

19 **1 Abstract**

20 Lyme disease is the most common wildlife-to-human transmitted disease reported
21 in North America. The study of this disease requires an understanding of the ecology of the
22 complex communities of ticks and host species involved in harboring and transmitting this
23 disease. Much of the ecology of this system is well understood, such as the life cycle of
24 ticks, and how hosts are able to support tick populations and serve as disease reservoirs, but
25 there is much to be explored about how the population dynamics of different host species
26 and communities impact disease risk to humans. One way in which we can study disease
27 effectively is through the use of theoretical models. These are powerful tools that allow in-
28 vestigation of complex species interactions before staging complicated and expensive stud-
29 ies that may not be productive. We construct a model to investigate how host population
30 dynamics can affect disease risk to humans. The model describes a tick population and a
31 simple community of three species in which mouse populations are made to fluctuate on an
32 annual basis. We tested the model under different environmental conditions to examine the
33 effect of environment on the interactions of host dynamics and disease risk. Results indi-
34 cate that host dynamics reduce mean nymphal infection prevalence and increase the yearly
35 amplitude of nymphal infection prevalence and the density of infected nymphs. Effects
36 were nonlinear and patterns in the effect of dynamics on amplitude in nymphal infection
37 prevalence varied across locations. These results highlight the importance of further study
38 of the effect of community dynamics on disease risk. This will involve the construction of
39 further theoretical models and collection of robust field data to inform these models. With
40 a more complete understanding of disease dynamics we can begin to better determine how
41 to predict and manage disease risk using these models.

42 **2 Introduction**

43 Emerging infectious diseases present an ever-growing threat to human health,
44 agriculture, and native flora and fauna [1–3]. In humans, around 75% of emerging diseases
45 are wildlife zoonoses. These diseases cause millions of cases each year, and many new
46 diseases are emerging or reemerging every year. Across species, diseases spread by vector
47 organisms are the most common, and a growing proportion of new emerging diseases in
48 humans are vector-borne [1, 2, 4, 5]. These disease systems involve complex dynamics
49 between wildlife hosts, vector organisms, and the disease itself. The study of vector-borne
50 diseases is of interest to many fields, including ecology, epidemiology, climatology, and
51 many other disciplines.

52 A key problem is understanding how to predict and control the proliferation of
53 vector-borne diseases. With the complexity of the different organisms involved in these
54 diseases, they are best thought of as webs of different interacting species [3, 6]. Study-
55 ing the community ecology of these disease systems is fundamental to understanding how
56 they develop and spread. The community ecology of disease is a rapidly developing field,
57 but still relatively in its infancy [3]. Much research has focused on "simpler" applications
58 of community and population ecology, such as single host-parasite interactions. The tools
59 of community ecology allow an understanding of the mechanisms leading to the particu-
60 lar community assemblages we see today and the dynamics that shape the systems across
61 time and space, including the emergence and spread of vector-borne diseases [4, 5]. This
62 has particular benefit in informing how host communities and parasite communities de-
63 velop and interact between levels and different scales, from within the individual to across
64 ecosystems.

65 In studies on vector-borne diseases, research has aimed to understand the impact
66 of different species and species communities on disease risk [6–15]. The impact of host

67 and community dynamics on Lyme disease risk is fairly uncertain considering the known
68 and widely implicated potential of these dynamics to have a significant impact on how *B.*
69 *burgdorferi* flows through wildlife communities. Rodents in particular are known to serve
70 as important reservoir hosts for *B. burgdorferi*, and populations of these species often un-
71 dergo extreme fluctuations in density driven by resources, predators, and environmental
72 effects [6, 16–18]. Other important host species undergo population dynamics and many
73 species are subject to changing populations and extinctions as a result of human caused
74 disturbances [6, 16]. Research has been equivocal in determining the effect of host dynam-
75 ics on disease, showing no effect or an inconsistent effect [19–22]. Understanding this pro-
76 cess will allow us to better understand how different systems experience disease risk, and
77 inform how disease may vary regionally or as species extinctions and invasions develop.

78 Lyme disease is one of the most prominent wildlife vector-borne diseases impact-
79 ing health in the United States, with estimates of as 476,000 cases per year [23]. Disease
80 models of Lyme disease treat hosts as static parameter values, with only a few models in-
81 vestigating the impact of using dynamic population data, and none investigating the impact
82 that dynamics have on this system [21, 24–27]. To investigate the impact of host dynamics
83 on Lyme disease risk, we have created a mathematical model which simulates the flow of
84 disease between a detailed tick population and a small community of hosts. To investigate
85 the impact of host dynamics on the model, densities of the primary rodent host, mice, are
86 varied on an annual basis. By changing the mean density and variation from the mean in a
87 collection of simulations, we will use our model to explore the impact that simple host dy-
88 namics have on disease risk measures of the model for a range of possible scenarios. This
89 model will direct future research into the effect of host dynamics and disease risk and high-
90 light the need for the further development of models and empirical studies investigating
91 this topic.

