A patchwork pathway of apparently recent origin enables degradation of the synthetic buffer compound TRIS in bacteria.

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Abstract

The widely used synthetic buffer compound TRIS was long considered to be biologically inert. Herein, we describe the discovery of a complete bacterial degradation pathway for TRIS. By serendipity, a *Pseudomonas* strain was isolated from sewage sludge that was able to grow with TRIS as only carbon and nitrogen source. Genome and transcriptome analyses revealed two adjacent gene clusters embedded in a mobile genetic element on a conjugative plasmid to be involved in TRIS degradation. Conjugational transfer of this plasmid into *P. putida* KT2440 enabled this strain to grow with TRIS, demonstrating that the complete TRIS degradation pathway can be transmitted by horizontal gene transfer. Heterologous gene expression revealed cluster I to encode a TRIS uptake protein, a TRIS alcohol dehydrogenase, and a TRIS aldehyde dehydrogenase, catalyzing the oxidation of TRIS into 2-hydroxymethylserine. Gene cluster II encodes a methylserine hydroxymethyltransferase and a D-serine dehydratase which plausibly catalyze the conversion of 2-hydroxymethylserine into pyruvate. Subsequent enrichments from wastewater purification systems led to the isolation of further TRIS-degrading bacteria from the *Pseudomonas* and *Shinella* genera carrying highly similar TRIS degradation gene clusters.

Our data indicate that TRIS degradation evolved recently via gene recruitment and enzyme adaptation from multiple independent metabolic pathways and database searches suggest that the TRIS degradation pathway is now globally distributed. Our study illustrates how engineered environments can enhance the emergence of new microbial metabolic pathways in short evolutionary time scales. This knowledge is valuable for developing new water purification processes in times of increasing water scarcity.

Introduction

Since the beginning of the industrial revolution hundreds of thousands of new pesticides, herbicides, medical drugs, and other industrial chemicals were synthesized to meet the growing demand of newly emerging agricultural, pharmaceutical, and industrial applications [1]. Many of these anthropogenic chemicals eventually end up in the environment, where their removal is initially restricted due to the absence of efficient biological degradation mechanisms. However, more and more degradation pathways for chemicals that were introduced in the early twentieth century are now being discovered in microbes, suggesting that functional mineralization pathways can evolve within decades after the initial exposure of the microbial community [2]. A prominent example is the degradation pathway for the anthropogenic herbicide atrazine, which has evolved in the genus *Pseudomonas* around fifty years after the large scale introduction of atrazine to the market, enabling these bacteria to exploit this compound as a carbon and nitrogen source for growth [3]. Similarly, catabolic mineralization pathways for other industrially relevant compounds that were introduced in the early twentieth century including herbicides, pesticides, food additives, and medical drugs are now found in different microorganisms isolated from habitats exposed to these chemicals [4–8]. Here, novel pathways can evolve as a result of sufficient positive selective pressure for either the detoxification of such chemicals or their use as a new nutrient source [2]. Typically, novel degradation pathways emerge in a process called patchwork assembly by the recruitment, adaptation, and mutation of one or more genes encoding promiscuous enzymes from other metabolic pathways, either from within the host organism via gene duplication or via horizontal gene transfer (HGT) from the microbial community [1, 2, 9]. If the mutations happen to increase the substrate affinity and catalytic activity towards the respective new substrate and if the new catalytic activity enables the host to transform the new substrate into metabolites of an already existing metabolic pathway, it will gain a selective advantage over competing microbes by being able to use this compound as a new nutrient and energy resource [1, 2]. The genetic information encoding such novel catabolic pathways is often located on mobile genetic elements, such as plasmids or transposons, enabling transfer of the novel pathway to other microbes via HGT [4, 10, 11]. Eventually, positive selection for the new catabolic trait can lead to a loss of mobility and manifestation of the new pathway in the genome of its host.

However, the environmental concentrations of synthetic compounds may be too low to provide enough selection pressure for the evolution of novel catabolic pathways. Especially in wastewater treatment plants where the availability of other, readily degradable organic

compounds is high, many low-concentrated synthetic compounds presumably reside too
shortly for the evolution of novel degradation pathways and remain in the purified wastewater
as so-called micropollutants [12, 13]. Even some synthetic and natural compounds for which
efficient biodegradation pathways exist, such as steroid hormones, are sometimes not fully
metabolized in wastewater treatment plants due to their low concentration and contribute to the
micropollutant pool. To enhance the biodegradation of such micropollutants, bioaugmentation
of water treatment facilities is increasingly being considered, which implies the establishment
of microorganisms in the water purification processes, which have been isolated for their ability
to efficiently degrade the respective organic compounds [14, 15].

The study presented here was originally set up to isolate bacteria capable of degrading
steroid hormones in low concentrations from activated sludge. To increase the chances of
isolating strains that can later be established for bioaugmentation in real wastewater treatment
plant conditions, we used an artificial wastewater medium to mimic typical wastewater nutrient
conditions. Surprisingly, our enrichment procedure resulted in the isolation of a bacterial strain
that did not grow with the provided steroid hormones, but with the synthetic medium buffer
TRIS (2-amino-2-(hydroxymethyl)propane-1,3-diol). Herein, we pursued this finding further
and present the identification and characterization of a bacterial degradation pathway for TRIS,
which has been considered to be biologically inert so far.

Results

Isolation and characterization of the TRIS-degrading strain *Pseudomonas hunanensis* Teo1

The first TRIS-degrading organism was isolated from an enrichment culture that was originally designed to isolate testosterone-degrading bacteria from activated sludge (**Supp. Methods**). We repeatedly observed growth in the enrichment cultures in TRIS-buffered artificial wastewater medium without removal of the provided carbon source testosterone (not shown). Plating of the enrichment cultures onto solid medium containing testosterone resulted in the isolation of a bacterial strain that was not able to grow with testosterone but showed growth in liquid artificial wastewater medium without the addition of any extra carbon or nitrogen source. Subsequent growth experiments in phosphate buffered minimal medium with different concentrations of TRIS confirmed a linear relationship between the provided TRIS concentration and the amount of biomass produced by the isolate (**Fig. 1A**), suggesting that TRIS was the growth substrate in these cultures. Growth experiments coupled with HPLC-MS analysis confirmed that TRIS was completely removed from the culture medium, suggesting its utilization as a carbon and nitrogen source for growth (**Fig. 1B-D**). The isolate had an exponential growth rate of 0.36 +/- 0.09 h^{-1} and a doubling time of 0.44 +/- 0.09 h (n = 6). The only detectable metabolite accumulating in the medium was ammonium, with around 40 % of the nitrogen introduced with the TRIS substrate being released as ammonium into the medium (**Fig. 1C-D**). The strain did not grow with the structurally similar amino alcohols serinol, 2-amino-2-methyl-1-propanol, and 2-amino-2-methyl-1,3-propanediol (**Supp. Fig. S1**), while it did grow with the quaternary ammonium compounds choline and betaine, and the secondary amine sarcosine, and with the amino acids D- and L-serine, but not with glycine. Based on its whole genome sequence (**Supp. Methods**), the isolate was taxonomically classified as a representative of the species *Pseudomonas hunanensis* and was named strain Teo1 (**Supp. Tab. ST1**). The genome of strain Teo1 consists of a circular chromosome (6.13 Mbp) and two large, circular plasmids, p1_Teo1 (156.3 kbp) and p2_Teo1 (65.0 kbp).

