Title: Thermal suppression of gametogenesis can explain historical collapses in larval

recruitment

Authors: Daniel K. Okamoto[1,2*], Nathan B. Spindel[2], Maya J. Mustermann[1,2], Sam Karelitz[1,2],

Brenna Collicutt[3], Iria Gimenez[3], Kate Rolheiser[3], Evan Cronmiller[3], Megan Foss[3], Natalie.

Mahara[3], Dan Swezey[4,5], Rachele Ferraro[2], Laura Rogers-Bennett[4,6] , and Stephen Schroeter[7]

Affiliations

1: Department of Integrative Biology, University of California, Berkeley, Berkeley California,

USA 94702

2: Department of Biological Science, The Florida State University, Tallahassee, Florida 32303

3: Hakai Institute, Quadra Island, British Columbia, Canada

4: Coastal Marine Science Institute and Bodega Marine Lab, University of California, Davis

California, USA 94923

5: Kashia Band of Pomo Indians of Stewarts Point Rancheria, Santa Rosa, CA

6: California Department of Fish and Wildlife

7: Marine Science Institute, University of California, Santa Barbara

*: Corresponding Author: dkokamoto@berkeley.edu

Abstract

Projections for population viability under climate change are often made using estimates

of thermal lethal thresholds. These estimates vary across life history stages and can be valuable

for explaining or forecasting shifts in population viability. However, sublethal temperatures can

also depress vital rates and shape fluctuations in the reproductive viability of populations. For

example, heatwaves may suppress reproduction, leading to recruitment failure before lethal

temperatures are reached. Despite a growing awareness of this issue, tying sublethal effects to

observed recruitment failure remains a challenge especially in marine environments. We

experimentally show that sublethal suppression of female gametogenesis by marine heatwaves

can partially explain historical collapses in urchin recruitment. These responses differ by sex but

are similar between animals from warmer or cooler regions of their range. Overall, we show

sublethal thermal sensitivities of reproduction can narrow the thermal envelope for population

viability compared to predictions from lethal limits.

Introduction

Predicting the impact of warming and extreme heatwaves on ecosystems and populations remains one of the most daunting but important tasks in ecology and conservation as climate change impacts intensify. Information about the future dynamics and viability of populations can be gleaned from measuring thermal lethal thresholds (or critical thermal limits - CTLs) in controlled settings, as well as from historical patterns of population performance in nature. For the latter, extrapolating historical patterns without understanding processes that shaped them remains dubious; meanwhile, expectations from CTLs often underpredict the abruptness and magnitude of population responses to warming in nature [1-3]. For example, warming and heatwaves may lead to recruitment failure in populations even when temperatures remain below lethal thresholds [4,5]. Thus, the realized thermal niche of many species (i.e., thermal regimes in which a population exhibits viability over time) is often narrower than expected based solely on CTLs for individual life stages (gametes, juveniles, adults, etc. [6]). Combining inference from a mixture of laboratory studies, historical patterns, and analysis of reproductive rates can provide a more comprehensive understanding of the impacts that climate events such as heatwaves have on recruitment and population dynamics.

Sublethal challenges of climate change and warming are impacting recruitment and subsequent population viability[7]. One pathway by which sublethal temperatures can alter population viability is by affecting gametogenesis, sexual maturation, and/or sterility. Changes in temperature can directly affect reproductive phenology for a diversity of plants and animals[8] which can significantly impact reproductive success, fitness, and productivity in a warming world. In plants, for example, warmer winters can lead to crop failure by disrupting vernalization pathways. This family of processes, in which plants require specific thermal cues, as well as other environmental patterns or thresholds for induction of flowering and seed/fruit development has long been studied and understood in agricultural and laboratory plant systems (e.g., [9,10]). Likewise, a diversity of animal taxa can exhibit infertility or low fecundity at sublethal temperatures [11-15].

For animals, and marine organisms with planktonic larvae in particular, it remains unclear if and to what extent patterns of recruitment failure in nature are shaped by sublethal

thermal suppression of reproduction. This knowledge gap remains for several reasons. First,

attributing causality to observed changes in year-class strength remains a challenge for animals

with highly dispersive reproductive propagules. Shifts in recruitment may be explained by

variation in food supply [16], trends in reproductive and mating success [17], shifts in dispersal [18], or

a host of other factors that may also be directly or indirectly related to warming. Second, thermal

limits are often estimated using constant temperatures; yet in the oceans, temperature is highly

variable, and the dynamics of this variation can shape patterns of reproduction. Thus, while it is

well known that sublethal thermal stress can limit reproduction in marine invertebrates [19,20], it

remains less well understood if heatwaves can induce such limitations and/or plausibly help

explain collapses in recruitment. If climate change and extreme events in nature lead to

suppressed gametogenesis well before lethal limits of organisms are reached (e.g., thermal stress

that decreases larval, juvenile, or adult survival), then declines in population viability and

contractions in distributions may occur and at less extreme temperatures than established under

critical thermal limits.

