Evolution in response to an abiotic stress shapes species coexistence

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Keywords: Experimental Evolution; Coexistence theory; Competition; Spider mites

Running title: Evolution shapes biotic interactions

Authorship Statement: IF, OG and SM conceived the original idea. IF, DG, OG and SM designed the experiments. IF, DG, LRR and MC performed the experiments and collected the data. IF performed the modelling work and analysed data, with advice from FZ and OG. IF produced the figures, with help from LRR. IF, SM and OG wrote the first draft of the manuscript, and all authors contributed substantially to revisions. SM supervised DG, IF, LRR

Abstract

Adaptation to abiotic stresses generally relies on traits that are not independent from those affecting species interactions. Still, the impact of such evolutionary processes on coexistence remains elusive. Here, we studied two spider mite species evolving separately on tomato plants that hyper-accumulated cadmium, a stressful environment for herbivores, or on plants without cadmium. Through experimental evolution and structural stability theory, we found that adaptation of both species to cadmium shifted predictions from exclusion to coexistence, but only in the cadmium environment. This shift occured due to a simultaneous increase in intra and a decrease in interspecific competitionin that environment. These predictions were further confirmed with complementary experiments of population dynamics, underscoring that evolution of single species in a new environment, even in absence of interspecific competitors, shapes species coexistence. Hence, population shifts to novel environments may have unforeseen evolutionary consequences for community composition and the maintenance of species diversity.

Introduction

Understanding how evolutionary dynamics shape and are shaped by long-term species persistence is key to predicting community composition and biodiversity maintenance. Character displacement in response to the presence of competitors is one of the classical examples of the role of natural selection in shaping trait evolution and species distribution (Brown & Wilson 1956; Lack 1947; Slatkin 1980). Theory predicts that evolving with competitors may lead to changes in competitive interactions (i.e., the negative per-capita effect of one species on another; Abrams & Matsuda 1994; Bernhardt *et al.* 2020; Sakarchi & Germain 2023), which in turn can affect species coexistence (Edwards *et al.* 2018; Germain *et al.* 2018; Vasseur *et al.* 2011; Yamamichi & Letten 2021). In line with these predictions,

empirical studies have shown that short term evolution in the presence of competitors can

modify competitive traits (Hart *et al.* 2019) and change coexistence patterns (Germain *et al.*

2020; Hiltunen *et al.* 2017; Lankau 2011; Sakarchi & Germain 2023; Zhao *et al.* 2016).

Although the abovementioned studies have demonstrated that short-term evolution is an

important force shaping ecological patterns, we may still be under-estimating the potential role

of evolution in species coexistence. An important yet overlooked possibility is that traits that

affect interactions between species may be selected even in the absence of competitors. For

example, species arriving in a vacant environment may rapidly adapt and, potentially

monopolize that environment (Nadeau *et al.* 2021). Likewise, prior adaptation to a given

environment can lead to a lower relative growth rate of late arriving species (Low-Décarie *et*

al. 2011). Moreover, trait evolution in a given environment may affect species community

composition in other environments (Fukano et al. 2022; Gallego & Narwani 2022; Limberger

& Fussmann 2021). However, to date no study has tested if and how trait evolution in response

to an abiotic selection pressure can change intra- and interspecific competition and its impact

on species coexistence.

Recent advances of the complementary theoretical frameworks of modern coexistence

theory and structural stability (Chesson 2000; Saavedra *et al.* 2017; Yamamichi *et al.* 2022)

provide clear mechanistic pathways into how competing species coexist in ecological

timescales. In particular, structural stability defines that coexistence is possible when the

feasibility domain (defined through species interactions) can accommodate the differences in

intrinsic growth rates between species (Godoy *et al.* 2018; Saavedra *et al.* 2017). Combining

this framework with experimental evolution allows unraveling whether trait evolution alters

the coexistence opportunities or differences in species´ performance, and enables identifying

specific mechanisms by which evolutionary changes affect ecological dynamics.

Here, we assessed how evolution in response to an abiotic stress affects species coexistence by combining two powerful tools: experimental evolution and structural stability, applied to two closely-related spider mite species, *Tetranychus urticae* and *T. evansi*. Spider mites rapidly adapt to novel host plants (Magalhães *et al.* 2007a; Sousa *et al.* 2019; Wybouw *et al.* 2015) and often form host races (Forbes *et al.* 2017; Magalhães *et al.* 2007b). However, they are also found on plants to which they are not adapted, given their passive dispersal (Bitume *et al.* 2011; Fronhofer *et al.* 2014), thus engaging in competitive interactions in less favorable environments. Competitive interactions between *T. urticae* and *T. evansi* have been documented on tomato plants (*Solanum lycopersicum*) (Fragata *et al.* 2022; Sarmento *et al.* 2011). These plants can hyper-accumulate cadmium in their shoots, which strongly reduces spider mite fecundity and survival, producing a strong selective pressure (Godinho *et al.* 2018, 2023, 2024). Here, we studied how adaptation of these two species to an abiotic selection pressure affected the probability of coexistence.