Identification of TRIS degradation genes

TRIS growth experiments and time-resolved analysis of TRIS degradation in cell suspensions of strain Teo1 showed that Tris degradation was induced in cells pre-grown with TRIS while it was not induced in pyruvate pre-grown cells (**Supp. Fig. S2**), suggesting that TRIS catabolism is regulated at the transcriptional level in strain Teo1. Based on this observation we used transcriptomics of TRIS-grown and pyruvate-grown Teo1 cells to identify

potential TRIS degradation genes. A total of 398 genes were differentially regulated in TRIS-

compared to pyruvate-grown cells (fold change (FC) > 4, p_{adj} < 0.05, **Fig. 1E**, **Dataset SD1**),

of which 65 were upregulated in TRIS-grown cells. Ten of the twelve most upregulated genes

in TRIS-grown cells were found in two adjacent gene clusters (**Fig. 1E-G**) localized on the

p1_Teo1 plasmid. The proteins encoded in cluster I were annotated as a solute binding protein,

an amino acid/polyamine transporter, a choline dehydrogenase, a NAD/NADP-dependent

aldehyde dehydrogenase, a $\Delta(1)$-pyrroline-2-carboxylate/$\Delta(1)$-piperideine-2-carboxylate

reductase, and an alpha-ketoglutaric semialdehyde dehydrogenase (**Tab. 1**). The proteins

encoded in cluster II were annotated as a serine hydroxymethyltransferase, a D-serine

dehydratase, and two proteins presumably involved in the transformation of methylene

tetrahydrofolate into tetrahydrofolate and formate. The clusters are separated by a gene that

encodes a helix-turn-helix (HTH)-type transcriptional regulator. Based on the similar

upregulation of their genes, both gene clusters appear to be expressed as operons (**Fig. 1G**).

Due to their strong upregulation and their colocalization in putative operons we proposed that

these genes or a subset of them are involved in TRIS degradation in strain Teo1.

Characterization of the initial TRIS degradation reactions

Based on the predicted protein functions encoded in the upregulated gene clusters, we

hypothesized that TRIS is initially oxidized by the proteins encoded in cluster I. To test this,

we cloned combinations of the genes encoding the putrescine importer-like protein, the choline

dehydrogenase-like protein and the aldehyde dehydrogenase into the expression plasmid

pUCp18 into *E. coli* (**Suppl. Methods**) and tested high density cell suspensions of the resulting

strains for TRIS transformation activity. While the *E. coli* strains that carry only the putrescine

importer gene or the choline and aldehyde dehydrogenase genes did not transform TRIS, the

strain with all three genes transformed TRIS almost completely into a new compound within

24 h (**Fig. 2A**). This compound was identified as 2-hydroxymethylserine (**Fig. 2B**) based on

its molecular weight, which was 14 Da heavier than that of TRIS, suggesting the oxidation of

one hydroxymethyl group into a carboxylic group. To test whether this presumable TRIS

degradation intermediate can be further utilized by strain Teo1, sterilized culture supernatants

containing 2-hydroxymethylserine were used as growth medium for Teo1, confirming that

strain Teo1 can grow with and fully degrade 2-hydroxymethylserine (**Fig. 2C**). These results

supported our hypothesis that TRIS is transported into the cells by the putrescine importer-like

protein and subsequently oxidized by the choline dehydrogenase-like protein and the aldehyde

dehydrogenase. Thus, we propose that these genes encode a TRIS uptake protein (TupA), and

two TRIS-oxidizing proteins, namely a TRIS alcohol dehydrogenase (TaoA) and a TRIS

aldehyde dehydrogenase (TaoB).

Isolation of further TRIS-degrading bacteria and genomic analysis of TRIS degradation

Subsequent targeted enrichment of TRIS-degrading bacteria resulted in the isolation of

seven *Pseudomonas* strains (Teo2, Teo3, Teo4, Teo6, Teo8, Teo10, and Teo11) and one

Shinella zoogloeoides strain (Teo12) (**Supp. Tab. ST1**). All isolates were all able to grow with

TRIS as only carbon and nitrogen source and completely removed TRIS from the culture

medium (**Supp. Fig. S3**). The *Pseudomonas* strains Teo2, Teo4, Teo6, Teo8, and Teo11 and

the *Shinella* isolate strain Teo12 were isolated from different municipal wastewater treatment

plant samples, *Pseudomonas sichuanensis* Teo3 from an activated charcoal basin of a water

purification plant, and *Pseudomonas* sp. strain Teo10 from a freshwater lake sample.

Enrichment cultures from two additional wastewater treatment plants and from five freshwater

habitats did not develop growth (**Supp. Tab. ST1)**.

We sequenced the genomes of five *Pseudomonas* isolates (**Supp. Methods**) and found

homologous TRIS degradation gene clusters to the ones identified in strain Teo1 in all of these

strains (**Fig. 3, Suppl. Fig. S4**). In *Pseudomonas* strains Teo2, Teo4, Teo6, and Teo8 the

clusters are also localized on plasmids (p1_Teo2, p1_Teo4, p1_Teo6, and p2_Teo8) and the

plasmids have high similarities to p1_Teo1 (**Fig. 3A**). In *Pseudomonas* strain Teo3, the clusters

are encoded in the chromosome (**Fig. 3B**). While all gene clusters II are highly similar, only

the *tupA*, *taoA*, and *taoB* genes are conserved in all cluster I sequences (**Fig. 3, Suppl. Fig. S4**),

suggesting that the Δ(1)-pyrroline-2-carboxylate/Δ(1)-piperideine-2-carboxylate reductase-

like protein and the alpha-ketoglutaric semialdehyde dehydrogenase-like protein are not

required for TRIS degradation. In all these strains, the TRIS degradation gene clusters are

flanked by sequences similar to known bacterial insertion elements (**Fig. 3B**), suggesting that

the TRIS degradation gene clusters are located within a mobile genetic element [11]. A search

in the NCBI nucleotide database identified one *Priestia* genome (accession CP065422.1) and

three *Pseudomonas* genomes (CP069081.1, CP097105.1, CP114115.1) with homologous gene

clusters (**Fig. 3A, Suppl. Fig. S4**). In *Priestia*, and two of the *Pseudomonas* strains the clusters

are also plasmid-borne, while they are encoded chromosomally in *Pseudomonas* sp. strain

GXZC (**Fig. 3B**).

