Serial dilution shapes genetic variation and defines conservation units in

Asian elephants

Anubhab Khan*[1,2,&], Maitreya Sil*[1,3], Tarsh Thekaekara[1,4], Ishani Sinha[1], Rupsy

Khurana[5], Raman Sukumar[5,&], Uma Ramakrishnan[1,&]

[1] National Centre for Biological Sciences, TIFR, GKVK Campus, Bangalore, India

[2] School of Biodiversity, One Health and Veterinary Medicine, University of Glasgow,

Glasgow, UK

[3] National Institute of Science Education and Research, Bhabaneshwar, India

[4] The Shola Trust, Gudalur, India

[5] Centre for Ecological Sciences, Indian Institute of Science, Bangalore, India

*These authors contributed equally

[&] Corresponding authors: anubhabkhan@gmail.com, rsuku@iisc.ac.in,

uramakri@ncbs.res.in

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Abstract

Megaherbivores are primary consumers who provide unique ecosystem services.

Given their body size, they are disproportionately threatened in the Anthropocene.

Asian elephants are the largest extant terrestrial megaherbivores native to Asia, with

60% of the population found in India. Despite their ecological and cultural

importance, the management/conservation units, genetic history, diversity and

threats remain understudied. We re-sequenced 31 whole genomes (between 11X -

32X) from all known elephant landscapes in India and identified five

management/conservation units corresponding to elephants in northern India, central

India and three in southern India. The genetic data reveal signatures of serial

colonisation, and a dilution of diversity from north to south of India. The northern

populations diverged from other populations more than 70,000 years ago, and have

higher genetic diversity, with low inbreeding/high effective size (Pi = 0.0016±0.0001;

$F_{ROH > 1MB}$= 0.09±0.03). Two of three populations in southern India have low diversity

and are inbred with much lower effective sizes than current populations sizes (Pi =

0.0014±0.00009and 0.0015±0.0001; $F_{ROH > 1MB}$= 0.25±0.09 and 0.17±0.02).

Additionally, future generations are expected to be more inbred since pairs of extant

elephants have large tracts of the genome that are already identical. Analyses of

genetic load reveals purging of potentially high-effect deleterious alleles in the

southern populations and potential dilution of all deleterious alleles from north to

south in India. However, southern Indian elephants are highly homozygous for all the

deleterious alleles that persist, despite dilution and purging. High homozygosity of

deleterious alleles, coupled with low neutral genetic diversity make them high priority

for conservation and management attention. Most surprisingly, our study suggests

that patterns of genetic diversity and genetic load can correspond to geographic

signatures of serial founding events even in large mobile endangered species.

Introduction

Megaherbivores are ecological engineers responsible for maintaining several
ecosystem functions (Owen-Smith 1987, Waldram et al. 2008, Coverdale et al. 2016,
Berzaghi et al. 2023). They are uniquely responsible for nutrient distribution, seed
dispersal and germination, and vegetation/habitat modification (Le Roux et al. 2018,
Harich et al. 2016). In the Anthropocene they have been affected disproportionately
and face various threats across their range (Enquist et al. 2020). In Asia,
megaherbivores have lost 56% (for gaur, *Bos gaurus*) to near 100% (for Javan rhino,
Rhinoceros sondaicus) of their historic range (Mahmood et al. 2021). They are more
inbred and lack genetic diversity compared to their small-bodied counterparts
(Brüniche-Olsen et al. 2018), and several megaherbivores have gone extinct in the
last few centuries. While megaherbivore genomes are being extensively studied in
African landscapes, such investigations for Asian counterparts have been lagging,
leading to challenges in identification of genetic threats relevant to their conservation
and management.

Asian elephants (*Elephas maximus*) are charismatic megaherbivores distributed
across South and Southeast Asia but are culturally important across the globe
(Sukumar 2011). They are found in a variety of natural ecosystems from tropical
evergreen forests to grasslands at various elevations. India harbours at least 60% of
the population of wild Asian elephants (Sukumar 2011, Menon and Tiwari 2019).
Increase in human footprint and land use change over the past two centuries has
impacted elephants significantly, resulting in population isolation even at regional
scales. Today, Asian elephant habitats are extremely fragmented and interspersed
with farmland, human settlement, commercial plantations and linear transport
infrastructure throughout their range (Liu et al., 2017; Padalia et al., 2020) leading to
extensive and, often, intense human-elephant conflicts (Sukumar, 1989, 2003; Gubbi
et al., 2014). Despite these challenges, camera trap data (Srinivasaiah et al., 2022;
Srinivasaiah et al., 2021) and radiotelemetry studies (Baskaran et al. 1995,
Venkataraman et al. 2005, Sukumar 2003, Sukumar et al. 2003) show that elephants
have annual home ranges of several hundreds of kilometres in India, often
encompassing dense human habitation. Such movement through human-dominated
areas might offset impacts of fragmentation, and the associated loss of genetic

variation and inbreeding (Parida et al. 2022). Hence, understanding their
phylogeographic history and current population genetic structuring, coupled with
possible effects of recent fragmentation could be useful in defining population units
for conservation and management.

Phylogeographic studies of elephants point towards a tangled history of
demographic change, range expansion/contraction, as well as population admixture
(Vidya et al., 2009). Using a few microsatellite and mitochondrial markers, Vidya et
al. (2005) reported four Indian elephant population genetic clusters, one each in
North-Northeastern India and Central India, and two in Southern India separated by
the Palghat gap, broadly corresponding to their regional population distributions.
However, a more recent study by De et al. (2021) suggest only three major genetic
clusters corresponding to Northern India, Northeastern India, and a combined
Southern and Central Indian population (with the Northern population appearing to
be admixed). The patterns of genetic diversity also varied across different
ecoregions. Vidya et al. (2005) showed that the Southern populations harboured
lower genetic diversity (haplotype diversity) than the other populations, while
according to De et al. (2021), the Northeast Indian populations showed low
heterozygosity. The only genome-wide study of elephants globally (Palkopoulou et
al., 2018), without range-wide sampling of Asian elephants, show old divergence
between populations in southern India across the Palghat Gap and also between
northern and the southern populations. The limited published genomic data from
Asian elephants, hence, does not allow unequivocal inference of population
structure, divergence times, conservation units and conservation challenges.

To address these gaps, we used whole genome sequences of wild-caught Asian
elephants from all the major landscapes, encompassing the known biogeographic
barriers in India, to assess their genetic structure, demographic history and genetic
variation. We infer the predominant factors that have shaped the observed genetic
diversity and its structuring across the country. Further, we use this information to
suggest population management units and challenges that could be important to
elephant conservation.