92 **3 Results**

93 **3.1 Site-specific effects of variance**

94 We examined model data grouped by simulations run with zero variance (con-
95 stant mouse densities) or the entire range of variances (1-99 mice) (S1 Table). *DIN* and
96 *DON* show similar patterns between groups, with the maximum and amplitude of each
97 being much larger in simulations with variance than those with constant mouse density.
98 Further analysis continued to show similar effects between *DIN* and *DON*, and *DIN* is of a
99 more direct concern, so we will focus on results for *DIN* as a proxy for both. *NIP* metrics
100 show lowered averages and medians for mean, min, and max *NIP*, and increased ampli-
101 tude between variance and no variance groups. Minimum and maximum values remained
102 constant for most *NIP* metrics.

103 Mean *DIN* shows a positive linear relationship with mean mouse density, and a
104 variable relationship with variance, ranging from a positive and slightly nonlinear relation-
105 ship at low means to a negative and slightly non-linear relationship at high density (Fig 1A).
106 Similar positive relationships with variance and mean mouse density are demonstrated for
107 amplitude, minimum, and maximum *DIN* (Figs 2A and 3A).

108 **Fig 1. Response of mean density of infected nymphs and nymphal infection**
109 **prevalence to variance in the mouse population and mean mouse density.** A) shows
110 the response of mean *DIN* to mouse variation and mean density. The relationship to mean
111 mouse density is positive and linear, while the relationship to variance is weakly nonlinear,
112 and shifts from a positive relationship to a negative relationship at low or high mean den-
113 sity, respectively. Contour lines plotted on the variance \times mean plane and coloring reveal
114 surface features. B) shows the response of mean *NIP* to mouse variation and mean density.
115 The relationship between mean *NIP* and mean mouse density is a positive saturating curve

116 under low variation in the mouse population, which becomes a near linear relationship for
117 high variance. The relationship between variance and *NIP* is negative, with higher variance
118 decreasing *NIP* at the given mean mouse density. At low mean densities, very high variance
119 begins to slow this decrease and even reverse it. Contour lines and coloration again reveal
120 surface features.

121 **Fig 2. Response of amplitude in density of infected nymphs and nymphal in-**
122 **fection prevalence to variance in the mouse population and mean mouse density.** A)
123 shows the response of amplitude in *DIN* to mouse variation and mean density. The relation-
124 ship to mean mouse density is positive and linear, while the relationship to mouse variance
125 appears close to linear and is positive. Contour lines plotted on the variance \times mean plane
126 and coloring reveal surface features. B) shows the response of amplitude in *NIP* to mouse
127 variation and mean density. The relationship between *NIP* and mean mouse density is a
128 positive saturating curve under low variation in the mouse population, which becomes lin-
129 ear as variance increases. The relationship is a positive saturating curve between variance
130 and amplitude in *NIP*. At high mean densities the relationship to variance is near linear, and
131 shows signs of beginning to become negative at very high means and variances. Contour
132 lines and coloration again reveal surface features.

133 **Fig 3. Response of minimum density of infected nymphs and nymphal in-**
134 **fection prevalence to variance in the mouse population and mean mouse density.** A)
135 shows the response of minimum *DIN* to mouse variation and mean density. The relationship
136 to mean mouse density is positive and linear, while the relationship to mouse variance ap-
137 pears weakly nonlinear, and shifts from a slight positive relationship at low mean density to
138 a slight negative relationship at high mean density. Contour lines plotted on the variance \times
139 mean plane and coloring reveal surface features. B) shows the response of minimum *NIP*
140 to mouse variation and mean density. There is a nonlinear positive relationship between
141 minimum *NIP* and mean mouse density. The relationship between variance and minimum

142 *NIP* is negative, with higher variance decreasing minimum *NIP* at the given mean mouse
143 density. At low mean densities, very high variance shows the slightest sign of beginning to
144 increase the minimum *NIP*. Contour lines and coloration again reveal surface features.

145 Measures for *NIP* show nonlinear relationships with both mean mouse density
146 and variance. These measures are positively related with mean mouse density (Figs 1B, 2B,
147 and 3B). At low variance this relationship is a saturating curve, while with higher variance
148 this relationship becomes linear. Mean *NIP* decreases with variance, but high variance at
149 low means begins to slow and eventually reverse this decrease (Fig 1B). The minimum
150 point of this shift changes to higher variance as the mean mouse density increases, with
151 no minimum appearing in the range of explored variances for higher mean densities. The
152 amplitude of *NIP* follows a saturating curve across variance levels, reaching a maximum
153 and decreasing for high variance and high mean mouse density (Fig 2B). The minimum
154 *NIP* shows a strong negative response to variance (Fig 3B), while the maximum shows an
155 unclear and relatively weak effect (S1 Fig), with a slight positive relationship at low mean
156 mouse density shifting to a slightly negative relationship at high density.