We also sequenced the genome of the *Shinella* isolate (**Supp. Methods**), which consists of

one circular chromosomes (3.54 Mbp) and three circular plasmids, p1_Teo12 (1.89 Mbp),

p2_Teo12 (221.54 kbp) and p3_Teo12 (140.01 kbp), which have no apparent similarity to the

TRIS degradation-encoding *Pseudomonas* plasmids. However, a DNA region on p2_Teo12

encodes homologs of all TRIS degradation genes in two adjacent gene clusters, except for the

genes encoding the solute binding protein, the uptake protein TupA, the Δ(1)-pyrroline-2-

carboxylate/Δ(1)-piperideine-2-carboxylate reductase-like protein, and the alpha-ketoglutaric

semialdehyde dehydrogenase-like protein (**Fig. 3B, Suppl. Fig. S4**). In these clusters, the *taoA*

and *taoB* gene homologs are encoded in the same orientation as in cluster I of the *Pseudomonas*

TRIS degradation gene clusters, while the genes are differently oriented in cluster II. A

homologous regulator gene to the one found in the *Pseudomonas* gene clusters is located

upstream of this cluster. A search in the NCBI nucleotide database identified a closely related

Rhizobium oryzihabitans strain M15 (CP048637.1), that carries a homologous DNA segment

on a plasmid (p5, **Fig. 3B**). The predicted TRIS degradation gene clusters in *S. zoogloeoides*

Teo12 and strain M15 are separated by a truncated transposase gene and flanked by additional

transposase genes and insertion elements.

A phylogenetic analysis of the TaoA proteins with closely related proteins of the glucose-

methanol-choline oxidoreductase family (GMC, InterPro IPR012132) revealed that the TaoA

proteins form a unique cluster that diverges from a clade containing choline dehydrogenase

protein sequences (**Supp. Fig. S5A**). In contrast, other choline dehydrogenase proteins of the

TRIS-degrading isolates encoded outside of the TRIS degradation gene clusters cluster within

the choline dehydrogenase clade, except for two choline dehydrogenase proteins from *S.*

zoogloeoides Teo12, which cluster at the base of the TaoA cluster. A phylogenetic analysis of

the TaoB proteins with closely related proteins containing an aldehyde dehydrogenase domain

(InterPro IPR015590**)** showed that the TaoB proteins also from a distinct cluster within the

dataset with closest relations to a reviewed aldehyde dehydrogenase (PuuC, Accession P23882)

that oxidizes aldehydes during putrescine catabolism in *E. coli* (**Supp. Fig. S5B**). Other PuuC-

like proteins encoded outside of the TRIS degradation gene clusters of the TRIS-degrading

isolates cluster in close proximity to the TaoB cluster.

Conjugational transfer of TRIS degradation plasmids

The TRIS degradation-encoding *Pseudomonas* plasmids have high similarities to the

previously described large plasmids pND6-2 from *Pseudomonas putida* ND6 and pS04-90

from *Pseudomonas aeruginosa* S04-90 [16, 17]. Both these plasmids lack the putative mobile

genetic element that encodes TRIS degradation (**Fig. 3A, Suppl. Fig. S4**). However, pS04-90

carries a carbapenem resistance-encoding class 1 integron at almost the same site in the pND6-

2 backbone as the TRIS degradation-encoding elements in the *Pseudomonas* plasmids. Both

pND6-2 and pS04-90 are conjugative plasmids that can be transferred from their hosts into
other *Pseudomonas* strains. Like pND6-2 and pS04-90, p1_Teo1 and its homologs carry genes
encoding type IV pili and a type IVB secretion system (**Fig. 3A, Suppl. Fig. S4**), suggesting
that these plasmids are also conjugative plasmids. To test this, we set up conjugation
experiments with strain Teo1 and strain Teo8 as plasmid donors and *Pseudomonas putida*
KT2440 as a non-TRIS-degrading recipient. To be able to differentiate between donor and
recipient cells, we labelled strain KT2440 with a chromosomally integrated gentamycin
resistance associated with a yellow fluorescent protein (YFP) gene. After the conjugation of
Teo1 and Teo8 with strain KT2440::*eyfp-gm* growth of fluorescing colonies was observed on
gentamycin-containing minimal medium plates with TRIS as only carbon and nitrogen source
(**Fig. 4A**), suggesting that conjugational transfer of the p1_Teo1 and p2_Teo8 plasmids also
transferred the ability to use TRIS as growth substrate to KT2440::*eyfp-gm*. The resulting
transconjugants were also able to grow with TRIS in liquid culture (**Fig. 4D**), confirming that
both gene clusters together encode the complete degradation pathway sufficient to use TRIS as
sole carbon and nitrogen source.
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Discussion

In this study we identify and characterize a degradation pathway for the synthetic buffer compound TRIS in several *Pseudomonas* strains and a *Shinella* strain, which were all isolated from wastewater purification systems with TRIS as only carbon source. Our results suggest that TRIS degradation proceeds via a pathway that originated recently via a combination of genes from different, formerly independent metabolic pathways and their assembly into two co-regulated gene clusters within a mobile genetic element. Although not all individual functions for the proteins encoded in clusters I and II were biochemically confirmed, conjugational transfer of the plasmid that contains both clusters into *P. putida* KT2440 co-transferred the ability to degrade TRIS, confirming that they together encode the complete degradation pathway sufficient to use TRIS as sole carbon and nitrogen source.