Material and methods

Details of the maintenance of outbred and experimental evolution populations, common garden procedures and all experiments performed are available in Supplementary Material and Methods.

Experimental evolution

Experimental evolution populations were initiated by transferring 220 adult mated females of *Tetranychus urticae* or *T. evansi* from outbred populations (Godinho et al. 2020, 2024) to an experimental box containing four tomato leaves and water. We established five replicate populations for each selection regime (Fig. S1A): *T. urticae* or *T. evansi*, exposed to leaves from plants watered either with a 2mM cadmium solution (the cadmium selection regime) or

with water (the no-cadmium selection regime). This cadmium concentration is highly detrimental to both spider mite species, (Godinho *et al.* 2018, 2023), a result we recapitulate here (cf.Supplementary Material). Every two weeks (circa one mite generation), 220 adult mated females were transferred to a new box containing four new tomato leaves, ensuringdiscrete generations. This setup mimics colonization of a new plant, as dispersal is mainly performed by young mated females (Li & Margolies 1993). If the total number of mated females did not reach 220, the transfer was complemented with females taken from the respective T-1 box or from the base population (cf. Godinho et al. 2020, 2024). Replicate 2 of the *T. urticae* cadmium selection regime was not tested because not enough females were available. Prior to the experiments, individuals from all regimes were placed in a common garden of cadmium-free tomato leaves for two generations, to equalize potential maternal effects and synchronize replicates (generation 40 and 42 for the competitive ability and population growth experiments, respectively).

Empirical estimation of competitive abilities and intrinsic growth rate

To test if evolution on plants with or without cadmium affected the probability of coexistence between the two species in these environments, we estimated the intrinsic growth rates and the strength of intra- and interspecific interactions of mites from each selection regime in each environment (Fig. S1B), following the methodology described in Hart *et al.* (2018). Briefly, after two generations of common garden, we placed one focal female alone or with 1, 3 or 9 females (i.e., competitors) on leaf disks from plants with or without cadmium. The focal females were exposed to competitor females either from the same (intraspecific) or a different (interspecific) selection regime (full factorial design) with matching population replicates. Females were left to oviposit for three days, then killed. Two weeks later, the number of adult females per patch was counted, representing the combined effect of competition on the females

133 and their offspring. Each experimental treatment (i.e. combination of density*selection history

134 of focal and competitor*environment) was replicated 10 times for each replicate populations

135 of the four selection regimes, making a total of 1260 replicates divided by six blocks.

Experimental validation of population dynamics in the presence of interspecific

competitors

To test if evolution on plants with or without cadmium affected the population growth rate

under competition, we placed six females of the two species from different selection regimes,

after two generations of common garden, on plants with or without cadmium (full factorial

design, Fig. S1C) and measured the number of adult females produced by each species after

two generations. Ten experimental replicates were tested per experimental condition, per

replicate population. This data was then compared with model predictions based on parameters

estimated from the intrinsic growth rate and competitive abilities (cf. Model validation section

below).

Theoretical estimation of competition and growth parameters

Data collected in the intrinsic growth rate and competitive ability experiment was used to

parameterize the Ricker competition model (Fowler 1981; Ricker 1954), used in a previous

study with spider mites (Bisschop *et al.* 2022). This model allows incorporating positive

interactions (i.e., facilitation, Bowler et al. 2022; Buche et al. 2025; Martyn et al. 2021; Ricker

1954; Stouffer 2022), which, based on initial data scrutiny, were likely to occur in our system.

The Ricker model is described by the following equation:

$$(1)\frac{N_{i,t+1}}{N_{i,t}} = \lambda_i e^{-\alpha_{ii}N_{i,t} - \alpha_{ij}N_{j,t}}$$

where $N_{i,t+1}$ is the number of individuals of species i in the next generation, λ_i the intrinsic

growth rate of species i in absence of competitors, α_{ii} and α_{ij} the strength of intra- and

interspecific competition, respectively, and $N_{i,t}$, $N_{j,t}$ the number of competitors of species i and

j, respectively, in the current generation. We used the "cxr" package (García-Callejas *et al.*