Results and Discussion

Population structure

We collected 31 blood samples from wild born captive elephants of known origin from almost all known elephant landscapes in India (Supplementary Table 1, Parida et al. 2022) and re-sequenced whole genomes (figure 1a). A set of 2,675,655 SNP markers revealed distinct population structure within elephant landscapes in India. The PCA plot revealed that populations separate from the south to north direction along PC1 axis (figure 1b). Elephants from Northern India (Terai) and Northeastern India (Ne & N) form a single cluster (figure 1b and c). Populations from Central India and Northern India cluster together in the PCA but resolve into separate clusters in the STRUCTURE plot (figure 1c) with the most optimum support for five population clusters (North-Northeast, Central and three clusters in the Western Ghats: North of the Palghat gap (NPG), South of the Palghat gap (SPG) and South of Shencottah gap (SSG), EvalAdmix, Supplementary figures 1 and 2). The admixture graph that includes a Bornean elephant (ERR2260499) as an outgroup with branch lengths qualitatively adjusted to the drift parameter (figure 1d) suggests that the central Indian lineage (SWB) derive their ancestry from a lineage of elephants in northern India (N&NE) and Myanmar. Similarly, in southern India (NPG, SPG and SSG) the populations derive their ancestry from a lineage closely related to the NPG population and the SSG population derives its ancestry from a lineage closely related to the SPG population. These patterns suggest a North-northeastern ancestry of elephants in India and colonisation of Central India, with a deeply diverged lineage and sequential colonisation of the Western Ghats and adjoining Eastern Ghats of Southern India.

Our results support previous conclusions that the North and Northeast Indian populations of elephants are different from other Indian populations (Vidya et al. 2005a,b; De et al. 2021) and that the Ganges river has acted as a potential barrier to geneflow (figure 1a). Consistent with movement ecology inferences (Koirala et al. 2016) the elephants in the Northern population (described as Northwest Indian population in De et al. 2021) are connected to the Northeast Indian elephants. This is a large landscape running west to east along the Himalayan foothills and potentially the elephant habitat connectivity here is fragile (Koirala et al. 2016) or even

completely broken in recent times (Sukumar 2011, Menon and Tiwari 2019). Our

genetic data also suggest that the Brahmaputra river acts as a barrier to geneflow in

the Northeast population. Previous studies suggested that the river acts only as an

incomplete barrier, with female-led family groups unable to cross except perhaps at

the higher reaches, but not a barrier to adult male elephants (Vidya et al., 2005).

Anecdotal information suggests that male elephants swimming across the

Brahmaputra (RS and AK independent personal observations), suggesting it is not a

barrier to dispersal and movement. De et al., 2021 suggest that the Torsa river is

also a barrier to elephant movement. However, radio-telemetry and observational

studies in this region have shown that elephants certainly cross the river (Sukumar et

al. 2003). We were unable to explicitly investigate the role of the Torsa river as a

barrier because of inadequate geographic sampling. While geographically

proximate, we find that Central Indian populations are genetically distinct as

suggested previously (Vidya et al. 2005).

Our data and analyses allow identification of a novel genetic cluster in Southern

India and suggest three genetically differentiated populations in the Western Ghats.

Along the Western Ghats, certain mountain passes divide the elephant population

with the Palghat gap being the most prominent barrier to elephant dispersal. Further

south, the Shencottah Gap also acts as a previously unknown impediment to

elephant movement (though anecdotal information suggests that elephants moved

across this gap until a few decades ago). Interestingly the population south of

Shencottah Gap (SSG) is indistinguishable from the population north of Shencottah

(SPG) when SNPs are filtered for LD (Supplementary figures 3-4). This probably

shows that the differentiation is potentially because of founder effects and inbreeding

combined with recent isolation and small population size (Khan et al. 2022).

Geneflow between north and south of the Shencottah Gap may have reduced further

recently (compared to across the Palghat gap) largely due to a railway line, a

highway and associated development along these transportation infrastructures.

While nested biogeographic implications of these gaps on population structure and

phylogeography has been highlighted for smaller species (e.g. montane birds, Robin

et al., 2015; bush frogs, Vijaykumar et al., 2017, geckoes, Chaitanya et al., 2019)

and some mammals (e.g. Lion Tailed Macaque, Ram et al., 2015), that they result in

such deeply divergent lineages in large, mobile species like the elephant is

surprising. Alternatively, founding events followed by minimal geneflow might have

resulted in the observed patterns of divergence. Additionally, elephant numbers and

densities south of the Shencottah Gap have always remained small (on the order of

100 individuals). Such small populations are subject to genetic drift, which could

accentuate the observed patterns. Pairwise F_{ST} supports these inferences

(Supplementary Table 2) and we find no significant geneflow between the clusters

based on F3 statistics (Supplementary Table 3).

Interestingly, the results obtained from the haplotype network analysis based on the

mitogenome paints a slightly different picture (Supplementary figure 5). Similar to the

nuclear genome, the Northern populations were embedded within the Northeastern

populations. However, the Central population was allied to the Southern populations,

unlike the results obtained from the nuclear data, more in line with some previous

studies (Vidya et al., 2009, De et al., 2021).

Figure 1 Population structure (a) sampling locations and the geographic

barriers to dispersal (b) PCA (c) ADMIXTURE plot at K=5 and (d) qpgraph with

no admixture events with the branch length qualitatively adjusted to the drift

parameter. NE&N (n=12) indicates samples from Northeastern India (n=9) and

Northern India (n=3), CI (n=5) from Central India, NPG (n=9) from north of

Palghat Gap, SPG (n=4) from south of Palghat Gap but north of Shencottah
Gap, and SSG (n=2) from south of Shencottah Gap.

Demographic History

We investigated recent demographic history for clusters with more than nine
samples (N&NE and NPG), and we find that they had undergone a recent bottleneck
around 1500-1000 years ago (figure 2a). Since the taming of elephants in the
subcontinent during Harappan times, at least four millennia ago, there has been
regular exploitation of wild stocks of the species for military and domestic use
(Sukumar 2011). The historical accounts suggest that the armies of ancient
kingdoms and republics in the north (the Gangetic basin) maintained several
thousand captive elephants since as early as the 3rd century BCE, suggestive of
overexploitation of wild populations for several hundred years (Sukumar 2011). This
is the first report of a bottleneck in Asian elephant populations in recent historical
times, mediated perhaps by human exploitation, unlike earlier bottlenecks of a
variety of larger mammals globally during Pleistocene glaciations (e.g. Menotti-
Raymond and O'Brien 1993, Dalui et al. 2021), Our results also suggest that both
populations (N&NE and NPG) may have started recovering from the bottlenecks
around 300-500 years ago. Often population structure can dictate the models of
demographic history (Mazet et al. 2016). Additionally, we find signatures of
population bottlenecks around 100,000 years ago (figure 2b). However, further
investigation with hPSMC (Cahill et al. 2016) suggests that this coincides with
populations differentiating from each other (figure 2c). We find that the Northern
elephant population (NE&N) separated from all other populations about 70,000-
100,000 years ago. Central Indian elephants separated from the rest around 50,000-
80,000 years ago, while the Southern Indian populations separated from each other
only around 20,000-30,000 years ago.