TRIS has long been considered biologically inert, because the bonding of its amino group to a tertiary carbon atom prevents its metabolism via transamination or oxidative deamination, which are the most common reactions in the degradation of amines. However, bacterial degradation of α-amino acids with a tertiary α-carbon atom is known for a long time and has been demonstrated for 2-hydroxymethylserine via a tetrahydrofolate-dependent methylserine hydroxymethyltransferase (mSHMT) with D-serine and formaldehyde as products [18]. All of our TRIS-degrading isolates encode a protein in cluster II with highest similarities to confirmed D-amino acid-producing mSHMTs [19] as well as a protein for the degradation of D-serine to pyruvate, suggesting that these proteins together can catalyze the degradation of 2-hydroxymethylserine, the product of TRIS oxidation by the TaoAB proteins. Thus, we propose a pathway for the catabolism of TRIS, in which TRIS is translocated into the cytoplasm by the TRIS uptake protein TupA and subsequently oxidized to 2-hydroxymethylserine by the TRIS alcohol dehydrogenase TaoA and the TRIS aldehyde dehydrogenase TaoB via a suspected aldehyde intermediate (**Fig. 5**). 2-Hydroxymethylserine is further transformed into D-serine by the mSHMT protein, which is then transformed into pyruvate by the D-serine dehydratase. The latter can be channelled into the tricarboxylic acid cycle via oxidative decarboxylation to acetyl-CoA. The formaldehyde-accepting tetrahydrofolate (THF) cofactor of the mSHMT protein can then be recycled by the bifunctional methylenetetrahydrofolate dehydrogenase/ methenyltetrahydrofolate cyclohydrolase FolD and the formyltetrahydrofolate deformylase PurU encoded in cluster II [20, 21].

The genetic context of the *tupA*, *taoA* and *taoB* genes is not found outside of cluster I, suggesting that these genes were assembled from multiple origins. The TaoA proteins have highest similarities to choline dehydrogenases (CDH), which catalyse the oxidation of choline into betaine aldehyde [22] and belong to the glucose-methanol-choline oxidoreductase family (GMC). Within a phylogenetic GMC protein tree, the TaoA proteins form a distinct clade that diverges from the CDH sequences with strong bootstrap support, suggesting that the TRIS alcohol dehydrogenase functionality evolved from enzymes with choline dehydrogenase activity. Strikingly, two CDH proteins encoded outside of gene cluster I in *S. zoogloeoides* Teo12 cluster basal to the TaoA protein clade, suggesting that they could be ancestral to TaoA. It is tempting to speculate that the TaoA proteins have originated in *Shinella* or a similar organism through the duplication and mutation of CDH-encoding genes. TupA and TaoB have highest similarities to proteins involved in the degradation of polyamines like putrescine [23, 24], suggesting that these proteins were recruited from bacterial polyamine degradation pathways [25]. The fact that Teo1 was not able to use other α-substituted amino alcohols such as 2-amino-2-methylpropanol and 2-amino-2-methylproandiol as growth substrates suggests that either the substrate uptake by TupA or the oxidation reactions catalysed by TaoAB evolved to be highly specific for TRIS.

Expression of *tupA* was strictly required for TRIS transformation in *E. coli* cells, suggesting that TRIS translocation is required for TRIS uptake in Gram-negative cells. With a pKa of 8.07 TRIS is mainly positively charged at neutral and acidic pH, preventing its diffusion across the cytoplasmic membrane. The absence of a *tupA* gene homolog in *S. zoogloeoides* Teo12 suggests that this gene was recruited later to the pathway. It remains unknow how TRIS is translocated by Teo12, however, alternative putrescine uptake transporters [23] are also encoded in the genome of Teo12, which might play a role in TRIS uptake in this strain.

The genetic context of the genes encoded in cluster II is also present in other *Pseudomonas* strains that do not encode cluster I, suggesting that this gene cluster may have been assembled prior to its recruitment into the TRIS degradation pathway. While SHMT and mSHMT proteins are often accompanied by THF co-factor recycling proteins [19, 26], their genetic association with D-amino acid degradation genes such as the D-serine dehydratase genes in cluster II in a putative operon was not reported so far. In pathogenic bacteria, D-serine dehydratase genes play important roles in the detoxification and catabolism of D-amino acids [27, 28]. Thus, we conclude that the second part of the TRIS degradation pathway was assembled by a combination of genes involved in the transformation of α-substituted amino acids and in the degradation of D-amino acids. The presence of mostly incomplete insertion

elements up- and downstream of the TRIS degradation gene clusters suggests that the whole
DNA segment was originally able to be mobilized as a composite transposon, allowing its
transposition between plasmids and the chromosome. The finding of an additional transposase
gene between the two gene clusters in *Shinella* and the closely related *Rhizobium* strain
corroborates our hypothesis that the pathway was assembled from two independent reactions
sequences encoded in clusters I and II.

TRIS is used in countless applications in research, industry, medicine, and household
settings. Its most prominent applications include its use as a buffer in molecular biology and
biochemistry labs, as an excipient in drug formulations, and as an active agent for the treatment
of metabolic acidosis and an enhancer for antibiotics and antiseptics for the treatment of Gram-
negative infections. Based on its first large scale production in the middle of the twentieth
century [29, 30], we hypothesize that the TRIS degradation pathway evolved within the last
thirty to fifty years, which is corroborated by the fact that TRIS degradation was not discovered
earlier despite its intense use in research laboratories. However, the organization of the TRIS
degradation genes in two compact gene clusters and the fact that expression of these
presumable operons was regulated in response to TRIS in strain Teo1 are signs for a highly
evolved pathway [5], suggesting its turn-around in the environment for a while. The regulator
gene between the two clusters is presumably involved in this regulation. Its protein product is
similar to the transcription factor GltC that regulates glutamate synthase expression in *Bacillus*
subtilis [31]. The first reports of bacteria being able to utilize TRIS as a growth substrate were
published thirty years ago from a group in Spain [32], however, no genetic or biochemical
analyses followed this observation.

Specific microbial habitats such as sinks and wastewater treatment facilities of large
research laboratories, hospitals, or from the TRIS-manufacturing industry could be locations
where the TRIS degradation pathway might have originated. First, such habitats encounter
significant TRIS concentrations on a regular basis. For example, the antibiotic fosfomycin,
which is produced and administered as a TRIS salt, was found at concentrations of 0.13 mM
in the influent and 0.015 mM in the effluent of a wastewater treatment plant of a fosfomycin
production plant [33], suggesting the presence of similar concentrations of its counterion TRIS.
Second, sinks and water purification systems are hotspots for horizontal gene transfer (HGT)
[11, 34, 35], which additionally favors the evolution of novel catabolic pathways.