157 The distributions of mean, minimum, and maximum *NIP* shows a negative shift
158 in the median *NIP* measure with increasing variance, with this shift being strongest in
159 minimum *NIP* (S1 Table). Median amplitude in *NIP* shows a positive shift with increasing
160 variance (S1 Table). The median of mean and minimum *DIN* remains fairly stable, while
161 amplitude and maximum *DIN* increase with variance (S1 Table). The median of mean *DIN*
162 remains nearly constant, while amplitude *DIN* increases, and the median of minimum and
163 maximum *DIN* decrease and increase, respectively with variance (S1 Table).

164 **3.2 Effects of variance across variable weather inputs**

165 We examined disease metrics across the selected locations to investigate patterns
166 between sites and general geographic regions. *DIN* measures showed somewhat of an in-

167 crease in mean, minimum, and maximum *DIN* under warmer environmental conditions, but
168 showed no discernible changes in the pattern of this effect. Mean, minimum, and maximum
169 *NIP* across locations shows very similar effects of mouse population variance, with the
170 range and interquartile range remaining similar while the median shifts down in response
171 to variance. This effect is apparent between geographic regions as well, though regions
172 differ in their distributions for both *DIN* and *NIP*. Amplitude of *DIN* has a consistent pat-
173 tern between locations and regions with or without variance (Fig 4A). *NIP* amplitude has a
174 markedly similar response to variance across locations and regions despite variability with
175 stable host populations (Fig 4B).

176 **Fig 4. Impact of variance in mouse population on amplitude of density of in-**
177 **fecting nymphs and nymphal infection prevalence under varying environmental con-**
178 **ditions.** A) displays the distributions of amplitude in *DIN* for locations with no variance
179 or variance for medium variance levels ($\sigma = 33 - 67$). The left column shows boxplots
180 for each location, while the right column shows boxplots for locations grouped by rough
181 geographic area, with "West" indicating sites on the western range of Lyme disease and
182 Eastern black-legged ticks. B) displays the same information for amplitude in *NIP*.

183 **3.3 Sensitivity Analysis**

184 Parameters for temperature induced activity and survival were important param-
185 eters for all measures of *DIN*, *DON*, and *NIP*. Parameters for maximum on-host survival
186 of immature ticks and hardening nymphs and engorgement index also were correlated with
187 disease measures. Weekly cumulative degree week thresholds also proved to be important,
188 as well as parameters related to host survival rates.

189 **4 Discussion**

190 Host dynamics appear to have little effect on the density of infected ticks, while
191 the proportion of infected ticks is affected. Our knowledge of the complexity of the species
192 involved in these disease systems indicates that the study of host dynamics is important to
193 our understanding of disease risk to humans. Yet, few studies have thoroughly investigated
194 this. Here, we have investigated this possibility by developing a Lyme disease model which
195 incorporates simulated host dynamics in rodent hosts, and investigating the impacts on dis-
196 ease risk predictions. The study of emerging infectious diseases continues to become more
197 important as increasing numbers of diseases emerge or reemerge. Vector-borne diseases
198 pose a particular threat with expanding ranges and the complex dynamics of spread.

199 **4.1 Associations Between Host Dynamics and Disease Risk**

200 Model results indicate that variance in the mouse population has only a small and
201 inconsistent effect on mean *DIN* and *DON*. While somewhat counterintuitive, this result is
202 not as surprising as it may seem. With the high association between small mammal den-
203 sity and *DIN* [16, 19], host density is a very reliable predictor of tick density. Tick density
204 tracks host density, rising and falling in response to corresponding changes in the host pop-
205 ulation. The result is that mean tick density trends to the average expected from the mean
206 host density in the model. Modeling studies have demonstrated that incorporating long-
207 term empirical data on host density yields more accurate predictions of tick density when
208 predicting disease risk at specific sites [21]. The effect on overall model output has not
209 been determined, with research focusing on the qualitative accuracy of predictions on short
210 timescales rather than multi-year simulations or investigation of model behavior. Studies
211 that have considered host dynamics have focused solely on the effect of host density on dis-
212 ease risk rather than the effect of changing host density on disease dynamics. The results of

213 our model still align with this research, and demonstrate the direct impact of host density
214 on *DIN* and *DON*.