2020) in R to estimate λ_i, α_{ii} and α_{ij} for each replicate population separately and with all

replicates of the same selection regime pooled (to increase statistical power). To include the

filtering effect of the cadmium environment and because the "cxr" package does not allow the

inclusion of zeros, we transformed our data by summing one to each of our datapoints (as the

logarithm of one is zero). Initial parameters were set at λ_i=1, α_{ii}=α_{ij}=0.1 and confidence

intervals were estimated from 1000 bootstrap iterations. Model fitting assessment was done by

visual inspection (availabe in the git repository) and by comparing the Euclidian distance

between observed and predicted values (Fig. S2). To assess the relative impact of the different

parameters on population growth, we e simulated offspring production of ten females using

only the intrinsic growth rate, the joint effects of intrinsic growth rate and of intraspecific

competition or the joint effects of those traits plus interspecific competition. Confidence

intervals were calculated using the lower and upper estimates obtained from the cxr package

(García-Callejas *et al.* 2020).

Structural stability approach to predict coexistence

Estimates obtained from the Ricker model (Fig. S3, S4, S5, S6) showed widespread weak

positive interspecific interactions (i.e., facilitation), which are not accounted for in common

approaches to study species coexistence (e.g. Chesson 2000). Therefore, we used the structural

stability framework (Allen-Perkins *et al.* 2023; Saavedra *et al.* 2017, 2020) to predict the

outcomes of species interactions. This approach utilizes the strength and sign of species

interactions to estimate the size of the feasibility domain, and allows incorporating stochasticity

and interactions other than direct competition (Rohr *et al.* 2014). Coexistence is possible if the

vector containing the intrinsic growth rate of both species falls within the feasibility domain.

In principle, the latter can include both negative and positive intrinsic growth rates. However,

we restricted coexistence predictions to positive values (i.e. bounded by the vectors $[0, \frac{\alpha_{22}}{\alpha_{12}}]$ and

$[0, \frac{\alpha_{21}}{\alpha_{11}}]$) to match information from our experimental system, as species can either have zero

or positive growth rates (Song *et al.* 2018). Finally, to account for uncertainty in the estimation

of model parameters, we estimated the feasibility domain bound by the upper (lower) 95% of

the vector $[0, \frac{\alpha_{22}}{\alpha_{12}}]$ and the lower (upper) boundaries of the $[0, \frac{\alpha_{21}}{\alpha_{11}}]$ to obtain the widest

(narrowest) conditions under which species are able to coexist.

To measure each species' vulnerability to exclusion, we tested how resistant coexistence is

to perturbations. For that we estimated the species exclusion distance (following Allen-Perkins

et al. 2023; Medeiros *et al.* 2021), which corresponds to the minimal distance between the

vector of intrinsic growth rates ($r = [\lambda_I, \lambda_j]$) and the edges of the feasibility domain (vectors

corresponding to either $[0, \frac{\alpha_{22}}{\alpha_{12}}]$ or $[0, \frac{\alpha_{21}}{\alpha_{11}}]$) as a proxy for how strong a perturbation must be to

change the coexistence outcome. To calculate the distance between the normalized vectors of

intrinsic growth rates and the edges of the feasibility domain, we applied the following formula,

(Allen-Perkins *et al.* 2023):

$$(2)\ dist_i = \cos^{-1} \begin{bmatrix} \lambda_i \\ \lambda_j \end{bmatrix} \cdot \begin{bmatrix} \alpha_{21} \\ \alpha_{11} \end{bmatrix}$$

Confidence intervals were obtained by estimating the distance between the normalized vectors

of intrinsic growth and the edges of the smallest or largest feasibility cones. The largest and

smallest feasibility cones were obtained by using the lower and upper parameter estimates,

respectively, to encompass potential trait variation underlying observed parameter estimates .

Model validation

We validated our approach by comparing the projections of abundances in the short term,

obtained from our model, with the empirical results of a population growth experiment after

two generations. For that, we used the Ricker model to predict the relative abundance of each

species after two generations, starting with the initial conditions of the experiment. Then, we

performed a general linear regression with a binomial error distribution to estimate the

correlation between the mean observed proportions of each competitor (using the cbind

function to join the number of *T. evansi* and *T. urticae* females) in all replicates of all treatments

with the model proportion estimates (Number of *T. evansi* females/Number of females)

obtained for each replicate.

Statistical analyses

To test how evolution on plants with cadmium affected population growth and competition in

the cadmium environment, we used general linear models with a gamma distribution for the

intrinsic growth rate, and with a normal distribution for intra- and interspecific competition.

All models were applied separately for both species, and normality and dispersion of the

residuals was inspected for all models. We included the parameters estimated with the cxr

package for each replicate as dependent variables and selection regime as a fixed factor (with

two levels, cadmium and no-cadmium). For the strength of interspecific competition, the model

included the selection regime of the focal and competitor species, and their interaction.