For populations of purple sea urchins (*Strongylocentrotus purpuratus* - an ecosystem

engineer) in Southern California, recruitment tends to collapse during warm El Niño events [21]

(Figure 1A). This phenomenon is remarkably consistent over time (extending back at least six

decades), even though temperatures during the reproductive season in Southern California (late

fall and early winter [21,22] Figure 1B) are almost always below those considered to be lethal for

adults (~25 °C [23], < 0.01% of daily observations since 1990 above this value). Moreover, winter

temperatures in the Southern California Bight when larvae are present (~ 11-16°C) are generally

well below the thresholds thought to impede multiple phases of larval survival [24] (fertilization

and early-phase survival of San Diego, Santa Barbara, and British Columbia animals declines to

near zero between 20-22°C - SK, unpublished data). Specifically, purple urchins tend to spawn

in the winter and spring, coinciding with favorable conditions (low temperatures and high food

availability) for larvae [21]. Larvae persist in the plankton from approximately 28 days to several

months depending on food availability and temperature [25,26]. Adult urchins build gonads over the

course of the summer and fall with gametogenesis occurring in the fall and early winter. El Niño

conditions in Southern California are characterized by warmer water temperatures in summer,

fall, and winter than during neutral years or during La Niña conditions, but still exhibit seasonal

cooling (Figure 1B). One proposed explanation for decreases in recruitment in Southern

California during El Niño is that temperatures experienced during these events suppress

gametogenesis even though energetic investment in gonads may be sufficient for reproduction [27-

29]. Specifically, Pearse [28] demonstrated that high temperatures decrease gamete production at

sublethal levels (i.e., at 21°C vs 14°C), but the higher temperature exceeded those experienced in

the spawning season. Moreover, animals used by Pearse were sourced from cooler regions of the

range that rarely experience such temperatures, while animals from southern regions are

frequently exposed to temperatures that exceed 21°C.

In this study, we used *S. purpuratus* to test whether historical collapses in larval supply

and year-class strength (i.e., abundance for a given young-of-year cohort) can, in part, be

explained by sublethal thermal suppression of reproduction. Specifically, we used two laboratory

experiments to ask whether marine heatwaves that have led to collapses in larval supply also

affect reproduction in terms of both gonad production and gametogenesis. For the first

experiment we compared effects of simulated heatwaves to simulated historical cool trends, as

well as responses under a gradient of constant temperatures. For the second experiment we

exposed animals from warm and cool regions to the same treatments in a common garden setting

to test whether source population (whether via local adaptation or acclimatization) affected

patterns of gametogenesis.

Results

Simulated El Niño and warmer conditions in the first experiment led to decreased gonad

production as well as suppressed gametogenesis in *S. purpuratus.* Declines in gonad production

associated with high temperatures were consistent for constant exposure to 20°C and the El Niño

treatment that shifted from 21 to 18°C with a mean of 20°C (Figure 4). In contrast, the La Niña

treatment (shifting from 18 to 14°C, with a mean of 16°C) led to reduced gonad production

compared to a constant mean temperature of 16°C. The effect of treatment on gametogenesis

differed depending on whether the thermal environment varied over time (El Niño and La Niña

versus constant temperatures) and by sex. With a mean of 16°C, we saw no difference in

gametogenesis between La Niña and constant temperatures for males or females. For females,

constant high temperature (20°C) and the El Niño treatment (21-18°C with a mean of 20°C) led

to a substantial reduction in oogenesis. For males, only the El Niño treatment affected

spermatogenesis, with no change observed across constant temperatures from 10 to 20°C. For the

second experiment, we found no differences in gametogenesis among source populations in

response to treatments, with qualitatively similar treatment effects as in the first experiment for

both males and females. Below, we describe the specific results of each experiment in detail.

Experiment 1: Effects of El Niño, La Niña, and a thermal gradient on gonad production and

gametogenesis

Gonad production

In the first experiment, exposure to the highest temperatures (for both variable and

constant treatments with mean of 20°C) negatively affected both male and female gonad

production (cumulative gonad mass, Figure 4). Specifically, we observed a 42% decrease (95%

highest probability density interval [HPD]: 28-55%) in gonad production at 20°C compared to a

constant temperature of 13°C (where peak gonad production was observed). Gonad production

did not differ for the El Niño (7.03g [95% HPD: 6.20g to 7.96g]) and constant treatment that

shared the same mean value of 20°C (7.2g, [95% HPD: 5.97g to 8.50g]). In contrast, the La Niña

treatment (declining from 18 to 14°C with mean of 16°C) showed a reduction in gonad mass

(26% - HPD: 15 to 38%) compared with a constant value of 16°C (8.69g [95% HPD: 7.68g to

9.77g] vs 11.84g [95% HPD 10.71g to 13.15g], respectively). The lowest temperature (10°C)

imposed a smaller, but still significant reduction in gonad mass (22% reduction, 95% HPD: 4-

39%) compared to the peak at 13°C. The reduction in gonad production from the La Niña to El

Niño (with means of 16°C and 20°C, respectively) was smaller than when comparing constant

temperature treatments of 16°C to 20°C (17% [HPD: 3.5 to 32%] vs 38% [HPD: 23 to 39%]

respectively) despite having the same mean temperatures. This reduced effect resulted from the

aforementioned decline in gonad size in La Niña versus equivalent but constant mean of 16°C.

