0	1	0	0	0	2	1	1	0	0	0	0	0	0	0	0	0

25	LEIARG	0	0	0	0	0	0	0	0	0	0	1	0	0	0	1	2	0	0	1	0
26	LEILUT	0	0	0	0	0	0	0	2	0	0	0	0	1	0	0	2	0	0	0	0
27	LIOBUG	0	0	0	0	1	0	0	0	0	0	0	0	0	0	0	2	0	0	0	0
28	MUSFER	0	0	0	0	0	0	0	0	0	0	2	2	0	0	1	0	0	0	0	0
29	MUSHOD	1	0	0	3	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
30	MYILEU	1	3	5	1	6	6	5	4	5	8	3	4	2	1	5	5	9	9	9	6
31	NILGRA	7	10	8	3	7	4	6	7	5	13	5	8	3	1	3	5	4	7	2	10
32	NILMAC	0	1	0	0	0	0	0	0	0	1	1	5	0	0	0	0	1	0	0	6
33	NILSUN	16	1	10	4	11	12	16	4	2	17	11	0	4	1	6	6	5	4	1	2
34	ORTCUC	0	2	0	0	1	0	4	0	0	0	0	4	1	3	3	2	4	3	0	0
35	PARNIP	17	33	14	8	7	17	7	6	11	26	33	7	8	1	16	18	32	19	13	26
36	PARRUF	0	0	9	1	2	5	2	0	2	0	5	1	0	0	1	0	1	1	0	4
37	PARSPI	2	0	5	0	1	0	1	0	2	1	0	0	1	0	0	1	0	0	0	3
38	PHYCHL	0	1	0	0	0	0	0	0	0	1	3	0	0	0	0	0	0	1	0	1
39	PHYREG	7	7	6	4	3	2	9	2	1	14	6	9	1	3	0	9	10	5	2	9
40	PITNIP	0	0	0	0	1	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
41	PN0PUS	0	0	0	0	0	0	0	0	0	2	0	0	0	0	0	0	0	0	0	0
42	PNOPUS	5	2	2	0	1	0	1	0	1	2	1	1	0	0	0	1	1	0	0	1
43	POMFER	0	1	2	1	0	3	1	4	0	2	1	2	2	2	3	0	1	0	0	1
44	POMRUF	3	4	0	2	2	1	3	2	1	2	6	3	0	1	4	3	8	5	2	7
45	PSADAL	0	0	0	0	0	0	0	1	0	1	0	0	0	0	0	0	0	0	0	0

46	RHIALB	2	2	2	1	0	4	3	2	2	9	8	3	3	1	0	1	5	5	4	3
47	RIMMAL	2	0	0	0	2	1	0	0	0	0	0	0	1	0	1	0	0	0	0	0
48	SASOCH	0	0	0	0	0	2	1	2	1	2	0	1	1	0	1	3	2	2	2	6
49	SEIAFF	9	8	8	2	6	6	7	5	6	15	12	9	3	2	6	6	4	4	2	13
50	SEIPOL	10	13	4	7	4	11	11	4	8	24	10	9	4	7	5	14	8	5	3	22
51	SPECAU	3	2	1	1	7	5	4	2	1	7	5	1	0	0	2	0	3	3	2	3
52	SPHHUM	1	0	0	0	0	0	0	0	0	0	0	1	0	0	0	0	0	0	0	0
53	STACHR	8	15	10	4	6	14	9	8	3	19	9	8	7	7	16	15	14	11	10	19
54	STANIG	0	0	0	0	0	0	0	0	0	0	0	1	0	0	0	0	0	0	0	0
55	STARUF	12	16	12	10	8	16	16	7	7	19	33	31	17	16	38	31	26	29	13	37
56	TESCAS	1	0	2	3	0	0	0	0	0	3	5	6	0	1	0	0	0	0	0	8
57	TESCYA	4	4	2	1	3	2	3	3	0	1	11	3	2	2	4	2	1	1	0	6
58	TESOLI	2	1	2	0	3	3	2	3	1	2	3	2	0	2	2	3	0	1	1	5
59	TICHOD	0	1	0	2	0	0	0	0	0	3	4	3	1	1	2	2	1	1	0	6
60	XIPSUP	0	0	2	0	2	1	0	2	2	5	3	5	0	0	1	0	1	1	2	5
61	YUHFLA	12	0	0	1	0	1	2	2	2	9	6	3	8	2	2	3	9	1	1	14