Additionally, our results emphasise the antiquity of the Northern populations of
elephants consistent with Vidya et al. (2005, 2009). Palkopoulou et al. (2018) had
suggested more recent divergence which could be due to the differences in the
mutation rates used for making inferences (Prado et al. 2023). Vidya et al. (2005,
2009) made inferences based on mitochondrial sequences which are expected to
provide older estimates of divergence times (Moriyama & Powell 1997, Hugall et al.

2007). However, consistent with Palkopoulou et al. (2018) we find that the Southern

populations of elephants split from each other only around 20,000 years ago, or the

time of the Last Glacial Maximum when southern India, in particular the Western

Ghats, was more arid (Sukumar et al. 1993, Rajagopalan et al. 1997). These findings

further support the recognition of five elephant Mus in India, emphasising their

antiquity and unique evolutionary histories.

Figure 2 Demographic history (a) recent demographic history within in the last

100 generations from GONE (b) old demographic history from 100 generations

to 40,000 generations from SMC++ and (c) population divergence estimated

from hPSMC. The point where population sizes start increasing exponentially

is the time to population split in hPSMC plot.

Genetic Diversity

But how do these management units compare with each other? Within each cluster

we estimated the genetic diversity as the number of average pairwise differences in

sequences (pi). We find that the clusters that are from southern India (SPG, SSG

and NPG) have lower average pi than the northern Indian and central Indian

clusters. Interestingly, there is no visible difference in pi within the southern and

between the northern and central (N&NE and SWB) clusters (figure 3a)

(Supplementary Table 4). This suggests that all the clusters within southern have a

similar number of haplotypes and the central and Northern cluster have a similar

number of haplotypes. About 75% of the discovered SNPs are shared, or present in

all populations (Supplementary table 5). However, the number of heterozygous sites

encountered per Mb is the highest for the northern population (NE&N) while it is

lowest for the southernmost population (SSG) (figure 3b). We find no difference in

heterozygous SNV per Mb for NPG and SWB. Overall, our results suggest that most

populations of elephants have similar nucleotide diversity (pi) but, in the population

south of Shencottah gap (SSG), similar nucleotides are often paired together

indicating lower effective population size, a pattern not investigated earlier, while in

the populations in Northern India (NE&N) most of the nucleotides pair with a different

one indicative of higher effective size.

Genetic variation is related to effective population size which, in turn, is dependent

not only on the total population size of a species but other variables such as the sex

ratio (Frankel and Soulé 1981). Although we cannot speculate on the historical

population sizes of Asian elephants across the various regions we have investigated,

a cursory look at the recent population sizes of the five management/conservation

units we have inferred from the genomic data shows that population size per se is

not indicative of the observed genetic variation (pi and heterozygosity)

(Supplementary Table 6). For instance, the genetic variation of the Northeastern

population is much higher than that of the Southern population north of the Palghat

Gap. Both these clusters have similar current population sizes on the order c. 10,000

individuals. Central Indian elephants have higher variation than Southern populations

even though the former has distinctly lower population numbers. This might be

indicative of recent bottlenecks in central India, as heterozygosity decays slowly

while southern populations, especially those south of Palghat gap may have had

historically smaller populations. The highly female-biased sex ratios in Southern

populations, especially those to the South of the Palghat Gap, in recent historical

times (1970s-1990s) from selective poaching of male elephants for ivory (Sukumar

1989, Ramakrishnan et al. 1998), could have decreased the effective population size

(compared to census size), but this is unlikely to have decreased genetic variation to

the extent observed. Elephants South of Shencottah Gap were also connected to the

SPG population in recent history. Overall, our results are thus also consistent with a

serial dilution of variation that could be the result of sequential colonization

(Hellenthal et al., 2008; Pierce et al., 2014; Pless et al., 2022) from north to south:

N&NE, Central, Southern, with North of PG, South of PG and South of Shencottah,

in that order.

Figure 3 Genetic diversity (a) boxplots of pairwise nucleotide differences per

site (pi) in the populations (error bars are variance in mean pi/site across

scaffolds) and (b) heterozygous SNV encounter rate per Mb of the genome.

We further test this by estimating F_{ROH}, the proportion of the genome in homozygous

stretches, length of which is an indicator of older inbreeding/recent bottlenecks. We

observe that individuals from south of Shencottah gap (SSG) have high average

$F_{ROH>0.1Mb}$ of 0.4 (40% of the genome is in homozygous stretches) while individuals

from northern India (NE&N) have the least average $F_{ROH>0.1Mb}$ of 0.2 (20% of the

genome is in homozygous stretches, Figure 4a). There is no significant difference

between individuals from SPG, NPG and SWB populations (average $F_{ROH>0.1Mb}$ is

0.25). Longer homozygous stretches are indicative of recent inbreeding (Curik et al.,

2014; Sumreddee, 2021). ROH longer than 1Mb and 10Mb shows few differences

between populations although, the northern population (NE&N) has the least recent

inbreeding. Additionally, the SPG and NE&N populations have lower variance in

F_{ROH} compared to other populations, potentially indicating geneflow and connectivity

within these large landscapes, which will result in lower variance in parental

relatedness. We caution that these results do not directly imply choosing to mate

between relatives, behavioral studies suggest ongoing inbreeding avoidance in

elephants (Sukumar 2006), but more likely suggest founder effect induced drift and

mating between relatives, especially south of the Shencottah gap. Founder effects

will result in a small set of individuals as founders, consequentially leading to the

individuals in the population soon being related. Mating within this set of individuals

then leads to inbreeding.

We measured the percent genome that is identical by descent (IBD) shared between

pairs of individuals in each population (Figure 4b). We observe that elephants in the

northern Indian cluster (NE&N) share very few IBD stretches of genome (on an

average 1.6% in more than 10Mb long and 6.6% in more than 0.1Mb long IBD

stretches) with each other while the two individuals we had from the south of

Shencottah Gap (SSG) shared about 43% of their genome in stretches longer than

10Mb and about 52% of their genome in stretches longer than 0.1Mb. Apart from the

one outlier pair in the SPG sample set we observe very negligible difference in SPG,

NPG and SWB population with regards to the proportion IBD stretches of genome

shared between pairs of individuals. On an average these three clusters share more

than 10% of their genome in IBD longer than 10Mb (NPG: 8%, SPG: 14%,

SWB:12%) and 19% of their genome in IBD longer than 0.1Mb (NPG: 15%, SPG:

25%, SWB:17%). These results indicate possible threats from loss of fitness due to

inbreeding in upcoming generations. Since the present individuals already share

large IBD stretches of genome, impacts on fitness will be maximal, especially if these

individuals have higher numbers of deleterious alleles. However, since there are a

few outlier pairs with lower amounts of shared genomic IBD stretches (figure 4b)

there is hope that maintaining gene flow within these population clusters could

sustain extant genetic variation in these populations into the future.