Notably, we saw no difference in gonad production by sex. The final model used for analysis

(temperature treatment but no sex or sex x treatment interaction) is supported by Bayes Factors

>100 for temperature treatment relative to both the null and to models with sex and interactions.

Bayes Factors > 100 are considered decisive evidence of one model over a second, values 10-100

are considered strong evidence [30], while values <1 favor the second model.

Gametogenesis

The El Niño treatment led to a significant reduction in gametogenesis for both males and

females, and this effect was more extreme for females compared to males. We scored

gametogenesis using a standard staging scale of I to IV where IV is fully mature and I is

immature (sensu [31]). Specifically, far fewer individuals in the El Niño treatment reached Stage

IV (most mature) compared to the La Niña treatment for both males and females (Figure 5). Yet

this effect was more extreme for females compared to males. Female urchins incubated in the La

Niña treatment actively produced mature gametes (89% [95% HPD: 77 to 94%] of females in

stage IV, Figure 5A). In contrast, only 15% [95% HPD: 4 to. 28%] of El Niño females were in

stage IV (Figure 5A), while most females were in stage I. This decline in gametogenesis

associated with El Niño amounted to an estimated 74% reduction for females compared to La

Niña (Figure 5A, HPD: 56 to 91% decline in females in stage IV). For male urchins, these results

were consistent but less extreme. 92% of La Niña males were stage IV [HPD: 77 to 99%] vs

44% of El Niño females [HPD: 23 to 67%]– a reduction of 48% [HPD: 23 to 72%] - Figure 5B).

Overall, models that included the treatment, sex, and interaction compared to the models with

only sex were supported Bayes Factors >100.

Our results from the first experiment also indicate that temperature dynamics, and not just

their mean values, can alter reproductive response depending on sex. Simulating a historical El

Niño trend resulted in a reduction in the proportion of males investing in spermatogenesis in

contrast to no change in spermatogenesis under a constant mean temperature. Specifically, male

animals subjected to the El Niño treatment (simulated seasonal peak and decrease from 21 to

18°C with a mean of 20°C) compared to a constant 20°C throughout showed substantially lower

rates of full gonad maturation (stage IV) with a reduction from 81.2 to 47.8% (a difference of

37% [95% HPD: 8% to 65%], Figure 5B, D). In contrast, we observed similarly low rates of full

maturity for females subjected to the El Niño simulation and constant 20°C treatment (15% for

El Niño vs 25% for constant 20°C, Bayesian P = 0.28, (Figure 5A, C).

Experiment 2: Effects of source population on effects of El Niño on gametogenesis

When comparing populations, we found a similar strong effect of temperature treatment

on gametogenesis but did not find a significant difference among populations in these responses

(Figure 6). Specifically, we did not find differences in gametogenesis in response to 20°C or El

Niño conditions (relative to 10°C) among animals sourced from different populations, including

those from warm regions (San Diego County and Santa Barbara County) or a cold upwelling

region (Sonoma County). Comparing models using Bayes Factors found no support for the

hypothesis that treatment effects varied by source population (Bayes Factor = 19 favoring the

reduced model [treatment, sex, and interaction] compared to the full model [with location,

treatment, sex and interactions]). The lack of location of origin effect is reinforced by analysis of

the full model, where we found no difference among locations in estimated treatment effects for

any stage (i.e., 95% credible intervals overlap zero for differences among populations in the log-

scale treatment effects for either 20°C or El Niño relative to 10°C by sex); however, we did find

consistent effects of each treatment across locations (>95% posterior probability of a difference

in proportion of stage IV animals for both 20°C and El Niño relative to 10°C for females and El

Niño relative to 10°C only for males). Thus, hereafter we describe analysis using the model

without location interactions. Overall, we estimate a combined 75% and 90% reduction in stage

IV females and males (respectively) for the El Niño treatment relative to the 10°C treatment

(95% HPD = 92% to 61% reduction for females and 99% to 65% reduction for males) and an

81% and 96% reduction in stage for IV females and males for the 20°C treatment relative to the

10 degree treatment (95% HPD = 96% to 68% reduction for females and 99 to 72% reduction for

males). The pooled results are shown as the thick lines/points in Figure 6G and H for females

and males, respectively.   While results from the first experiment were conducted with animals

sourced from cooler water, gametogenesis trends are similar when comparing animals from

warmer versus cooler environments.

Discussion

Global warming and extreme heatwaves present a significant threat to ecosystems

worldwide. Characterizing temperature thresholds that cause mortality across life stages can

provide essential insight into the impacts of heatwaves. It is well understood that range limits can

contract and abundances can collapse well before organismal lethal limits are reached.