Supp. Table S3: Similarity percentage (SIMPER) routines of species that contribute to 50% cumulative dissimilarity between the two habitats.
Note: *Schoeniparus cinereus* and *Tesia castaneocoronata* have not been included in Figure 2 (NMDS)

Species	Species code	average	sd	ratio	ava	avb	cumsum	*p* value	significance
Niltava sundara	NILSUN	0.009681	0.006145	1.5754	0.2031	0.1182	0.038	0.008	**
Stachyris ruficeps	STARUF	0.009032	0.004001	2.2573	0.249	0.3419	0.074	0.001	***
Yuhina flavicollis	YUHFLA	0.007835	0.005886	1.3311	0.0912	0.1336	0.105	0.227	
Lioparus chrysotis	ALCCHR	0.007678	0.004217	1.8208	0.2292	0.3063	0.136	0.001	***
Phyllergates cuculatus	ORTCUC	0.007159	0.005321	1.3454	0.0304	0.0818	0.164	0.048	*
Psittiparus ruficeps	PARRUF	0.006938	0.005376	1.2905	0.0792	0.0472	0.192	0.239	
Schoeniparus cinereus	ALCCIN	0.00692	0.004932	1.4032	0.3324	0.3265	0.219	0.879	
Tickellia hodgsoni	TICHOD	0.006553	0.003706	1.7682	0.0277	0.0825	0.245	0.005	**
Suthora nipalensis	PARNIP	0.00652	0.005306	1.2288	0.2616	0.2528	0.271	0.944	
Spelaeornis caudatus	SPECAU	0.006052	0.004935	1.2263	0.1216	0.07	0.295	0.033	*
Abroscopus schisticeps	ABRSCH	0.005949	0.005781	1.029	0.1275	0.1021	0.318	0.685	
Tesia castaneocoronata	TESCAS	0.005933	0.005626	1.0546	0.0413	0.0508	0.342	0.811	
Pomatorhinus superciliaris	XIPSUP	0.005774	0.004417	1.3074	0.0651	0.0655	0.365	0.842	
Pomatorhinus ferruginosus	POMFER	0.00576	0.004343	1.3263	0.0688	0.0625	0.388	0.922	
Brachypteryx leucophris	BRALEU	0.005728	0.004386	1.3059	0.1109	0.1312	0.41	0.433	
Anthipes monileger	FICMON	0.005618	0.004838	1.1611	0.1383	0.0966	0.433	0.121	
Machlolophus spilonotus	PARSPI	0.005486	0.004884	1.1234	0.0589	0.023	0.455	0.183	
Sasia ochracea	SASOCH	0.005414	0.004066	1.3315	0.0459	0.0739	0.476	0.234	
Ficedula hyperythra	FICHYP	0.005386	0.004097	1.3149	0.245	0.2015	0.497	0.038	*
Niltava grandis	NILGRA	0.005301	0.003144	1.6863	0.1875	0.1365	0.518	0.001	***

Significance codes: '***' = p < 0.001, '**' = p < 0.01, '*' = p < 0.05
average = average contribution of the species to between-group dissimilarity
ava, avb = average abundances of each habitat (primary and logged)

Figure S1: Highest breeding elevation range of the species that contributed to 50% dissimilarity in each habitat. The box shows the middle 50% of the data, with the vertical line inside the box indicating the median value. The whiskers of the boxplot extend from the box to show the minimum and maximum values of the data. Each dot represents individual data points and is plotted with a jitter to avoid overlap.