215 Long term measurements of interannual variation in infection prevalence is lack-
216 ing, and few modeling studies report yearly predicted *NIP* e.g. ????? [21]. Available re-
217 sults show little variation in mean *NIP* with a constant host population, while simulations
218 incorporating field data on host populations demonstrate lowered and more variable *NIP*.
219 Empirical studies demonstrate this same variability in annual *NIP*, [19, 28, 29] which our
220 results align with. Studies which examine variation in the host population also focus on
221 the predictive ability of host density for *NIP*. The density of small mammal hosts has been
222 found to be a poor predictor of disease risk [16, 19]. On the other hand, resources such
223 as acorns can be a good predictor of mean *NIP*. This likely results from direct changes in
224 the proportion of competent host species in the system, e.g. seed predators. The density of
225 hosts does not necessarily indicate changing host proportions, and may instead result from
226 community-wide effects that change overall host density. In this model, we have demon-
227 strated the scenario in which the proportion of competent hosts is changing by directly
228 changing the density of a single host, and thus the proportional density, of hosts in the
229 model system. The resulting effect on mean *NIP* is apparent in the results, and demon-
230 strates the importance of considering relative host densities of competent and incompetent
231 species rather than focusing on specific hosts or total host density in a system.

232 Results further show that the strength of host density variability is important, with
233 a parabolic response of *NIP* to variance apparent in model results (Fig 1B). This likely re-
234 sults because all means $0 - 99 \text{ mice}^{-ha^2}$ are simulated, and for each mean, variance from
235 $0 - 99 \text{ mice}^{-ha^2}$ is also simulated. For mean host densities < 99 , the range of variance will
236 overlap with 0 from $0 - 99$. Mouse density is bounded by 0, as negative density of course
237 makes no sense. With greater variance the mean density of mice in a simulation will be
238 higher than the mean of the distribution. The decrease of *NIP* in response to variance be-

239 gins to flatten out with boost in mouse density, and eventually even starts to increase *NIP*
240 as the proportion of mice shifts. Figs 1-3 could rather show the relation between true simu-
241 lated mouse density, variance, and *NIP*, but this neglects to show the nature of the response
242 of *NIP* to variance at a simulated mean. The variance must exceed the mean mouse den-
243 sity before mean *NIP* begins to rise again (Fig 1B), showing that the high variance has a
244 strongly negative effect on *NIP* that remains even after mean host density has been boosted
245 significantly.

246 **4.2 Temporal Dynamics in Disease Risk**

247 In the model, what we have defined as amplitude serves as an indicator of the
248 range of a disease measure, and thus disease risk, over the course of a year. This quantifies
249 how variance and mean host density affects disease predictions in the model. A greater
250 amplitude indicates greater contrast between least and most risky seasons, and a smaller
251 range indicates similar risk between seasons. Understanding this range gives a sense of
252 the pattern of disease as predicted by a model. The density of infected nymphs and total
253 nymphs fluctuates seasonally, with both measures typically peaking in late spring or early
254 summer to later in the year depending on how *DIN* is measured in a study. Changes in the
255 amplitude of *DIN/DON* represent a change in the regular pattern of disease. The positive
256 shift in amplitude for *DIN* shows a change in this pattern, and is a result of high host
257 burdens supporting increased ‘crops’ of ticks in high host density years. The minimum
258 unsurprisingly shifts very little, as this represents the sharp drop in tick densities as a result
259 of mortality during periods of low activity and cold temperatures during the winter.

260 Intraannual variation data for *NIP* is very lacking, but the few studies that have
261 been conducted indicate that seasonal *NIP* variability is indeed present in wild popula-
262 tions, and thus may present an important consideration when forecasting and managing
263 disease risk through human behavioral changes [30, 31], and might also have value in sea-

264 sonally focused wildlife management and disease prediction [32]. Mathematical models
265 of Lyme disease, which often operate on sub-yearly time scales, predict this seasonality.
266 This behavior is a feature of modeled interactions between populations of susceptible and
267 infected ticks and hosts, which fluctuate throughout the course of a year in response to host
268 births, deaths, and infection rates, which are related to periods of tick activity. Our model
269 shows that host variation affects the amplitude of *NIP* in a year. The minimum range of
270 *NIP* was most strongly affected, which suggests that host variation may be important in
271 determining the magnitude of periods of low and high risk, likely in response to changes
272 in the birth/death ratio throughout a year. Prior modeling studies have suggested that co-
273 occurrence of host density increases and high questing tick activity may boost tick density
274 [32], but it is not known how this would affect *NIP*, and this was not explored in our model.