Figure 4 Present and future inbreeding (a) inbreeding measured as FROH of

individuals in each population based on ROH stretches longer than 0.1, 1 and

10 Mb of genome and (b) identical by descent (IBD) stretches of genome

longer than 0.1Mb vs 10Mb shared between pairs of individuals in each

population.

Mutation load

We attempted to understand whether individuals, from certain populations, that have

high homozygosity could suffer fitness effects due to inbreeding depression by

investigating mutation load, or putative deleterious mutations. This depends on the

number of derived deleterious mutations harboured across the genome, how many

are homozygous, and the putative magnitude of their fitness effects (indel Loss of

function > Loss of function > missense mutation). We normalised this by derived

SNPs in neutral/intergenic regions, to account for possible missingness/uneven

coverage across genomes. As expected, populations with lowest genetic variation

(SSG) had fewer LOF mutations (5a, x axis). This could be because of the serial

founding demographic history from north to south (see Figure 1d). However, LOFs in

SSG were not entirely a subset of those in N&NE, supportive of emergence of *de-*

novo mutations in the SSG population due to long history of isolation between the

two populations. SSG population has the highest homozygosity for the LOF

mutations. Interestingly, SPG also had fewer homozygous LOF mutations (Figure

5a), potentially due to purging and *de-novo* mutation accumulation due to historical

isolation.

To better understand whether populations like SSG are endangered by imminent

fitness effects, we compared indel LOFs. While the total indel LOFs did not vary

considerably between populations (x axis, Figure 5b). The observation that all

populations have a similar indel LOF load (compared to LOF load, Figure 5a, x axis)

suggests that they are under strong purifying selection, and that these putative high

negative effect mutations could have already been purged from these populations.

In contrast, the homozygous indel LOFs remain higher in SSG. This suggests that

despite purging, elephants south of the Shencottah Gap could experience negative

fitness consequences of indel LOFs first. This population may be threatened with

inbreeding depression, or lower fitness of individuals, ensuing negative feedback on

population growth rates, and a potential for further increase in homozygosity due to

inbreeding in the future generations (Robinson et al. 2023). However, African

elephant populations (such as at Addo National Park) grew rapidly with no apparent
deleterious effects after a severe bottleneck only about a century ago (Whitehouse
and Hall-Martin 2000). Genetically small populations are expected to be especially
efficient at purging strongly deleterious alleles (Kyriazis et al. 2023, Khan et al.
2021), this does seem to be happening in the isolated population south of the
Shencottah Gap. But the question that remains is whether these endangered
populations with long generation times and low population growth rates can tolerate
the selection pressure brought about by highly deleterious alleles.

Southern populations (NPG, SPG and SSG) have fewer derived missense mutations
than the central (SWB) and the northern (NE&N) populations (Supplementary figure
6) but the populations south of Palghat Gap (SPG and SSG) have higher
homozygous missense mutation load than the NPG, SWB and NE&N populations,
suggesting lack of purging. The fewer number of mildly deleterious missense alleles
in the southern Indian population north of Palghat Gap (NPG) compared to SWB and
N&NE is counterintuitive, as current literature suggests larger populations host high
numbers of mildly deleterious alleles (Kyriazis et al. 2023, Khan et al. 2021). We
suggest serial dilution of missense variants combined with a long isolation from the
larger and connected north Indian population (NE&N) may have led to restricted
immigration of mild impact deleterious alleles in the NPG population while the LOF
are selected out equally well in the large NPG, SWB and NE&N populations.
Additionally, the low heterozygosity in the NPG, SWB and NE&N populations
cautions that all populations have potential for inbreeding depression (Robinson et
al. 2022).

Functionally, the derived missense mutations are related to sensory perception,
detection of chemical stimulus and especially olfaction (Supplementary figure 7).
This could also be due to independent evolution of olfactory genes in Asiatic
elephants (Reddy et al. 2015); however, since we also polarized the alleles with
Bornean elephant genome, the missense mutations can be confidently classified to
be recently derived and maybe potentially deleterious. The loss-of-function mutations
mostly affect protein, ions and nucleic acid binding abilities along with transferases
and transporters (Supplementary figure 8). Given that elephants rely heavily on
chemical signals, the deleterious mutations may impair crucial olfactory functions in

the elephants. Although inbred individuals are more homozygous for these mutations
and maybe expected to be severely affected, it is possible that sociality and herd
living allows reduced sensory abilities in individuals. However, this needs further
investigation.

Figure 5 Mutation load measured as a function of homozygosity vs number of
derived (a) indel mutations and (b) loss-of-function (lof) mutation. The number
of derived deleterious alleles/number of derived neutral alleles is a proxy for
number of deleterious alleles. The error bars indicate standard deviations.

Conservation implications

Overall, our results support five management units of Asian elephants in India.
Elephants in the Himalayan foothills from the Northern population (in the west) to the
Northeast of India are a single cluster that diverged from other Indian populations
more than 70,000 years ago. This makes the N&NE cluster the oldest and, hence,
the most evolutionarily unique population. Elephant populations south of the
Shencottah Gap need conservation attention through detailed genomic research at
fine spatial scales, conservation action to protect remaining habitats and minimise
conflict associated mortality, and serious consideration about translocations into and
out of this population. Our results suggest that further decline in population numbers
here could result in inbreeding depression. Long-term studies on inbreeding and
possible associated phenotypes will be important to understand future fitness
trajectories here. Populations north and south of the Palghat Gap are distinct
management units, and animals should be moved across this biogeographic divide
only after very careful evaluations of the consequences. Further, given the high
missense mutation load in these two management units, connectivity must be
maintained within these two landscapes. The Ganga and Brahmaputra river in West

Bengal function as biogeographic divides, making elephants south of these rivers a

unique management/conservation unit. Elephants here are facing significant

challenges because of negative human impacts, and their conservation must take

into account their unique 50,000 year old evolutionary history.

Our descriptive population genomic approach allowed for unique and important

conservation insights. To the best of our knowledge, elephants in India present one

of the first clear examples of serial founding of a large landscape at a subcontinental

scale, with decreasing founder size from the source. Our data allows us to detect

recent declines, potentially mediated by historic elephant captures on a large scale.

Further, elephants provide an excellent system to better understand the interplay

between mutation load and inbreeding in a set of five populations, a rare set up in

endangered species. While on-ground conservation challenges for elephants are

mitigation of human-associated conflict and human infrastructure associated

mortality, conservation genomics insights provide long term conceptual guidance for

future survival of these populations.

Methods

Sample collection: State Forest Department databases of captive elephants were

examined from 6 states, to locate individual elephants that were captured or rescued

from all of the known four major disjunct populations across India. Within these

populations, individuals representing habitats across major barriers such as Palghat

and the Shencottah Gaps in Southern India and the Ganges and the Brahmaputra

rivers in northern and Northeast India were also located. Blood samples and exact

capture locations were collected from 28 such individuals and included in our final

dataset (see Supplementary Table 1). Out of those samples, 12 were collected from

the Southern eco-region, four were collected from the Central eco-region, three were

collected from the Northern eco-region, and lastly nine were collected from the

Northeastern eco-region. Out of the 12 samples collected from the Southern eco-

region, six were collected from North of Palghat Gap, four from South of Palghat Gap

(but north of Shecottah Gap), and two from South of Shencottah Gap. We

additionally included genomic data from five more individuals obtained from online

sources — one individual each from Borneo, Myanmar and Northeastern India, and

three individuals from Southern India (two North of Palghat Gap and one South of
Palghat Gap) (See Supplementary Table 1).