Whenever the interaction term was significant, we applied contrasts for all combinations of

selection regimes. Similar models were applied to test the impact of evolution in cadmium on

the performance in the no-cadmium environment. Finally, we estimated the likelihood of

finding differences between environments/selection regimes using 1000 bootstrap iterations.

For that, we randomized (with replacement) the parameter estimates within

environments/selection regimes (depending on the model) and applied the same models as

above. Then, we estimated the probability of randomly obtaining a significant p-value (below

or equal to 0.05) with our data set by computing the number of times a significant result was

obtained.

All analyses were done using the package "glmmTMB" (Brooks *et al.* 2017) in R 4.2.1

version (R Core Team 2022). Contrasts were performed using the "emmeans" package (Lenth

2024), analyses of residuals were done for each model using the DHARMa package (Hartig

2022), and graphical representation was done using "ggplot2" (Wickham 2016). All data and

scripts are available in the repository https://figshare.com/s/f001d9f699a4027d7b62.

Results

Evolution in cadmium changes population growth and the interactions within and between species

In the cadmium environment, the intrinsic growth rate of both species evolving in that

environment for 40 generations was higher than that of populations evolving on plants without

cadmium (*T. urticae*: $\chi^2_{1,6}$=6.187, P-value=0.013, *T. evansi*: $Z_{1,7}$= 4.483, P-value=0.034, Fig

1, Fig. S7, Table S1, S2). In addition, the strength of intraspecific competition increased for *T.*

evansi ($\chi^2_{1,7}$ = 8.570, P-value=0.003, Fig S3A, Table S1, S2), but not for *T. urticae* ($\chi^2_{1,6}$ =

0.131, P-value=0.717, Fig S3B, Table S1, S2). This resulted in a decreased predicted number

of offspring produced by *T. evansi* in competition (relative to when growing alone) but only

when *T. evansi* evolved in the cadmium environment (Fig. 1A, λ vs $\lambda+\alpha_{ii}$ for the yellow and

blue vs the green and red dot). Moreover, evolving on plants with cadmium made *T. evansi*

more sensitive to competition with *T. urticae* from the no-cadmium selection regime (Fig. S4A,

Table S3A, S3B, contrasts: T ratio$_{1,13}$= -0.053, P-value=0.003), reducing the predicted number

of offspring produced after one generation of competition (Fig. 1A, $\lambda+\alpha_{ii}+\alpha_{ij}$, comparing

yellow and red dots). Conversely, evolution in cadmium did not affect the sensitivity to

interspecific competition in *T. urticae* (Fig. S4B, Table S3A). We observed weak positive

interactions within and between species in several replicates (i.e., negative alpha values) across

selection regimes (Fig S3, S4), which reinforces the added value of using the structural stability

framework.

In the no-cadmium environment, no significant differences in the intrinsic growth rate

(Fig. S8, T. *evansi*: $\chi^2_{1,7}$=2.628, P-value=0.105; T. *urticae*: $\chi^2_{1,6}$= 0.121, P-value=0.728, Table

S1, S2), intraspecific competition (Fig. S5, *T. evansi:* $\chi^2_{1,7}$= 1.445, P-value=0.229; *T. urticae*:

$\chi^2_{1,6}$= 0.003, P-value=0.998, Table S1, S2) or interspecific competition (Fig. S6, Table S1, S2,

S4) were observed between populations evolving in that environment and those evolving in the

environment with cadmium. Weak intra- and interspecific facilitation was found for several

replicates across selection regimes (Figs. S5, S6), as in the cadmium environment.

Evolution of both competitors increases the probability of long-term coexistence in the

cadmium environment

In the cadmium environment, coexistence was more likely when both species had evolved in

that environment (Fig. 2), due to an increase in the size of the feasibility domain via increased

intraspecific competition for the better competitor (*T. evansi*, Fig S3). The prediction of

coexistence of both species after evolution in cadmium was robust to parameter variation

because it holds even after taking uncertainty (95% CI) into account when estimating the

feasibility domain (Fig 2, light blue region), and the differences of species intrinsic growth

rates (Fig. 3, S9).

In the cadmium environment, coexistence was not possible when only one species

evolved in that environment (Figs 2, 3). In fact, we predicted competitive exclusion for

whichever species evolved in cadmium (Figs. 2, 3, S9). The exclusion of *T. evansi* after

evolving in cadmium can be explained by two factors: increased intraspecific competition (Fig

S3) and increased sensitivity to competition (i.e., stronger response to competition) from *T.*

urticae of the no-cadmium selection regime (Fig S4). In the reverse case, the exclusion of *T.*

urticae that evolved in cadmium (Fig. 4), can be explained by an assymetric effect of

interspecific competition. *T. urticae* barely affecting *T. evansi* from the no-cadmium selection

regime, but the latter strongly affecting *T. urticae* (Fig. S4). Again, these predictions, were

robust to variation in parameter estimation when accounting for lower and upper 95%

confidence intervals.