Mechanistically explaining how and why such sublethal thermal regimes affect populations

remains a challenge. In this study, we show how thermal suppression of reproduction can, in

part, provide a plausible explanation for historical collapses in larval supply associated with

warming events. Our results provide two key general conclusions. First, sublethal impacts of

warming events on reproduction may lead to recruitment failure and population declines before

lethal upper thermal limits are experienced. This insight is derived from the results of our lab

experiments in the context of our historical observations of larval supply. We found these results

to be consistent for animals sourced from both warmer and cooler parts of their range. Second,

variable temperature treatments that reflect historical observations of heatwaves, in comparison

to constant temperatures, exert a more pronounced effect on reproductive suppression. Therefore,

ignoring the dynamics of thermal regimes may underpredict the magnitude of sublethal thermal

effects. Importantly, these results raise questions regarding physiological processes by which

temperature shapes gametogenesis, and how such effects may vary in and among populations

over time. These effects will almost certainly interact with other stressors such as food

availability for adults (sensu [32]) or shifts in physical dispersal, gametogenesis, or heatwave

conditions.

Organismal thermal lethal limits can underpredict observed shifts in population-level

vital rates. Our experimental results provide plausible evidence that sublethal reproductive

failure contributes to collapses in year-class strength. These insights hinge upon a combination of

long-term monitoring, historical observational data collection, and incremental prior

experimentation. Specifically, our historical data collection in Southern California shows larval

supply of purple urchins consistently collapses during marine heatwaves[21]. Prior field

observations proposed that temperatures above a 17°C threshold led to suppressed reproduction

during a marine heatwave (assessed observationally via seasonal gamete extrusion [29] and

histology [27]). Additionally, the prior experiment by Pearse [28] in Monterey Bay, California

showed that urchins exposed to constant 21°C suppressed male and female gametogenesis

compared to constant 14°C. However, 21°C is outside the range of temperatures generally

experienced during the spawning season, even during marine heatwaves in Southern California

(Figure 1A, Figure 7A,B). Our results suggest a higher limit than 17°C but still support the

notion that El Niño can suppress gametogenesis. Specifically, replicating historical marine

heatwaves (shifting from 21°C and ending at 18°C with a mean of 20°C) can suppress

gametogenesis for both males and females sourced from across a wide latitudinal range. For

females, this suppression appears to occur above 18°C but at or below 20°C. Females exposed to

a constant 18°C in the first experiment proceeded with vitellogenesis and produced mature

gametes, whereas females generally did not when under the El Niño treatment (ending at 18°C)

or the constant 20°C treatment. Similarly, males exposed to a constant 20°C proceeded with

spermatogenesis and produced mature gametes but not those in the El Niño treatments. In

contrast to gametogenesis, overall treatment effects on gonad production (mass – a metric of

potential fecundity) did not differ by sex. These disparities in gametogenesis patterns among

sexes raise several hypotheses that require further investigation. For example, sperm and eggs

may differ in their thermal constraints, time required to complete gametogenesis, and/or costs

associated with production and maturation. Despite the aforementioned differences, El Niño

treatments that ended at 18°C yielded consistent results among males and female gametogenesis.

This final temperature is several degrees lower than what is considered to be lethal for adults and

larvae (23-25°C and 20-22°C, respectively). As a result, inferences derived solely from larval or

adult survival under laboratory conditions would not predict collapses in year-class strength in

association with marine heatwaves.

Heatwave events in both marine and terrestrial settings are often highly dynamic,

characterized by rapid fluctuations in temperature. We've shown how experiments that test

constant temperatures and ignore such fluctuations may yield erroneous conclusions. When

comparing variable versus constant temperature treatments, gonad production (and by extension

potential male and female fecundity) differed between La Niña and constant temperatures of the

same mean. These disparities in gonad production likely result from nonlinear averaging (e.g.,

Jensen's inequality) associated with dynamics of feeding, metabolism, and/or structural growth

allocation[33]. With regards to gametogenesis, male but not female patterns of gametogenesis

consistently differed between constant versus El Niño temperature profiles that shared the same

mean value (20°C). Thus, the variability, specific time course, and extremes of environmental

exposure have the potential to alter reproductive responses to heatwaves and can differ by sex.

Importantly, our studies ended in December, when temperatures approach winter lows (Figure 1,

Figure 7). Even in El Niño years, winter temperatures are likely sufficient to facilitate

reproduction over the long term (Figure 1). Thus, rather than inducing skip-spawning, thermal

suppression of gametogenesis in fall and early winter may instead lead to a shortened spawning

season. Because urchin settlement timing aligns with peak phytoplankton production[21], such

effects may lead to other negative impacts on recruitment such as match-mismatch dynamics

with poor survival [34]. Moreover, our variable treatments are limited to a specific set of

trajectories and our constant treatments do not include 21°C or above (the maximum temperature

of the El Niño treatment); thus, future work may benefit from assessing how brief or consistent

incursions into the 21°C range, and the timing of such incursions, affect both male

spermatogenesis and female oogenesis. As a result, future studies may benefit from documenting

gonad production and gametogenesis under different real-world and forecasted conditions in

both controlled settings and among populations across time and space.