DNA extraction, library preparation and sequencing

450   Genomic DNA was extracted from blood samples using Qiagen DNeasy Blood &
451   Tissue Kit. The library preparation and whole genome resequencing was carried out
452   at Medgenome INC. The DNA extraction and sequencing were carried out following
453   Khan et al., 2020.

Variant calling and filtering

The raw sequencing reads were trimmed using the default settings in TrimGalore-
0.4.5. The trimmed reads were mapped to the *Elephas maximus* reference genome
(https://dnazoo.s3.wasabisys.com/index.html?prefix=Elephas_maximus/) using
default settings of BWA *mem* (https://github.com/lh3/bwa). The mapped reads were
converted to binary format and sorted using Samtools-1.9 (Li et al. 2009). The PCR
and optical duplicates were marked using Picardtools MarkDuplicates
(http://broadinstitute.github.io/picard) or Samtools-markdup. SNPs were identified
using Strelka variant caller 2.9.10 (Saunders et al. 2012). The variants were filtered
using VcftoolsV13 (Danecek et al. 2011). We removed indels and retained only SNP
loci with minimum phred scaled base quality 30, minimum genotype quality 30, a
minimum minor allele count of 3, did not deviate from Hardy-Weinberg equilibrium
with chi square p value of 0.05. All sites with the FILTER flag other than PASS were
removed. Sites that showed mean depth across individuals more than the 97.5th
percentile and less than 2.5th percentile or were missing in more than 20%
individuals , were removed. We identified the sex chromosome scaffold as described
in Armstrong et al. (2021) and removed the sex chromosome scaffolds.

Population genetic structure:

PCA
We first employed PCA to partition the data along their main axes of variation. The
PCA was carried out using the *--pca* function in Plink (Purcell et al. 2007) based on
the final filtered variants. The resulting eigenvectors were plotted in R using ggplot2.

STRUCTURE

Thereafter, a maximum likelihood-based method, ADMIXTURE (Alexander et al.

2009) was employed to investigate the population genomic structure of Asian

elephants. We estimated the number of clusters across *K* values ranging from 1 to 6,

based on the number of clusters obtained from PCA. The evaluation of the most

optimal number of clusters was carried out using EvalAdmix (Garcia-Erill et al. 2020).

EvalAdmix estimates the correlation of the residual matrices of the individuals from

the Admixture analyses. The resulting correlation matrices were plotted in R.

qpgraph

We assigned a population to each individual based on the results from admixture.

We converted the .vcf file to .ped format using VCFtools and then .ped to .bed format

using Plink. Employed the *find_graphs* function in ADMIXTOOLS2

(https://github.com/uqrmaie1/admixtools) to automatically find the optimum graphs

using various values for the number of admixture events. The model with zero

admixture had the best statistical support as tested using the protocol suggested in

ADMIXTOOLS2 with functions *qpgraph_resample_multi* and *compare_fits*.

Demographic history

Recent Demographic History

We used the LD based GONE (Santiago et al. 2020) for estimating recent

demography history up to a couple of hundred generations ago. We used the default

settings of the parameters and set *PHASE* as 0. We plotted the results assuming a

generation time of 31 years. This is a multiple sample analysis and requires several

samples for accurate estimates. Hence, we performed this analysis only for the

NE&N and the NPG cluster which have at least nine samples. The results from this

analysis better predict recent events but are not very reliable for estimating events

older than a hundred generations (Santiago et al. 2020).

Old demographic History

We used the coalescent based SMC++ (Terhorst et al. 2017) method for estimating

older demographic history. We used the protocol described for the implementation of

the method (https://github.com/popgenmethods/smcpp) as is. We set the mutation

rate to 5.3*10^-9/base/generation (Prado et al. 2023) and a generation time of 31

years. The estimates from this method are reliable for estimating demographic

events between a hundred generations ago to forty thousand generations ago.

Divergence time

We estimate divergence times between populations using hPSMC (). We used a

single individual selected at random from each population (Supplementary table 1).

We created a pseudo-diploid individual using samples from two populations for each

scaffold except for the sex chromosome scaffold (https://github.com/jacahill/hPSMC).

Then performed PSMC with the default settings. We plotted the results assuming a

mutation rate of 5.3*10^-9/base/generation (Prado et al. 2023) and a generation time

of 31 years. The time point where the effective population size estimated from the

pseudo-diploid individual rises exponentially is the point where the two haplotypes do

not coalesce and hence signify population divergence.

Genetic diversity

Nucleotide diversity: pi

We randomly subsampled four individuals from N&NE, SWB, NPG and SPG sample

set and retained the SSG dataset (n=2) as is. We estimated the average number of

pairwise differences per site for each population cluster as described in Wang et al.

(2020). Briefly, we set the function *dosaf* to 1 ANGSD (Korneliussen et al. 2014) to

estimate allele frequency likelihood for each site. The we used the used the function

realSFS and estimated a folded site frequency spectrum. We set the function

doThetas to 1 and estimated pi per site and used the *ThetaStat* function the

summarise the pi value for each scaffold. We used only the chromosomal scaffolds

from this summary and divided the "tP" values by the number of sites for each

scaffold to obtain the average pi per site. Furthermore, we statistically compared the

significance of the estimated values between populations (see Supplement).

Heterozygous SNV encounter rate

533 We estimated the number of heterozygous SNPs for each individual using the

534 *vcfstats* function of RTGtools (https://github.com/RealTimeGenomics/rtg-tools). We

535 then divided the number of heterozygous sites by the total number of sites

536 genotyped for the individual and multiplied by 10^6 to obtain SNV encounter rate/Mb.

Inbreeding

FROH

We estimated runs of homozygosity (ROH) using the *roh* function in BCFtools

version 1.3 (Narasimhan et al. 2016) as described previously (Shukla et al. 2023).

We classified the ROH stretches into three size classes: the more than 0.1 Mb (100

Kb) stretches show cumulative inbreeding due to old and recent bottlenecks, the

more than 1Mb stretches show cumulative inbreeding in the recent past and the

more than 10Mb stretches show recent inbreeding. FROH was estimated as

described previously (Khan et al. 2021).

Shared IBD stretches

547 We used IBDseq (Browning & Browning, 2013) to detect the stretches of genome

548 that are identical-by-descent in pairs of individuals. For each pair of individuals in a

549 population, we summed the lengths of IBD stretches longer than 0.1Mb and 10Mb

550 and divided by the total autosomal length.