In the no-cadmium environment, the two species were generally predicted to coexist

except when *T. urticae* from the cadmium selection regime competed with *T. evansi* from the

no-cadmium selection regime, in which case exclusion of *T. urticae* by *T. evansi* was predicted

(Fig. 2). However, the minimum distance to the edges of the feasibility domain was short in all

cases (Fig. 3, S9), especially when considering those with lower confidence intervals,

indicating that small perturbations may shift the prediction from coexistence to exclusion and

vice versa. Importantly, there was high heterogeneity among replicates but little effect of the

evolutionary history of both species (Fig. 3). Thus, evolving in the cadmium environment did

not affect the range of opportunities for species to coexist in the no-cadmium environment (Fig.

2, S9).

Simulated competitive dynamics partially recover empirical patterns

To validate our approach, we compared theoretical predictions with a two-generation

population dynamic experiment with the two species from all combinations of selection

regimes. (Fig. S1). Results show that we consistently underestimated the proportion of *T.*

evansi in the populations. Still, the model was able to recapture the tendency observed in the

data (Fig. 4, slope pooled replicates: 0.707, P-value<0.001), despite some heterogeneity

between replicate populations, especially in the no-cadmium environment (slope with

replicates: 0.631, P-value<0.001, Table S5).

Discussion

We used a combination of experimental evolution and structural stability theory to predict the

impact of evolution of two spider mite species in response to an abiotic selection pressure

(cadmium) on species persistence, in an environment with or without cadmium. To obtain

theoretical predictions, we quantified the intrinsic growth rate of each species and the strength

of their intra- and interspecific competitive interactions across different environmental and

selection scenarios. While evolutionary history had little effect on coexistence patterns in the

environment without cadmium, the independent evolution of both species in response to

cadmium led to a change from competitive exclusion to coexistence in the cadmium

environment. This was due to differential changes in intrinsic growth rates, which equalized

fitness between species, coupled with an increase in intraspecific competition in the superior

competitor, which led to increased structural niche differences, and therefore to an increased

range of fitness differences compatible with persistence of both species. In sum, we show that

evolution in response to an abiotic selection pressure can change interactions between and

within populations and affect the probability of coexistence, even without a direct selection

pressure posed by the presence of competitors as showed previously (Germain *et al.* 2020;

Sakarchi & Germain 2023; Vasseur *et al.* 2011; Zhao *et al.* 2016).

Both species adapted to the cadmium environment within 42 generations of experimental

evolution. This is supported by the observed increase in the intrinsic growth rate in the

cadmium environment of populations from the cadmium selection regime, when compared to

the no-cadmium regime. A previous study reported that these populations of *T. evansi* did not

shown signs of adaptation at the 33rd generation of evolution (Godinho et al. 2024). Therefore,

the adaptation reported here occurred at a later time point. Moreover, this previous study

revealed lower performance of *T. evansi* populations evolving with cadmium, on the

environment without cadmium, compared to populations evolving without cadmium (Godinho

et al. 2024). We do not recover this result, suggesting that extending time for evolution to

operate led to reduced costs of adaptation. This adaptation pattern, together with the lack of

initial genetic variation for traits associated to performance in cadmium, is compatible with the

existence of cryptic genetic variation that is released upon evolution in the cadmium

environment, for example via newly formed epistatic interactions (Paaby & Rockman 2014).

As the traits we are measuring are likely polygenic, this leads to more allele combinations,

thereby increasing the likelihood of such cryptic genetic variation to arise (Chandler 2010;

Lauter & Doebley 2002). Still, adaptation may be too slow to enable the establishment of mites

in cadmium-contaminated sites. Indeed, assuming similar generation time in the lab and in the

field, a tomato growing season roughly corresponds to 15 generations for spider mites, which

is not sufficient to lead to genetic changes allowing adaptation to cadmium.

Evolution of *T. evansi* in cadmium also led to stronger sensitivity to intraspecific

competition in that environment, as compared to mites evolving on plants without cadmium.

Evidence that intraspecific competitive ability may evolve during adaptation to a novel abiotic

environment has been accumulating (Bernhardt *et al.* 2020; Bono *et al.* 2015; Limberger &

Fussmann 2021). There are many reasons to expect such evolution (Siepielski *et al.* 2016). For

example, individuals may become more efficient at extracting resources when these are

limiting (Bernhardt *et al.* 2020), or higher population growth may lead to higher densities being

reached earlier, thus increasing intraspecific competition. This effect is expected to be stronger

for *T. evansi*, given that its intrinsic growth rate is higher than that of its competitor. Measuring

intraspecific competition should thus be mainstreamed in experimental evolution studies,

which typically measure only individual life-history traits such as fecundity and survival (Kawecki *et al.* 2012).