Many species exhibit spatial and/or population level variation in response to temperature.

Purple urchins, for example, can display significant differences in environmental sensitivities

among populations (e.g., [35]). Our results suggest that, at least for the populations sampled and

treatments applied, impacts of marine heatwaves are consistent between sea urchins sourced

from cool and warm regions of the range. These results reflect only these specific contrasts and

populations, and thus cannot rule out variation to different contrasts or variation in the response

in other populations. Yet with animals ranging from British Columbia to Southern California, we

show that marine heatwave effects on sea urchin gametogenesis are likely significant in those

regions exposed to warm waters. However, we acknowledge that linking these results to

outcomes in the field definitively, rather than logically, requires documenting the spatial and

temporal patterns of reproductive phenology and gametogenesis in nature and any association

with larval supply. Moreover, the physiological and regulatory processes shaping these patterns

have yet to be systematically characterized. Thus, examining the genetic and plastic basis of how

reproduction and gametogenic cycles may respond to changes in temperature regimes remains an

important objective both in this species and more broadly.

These experiments present plausible evidence for how and why historical collapses in

larval supply occur during marine heatwaves. Yet there are likely myriad factors that occur

during marine heatwaves that may shape recruitment in addition to, or in synergy with

suppression of reproduction. First, food stress occurs during El Niño events in Southern

California [36]. Such trends might exacerbate the effects of temperature on gamete production as

food limitation for adults may alter thermal energetic performance or allocation (e.g., via

"metabolic meltdown" [32]). In these experiments, we focused on well-fed adult animals to avoid

confounding effects of food availability. The results shown here raise the question of whether

declines in availability and productivity of macroalgae can exacerbate responses of

gametogenesis in sea urchins. Second, El Niño events are thought to be coupled with shifts in

ocean circulation that might alter patterns of phytoplankton production (i.e., larval food) as well

as the delivery of planktonic larvae to suitable settlement sites. These trends may also affect

reproduction as urchin spawning may be sensitive to changes in phytoplankton-derived

chlorophyll [37,38]. Thus, even in the cases where offspring are produced from successful

maturation and spawning, shifts in larval survival and delivery may further compound effects of

heatwaves. Finally, temperature [39] and animal density may increase disease prevalence and

disease related mortality [40,41] in nature in addition to the reproductive effects observed here.

Therefore, our results focus on one set of key factors among many that may simultaneously

respond to abiotic and biotic conditions during extreme events known to alter dynamics of

populations.

Marine heatwaves have led to substantial reorganization of ecosystems. This trend has

been particularly apparent in recent years in temperate rocky reefs [42-48] with resulting, negative

effects on ecosystem services and biodiversity [49]. While extreme events can impose direct and

observable effects on mortality, sublethal effects, when present, may be far more frequent and

insidious because they occur below lethal thresholds and may have less immediately visible or

observable impacts. This suggests that marine heatwave survivors maybe impaired and

reproductively dysfunctional despite being alive [19]. This study is among the first to show that

widespread, historical collapses in sea urchin larval supply in the field can at least partially be

explained by sublethal suppression of gametogenesis resulting from marine heatwaves.

Moreover, our study demonstrates the value of long-term population studies and quantifying

non-lethal physiological and reproductive sensitivities when considering sublethal thermal

envelopes or thermal fertility limits [50] and the population-level impacts of global warming and

heatwaves.

Methods

To quantify how different thermal regimes affect investment in gonads and development of gametes in male and female urchins, we first conducted a 10-week experiment in which 300 animals were incubated in replicate 350L mesocosms that simulated El Niño (N = 4 mesocosms, 60 animals per treatment) or La Niña (N = 4 mesocosms, 60 animals) conditions based on historical, empirical benthic temperature time series from Scripps Pier in La Jolla, California[51] (trends: Figure 1B, map: Figure 1C) that coincide with historical collapses in larval supply in Southern California. We paired these treatments with a range of fixed temperature incubations (10, 13, 16, 17, 18, 20 °C, N = 2 mesocosms, 30 animals per treatment), two of which matched the mean temperature of the El Niño (20 °C) and La Niña (16 °C) (Figure 1). We chose this benthic time series rather than satellite derived sea surface temperature information because while the latter provides high quality spatial representations of sea surface temperature but surface temperatures can be markedly different than temperatures experienced at depth by benthic organisms[52]. Experiments were conducted at the Marna Lab at the Hakai Institute's Quadra Island Ecological Observatory in Heriot Bay, British Columbia due to availability of sophisticated seawater systems for careful, replicated temperature manipulations.