Genetic load

Filtering variant sites

We filtered the raw variants obtained from Strelka variant caller 2.9.10 again in

VCFtools using the parameters with minimum phred scaled base quality 30,

minimum genotype quality 30, minor allele count 1, FILTER flag as PASS. We also

removed sited with -0.5>Fis>0.95 as described previously (Khan et al. 2021) and

removed sites missing in more than 20% of the individuals. We filtered out the sex

chromosomes as well. We call this set of loci as deleterious_allele_set.

Identifying ancestral alleles

We defined the ancestral allele as the allele present most commonly in sister species

of Asiatic elephants. For this we used the genome of African elephant (genbank

accession: GCA_000001905.1) and dugong (genbank accession: GCA_015147995).

We converted these assemblies to 100bp long FASTQ reads and mapped to the

Asiatic elephant reference genome and removed reads mapping to multiple regions

as described in Khan et al. (2021). We converted the mapped reads to consensus

fasta files and estimated the majority allele at a locus. We further filtered these

alleles with the genome of Bornean elephant (SRA accession: ERR2260499). Any

loci where the Bornean elephant was not homozygous for the ancestral allele was

removed from the analysis thus ensuring that the derived alleles are new to the

population which is an important assumption for the identification of deleterious

alleles (Khan 2023).

Derived indel load

We further filtered the loci in the deleterious_allele_set using the *--keep-only-indels*

tag in VCFtools. We annotate this set of loci with the Asiatic elephant genome

annotation (https://dnazoo.s3.wasabisys.com/index.html?prefix=Elephas_maximus/)

using Ensembl Variant Effect Predictor (VEP, McLaren et al. 2016). For indels, the

ancestral state cannot be determined using distant species. Hence, the indel allele

that was homozygous in the Bornean elephant was referred to as the ancestral allele

and any indel site where the Bornean elephant was not homozygous, was removed

from the analysis. We then estimated the number of indels predicted to cause

transcript_ablation, splice_donor_variant, splice_acceptor_variant, stop_gained,

frameshift_variant, inframe_insertion, inframe_deletion, splice_region_variant were

classified as lof causing indels. We also counted the derived indels in the intergenic

regions of the genome. We divided the number of lof indels with the number of

intergenic indels to control for differences in depth leading to missingness in the

data.

Derived lof load

We filtered the loci in the deleterious_allele_set to remove indels and that had mean

depth across individuals more than the 97.5th percentile and less than 2.5th

percentile. This set was annotated on VEP and lof mutations were identified as

described earlier. We then counted the number of derived lof alleles and divided
them with the number of synonymous mutations.

Derived missense load

We used the same set of loci used for estimating the lof load but chose those set of
loci that caused non-synonymous changes in the genome and followed the same
procedure as described for the lof load.

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614
615 **Data availability**

616
617 All sequencing data have been deposited in BioProject SUB13810487 in NCBI.
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References

Alexander, D.H., Novembre, J. and Lange, K., 2009. Fast model-based estimation of ancestry in unrelated individuals. Genome research, 19(9), pp.1655-1664.

Baskaran, N., Balasubramanian, M., Swaminathan, S. and Desai, A.A., 1995. Home range of elephants in the Nilgiri Biosphere Reserve, south India. A week with elephants. Bombay Natural History Society and Oxford University Press, Bombay, pp.296-313.

Berzaghi, F., Bretagnolle, F., Durand-Bessart, C. and Blake, S., 2023. Megaherbivores modify forest structure and increase carbon stocks through multiple pathways. *Proceedings of the National Academy of Sciences*, *120*(5), p.e2201832120.

Browning, B.L. and Browning, S.R., 2013. Detecting identity by descent and estimating genotype error rates in sequence data. The American Journal of Human Genetics, 93(5), pp.840-851.

Brüniche-Olsen, A., Kellner, K.F., Anderson, C.J. and DeWoody, J.A., 2018. Runs of homozygosity have utility in mammalian conservation and evolutionary studies. *Conservation Genetics*, *19*(6), pp.1295-1307.

Cahill, J.A., Soares, A.E., Green, R.E. and Shapiro, B., 2016. Inferring species divergence times using pairwise sequential Markovian coalescent modelling and low-coverage genomic data. Philosophical Transactions of the Royal Society B: Biological Sciences, 371(1699), p.20150138.

Chaitanya, R., Giri, V.B., Deepak, V., Datta-Roy, A., Murthy, B.H.C.K. and Karanth, P., 2019. Diversification in the mountains: a generic reappraisal of the Western Ghats endemic gecko genus Dravidogecko Smith, 1933 (Squamata: Gekkonidae) with descriptions of six new species. Zootaxa, 4688(1), pp.1-56.

Coverdale, T.C., Kartzinel, T.R., Grabowski, K.L., Shriver, R.K., Hassan, A.A., Goheen, J.R., Palmer, T.M. and Pringle, R.M., 2016. Elephants in the understory: opposing direct and indirect effects of consumption and ecosystem engineering by megaherbivores. Ecology, 97(11), pp.3219-3230.

Curik, I., Ferenčaković, M. and Sölkner, J., 2014. Inbreeding and runs of homozygosity: A possible solution to an old problem. Livestock Science, 166, pp.26-34.

658 Danecek, P., Auton, A., Abecasis, G., Albers, C.A., Banks, E., DePristo, M.A.,
659 Handsaker, R.E., Lunter, G., Marth, G.T., Sherry, S.T. and McVean, G., 2011. The
660 variant call format and VCFtools. Bioinformatics, 27(15), pp.2156-2158.
661
662
663 De, R., Sharma, R., Davidar, P., Arumugam, N., Sedhupathy, A., Puyravaud, J.P.,
664 Selvan, K.M., Rahim, P.A., Udayraj, S., Parida, J. and Digal, D.K., 2021. Pan-India
665 population genetics signifies the importance of habitat connectivity for wild Asian
666 elephant conservation. Global Ecology and Conservation, 32, p.e01888.
667
668 Enquist, B.J., Abraham, A.J., Harfoot, M.B., Malhi, Y. and Doughty, C.E., 2020. The
669 megabiota are disproportionately important for biosphere functioning. *Nature
670 Communications*, *11*(1), p.699.

671 Frankel, O.H. and Soulé, M.E. 1981. Conservation and evolution. Cambridge

672 University Press, Cambridge, U.K.

673
Harich, F.K., Treydte, A.C., Ogutu, J.O., Roberts, J.E., Savini, C., Bauer, J.M. and
Savini, T., 2016. Seed dispersal potential of Asian elephants. Acta Oecologica, 77,
pp.144-151.

Gubbi, S., Swaminath, M.H., Poornesha, H.C., Bhat, R. and Raghunath, R., 2014.

An elephantine challenge: human–elephant conflict distribution in the largest Asian

elephant population, southern India. Biodiversity and conservation, 23, pp.633-647.