The identification of identical TRIS degradation gene clusters in bacteria that have been
isolated in Germany, the Republic of Korea, and in China corroborates our findings, that TRIS
degradation-encoding plasmids can be readily transferred among bacteria by conjugation-

driven HGT. Our isolation studies suggest that the TRIS degradation pathway is frequently present in the microbiome of municipal wastewater treatment plants but less frequent in freshwater habitats, indicating that environments with high carbon and nitrogen loads better support the maintenance and dissemination of the TRIS degradation pathway than more oligotrophic environments. Although the average TRIS concentration in the sampled wastewater treatment plants is presumably very low, its metabolism as an additional nitrogen and carbon source alongside other recalcitrant molecules might provide sufficient positive selection pressure for bacteria with broad substrate ranges such as *Pseudomonas* and *Shinella* [36–39] to maintain the pathway. Our results corroborate the importance of HGT and mobile genetic elements in the evolution of novel catabolic pathways. With this, they provide additional knowledge about the basic structure of such elements, which can be used for future use in engineering bacteria for bioaugmentation [11].

Experimental procedures

Additional experimental methods can be found in the supplemental information.

Targeted enrichment and growth cultures

TRIS-degrading strains were isolated and grown in a phosphate-buffered minimal medium (pH 7.0) containing 0.01 mM $CaCl_2$ x 2 H_2O, 1.5 mM $MgSO_4$ x 7 H_2O, 35 mM K_2HPO_4 x 3 H_2O, 15 mM NaH_2PO_4 x H_2O, trace elements and 10 mM TRIS. For solid media plates, 1.5 % agar was added prior to autoclaving. Enrichment cultures (5 ml) were inoculated with fresh samples (1 ml) from local wastewater treatment plants (activated sludge or activated carbon), a water purification plant (activated carbon), or different freshwater habitats (**Supp. Tab. ST1**). When enrichment cultures showed growth, 100 µl of the culture supernatant were transferred to 5 ml fresh minimal medium with TRIS, and this was repeated two to three times before cultures were spread onto solid minimal medium plates containing to isolate individual TRIS-degrading colonies. Growth experiments in liquid cultures were carried out at 30°C and 200 rpm in 1 ml or 5 ml medium in 24-well plates or in reaction tubes, respectively. Experimental cultures were inoculated from washed, over-night starter cultures grown with TRIS. Growth was followed by measuring the optical density at 600 nm (OD_{600}). For transcriptome experiments, TRIS was replaced with 20 mM pyruvate and 20 mM NH_4Cl were added as nitrogen source. For substrate tests, TRIS was replaced with 10 mM D- or L-serine, 10 mM choline, 10 mM betaine, 10 mM 2-amino-2-methylpropanol, 10 mM 2-amino-2-methylproandiol, 10 mM sarcosine, or 10 mM serinol, and 20 mM NH_4Cl were added as nitrogen source. *Pseudomonas putida* KT2440 and its derivatives were grown in the same minimal medium with succinate as carbon source at 30°C. *E. coli* strains were grown in half-concentrated lysogeny broth (LB) at 37°C. If required, kanamycin (20-50 µg/ml), tetracycline (10 µg/ml), or gentamycin (20-40 µg/ml) antibiotics were added to the medium after autoclaving. To test whether 2-hydroxymethylserine can be used as a growth for Teo1, cell suspensions of *tupA*, *taoB* and *taoA* expressing *E. coli* cells that had produced 2-hydroxymethylserine from TRIS were centrifuged and sterile filtered (0.2 µm) and the culture supernatant was inoculated with strain Teo1 cells.

TRIS quantification

TRIS, its degradation intermediates, and ammonium were identified and quantified by high performance liquid chromatography (HPLC) coupled to a mass spectrometer and a UV/Vis detector [40] after derivatization of their amino group with diethyl ethoxymethylenemalonate (DEEMM, [41]). For this, 100 µl culture supernatants were mixed with 130 µl borate buffer (1

M, pH 9.0), 75 µl methanol, and 3 µl DEEM reagent, followed by 30 min sonication and 60

min incubation at 70°C before injection into the HPLC. A gradient (flow rate 0.3 ml/min) of

ammonium-acetate buffer (10 mM, pH 3.0, with 0.1 % (v/v) formic acid, eluent A) and

acetonitrile (eluent B) were used for product elution starting with 10 % eluent B for 2 min,

increasing to 90 % eluent B within 22 min, remaining at 90 % eluent B for 2 min, and returning

to 10 % eluent B within 1 min, followed by an equilibration of 5 min. Products were identified

based on their mass spectra after electrospray ionization in positive ion mode. The capillary

voltage was 2500 V, the end plate offset 500 V and the capillary temperature 300°C. Nebulizer

pressure was 22.5 psi, dry gas flow 12 l/min and dry temperature 200°C. Quantification was

performed based on the product peak areas at 280 nm UV absorbance, and concentrations were

calculated based on concentration curves of authentic standards treated in the same way. The

concentration of 2-hydroxymethylserine was calculated using the same standard curve as for

TRIS.

Transcriptome analysis

P. hunanensis Teo1 was grown in minimal medium with TRIS as only carbon and nitrogen

source or with pyruvate and ammonium and cultures were harvested during mid-log phase.

Harvested cells were re-suspended in 800 µl RLT buffer (RNeasy Mini Kit, Qiagen) with β-

mercaptoethanol (10 µl ml^{-1}) and cell lysis was performed using a laboratory ball mill.

Subsequently, 400 µl RLT buffer (RNeasy Mini Kit Qiagen) with β-mercaptoethanol (10 µl

ml^{-1}) and 1200 µl 96% (vol./vol.) ethanol were added. For RNA isolation, the RNeasy Mini Kit

(Qiagen) was used as recommended by the manufacturer, but instead of RW1 buffer RWT

buffer (Qiagen) was used in order to isolate RNAs smaller than 200 nucleotides also. For

sequencing, the strand-specific cDNA libraries were constructed with a NEB Next Ultra II

Directional RNA library preparation kit for Illumina and the NEB Next Multiplex Oligos for

Illumina (New England BioLabs, Frankfurt am Main, Germany). Sequencing was performed

by using the HiSeq2500 instrument (Illumina Inc., San Diego, CA, USA) using the HiSeq

Rapid SR Cluster Kit v2 for cluster generation and the HiSeq Raid SBS Kit (50 cycles) for

sequencing in the single-end mode and running 1x50 cycles. Raw reads have been deposited

in the Sequence Read Archive under study SRP451968 (SRR25447477-SRR25447482). More

detailed information can be found in the Supplementary Information.