275 **4.3 Varying Patterns in Disease Risk Under a Changing Envi-** 276 **ronment**

277 While this model is not intended to be predictive of the specific dynamics of a
278 particular area, qualitative changes in the response to host dynamics under different en-
279 vironmental conditions are of interest. We used the environmental conditions of different
280 geographic locations to explore this idea. Model results show a latitudinal gradient in the
281 magnitude of *DIN* and the pattern of the response of *NIP*. The increase in *DIN* measures at
282 locations with higher annual temperatures is expected, as warmer winters result in greater
283 tick survival and activity. The consistency of the response between locations for most mea-
284 sures is rather surprising, and suggests that similar communities of species will have similar
285 disease dynamics at different locations, though the disease risk may be different between
286 sites. The changing pattern in the effect of variance on the amplitude of *NIP* at differ-
287 ent locations, however, demonstrates the importance of considering host dynamics. It is

288 apparent that environmental conditions may alter how disease risk interacts with host den-
289 sity and host variance at different sites. Incorporation of host dynamics should allow for
290 greater understanding of the particular disease dynamics when forecasting disease risk at
291 specific locations using mathematical models. Further study of how different kinds of host
292 dynamics affects the prediction of disease risk will be valuable for better understanding our
293 models of disease as well as patterns in disease under real world patterns of host dynamics.

294 **4.4 Further Directions**

295 Our results suggest that host dynamics, and more specifically the magnitude of
296 variation in host population densities is as an important consideration when modeling Lyme
297 disease. More research is needed to understand how these dynamics may affect the use
298 of theoretical models as exploratory and predictive tools. This might include simple to
299 detailed modeling of host dynamics, or direct use of long term population data. Predictive
300 models should rely on either detailed modeling of host dynamics or, preferably, long-term
301 host data to investigate location-specific disease dynamics. This will allow these models to
302 more accurately describe disease at specific locations and will further the use of theoretical
303 studies as investigative tools for management of disease risk.

304 In our model, we assume instantaneous births at the beginning of a year, which
305 will impact the ratio of infected to susceptible hosts when the host density is increasing.
306 While this is a biologically unrealistic scenario, it presents a situation in which this shift
307 in infected host prevalence drops as a result of reproduction, one of the important ways
308 by which host dynamics are likely to impact disease risk, and especially the prevalence of
309 infection. Any process that significantly shifts the proportion of susceptible hosts is likely
310 to have repercussions on disease dynamics. This might occur in the real world as a result of
311 normal population dynamics such as overwintering deaths in host species, seasonal periods
312 of reproduction, or as a result of increased resource levels. In future models, host population

313 dynamics might be modeled on a shorter timescale, reflecting seasonal birth and death rates
314 or the response to resource levels such as seed masting.

315 An important point to investigate is how the density of hosts in one year impacts
316 disease measures in successive years in a host dynamics simulation. This can be explored
317 using current model data to investigate how the relationship between host density and dis-
318 ease changes at different host variance levels. It will also be important to determine how
319 the timing of host population changes might affect the timing of peaks in disease risk. It
320 has been shown that this can affect tick density [32], but the interactions of patterns in host
321 density with *DIN* and *NIP* have not been explored. This can be examined in our model by
322 changing the week in which the new mouse population is chosen or by introducing a more
323 complex model of mouse population dynamics. As has been suggested with *DON* and the
324 interaction between host density and tick density peaks, the timing of tick activity, quest-
325 ing, and life stage peaks is likely to factor into further layers by which hosts may influence
326 disease dynamics.

327 This project has demonstrated the importance of considering population dynam-
328 ics in tick hosts when modeling Lyme disease. This model only begins to touch on potential
329 outcomes on model behavior with the incorporation of very simple host dynamics. Poten-
330 tial avenues of expansion and further exploration with this model are many, and they offer
331 strong potential to further our understanding of Lyme disease. With a complete understand-
332 ing of how host, tick, and disease dynamics interact, we can begin to understand when and
333 when not to emphasize different pieces of this complex system. Further exploration of the
334 impact of host dynamics on disease risk will hopefully increase our knowledge of how
335 Lyme disease spreads and behaves, and aide the development of models which are able to
336 more accurately study and predict disease risk.

337 **5 Methods**

338 **5.1 Tick Population Structure**

339 The model consists of a collection of discrete-time difference equations repre-
340 senting the populations of a tick species and three hosts (S1 File). The tick population is
341 stage structured and details each of the main life stages and substages through which an
342 *Ixodid* tick will evolve (Fig 5, S2 Table).

343 **Fig 5. Flow of ticks through different developmental stages and infection.** The
344 structure and flow of the tick population through different developmental stages is shown.
345 Infected ticks go through the same stages but are kept as a separate population. Infection
346 occurs during blood meals on hosts, which can transmit to and from ticks and their hosts.
347 New eggs from infected and uninfected adults produce the next generation of uninfected
348 ticks.

349 The tick population is highly detailed to ensure realistic dynamics in vector abun-
350 dance and disease transmission. Ticks that enter a stage in a given week are tracked as a
351 cohort and undergo the appropriate processes for that stage (Fig 5, S3 Table). Ticks of a
352 stage in one week are a function of those that have survived from the previous week, with a
353 gain or loss of density in one stage from development, survival, host finding, and infection
354 of susceptible ticks.