Hedrick, P.W., 1999. Perspective: highly variable loci and their interpretation in
evolution and conservation. Evolution, 53(2), pp.313-318.
682
Hellenthal, G., Auton, A. and Falush, D., 2008. Inferring human colonization history
using a copying model. PLoS genetics, 4(5), p.e1000078.

Khan, A., Krishna, S.M., Ramakrishnan, U. and Das, R., 2022. Recapitulating whole

genome based population genetic structure for Indian wild tigers through an ancestry

informative marker panel. Heredity, 128(2), pp.88-96.

Khan, A., Patel, K., Shukla, H., Viswanathan, A., van der Valk, T., Borthakur, U.,

Nigam, P., Zachariah, A., Jhala, Y.V., Kardos, M. and Ramakrishnan, U., 2021.

Genomic evidence for inbreeding depression and purging of deleterious genetic

variation in Indian tigers. *Proceedings of the National Academy of Sciences*, *118*(49),

p.e2023018118.

Khan, A., Patel, K., Bhattacharjee, S., Sharma, S., Chugani, A.N., Sivaraman, K.,
Hosawad, V., Sahu, Y.K., Reddy, G.V. and Ramakrishnan, U., 2020. Are shed hair
genomes the most effective noninvasive resource for estimating relationships in the
wild?. *Ecology and Evolution*, *10*(11), pp.4583-4594.

Khan, A., 2023. The year of the tiger and the year of tiger genomes!. *Molecular*
Ecology Resources, *23*(2), pp.327-329.

Koirala, R.K., Ji, W., Aryal, A., Rothman, J. and Raubenheimer, D., 2016. Dispersal
and ranging patterns of the Asian Elephant (Elephas maximus) in relation to their
interactions with humans in Nepal. Ethology Ecology & Evolution, 28(2), pp.221-231.
702
703 Korneliussen, T.S., Albrechtsen, A. and Nielsen, R., 2014. ANGSD: analysis of next
704 generation sequencing data. *BMC bioinformatics*, *15*(1), pp.1-13.
705
706 Kyriazis, C.C., Beichman, A.C., Brzeski, K.E., Hoy, S.R., Peterson, R.O., Vucetich,
707 J.A., Vucetich, L.M., Lohmueller, K.E. and Wayne, R.K., 2023. Genomic
708 underpinnings of population persistence in Isle Royale moose. Molecular biology and
709 evolution, 40(2), p.msad021.
710
711 Le Roux, E., Kerley, G.I. and Cromsigt, J.P., 2018. Megaherbivores modify trophic
712 cascades triggered by fear of predation in an African savanna ecosystem. *Current*
713 *Biology*, *28*(15), pp.2493-2499.
714
715 Li, H., Handsaker, B., Wysoker, A., Fennell, T., Ruan, J., Homer, N., Marth, G.,
716 Abecasis, G., Durbin, R. and 1000 Genome Project Data Processing Subgroup,
717 2009. The sequence alignment/map format and SAMtools. bioinformatics, 25(16),
718 pp.2078-2079.
719
Liu, P., Wen, H., Harich, F.K., He, C., Wang, L., Guo, X., Zhao, J., Luo, A., Yang, H.,
Sun, X. and Yu, Y., 2017. Conflict between conservation and development: cash
forest encroachment in Asian elephant distributions. Scientific Reports, 7(1), p.6404.
Mahmood, T., Vu, T.T., Campos-Arceiz, A., Akrim, F., Andleeb, S., Farooq, M.,
Hamid, A., Munawar, N., Waseem, M., Hussain, A. and Fatima, H., 2021. Historical
and current distribution ranges and loss of mega-herbivores and carnivores of Asia.
PeerJ, *9*, p.e10738.
727

Mazet, O., Rodríguez, W., Grusea, S., Boitard, S. and Chikhi, L., 2016. On the
importance of being structured: instantaneous coalescence rates and human
evolution—lessons for ancestral population size inference?. Heredity, 116(4),
pp.362-371.

McLaren, W., Gil, L., Hunt, S.E., Riat, H.S., Ritchie, G.R., Thormann, A., Flicek, P.
and Cunningham, F., 2016. The ensembl variant effect predictor. Genome biology,
17(1), pp.1-14.

Menon, V. and Tiwari, S.K., 2019. Population status of Asian elephants Elephas
maximus and key threats. *International Zoo Yearbook*, *53*(1), pp.17-30.
737
738
Menotti-Raymond, M. and O'Brien, S.J., 1993. Dating the genetic bottleneck of the
African cheetah. *Proceedings of the National Academy of Sciences*, *90*(8), pp.3172-
3176.

Narasimhan, V., Danecek, P., Scally, A., Xue, Y., Tyler-Smith, C. and Durbin, R.,
2016. BCFtools/RoH: a hidden Markov model approach for detecting autozygosity
from next-generation sequencing data. Bioinformatics, 32(11), pp.1749-1751.

Owen-Smith, N., 1987. Pleistocene extinctions: the pivotal role of megaherbivores.
Paleobiology, 13(3), pp.351-362.
747
748 Padalia, H., Ghosh, S., Reddy, C.S., Nandy, S., Singh, S. and Kumar, A.S., 2019.
749 Assessment of historical forest cover loss and fragmentation in Asian elephant
750 ranges in India. Environmental Monitoring and Assessment, 191, pp.1-13.
751
752 Palkopoulou, E., Lipson, M., Mallick, S., Nielsen, S., Rohland, N., Baleka, S.,
753 Karpinski, E., Ivancevic, A.M., To, T.H., Kortschak, R.D. and Raison, J.M., 2018. A
754 comprehensive genomic history of extinct and living elephants. Proceedings of the
755 National Academy of Sciences, 115(11), pp.E2566-E2574.
756
757 Parida, J., Sharma, R., De, R., Kalam, T., Sedhupathy, A., Digal, D.K., Reddy, P.A.,
758 Goyal, S.P., Puyravaud, J.P. and Davidar, P., 2022. Genetic characterisation of
759 fragmented Asian elephant populations with one recent extinction in its eastern-
760 central Indian range. Ecological Genetics and Genomics, 24, p.100132.
761
762 Pierce, A.A., Zalucki, M.P., Bangura, M., Udawatta, M., Kronforst, M.R., Altizer, S.,
763 Haeger, J.F. and de Roode, J.C., 2014. Serial founder effects and genetic
764 differentiation during worldwide range expansion of monarch butterflies. Proceedings
765 of the Royal Society B: Biological Sciences, 281(1797), p.20142230.
766
767 Pless, E., Powell, J.R., Seger, K.R., Ellis, B. and Gloria-Soria, A., 2022. Evidence for
768 serial founder events during the colonization of North America by the yellow fever
769 mosquito, Aedes aegypti. Ecology and Evolution, 12(5), p.e8896.
770

772 Prado, N.A., Armstrong, E.E., Brown, J.L., Goldenberg, S.Z., Leimgruber, P.,
773 Pearson, V.R., Maldonado, J.E. and Campana, M.G., 2023. Genomic Resources for
774 Asian (Elephas maximus) and African Savannah Elephant (Loxodonta africana)
775 Conservation and Health Research. *bioRxiv*, pp.2023-02.
776

Purcell, S., Neale, B., Todd-Brown, K., Thomas, L., Ferreira, M.A., Bender, D.,

Maller, J., Sklar, P., De Bakker, P.I., Daly, M.J. and Sham, P.C., 2007. PLINK: a tool

set for whole-genome association and population-based linkage analyses. The

American journal of human genetics, 81(3), pp.559-575.