Conjugation experiments

Conjugation experiments were carried out using strains Teo1 and Teo8 as donors and gentamicin-resistant *P. putida* KT2440::*eyfp-gm* as recipient. Approximately 1 x 10^9 donor cells and 3 x 10^9 recipient cells were mixed and spotted onto sterile cellulose acetate filter disks (ca. 3 cm^2, 0.45 µm pores) on LB agar plates. Conjugation plates were incubated at 30°C for 24 h before cells were washed off the filter disks with 50 mM phosphate buffer (pH 7.2). Decimal dilutions were plated onto minimal medium agar plates with TRIS as only carbon and nitrogen source containing 40 µg/ml gentamicin. Transconjugants were screen using a ChemiDocTM Imaging System (Bio-Rad) using blue epi illumination and an emission filter at 530 nm. The conjugation rate was calculated by dividing the number of positive transconjugants by the number of recipient cells in each assay. The successful transfer of the p1_Teo1 and p2_Teo8 plasmids was confirmed by colony PCR amplifying the *tupA*, *taoA*, *taoB*, *glyA*, and *dsdA* genes (**Suppl. Tab. ST2**).

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Competing Interest Statement

The authors declare no conflict of interest.

Data availability

The whole-genome shotgun projects have been deposited at DDBJ/ENA/GenBank under the BioProject accession numbers PRJNA999090, PRJNA999091, PRJNA999092, PRJNA999094, PRJNA999097, PRJNA999137, PRJNA999141 and the transcriptome data under the BioProject accession number PRJNA999090 with the SRA accession numbers SRR25447477 - SRR25447482.

References

1. Kolvenbach BA, Helbling DE, Kohler H-PE, Corvini PF-X. Emerging chemicals and the evolution of biodegradation capacities and pathways in bacteria. *Curr Opin Biotechnol* 2014; **27**: 8–14.

2. Copley SD. Evolution of efficient pathways for degradation of anthropogenic chemicals. *Nat Chem Biol* 2009; **5**: 559–566.

3. Udiković-Kolić N, Scott C, Martin-Laurent F. Evolution of atrazine-degrading capabilities in the environment. *Appl Microbiol Biotechnol* 2012; **96**: 1175–1189.

4. Top EM, Springael D, Boon N. Catabolic mobile genetic elements and their potential use in bioaugmentation of polluted soils and waters. *FEMS Microbiol Ecol* 2002; **42**: 199–208.

5. Johnson GR, Spain JC. Evolution of catabolic pathways for synthetic compounds: bacterial pathways for degradation of 2,4-dinitrotoluene and nitrobenzene. *Appl Microbiol Biotechnol* 2003; **62**: 110–123.

6. Aulestia M, Flores A, Acosta-Jurado S, Santero E, Camacho EM. Genetic characterization of the ibuprofen-degradative pathway of *Rhizorhabdus wittichii* MPO218. *Appl Environ Microbiol* 2022; **88**: e00388-22.

7. Bonatelli ML, Rohwerder T, Popp D, Liu Y, Akay C, Schultz C, et al. Recently evolved combination of unique sulfatase and amidase genes enables bacterial degradation of the wastewater micropollutant acesulfame worldwide. *Front Microbiol* 2023; **14**: 14:1223838.

8. Rios-Miguel AB, Smith GJ, Cremers G, Van Alen T, Jetten MSM, Op Den Camp HJM, et al. Microbial paracetamol degradation involves a high diversity of novel amidase enzyme candidates. *Water Res X* 2022; **16**: 100152.

9. Copley SD. Evolution of a metabolic pathway for degradation of a toxic xenobiotic: the patchwork approach. *Trends Biochem Sci* 2000; **25**: 261–265.

10. Springael D, Top EM. Horizontal gene transfer and microbial adaptation to xenobiotics: new types of mobile genetic elements and lessons from ecological studies. *Trends Microbiol* 2004; **12**: 53–58.

11. Rios Miguel AB, Jetten MSM, Welte CU. The role of mobile genetic elements in organic micropollutant degradation during biological wastewater treatment. *Water Res X* 2020; **9**: 100065.

12. Schwarzenbach RP, Escher BI, Fenner K, Hofstetter TB, Johnson CA, von Gunten U, et
al. The challenge of micropollutants in aquatic systems. *Science* 2006; **313**: 1072–1077.

13. Petrie B, Barden R, Kasprzyk-Hordern B. A review on emerging contaminants in
wastewaters and the environment: Current knowledge, understudied areas and
recommendations for future monitoring. *Water Res* 2015; **72**: 3–27.

14. Benner J, Helbling DE, Kohler H-PE, Wittebol J, Kaiser E, Prasse C, et al. Is biological
treatment a viable alternative for micropollutant removal in drinking water treatment
processes? *Water Res* 2013; **47**: 5955–5976.

15. Herrero M, Stuckey DC. Bioaugmentation and its application in wastewater treatment: A
review. *Chemosphere* 2015; **140**: 119–128.

16. van der Zee A, Kraak WB, Burggraaf A, Goessens WHF, Pirovano W, Ossewaarde JM, et
al. Spread of carbapenem resistance by transposition and conjugation among
Pseudomonas aeruginosa. Front Microbiol 2018; **9**: 2057.

17. Wang S, Li S, Du D, Wang D, Yan W. Conjugative transfer of megaplasmids pND6–1 and
pND6–2 enhancing naphthalene degradation in aqueous environment: characterization
and bioaugmentation prospects. *Appl Microbiol Biotechnol* 2020; **104**: 861–871.

18. Wilson EM, Snell EE. Metabolism of α-methylserine. *J Biol Chem* 1962; **237**: 3171–3179.

19. Nozaki H, Kuroda S, Watanabe K, Yokozeki K. Cloning of the gene encoding α-
methylserine hydroxymethyltransferase from *Aminobacter* sp. AJ110403 and *Ensifer* sp.
AJ110404 and characterization of the recombinant enzyme. *Biosci Biotechnol Biochem*
2008; **72**: 3002–3005.

20. D'Ari L, Rabinowitz JC. Purification, characterization, cloning, and amino acid sequence
of the bifunctional enzyme 5,10-methylenetetrahydrofolate dehydrogenase/5,10-
methenyltetrahydrofolate cyclohydrolase from *Escherichia coli. J Biol Chem* 1991; **266**:
23953–23958.

21. Nagy PL, Marolewski A, Benkovic SJ, Zalkin H. Formyltetrahydrofolate hydrolase, a
regulatory enzyme that functions to balance pools of tetrahydrofolate and one-carbon
tetrahydrofolate adducts in *Escherichia coli. J Bacteriol* 1995; **177**: 1292–1298.