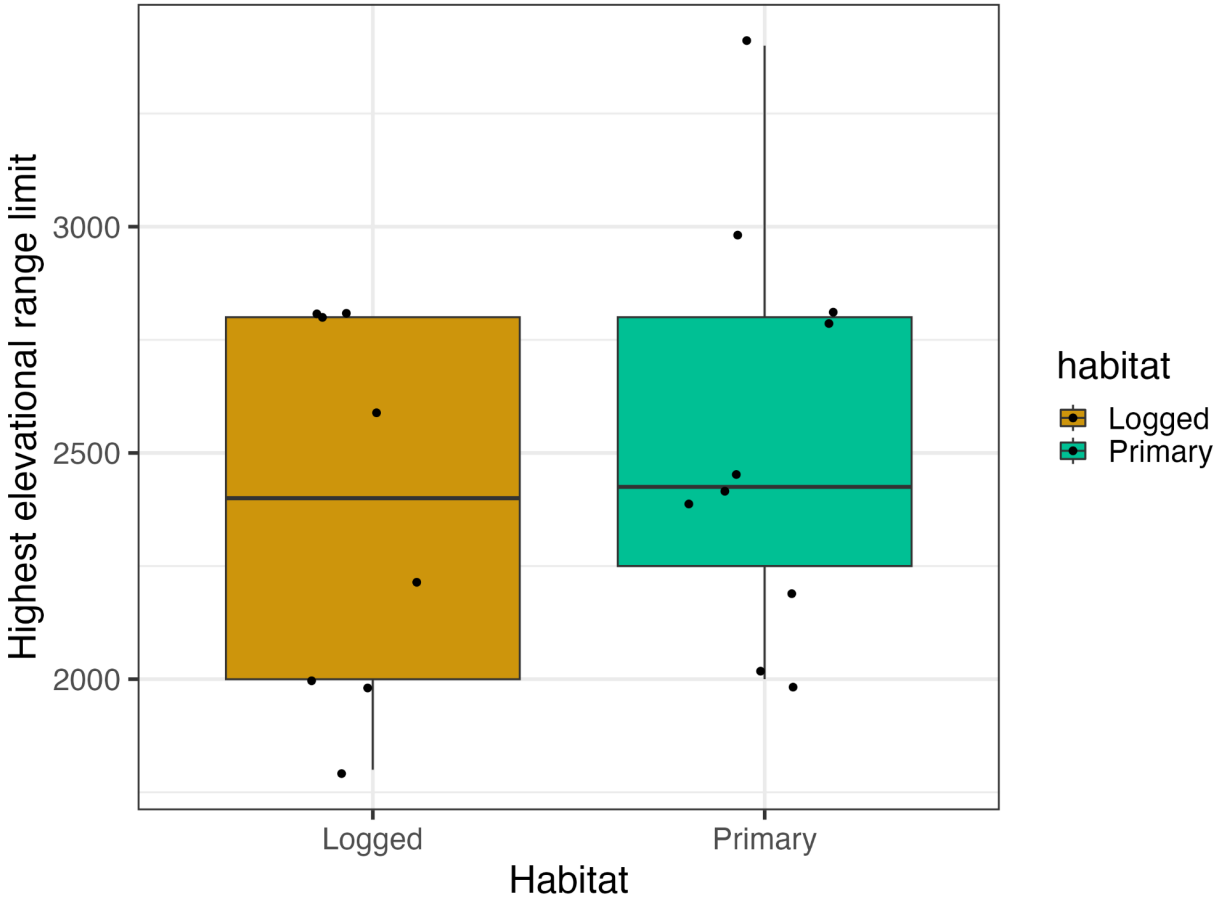
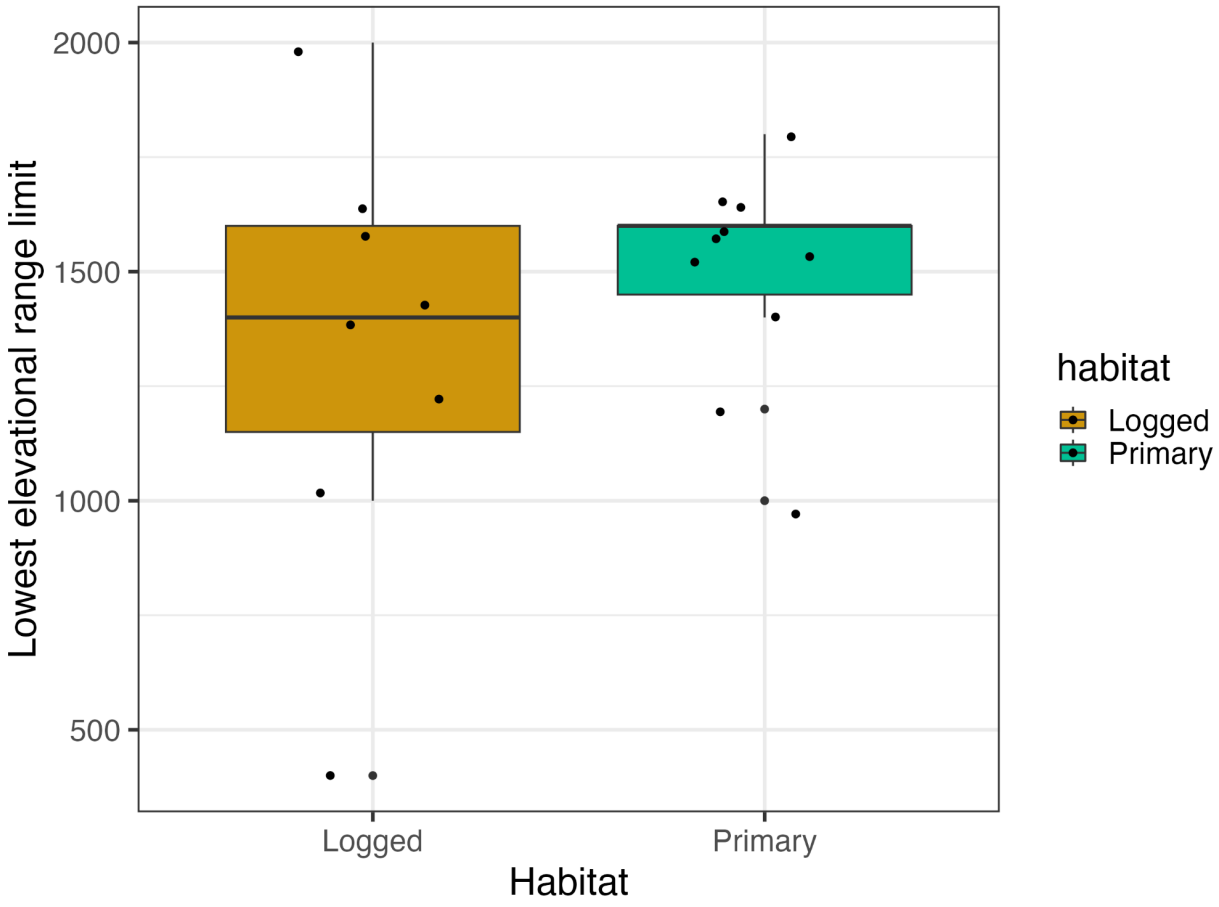


Figure S2: Lowest breeding elevation range of the species that contributed to 50% dissimilarity in each habitat. The box shows the middle 50% of the data, with the vertical line inside the box indicating the median value. The whiskers of the boxplot extend from the box to show the minimum and maximum values of the data. Each dot represents individual data points and is plotted with a jitter to avoid overlap.



Supp Table S4: Recapture probabilities for "small" and "large" species in primary and logged forest.

	2012	2013	2014	2015	2016	2017	2018	2019	2021
intact.large	0.12	0.38	0.12	0.16	0.29	0.47	0.48	0.29	0.62
logged.large	0.09	0.07	0.10	0.18	0.21	0.30	0.15	0.21	0.38
intact.small	0.56	0.39	0.32	0.27	0.44	0.51	0.45	0.30	0.90

logged.small	0.42	0.30	0.36	0.40	0.38	0.39	0.44	0.27	0.60