Rajagopalan, G., Sukumar, R., Ramesh, R., Pant, R.K. and Rajagopalan, G., 1997.
Late Quaternary vegetational and climatic changes from tropical peats in southern
India–an extended record up to 40,000 years BP. *Current Science*, pp.60-63.

Ram, M.S., Marne, M., Gaur, A., Kumara, H.N., Singh, M., Kumar, A. and Umapathy,

G., 2015. Pre-historic and recent vicariance events shape genetic structure and

diversity in endangered lion-tailed macaque in the Western Ghats: implications for

conservation. PloS one, 10(11), p.e0142597.

Robin, V.V., Vishnudas, C.K., Gupta, P. and Ramakrishnan, U., 2015. Deep and

wide valleys drive nested phylogeographic patterns across a montane bird

community. Proceedings of the Royal Society B: Biological Sciences, 282(1810),

p.20150861.

Robinson, J., Kyriazis, C.C., Yuan, S.C. and Lohmueller, K.E., 2023. Deleterious

variation in natural populations and implications for conservation genetics. Annual

review of animal biosciences, 11, pp.93-114.

Santiago, E., Novo, I., Pardiñas, A.F., Saura, M., Wang, J. and Caballero, A., 2020.

Recent demographic history inferred by high-resolution analysis of linkage

disequilibrium. Molecular Biology and Evolution, 37(12), pp.3642-3653.

Saunders, C.T., Wong, W.S., Swamy, S., Becq, J., Murray, L.J. and Cheetham, R.K.,

2012. Strelka: accurate somatic small-variant calling from sequenced tumor–normal

sample pairs. Bioinformatics, 28(14), pp.1811-1817.

Srinivasaiah NM, Vaidyanathan S, Sukumar R, Sinha A. (2021). Elephants on the
move: Implications for human–elephant interactions. In: *The Crisis of Climate*
Change: Weather Report. (eds. Agarwal R and Goyal O). Routledge, Abingdon,
United Kingdom and New York, pp. 92-103

Srinivasaiah NM, Vaidyanathan S, Sukumar R, Sinha A. (2022). The rurban
elephant: Behavioural ecology of Asian elephants in response to large-scale land
use change in a human-dominated landscape in peri-urban southern India. In: *New*
Forms of Urban Agriculture: An Urban Ecology Perspective (eds. Diehl J A and Kaur
H). Springer Nature, Singapore, pp. 289-31

Sukumar, R. 1989. The Asian elephant: ecology and management. Cambridge
University Press, Cambridge, U.K.
813
814   Sukumar, R., Ramakrishnan, U. and Santosh, J.A., 1998, November. Impact of
815   poaching on an Asian elephant population in Periyar, southern India: a model of
816   demography and tusk harvest. In *Animal Conservation forum* (Vol. 1, No. 4, pp. 281-
817   291). Cambridge University Press.
818
819   Sukumar, R., Ramesh, R., Pant, R.K. and Rajagopalan, G., 1993. A $\delta13C$ record of
820   late Quaternary climate change from tropical peats in southern
821   India. *Nature*, *364*(6439), pp.703-706.
822
823   Sukumar, R. 2003. The Living Elephants: Evolutionary Ecology, Behavior and
824   Conservation. Oxford University Press, New York
825
826   Sukumar, R., 2006. A brief review of the status, distribution and biology of wild Asian
827   elephants Elephas maximus. International Zoo Yearbook, 40(1), pp.1-8.
828
829   Sukumar, R., 2011. *The story of Asia's elephants*. Mumbai: MARG foundation.
830
831   Shukla, H., Suryamohan, K., Khan, A., Mohan, K., Perumal, R.C., Mathew, O.K.,
832   Menon, R., Dixon, M.D., Muraleedharan, M., Kuriakose, B. and Michael, S., 2023.
833   Near-chromosomal de novo assembly of Bengal tiger genome reveals genetic
834   hallmarks of apex predation. GigaScience, 12, p.giac112.
835
836   Sumreddee, P., Hay, E.H., Toghiani, S., Roberts, A., Aggrey, S.E. and Rekaya, R.,
837   2021. Grid search approach to discriminate between old and recent inbreeding using
838   phenotypic, pedigree and genomic information. BMC genomics, 22, pp.1-17
839

840 Terhorst, J., Kamm, J.A. and Song, Y.S., 2017. Robust and scalable inference of
841 population history from hundreds of unphased whole genomes. Nature genetics,
842 49(2), pp.303-309.
843
844 Venkataraman, A.B., Saandeep, R., Baskaran, N., Roy, M., Madhivanan, A. and
845 Sukumar, R., 2005. Using satellite telemetry to mitigate elephant–human conflict: an
846 experiment in northern West Bengal, India. Current Science, pp.1827-1831.
847
848 Vidya, T.N.C., Fernando, P., Melnick, D.J. and Sukumar, R. 2005a. Population
849 differentiation within and among Asian elephant (Elephas maximus) populations in
850 southern India. Heredity, 94(1), pp.71-80.
851
Vidya, T.N.C., Fernando, P., Melnick, D.J. and Sukumar, R., 2005b. Population
genetic structure and conservation of Asian elephant (*Elephas maximus*) across
India. Animal Conservation, 8: 377-388.

Vidya, T.N.C., Sukumar, R. and Melnick, D.J., 2009. Range-wide mtDNA
phylogeography yields insights into the origins of Asian elephants. Proceedings of
the Royal Society B: Biological Sciences, 276(1658), pp.893-902.

Vijayakumar, S.P., Menezes, R.C., Jayarajan, A. and Shanker, K., 2016. Glaciations,
gradients, and geography: multiple drivers of diversification of bush frogs in the
Western Ghats Escarpment. Proceedings of the Royal Society B: Biological
Sciences, 283(1836), p.20161011.

Whitehouse, A.M. and Hall-Martin, A.J., 2000. Elephants in Addo Elephant National
Park, South Africa: reconstruction of the population's history. Oryx, 34(1), pp.46-55.

Waldram, M.S., Bond, W.J. and Stock, W.D., 2008. Ecological engineering by a
mega-grazer: white rhino impacts on a South African savanna. Ecosystems, 11,
pp.101-112.

Wang, X., Bernhardsson, C. and Ingvarsson, P.K., 2020. Demography and natural
selection have shaped genetic variation in the widely distributed conifer Norway
spruce (Picea abies). *Genome biology and evolution*, *12*(2), pp.3803-3817.

871