22. Wargo MJ. Homeostasis and catabolism of choline and glycine betaine: Lessons from
Pseudomonas aeruginosa. Appl Environ Microbiol 2013; **79**: 2112–2120.

23. Krysenko S, Wohlleben W. Polyamine and ethanolamine metabolism in bacteria as an
important component of nitrogen assimilation for survival and pathogenicity. *Med Sci*
2022; **10**: 40.

24. Luengo JM, Olivera ER. Catabolism of biogenic amines in *Pseudomonas* species. *Environ*
Microbiol 2020; **22**: 1174–1192.

25. Nemoto N, Kurihara S, Kitahara Y, Asada K, Kato K, Suzuki H. Mechanism for regulation
of the putrescine utilization pathway by the transcription factor PuuR in *Escherichia coli*
K-12. *J Bacteriol* 2012; **194**: 3437–3447.

26. Shetty S, Varshney U. Regulation of translation by one-carbon metabolism in bacteria and
eukaryotic organelles. *J Biol Chem* 2021; **296**: 100088.

27. Connolly JPR, Roe AJ. When and where? Pathogenic *Escherichia coli* differentially sense
host D-serine using a universal transporter system to monitor their environment. *Microb*
Cell 2016; **3**: 181–184.

28. Li G, Lu C-D. The cryptic *dsdA* gene encodes a functional D-serine dehydratase in
Pseudomonas aeruginosa PAO1. *Curr Microbiol* 2016; **72**: 788–794.

29. Bourguignon J, Sion M-X, Moreau M. Preparation of Tris(hydroxymethyl)aminomethane.
US Patent No. US4233245A, U.S. Patent and Trademark Office 1980.

30. McMillan GW. Electrolytic production of aminoalcohols. US Patent No. US2485982A,
U.S. Patent and Trademark Office 1949.

31. Bohannon DE, Sonenshein AL. Positive regulation of glutamate biosynthesis in *Bacillus*
subtilis. *J Bacteriol* 1989; **171**: 4718–4727.

32. Fábregas J, Vázquez V, Cabezas B, Otero A. Tris not only controls the pH in microalgal
cultures, but also feeds bacteria. *J Appl Phycol* 1993; **5**: 543–545.

33. Zeng P, Xie X, Song Y, Liu R, Zhu C, Galarneau A, et al. Ion chromatography as highly
suitable method for rapid and accurate determination of antibiotic fosfomycin in
pharmaceutical wastewater. *Water Sci Technol* 2014; **69**: 2014–2022.

34. Abe K, Nomura N, Suzuki S. Biofilms: hot spots of horizontal gene transfer (HGT) in
aquatic environments, with a focus on a new HGT mechanism. *FEMS Microbiol Ecol*
2020; **96**: fiaa031.

35. Diorio-Toth L, Wallace MA, Farnsworth CW, Wang B, Gul D, Kwon JH, et al. Intensive
care unit sinks are persistently colonized with multidrug resistant bacteria and
mobilizable, resistance-conferring plasmids. *mSystems* 2023; **8:** e00206-23.

36. Moore ERB, Tindall BJ, Martins Dos Santos VAP, Pieper DH, Ramos J-L, Palleroni NJ.
Nonmedical: Pseudomonas. In: Dworkin M, Falkow S, Rosenberg E, Schleifer K-H,
Stackebrandt E (eds). *The Prokaryotes*. 2006. Springer New York, New York, NY, pp
646–703.

37. Jiang HJ, Ma Y, Qiu GJ, Wu FL, Chen SL. Biodegradation of nicotine by a novel strain
Shinella sp. HZN1 isolated from activated sludge. *J Environ Sci Health Part B* 2011; **46**:
703–708.

38. Hesse C, Schulz F, Bull CT, Shaffer BT, Yan Q, Shapiro N, et al. Genome-based
evolutionary history of *Pseudomonas* spp. *Environ Microbiol* 2018; **20**: 2142–2159.

39. Huang Y, Yu Z, Liu L, Che Y, Zhang T. Acesulfame anoxic biodegradation coupled to
nitrate reduction by enriched consortia and isolated *Shinella* spp. *Environ Sci Technol*
2022; **56:** 13096-13106.

40. Holert J, Yücel O, Jagmann N, Prestel A, Möller HM, Philipp B. Identification of bypass
reactions leading to the formation of one central steroid degradation intermediate in
metabolism of different bile salts in *Pseudomonas* sp. strain Chol1. *Environ Microbiol*
2016; **8**: 3373–3389.

41. Alaiz M, Navarro JL, Girón J, Vioque E. Amino acid analysis by high-performance liquid
chromatography after derivatization with diethyl ethoxymethylenemalonate. *J
Chromatogr A* 1992; **591**: 181–186.

Figure and table legends

Tab. 1: Genes potentially involved in TRIS degradation in *Pseudomonas hunanensis* Teo1.

Gene	Locus tag	Annotated function	Confirmed function and new gene ID	Protein family (InterProScan)	Best hit (Swiss-Prot)	% identity	
purU	PshuTeo1_56140	Formyltetrahydrofolate deformylase		Formyltetrahydrofolate deformylase (IPR004810)	Q46339, Formyltetrahydrofolate deformylase, *Corynebacterium* sp.	46.8	Cluster II
folD	PshuTeo1_56150	Bifunctional protein FolD		Tetrahydrofolate dehydrogenase/ cyclohydrolase (IPR000672)	A6UB98, Bifunctional protein FolD 2, *Sinorhizobium medicae*	73.0	
dsdA	PshuTeo1_56160	D-serine dehydratase		D-serine ammonia-lyase (IPR011780)	B3PYT5, Probable D-serine dehydratase, *Rhizobium etli*	66.1	
glyA	PshuTeo1_56170	Serine hydroxy-methyltransferase 2		Serine hydroxymethyltransferase (IPR001085)	B2DEU8, 2-methylserine hydroxymethyltransferase, *Aminobacter* sp.	66.6	
gltC	PshuTeo1_56180	HTH-type transcriptional regulator		non predicted	P71318, Hydrogen peroxide-inducible genes activator, *Pectobacterium carotovorum*	26.9	
56190	PshuTeo1_56190	hypothetical protein		none predicted	P45766, Putative amino-acid ABC transporter-binding protein YhdW, *Escherichia coli*	68.9	Cluster I
puuP	PshuTeo1_56200	Putrescine importer	TRIS uptake protein, *tupA*	Amino acid/polyamine transporter I (IPR002293)	P76037, Putrescine importer PuuP, *Escherichia coli*	41.8	
puuC	PshuTeo1_56210	NADP/NAD-dependent aldehyde dehydrogenase	TRIS aldehyde dehydrogenase, *taoB*	none predicted	P23883, NADP/NAD-dependent aldehyde dehydrogenase PuuC, *Escherichia coli*	46.0	
betA	PshuTeo1_56220	Oxygen-dependent choline dehydrogenase	TRIS alcohol dehydrogenase, *taoA*	Glucose-methanol-choline oxidoreductase (IPR012132)	Q9L4K0, Oxygen-dependent choline dehydrogenase, *Chromohalobacter salexigens*	48.3	
lhpD	PshuTeo1_56230	Δ(1)-Pyrroline-2-carboxylate/Δ(1)-piperideine-2-carboxylate reductase		Malate/L-lactate dehydrogenase-like (IPR003767)	Q9I492, Δ(1)-pyrroline-2-carboxylate/Δ(1)-piperideine-2-carboxylate reductase, *Pseudomonas aeruginosa*	67.0	
aldH	PshuTeo1_56240	Alpha-ketoglutaric semialdehyde dehydrogenase		Alpha-ketoglutaric semialdehyde dehydrogenase (IPR044151)	Q6FFQ0, Alpha-ketoglutaric semialdehyde dehydrogenase, *Acinetobacter baylyi*	58.5	