355 The maximum time which a tick may spend in a life stage is unclear from labora-
356 tory [33, 34] and field studies [35–37], and models have incorporated a variety of assump-
357 tions to give life stage limits [21, 24, 25, 27]. We have set 52 weeks as a maximum limit
358 that ticks of a cohort are tracked. We used cumulative degree weeks (CDW) to determine
359 development rates between major life stages (egg, larvae, nymph, adult) as CDW has been
360 shown to relate well to tick phenology and is a standard in many models [21, 25, 38, 39].

361 The cuticle hardening period undergone by free-living stages is a final point of confusion,
362 with estimates ranging from 0-4 weeks, and some models only considering hardening in
363 larvae. We have chosen a hardening period of one week for all free-living stages.

364 **5.2 Host Population Structure**

365 There are three hosts in the model: mice, a "medium" mammal species, and deer.
366 Medium mammals and deer remain at constant densities of 4 and 0.4 ind. ha⁻¹, respec-
367 tively. These values were selected from the literature to keep tick densities within that
368 predicted by previous models and aide in potential comparisons [21]. Both species have a
369 constant survival rate to provide lifespans of 2 and 3 years, respectively. This rate allows
370 infection to "clear" from the population as offspring are assumed to be born susceptible.
371 Mice have a clearing rate for a 1 year lifespan. Lifespans were estimated from literature
372 for mice [40–42] and deer [43], and 2 years was chosen for medium mammals to provide a
373 suitable intermediate host.

374 Mouse densities (M) are drawn from a normal distribution (N) to simulate pop-
375 ulation dynamics. Density values are bounded by zero, and by varying the mean (μ) and
376 variance (σ^2) of the distribution a range of population dynamics can be simulated, including
377 $\sigma^2 = 0$, or constant mouse density (Fig 6). For an increase in mice, the difference ($\pm\Delta_M$)
378 is added to the population of susceptible (S) individuals, while a decrease proportionately
379 affects both susceptible and infected (I) individuals.

380 **Fig 6. Sample time series of model output with or without mouse variance.**

381 Data from two simulation runs are shown in the left and right columns, with a mean density
382 of mice $\mu = 30$ individuals per hectare and zero variance ($\sigma = 0$) or a standard deviation of
383 25 individuals per hectare around this same mean. All other input and parameters remain
384 the same. Shading indicates each segment of 52 weeks or one year. (a) and (d) show the
385 total density of eggs and questing larvae, nymphs, and adults. As is apparent, the quest-

386 ing populations of nymphs and adults are significantly smaller than the total population of
387 questing larvae, as most larvae do not manage to survive and find hosts. Populations fluctuate
388 on an annual basis but with different phases. In (d) differences in density are apparent
389 as a result of variable host populations. (b) and (e) show the populations of on host larvae,
390 nymphs, and adults over time. Again, these densities cycle yearly, and changes in on host
391 density are apparent with $\sigma = 25$ in (e). (c) and (f) demonstrate the difference in mouse
392 density for a simulation without variation (c) and with variation (f) in mouse density. At
393 the beginning of each year the population of mice is changed via a random distribution.

$$394 \quad M_{y+1} = N(\mu, \sigma^2) \quad (2.2.1)$$

$$395 \quad \Delta_M = M_{y+1} - (S_y + I_y) \quad (2.2.2)$$

$$396 \quad S_{y+1} = S_y + \Delta_M, \quad \Delta_M > 0 \quad (2.2.3)$$

$$397 \quad S_{y+1} = S_y + \Delta_M \times \frac{S_y}{S_y + I_y}, \quad I_{y+1} = S_y + \Delta_M \times \frac{S_y}{S_y + I_y}, \quad \Delta_M < 0 \quad (2.2.4)$$

398 **5.3 Activity and Host Finding**

399 Questing ticks seek out hosts each week. The total rate at which hosts are found
400 is determined by the host finding rate and the tick activity rate. The host finding rate (F_x) for
401 a host species (x) is modeled as in other discrete-time tick models [21, 25] as this strategy
402 yielded realistic rates of host finding for a sustained tick population. In Equation ??, a is a
403 host species and tick life stage (i) dependent coefficient, which relates the base finding rate
404 of a tick to host density (H_x). The exponent of 0.515 scales down the rate of host finding to
405 produce realistic densities [25]. The host finding rate is scaled via the level of tick activity,
406 with F_x being the maximum host finding rate assuming 100% tick activity. Tick activity
407 (A) is calculated with a normal distribution (N) centered on an optimal activity temperature

408 (t_{opt}), with a variance (σ^2), both determined from empirical measurements of activity levels
409 across temperature [44, 45]. The final number of questing ticks (Q) which find a host x in a
410 given week is the product of these rates and the total questing density. Each host species has
411 a maximum host burden. If new questing ticks will cause hosts to exceed their burden in the
412 successive week, ticks which exceed this threshold will remain in the questing population.