An overlooked logical follow-up of our observations is that adaptation to abiotic selection pressures may affect interspecific interactions as well. Here, we shed light on this understudied hypothesis by documenting that, when spider mites evolved in cadmium, they affected their competitors less or equally than mites that evolved on plants without cadmium. This counter-intuitive result may be due to a growth-competition trade-off, given that both species also increased their growth rate on plants with cadmium upon evolution in that environment. Such trade-off has been assumed in theoretical models (Wittmann & Fukami 2018), and experimentally demonstrated in *Brassica nigra* (Lankau 2008; Lankau & Strauss 2007). Moreover, algae adapt more rapidly to abiotic selection pressures when they evolve alone *vs.* when in presence of competitors (Limberger & Fussmann 2021), which is compatible with the existence of such trade-off. In sum, we show that interspecific competitive ability evolves in absence of competitors, a result similar to when competing species coevolve (Fukano *et al.* 2022; Germain *et al.* 2020; Hart *et al.* 2019; Sakarchi & Germain 2023). The fact that we recover such patterns even in the absence of a competitor in the environment implies both that causality needs to be scrutinized in eco-evolutionary studies and that evolution has consequences for community structure that are much more far reaching than initially thought.

Empirical studies have shown that adaptation to an abiotic selection pressure can change species interactions (Bach *et al.* 2018; Fukano *et al.* 2022; Limberger & Fussmann 2021) but none have addressed how it affects species coexistence. In fact, the only studies applying coexistence theory to evolutionary data concern adaptation to the presence of a competitor (Germain *et al.* 2020; Hart *et al.* 2019; Zhao *et al.* 2016). Here, we found that, in the cadmium environment, species are predicted to coexist only when both competitors evolved separately in cadmium. Given the mosaic structure of spider mite populations in the field (Magalhães *et*

al. 2007b), their rapid adaptation to novel host plants (Sousa *et al.* 2019) and the passive nature of their dispersal (Smitley' & Kennedy 1988) this scenario is quite plausible.

The change we observed from competitive exclusion to coexistence was not due to the large change of a specific parameter but rather a small but full reorganization of how species grow and interact with each other. Namely, coexistence was due to a combination of a stronger self-limitation of the superior competitor with a reduced negative competitive effect on the inferior competitor. Although the ultimate mechanisms of these changes are unclear, we speculate that evolution improved the ability of spider mites to feed on tomato plants that accumulated cadmium, likely by increasing resource uptake, which in turn decreased resources available to others (thus increasing intraspecific competition). This positive covariance between an increase in fitness and an in self-limitation is congruent with theoretical predictions (Carroll *et al.* 2011), previously reported in other systems (e.g. Angert et al. 2009). Additionally, the observed reduction in interspecific competition could be due to plant defence suppression, documented in these populations (Fragata *et al.* 2022). Such suppression may make resources more available to both species. Another possible mechanism affecting competition in our system is resource heterogeneity among leaves, which, together with differences in arrival, have been shown to shape coexistence between these two species (Fragata *et al.* 2022). Curiously, studies addressing how evolution in the presence of a competitor affected species coexistence have not found such a radical change from exclusion to coexistence. Instead, they found that coexistence is maintained when species coevolve, albeit by mechanisms different than those operating in the absence of such evolution (Germain *et al.* 2020; Hart *et al.* 2019; Zhao *et al.* 2016). Given that habitats are increasingly fragmented in space and changing in time, organisms are often exposed to different abiotic challenges. We show that they may rapidly adapt to these changes, fostering their ability to persist in those environments, both directly via coping with abiotic challenges, and indirectly, via the effect of

such adaptation on species interactions. Therefore, adaptation to an abiotic selection pressure

may affect community composition to an unprecedent degree, higher than direct adaptation to

the presence of competitors.

We also tested the accuracy of the theoretical predictions, by comparing model predictions

based on parameters estimated from competition experiments with a short-term population

dynamics experiment. Despite the fact that our model consistently underpredicted the

proportion of *T. evansi* in the population (as the slope was lower than 1), we still obtain a good

overlap between predicted and observed relative abundances of the two species. Some

mismatch was found when only one of the populations evolved in the cadmium environment.