Second, to quantify the degree to which using animals from a northern population may have affected estimates of effects compared to animals from warmer regions, we conducted a second 12-week experiment comparing animals from the warmer Southern California bight including San Diego and Santa Barbara with the cool upwelling region of Sonoma County. Seasonal daily temperature data from 2009 (the earliest record at Bodega Marine Lab) for all sites are shown in Figure 7. Seasonal averages from September into December since 2009 are 19 °C, 16.8 °C, and 13.5 °C from Scripps Pier (San Diego), Arroyo Burro (Santa Barbara), and Bodega Marine Lab (Sonoma), respectively. The percent of daily observations above 19°C at these sites since 1990 are 47%, 10%, and <1% at Scripps Pier, Arroyo Burro, and Sonoma, respectively. In this experiment, we exposed animals to the same heatwave treatments (21-18°C), 10°C, and 20°C in a split plot design. See below for experimental details.

For both experiments we used the same general methodologies (food source, feeding regime, containers) but in different settings (Quadra Island Ecological Observatory in experiment 1 and Bodega Marine Lab in experiment 2).

Field Collections and Acclimation

For the first experiment, we collected sea urchins by hand on SCUBA in the vicinity of

Ucluelet, British Columbia, Canada (48.94˚N, 125.56˚ W) from a depth of 7-8 m relative to

mean low tide in September 2021 and transported them to the Marna Lab via truck in seawater

filled coolers with bubblers in less than 24 hours. We transferred sea urchins to flow-through sea

tables and allowed them to recover for a period of one week before placing animals into the

mesocosm system. We selected healthy individuals within a constrained size range for

incubations (n = 300, mean test diameter = 56.09 mm, range test diameter = 42.12 – 69.46 mm).

Finally, we assigned animals to mesocosms at random at ambient temperature and exposed each

assigned mesocosm to a temperature ramp, where the ramp reached target temperatures after two

weeks from the initial incoming, ambient temperature (mean across all tanks of $13.3˚C$, SD =

$0.3˚C$) to avoid thermal shock. Once initial target temperatures were reached, they were

maintained or, for the variable treatments, were manually adjusted daily in the AM (~8am each

day) as needed by $0.5\,^{\circ}C$ increments in a scheduled manner to match historical mean El Niño and

La Niña daily temperature trends shown in Figure 1B.

For the second experiment, we collected animals from three locations in California

including Stillwater Cove in Sonoma County, Mohawk Reef in Santa Barbara, and Point Loma

in San Diego using SCUBA from 3-5 meters mean low water. Urchins were dry transported

layered between kelp and transferred to ambient flow through tanks at Bodega Marine Lab

within 6 hours. Temperature ramps and acclimation periods were manually adjusted in 1ºC

increments.

Mesocosm Systems

For the first experiment, we placed urchins in a custom-built array of twenty replicated

214 L [90(L) x 59.5(W) x 40(H) cm] acrylic mesocosms supplied with flow-through UV

sterilized and filtered seawater (Figure 3:). Each mesocosm was capable of independent control

of temperature and animals were provided a lighting regime for all mesocosms using LED

fixtures (Aquamaxx, CA, USA) programmed to provide 10L:14D with two-hour linear light

intensity transition periods for dawn and dusk (0-100% from 07:00 to 09:00 "dawn", and 100-

0% from 17:00 to 19:00 "dusk"). Each mesocosm independently maintained temperature

treatments using a heat exchanger fitted with a titanium coil regulated by a dual stage digital

temperature controller (Resolution = 0.1℃, Dwyer Instruments, LLC.©, Michigan City, IN,

USA). The mesocosm system employed central cooling (Aermec Mits Airconditioning Inc.,

Mississauga, ON, Canada) and heating (boiler array, Viessmann Manufacturing Company Inc.,

Warwick, RI, USA) to supply independent heat exchangers with on-demand cold and warm

glycol loops for down- and up-regulation of water temperature, respectively. We manually

checked and re-calibrated sensors, as needed, using digital traceable thermometers twice daily to

control potential temperature sensor drift. We randomly assigned mesocosms to the specified

treatments.

For the second experiment, we placed individuals in a custom-built experimental array at

the Bodega Marine Laboratory in which individuals from each population were placed in a split

plot design in replicated (N = 4 each 140L - 73 (W) x 66 (H) x 32 (D) cm) acrylic tanks per

treatment (Oceans Design Inc). Tanks were fed by sumps fixed with both 0.25hp chillers

(Aqualogic Delta Star®)and 1000W heat sticks (TSHTCE-1000S) and temperature controllers to

regulate temperature in a partially recirculating flow through system with fresh seawater

allocated to each system at a rate of 0.5L/min (approx. 5x turnover per day). Each sump was

affixed with protein skimmers, UV filter, and bio-ball filters such that water was first filtered and

then UV sterilized upon recirculation. Temperatures were set by hand each day for the heatwave

treatments and checked with temperature probes for all treatments. We ran experiments from

September 15, 2023, to December 19, 2023.