Fig. 1: Characteristics of TRIS metabolism in *Pseudomonas hunanensis* Teo1 and identification of potential TRIS
degradation genes. (**A**) Maximal biomass production of strain Teo1 in relation to TRIS concentration. Error bars
indicate standard deviation (n = 3). (**B**) Growth of strain Teo1 with 10 mM TRIS as only carbon and nitrogen
source (n = 6). (**C**) Removal of TRIS (straight lines) and formation of ammonium (dashed lines) in Teo1 growth
cultures. (**D**) Representative HPLC chromatograms of a strain Teo1 culture growing with 10 mM TRIS. TRIS is
depleted from the culture while ammonium is released. (**E**) Volcano plot of transcriptome data showing
significantly up-regulated genes (red) under TRIS-grown (positive fold changes) and pyruvate-grown conditions
(negative fold changes) in Teo1. Genes predicted to be involved in TRIS degradation are labelled. (**F**) Transcript
abundances and log$_2$(fold changes) of the predicted TRIS degradation genes in Teo1 in TRIS- and pyruvate-grown
cells (three biological replicates of each growth substrate, horizontal bar indicates median, corrected p values of
all selected genes were < 0.05). (**G**) mRNA reads (red: positive strand, blue: negative strand) of a representative
TRIS-grown culture mapped onto the DNA section containing the predicted TRIS degradation gene clusters in
strain Teo1.

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Fig. 2: TRIS transformation activity of *E. coli* strains carrying different combinations of the putrescine importer-
like protein-encoding gene (*puuP,* now *tupA*), the aldehyde dehydrogenase-encoding gene (*puuC,* now *taoB*), and
the choline dehydrogenase-like protein-encoding gene (*betA,* now *taoA*) on the expression plasmid pUCp18. (**A**)
Representative HPLC-UV/Vis chromatograms of TRIS-containing cell suspension supernatants of *E. coli* strains
that carried either the *tupA* gene, the *taoB* and *taoA* genes, or all three genes. TRIS (peak I, red) was transformed
into a new compound (peak II, orange) in the *E. coli* strain that carries all three genes only. The cloned genes were
under the activity of the *lac* promotor (grey arrow). (**B**) Mass spectra of TRIS (peak I) and the new product (peak
II). The new product was identified as 2-hydroxymethylserine based on the molecular mass of its diethyl
ethoxymethylenemalonate (DEEMM, grey molecule moiety) derivative of 305 Da, which is 14 Da heavier than
the molecular mass of 291 Da of the DEEMM derivative of TRIS. (**C**) Growth of strain Teo1 (blue circles) in
sterilized culture supernatants containing 2-hydroxymethylserine produced from TRIS and removal of 2-
hydroxymethylserine (white circles). Average and standard deviation of three independent biological replicates
are shown.
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Fig: 3: TRIS degradation gene clusters in the genomes of the sequenced TRIS-degrading isolates from this study
and of homologous gene clusters found in the NCBI nr nucleotide database. (**A**) Most TRIS degradation gene
clusters are located on large circular plasmids with high sequence similarities to the *Pseudomonas* plasmids
pND6-2 and pS04-90, which lack the TRIS degradation clusters. The outer black line indicates the location of the
putative mobile genetic element including the TRIS degradation genes. Like pND6-2 and pS04-90, the TRIS
degradation-encoding plasmids encode orthologs to mobile elements groups (inner circle) including a type IVB
secretion system and a type IV pilus for plasmid transfer as well as a type II toxin-antitoxin system. (**B**) All TRIS
degradation gene clusters, including plasmid- and chromosome-encoded clusters, are flanked upstream and
downstream by transposon- or insertion element-like structures. The plasmid p2_Teo12 from *Shinella* sp. strain
Teo12 and the homologous plasmid p5 in a *Rhizobium* strain have no similarity to the other TRIS degradation-
encoding plasmids, but also carry two gene clusters which encode homologs to the other TRIS degradation
proteins.
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Fig. 4: Conjugative transfer of the TRIS degradation pathway-encoding plasmids p1_Teo1 and p2_Teo8. (**A**)
Growth of YFP-fluorescing colonies on TRIS-containing agar plates after conjugation of *P. putida* KT2440::*eyfp-gm* with Teo1 (left panel) and Teo8 (right panel) as donor strains. (**B**) Conjugation rate of KT2440::*eyfp-gm* with p1_Teo1, p2_Teo8, and of a control without donor cells. Black lines indicate median rates of three independent conjugation experiments, boxes indicate the interquartile range, and dashed lines the maximum and minimum range. Transconjugants of up to seven replicates from three independent experiments were counted. Datapoints are coloured based on the donor strain. Outliers are not filled. (**C**) Degradation of TRIS (dashed lines) and growth (solid lines) of *P. putida* KT2440::*eyfp-gm* p1_Teo1 and p2_Teo8 transconjugants in liquid medium. The recipient strain *P. putida* KT2440::*eyfp-gm* without the conjugated plasmids was neither able to grow with, nor to degrade TRIS.

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607 **Fig. 5**: Proposed TRIS degradation pathway. Confirmed reactions and intermediates are in black, proposed
608 reactions and intermediates are in grey, THF = tetrahydrofolate, mTHF = methenyltetrahydrofolate, fTHF =
609 formyltetrahydrofolate.
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