$$413 \quad F_x = a_{i,x}(H_x)^{0.515} \quad (2.3.1)$$

$$414 \quad A = N(t_{opt}, \sigma^2) \quad (2.3.2)$$

$$415 \quad Q_x = F_x \cdot A \cdot Q \quad (2.3.3)$$

416 **5.4 Survival**

Eggs, questing, and engorged ticks exhibit environmentally dependent survival (S_e), modeled with normal distributions for temperature (T_s) and precipitation index (P_s). Optimal temperature (t_s) and precipitation index (p_s) values were chosen from the literature and the variance of distributions were adjusted to produce reasonable tick densities [21, 25, 35, 37]. This eliminates assumptions made in previous models, as exact relationships between environmental conditions and survival have not been determined.

$$T_s = N(t_s, \sigma^2), \quad P_s = N(p_s, \sigma^2) \quad (2.4.1)$$

$$S_e = T_s \cdot P_s \quad (2.4.2)$$

417 On-host ticks exhibit density dependent survival. Studies demonstrate this may
418 result from true density dependence or from density dependent host grooming behaviors
419 as a result of tick exposure. To implement this, an exposure index (EI) is calculated which
420 measures the total number of ticks on each host type (x), scaled proportionately by mass

421 of each life stage for the previous 8 weeks, with a loss of exposure of 0.44 per week.
422 For each host type and tick life stage, there are given estimated minimum and maximum
423 survival rates for high and low exposure rates. In between these bounds, on-host survival is
424 determined by Equation ??, which yields a linear decrease in survival for increased EI .

$$EI = \sum_{i=1}^9 0.44^{i-1} \cdot (0.0021L_{(t-i)} + 0.014N_{(t-i)} + A_{(t-i)}) \quad (2.4.3)$$

$$S_o = \frac{S_{min} - S_{max}}{EI_{max} - EI_{min}} \cdot (EI - EI_{min}) + S_{max} \quad (2.4.4)$$

425 Ticks in the hardening stages are modeled with a constant survival, as molting
426 success becomes the primary determinant of survival. Parameterization is based off of pre-
427 vious models to ensure realistic rates [21, 25].

428 **5.5 Infection**

429 Transmission occurs during the on-host stages, starting in larvae, and passes up-
430 wards through life stages, but will not transmit from adult ticks to eggs [46]. Ticks disperse
431 evenly between infected and susceptible hosts of the same species, and the rate of infection
432 from hosts to ticks is determined by the competency of each host species, the proportion
433 of infected hosts of each species, and the density of ticks per host. Competency for mice
434 is set to 75%, and deer to 0% [6, 13, 14, 19, 47]. The intermediate host serves as a generic
435 species to maintain infection in the system in the absence of mice, and has a competency
436 of 50%, in line with other small to medium sized mammal species [6, 13, 16].

437 Host infection is slightly more complicated. Infected ticks distribute evenly among
438 hosts, but infected tick burdens may be in the range of only a few to no ticks per host
439 throughout parts of the year. During these periods, it is expected that some hosts may have

440 several ticks while others have none. We used a modification of a method established previ-
441 ously [21] which uses a Monte Carlo simulation to predict the rate of host infection, given
442 the number of infected ticks per host (ITH) and assuming an infection rate of 100%, which
443 is then scaled by the expected infection rate of ticks. This calculation is made several hun-
444 dred times to ensure a robust calculation for infected tick burdens from 0.01 to 7.5 ITH ,
445 the upper range of which is sufficient to provide $> 99.9\%$ chance of infection. The ITH is
446 scaled by an infection rate of 0.9 to provide a 90% chance of infection per infected tick to
447 host. For sufficiently high ITH there will be an effectively 100% infection rate.

448 **5.6 Environmental Data**

449 Average temperature and relative humidity are indicated as the most significant
450 environmental factors which affect survival and development in *Ixodid* ticks [35–37, 48].
451 In the northeastern United States, where Lyme disease is most prevalent, humidity does not
452 appear to play as significant a role, as humidity ranges do not typically fall outside of those
453 optimal for tick survival [27]. As long term humidity data is less easily available, precipita-
454 tion is sometimes used as a stand-in, which we have chosen to do. The average temperature
455 is calculated as the average of the minimum and maximum temperature recordings for
456 a day, and precipitation is calculated as an index, (PI), measured as $1/10$ *th* the current
457 week's ($w + 1$) rainfall (R) in *mm* with a loss of 65% from the previous week's (w) index.

$$PI_{w+1} = R_{w+1} + 0.65PI_w \quad (2.6.1)$$

458 We obtained daily measurements of minimum and maximum temperature and
459 precipitation from 22 sites throughout North America [49]. Data were selected to cover a
460 range of locations and climate conditions from 1971-2021. Data were converted to weekly

461 measurements and mean temperature and precipitation indices were calculated. Portland,
462 Maine weather data was used as the primary location of focus for investigating model
463 behavior.