Animal husbandry

For both experiments, animals were fed on the same feeding regime and the same food as

above, also housed in the same density and same trays for consistency. We fed individuals

uniform dry pellets combining several macroalgal species formulated for the aquaculture of *S.*

purpuratus (Urchinomics Canada Inc., Halifax, NS, Canada). Animals in mesocosms were fed

twice per week and we removed uneaten food and refuse every 72 h. Food rations were

determined by trial and error in a prior pilot study to ensure all animals had consistent access to

food over time. To optimize access for all mesocosm inhabitants to abundant food, we enclosed

subjects and food in aquaculture baskets (two baskets per mesocosm, 7 or 8 animals per basket,

Thunderbird Plastics 48 x 33.5 x 10 cm Fish Farm Tray) such that food was always readily

accessible, and movement was not impeded. Each animal was supplied approximately 2.7 grams

of pelleted food twice per week (either 19 or 21 grams per basket for the baskets with 7 and 8

individuals, respectively in each mesocosm) for the duration of the experiment.

Histological and gonad assays

At the end of the first experiment, we measured all individuals to test for changes in

height and diameter (using precision digital calipers) and wet mass (to the nearest 0.1 g). Full

growth measurements are reported by Spindel (2023)[53]. Animals were then sacrificed to measure

gonad and histological properties. After opening urchin tests, we immediately removed gonads

for sampling. We excised one gonad from each animal for histological analyses; a second gonad

was excised and carefully weighed to the nearest 0.01 g. Using a clean, sterile scalpel we excised

an approximately 2 mm cross section from the first gonad which we immediately placed in a

histological cassette, preserved in Hartmann's fixative for 24 hours, and transferred to 70%

EtOH. Preserved gonads were embedded in paraffin, sliced, stained using eosin and hematoxylin,

and mounted. We assessed gonad samples for sex and developmental stage using four visual

subsections and the entire sample collectively to ensure agreement among subsamples.

Histological slides were scored on a scale of I to IV (sensu [31]), where representative stages are

depicted in Figure 2. For the second experiment, we only report analyses of histological data in

this paper.

Statistical Analyses

For the first experiment, we estimated the probability of individuals in each reproductive

stage as a function of mean temperature by sex and in response to El Niño and La Niña

treatments using a Bayesian ordinal logistic regression model. We modeled the response to mean

temperature using a non-parametric Gaussian process smoother. To account for treatment effects,

we included a categorical effect of treatment (constant, El Niño, La Niña). Gonad mass was

modeled with a Gamma likelihood and log link with the log of test diameter included as a

covariate to account for gonad-body size allometry. Because gonad mass was expected to exhibit

a concave response to temperature, a 3rd order polynomial was used for mean temperature. For

both analyses, intercepts were allowed to vary randomly by mesocosm. For the second

experiment, we used the same model structure for reproductive stage but with no smoother

effects (only categorical terms for constant 10˚C, constant 20˚C, or El Niño) and included

location of collection as a random effect.

For both datasets, we directly compared parameter posteriors within models and to ensure

that comparing parameters within models was robust to model (structural) uncertainty, we

conducted model comparison using Bayes Factors[54] (as hypothesis tests of the full versus null

models). For Bayes Factors we estimated the marginal likelihood using bridge sampling[55] via the

R package bridgesampling[30]. Because Bayes Factors can be sensitive to prior construction, we

ensured our analyses are robust by simulating 200 datasets from both the full and null model

under the same prior construction to ensure that Bayes Factor would consistently reject the null

when simulating from the full model and vice versa. We sampled posteriors using brms in R [56,57].

We ran models with vague priors, and conducted sensitivity analyses to ensure varying prior

distributions and hyperparameters did not qualitatively affect the results. We ran models for

12000 iterations for each of four chains after a 2000 iteration warmup period and checked

convergence visually and checked that parameters had convergence diagnostics (Rhat) less than

1.001. Data and code from this study are available at https://www.bco-dmo.org/project/818918.

Author contributions:

DKO, NBS, MJM, DS, BC, LRB and SS designed the research. DKO, NBS, MJM, BC, SK, KR,

DS, IG, EC, MF, and NM conducted mesocosm experiments. RF, MM, and DKO oversaw

conducted histological scoring. DKO conducted statistical analyses. DKO wrote the initial draft

of the manuscript, and all authors contributed to revisions.

Acknowledgements:

This study was funded with generous support from the Hakai Institute, a program of the

Tula Foundation, a grant from the US National Science Foundation to DKO and LRB (OCE nos.

2023649 and 2023664), as well as from the Santa Barbara Coastal (SCB) LTER (NSF OCE

no.1831937). We thank Eric Peterson, Christina Munck, as well as the facilities staff at the

Quadra Island Ecological Observatory for their generous support. Ellington Chen, Janelle

Beduya, and Rebecca Ward-Diorio assisted with imaging histological slides. Histology

Consultants Inc. conducted histological sectioning. We thank the Coastal Marine Science Inst.