This is probably due to stochasticity, as the experiment started with a small number of

individuals. We also obtain large confidence intervals for the competitive interactions

estimates, which may be due to low statistical power (especially when estimating per replicate

population), but can also be caused by non constant effect of competitive interactions that is

not captured by our current approach. In the future, it will be important to expand the structural

stability approach to include different types of response to competiton. Still, the strong

correlation between observed and predicted values suggests that our model parameterization is

robust to changes in the number of individuals, resource and space availability, indicating that

the combination of theoretical modelling and experimental estimation of competitive

interactions has a high predictive power.

Another strength of our study is that, unlike previous studies, we applied a full factorial

design, in which we explore how different combinations of evolutionary histories of both

competitors affect the probability of coexistence. We found that coexistence in an environment

with cadmium is only possible when both species evolved in that environment. In fact, when

competition occurred between one species that evolved without cadmium and one that evolved

in cadmium, the latter was more likely to be excluded in the cadmium environment. This

unintuittive outcome may be explained by a combination of increased self-regulation (i.e.

increased intraspecific competition) by the cadmium-evolved population and the weaker

interspecific competition exerted by the cadmium populations. Hence, community assembly

does not follow a linear path, being contingent on the evolutionary history of the two species.

This suggests that asynchronies in the arrival to cadmium-contaminated sites may lead to

species exclusion, due to evolution of the species first colonizing that environment.

Coexistence is only possible when the two species have both evolved in the cadmium

environment, which can be achieved either if they arrive simultaneously to a site with

cadmium, then adapt at a similar pace, or if they are already adapted upon arrival. Hence, as

previously shown for short-term differences in arrival time (Fragata *et al.* 2022), we here show

that historical contingencies affect species coexistence also via their effect on evolution.

As found in studies with no evolution, we found that interactions are specific to a given

environment (Grainger *et al.* 2019; Granjel *et al.* 2023; Matías *et al.* 2018; Wainwright *et al.*

2019). In general, our results align with previous work, not considering evolution, showing that

stable coexistence is fostered in stressful environments via a reduction in growth rate that

equalizes fitness differences, and by a shift from interspecific to intraspecific competition,

which increases the range of coexistence opportunies (i.e. the size of the feasibility domain)

(Grainger *et al.* 2019; Granjel *et al.* 2023; Matías *et al.* 2018; Song *et al.* 2020b; Wainwright

et al. 2019), and stabilize the population dynamics of interacting species. However, our study

tempers this statement by the finding that this is only true when both species have evolved in

that stressful environment. Moreover, the use of the structural stability framework allows

incorporating sensitivity to changes in environmental conditions in these coexistence

predictions (Allen-Perkins *et al.* 2023; Song *et al.* 2020a; Tabi *et al.* 2020). In the environment

without cadmium, we found a small distance to the edge (i.e., low robustness) across all

selection regimes, suggesting that changes in competitive outcomes are likely to occur due to

stochastic events. Instead, in the cadmium environment, the distance to the edge was higher in

all cases, suggesting that communities in that environment are more long-lasting. Thus, our

results suggest that communities in cadmium-free environments will be modulated by small

environmental changes, whereas those in environments with cadmium will be more shaped by

evolution.

Our study highlights the added value of combining experimental evolution and controlled

experiments with ecological theory, by providing novel insights of how species adaptation to

an abiotic stressor affects their ability to coexist. This work provides significant advances in

both evolutionary and ecological fields. On the one hand, it shows that traits such as the strength

of intra- and interspecific competition should be incorporated in the characterization of species

adaptation to novel environments. This is particularly true in experimental evolution studies,

which generally quantify evolution by measuring classical life-history traits of single

individuals (Garland & Rose 2009; Kawecki *et al.* 2012), thus ignoring traits associated to

evolving in the presence of others (Chippindale *et al.* 2003). Conversely, ecological studies

benefit from incorporating past evolutionary history, as such approach has the capacity to refine

our understanding of how species interact and coexist (Leibold *et al.* 2019; Urban & De

Meester 2009; Wittmann & Fukami 2018; Yamamichi *et al.* 2020). Thus, we highlight the need

to combine ecological and evolutionary perspectives and methodologies to understand

community composition.

Acknowledgments:

This work was financed by an ERC (European Research Council) consolidator grant

COMPCON, GA 725419 attributed to SM, an ERC starting grant DYNAMICTRIO, GA

101042392 attributed to IF, and by FCT (Fundação para Ciência e Tecnologia) with the Junior

researcher contract (CEECIND/02616/2018) attributed to IF. OG acknowledges financial

support provided by the Spanish Ministry of Economy and Competitiveness (MINECO) and

by the European Social Fund through the UCCO project (CNS2023-144337). This is

contribution ISEM-2023-XXX of the Institute of Evolutionary Science of Montpellier (ISEM).