464 **5.7 Simulation**

465 All Simulations were run in Julia version 1.6.1 [50]. A variety of packages were
466 used in the process of conducting simulations, analyzing data, and plotting in Julia [51–60]
467 and in R [61–63]. To investigate the impact of host population dynamics on disease flow in
468 the model, 10,000 simulations were run with mean and variance of the density distribution
469 for all combinations of each variable ranging from 0 – 99 in increments of 1. Ranges were
470 chosen to cover empirical measurements of density in the white-footed mouse, *Peromyscus*
471 *leucopus*, and potential variation in these densities [17]. Simulations were repeated for each
472 of the 22 chosen locations to investigate how changing environmental data affects model
473 behavior. Simulation output consists of weekly time-series data for all tick and host stage
474 classes within the model. A single simulation involves calculating the output of the discrete
475 difference equations at each time step of one week, and using these result to calculate
476 again for each successive time step until the completion of the desired number of weeks of
477 simulation. The result is the time series data as described above, stored in large matrices
478 for each life stage and host type.

479 **5.8 Model Analysis**

480 Analysis of model data was conducted in Julia version 1.6.1, and plotting con-
481 ducted with Julia and R version 4.0.5. To determine model prediction of disease risk, we
482 calculated the mean, minimum, and maximum densities of infected and susceptible nymphs
483 per year for the last 10 years of each simulation. These data were used to calculate standard

484 disease risk measures. The density of infected nymphs, *DIN* was calculated as the mean,
485 minimum, and maximum density of questing infected nymphs over the course of a year,
486 with minimum and maximum density not necessarily aligned with the density of questing
487 susceptible and infected nymphs. Density of nymphs, *DON* was calculated as the mean,
488 minimum, and maximum density of the combined population of susceptible and infected
489 nymphs. Nymphal infection prevalence, *NIP* was calculated as the mean, minimum, and
490 maximum measures of the infected population divided by the combined population of sus-
491 ceptible and infected nymphs. Amplitude is measured as the yearly minimum subtracted
492 from the maximum for these disease metrics. The mean over 10 years for each metric was
493 used when analyzing results.

494 We completed a sensitivity analysis to determine the sensitivity of results to par-
495 ticular variables. Parameters were varied with a one at a time approach across a range from
496 -10% to 10% of the baseline value. Sensitivity was calculated as the percent change in
497 the output measures (as are described above) divided by percent change in the parameter.
498 Portland, Maine environmental data was used as well as a constant mouse density of 50.

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724 **7 Supporting Information**

725 **S1 Fig. Response of maximum density of infected nymphs and nymphal in-**
726 **fection prevalence to variance in the mouse population and mean mouse density.** A)
727 shows the response of maximum *DIN* to mouse variation and mean density. The relation-
728 ship to mean mouse density is positive and linear, as is the relationship to mouse variance.
729 Contour lines plotted on the variance \times mean plane and coloring reveal surface features.
730 B) shows the response of maximum *NIP* to mouse variation and mean density. There is
731 a nonlinear positive relationship between maximum *NIP* and mean mouse density. The
732 relationship between mouse population variance and maximum *NIP* is inconsistent, with
733 higher variance increasing maximum *NIP* at the low mean mouse density. At higher mean
734 densities, maximum *NIP* transitions to decreasing with variance. Contour lines and col-
735 oration again reveal surface features.

736 **S1 Table. Summary Statistics, Grouped by Variance, DIN, NIP** This table
737 presents summary statistics for disease risk metrics, grouped by different levels of mouse
738 population variance.

739 **S2 Table. Tick life stages.** Tick life and developmental stages represented in the
740 model.

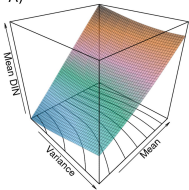
741 **S3 Table. Demographic Tick Processes.** Tick demographic processes repre-
742 sented in the model.

743 **S4 Table. Parameters used in the model.**

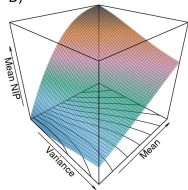
744 **S5 Table. Parameters used in the model, continued.**

745 **S1 File. Model equations.** The equations representing tick, host, and infection
746 processes in the model are presented here. See S4 and S4 Tables for parameter names and
747 purposes.

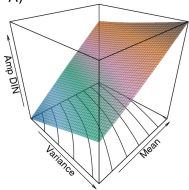
A)



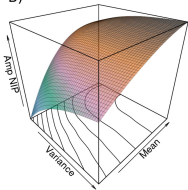
B)



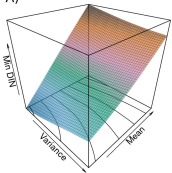
A)



B)



A)



B)