The authors acknowledge stimulating discussion with all members of the Mite Squad and Raul

Costa-Pereira, which have significantly improved the experimental design and interpretation

of the results, and Marta Artal and Liesbeth de Jong for invaluable infrastructure for meetings

in Sevilla. For the purpose of Open Access, a CC-BY 4.0 public copyright licence has been

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up to the Author Accepted Manuscript arising from this submission.

Competing interests: Authors declare no competing interests.

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Figures

Figure 1 – Predicted number of female offspring after one generation for A) *T. evansi* (Te) and B) *T. urticae* (Tu), to estimate the relative effect of competition on the growth rate of the different populations. To represent the cumulative impact of each parameter on offspring production, predictions were done using only the intrinsic growth rate (λ), the intrinsic growth rate plus intraspecific competition ($\lambda + \alpha_{ii}$) and the latter plus interspecific competition ($\lambda + \alpha_{ii} + \alpha_{ij}$) in an environment with or without cadmium (left and right panels, respectively). Note that, since two of the four treatments share the same focal selection regime (e.g. for *T.*

urticae the orange and red dots correspond to the focal control regime "Tu no-cadmium"), two colours are duplicates for the metrics without interspecific competition (*i.e.*, λ and $\lambda + \alpha_{ii}$). Starting conditions for simulations: 10 individuals of each species. Treatments correspond to combinations of *T. evansi* (Te) and *T. urticae* (Tu) selection regimes (cf. colour code). Note the different scales on the X-axis between the two species (A vs B). Confidence intervals for each prediction were estimated using the lower and upper estimates of the intrinsic growth rate and intra and interspecific competition indexes obtained from the model.

Figure 2 – Feasibility domain predicted for enviroments with (A) or without (B) cadmium for the different combinations of cadmium and no-cadmium selection regimes in which *T. urticae* (Tu) or *T. evansi* (Te) have evolved. The x and y axis correspond to the positive intrinsic growth rate for *T. urticae* and *T. evansi*, respectively. The arrow represents the vector of the intrinsic growth for each of the two species. The black lines delimiting the grey cone indicate the area under which the isoclines cross at positive abundances (i.e., the feasibility domain, in which coexistence is possible). Dashed lines surrounding the arrows represent the 95%

confidence interval of the intrinsic growth rate. Arrows that fall outside of the cone indicate

that Te excludes Tu, except in the upper-right panel of figure A (Te cadmium- Tu no-

cadmium), in which Tu excludes Te. The dark grey region indicates the feasibility domain

obtained from the ratio of intra and interspecific competitive abilities estimated with data from

all replicates pooled (see methods). This region has been evaluated only under the conditions

in which both species show positive intrinsic growth rates. The light blue and light red regions

delimitate the smallest and largest feasibility domains obtained from the lower and upper 95%

confidence interval of the competitive ability parameters (the light blue region is always

contained within the feasibility cone in grey). White regions denote the unfeasibility domain

(region for exclusion). Confidence intervals were obtained from the upper and lower

parameters estimates from the data of all experimental replicates pooled.

Figure 3 – Distance between the vector of intrinsic growth rates and the closest edge of the

feasibility domain in environments with (left panel) or without (right panel) cadmium for the

different combinations of no-cadmium or cadmium selection regimes for *T. evansi* (Te) and for

T. urticae (Tu) (cf. colour code). Positive distances indicate that the vector of growth rates is

inside the feasibility domain (i.e., coexistence is possible), and negative distances indicate that

the vector is outside the feasibility domain (i.e., exclusion is predicted). Circles correspond to

the distance calculated with all data pooled togetherand triangles to distances calculated per

replicate. Confidence intervals describe the difference between the vector of intrinsic growth

rates and the edges of the largest or smallest feasibility domain, which were obtained from the

lower and upper estimates of the parameters as described in Figure 2. The pooled data estimates

were obtainedfrom data from all experimental evolution replicates to increase statistical power

and to account for the variation between replicates.

Figure 4 – Correlation between predicted and observed relative abundance of *T. evansi* (Number of *T. evansi* females/Number of females of both species) after 2 generations in an independent population growth experiment in which the two species are released at equal densities. Squares indicate data from the cadmium environment and diamonds data from the no-cadmium environment. The full line represents the slope obtained from a general linear model (see methods) and the shaded area corresponds to the confidence interval for the slope obtained from the model. Treatments correspond to combinations of selection regimes to which *T. evansi* (Te) and *T. urticae* (Tu) were exposed (cf. colour code). Predicted error bars were calculated based on the upper and lower parameter estimates of all data pooled and observed error bars were obtained from the standard error of the replicate populations. Dashed black line indicates 1:1 ratio between predicted and observed values.