Figure 1: A) Historical mean annual (large points) and monthly mean (small points) larval settlement on standardized brush samplers across six sites in Southern California vs Multivariate El Niño Index (MEI) since 1990 (reproduced from Okamoto et al. [21]). Colors correspond to the x-axis representing cooler (negative) and warmer (positive) temperatures associated with the MEI in Southern California. B) Historical benthic temperatures from Scripps Pier [51]. Small points: daily means. Black line: historical seasonal mean. Red and blue lines: historical El Niño & La Niña seasonal trends, respectively, from the data that were simulated in mesocosms (the period simulated in the experiment is indicated by the vertical grey bar). Horizontal dotted lines: constant mean temperatures simulated in replicate mesocosms. C & D) Mean satellite derived sea surface temperature from September through December in four strong (C - 1993, 1996, 1999, & 2010) and poor (D - 1994, 1997, 2004, & 2015) years for larval settlement from 1990 to 2016 [52]. Note that C & D are visualized to show comparative spatial trends, though satellite-derived temperature data may be biased relative to in situ benthic temperature data. Sites in C represent the six historical settlement collection locations that all show negative correlations with the MEI.

Figure 2: Gonad sections stained with hematoxylin and eosin with scale bar representing
500 μm. A-B Female ovaries in Stage I (A) and Stage IV (B) arrows indicate small
previtellogenic oocytes, vitellogenic oocytes in the process of meiosis, and mature ova. C-D.
Females in stage I are largely devoid of developing and vitellogenic oocytes with substantial
reserves invested in nutritive phagocytes (light pink), Stage II has few developing oocytes, Stage
III has some developing oocytes with few, scattered mature eggs, and Stage IV has many mature
eggs (dark, solid circles) and some developing oocytes (dark circles with visible, central
germinal vesicle). Male testes showing stage I (C) with few developing columns of
spermatocytes, and stage IV (D) with large sections of the gonad converted to mature
spermatozoa. Males in stage I have few visible developing spermatocytes (columns of dark
purple around the margins), males in stage II have dense columns of spermatocytes, males in
stage III have few small pockets of spermatozoa, and males in stage IV have large pockets of
spermatozoa and rapidly disappearing somatic tissue (light pink).

Figure 3: A) Mesocosms at the Hakai Institute's Quadra Island Ecological Observatory with trays holding experimental subjects. B) Experimental subjects consuming controlled, manufactured Urchinomics algal feed pellets.

512   .

Figure 4: Final gonad mass vs mean temperature. Black circles and the smooth trend represent
fixed temperature treatments; the red diamond and blue square represent El Niño and La Niña
thermal treatments, respectively. Error bars and bands represent the 90, and 95% highest
probability density interval of the posterior. Results are shown for males and females combined
with no evidence for a difference between sexes in the response (see Results).

Figure 5: Proportion of adult females (A, B) and males (C, D) in each reproductive stage by
treatment (El Niño (squares), La Niña (diamonds), or constant temperature (circles)) and mean
temperature. A) Proportion of females in stage IV. B) Proportion of females in stages I (dotted
red lines, red points), II (dashed grey lines, grey points, and III (solid purple lines, purple points).
C) Proportion of males in stage IV, and D) proportion of males in stages I, II, and III (same
symbology as females). Circles represent empirical means for constant temperature treatments,
while El Niño and La Niña points represent posterior mean estimates. Constant temperature
treatments had two replicate mesocosms per treatment and 15 animals per mesocosm. El Niño
and La Niña had four replicate mesocosms and 15 animals per mesocosm. Uncertainty
bands/intervals represent 95% (light/thin) and 90% (dark/thick) from the ordinal logistic
regression.

Figure 6: Estimated proportions of each stage for each sex for each treatment from the
population comparison experiment (Sonoma, Santa Barbara, San Diego). Points represent mean
and error bars 90/95% credible intervals. The model with no location interaction is indicated by
black circles while other symbols/colors reflect the model with location interaction with
sex/treatment (red circles - SB = Santa Barbara County, pink triangles - SD = San Diego, and
blue squares points - SON = Sonoma). All estimates are derived from Bayesian ordinal logistic
regressions with random effects of location and vague priors on the estimated means.

Figure 7: Daily benthic temperatures from San Diego (Scripps Pier), Santa Barbara (Arroyo
Burro SBC LTER), Sonoma (Bodega Marine Lab), and Quadra Island Ecological Observatory.
For consistency, data are shown only since 2015 (the inception of the Quadra Island time series).

References

1   Braschler, B., Chown, S. L. & Duffy, G. A. Sub-critical limits are viable alternatives to
critical thermal limits. *J Therm Biol* **101**, 103106 (2021).
https://doi.org:10.1016/j.jtherbio.2021.103106 PMID - 34879920

2   Kinzner, M. T. *et al.* Is temperature preference in the laboratory ecologically relevant for
the field? The case of *Drosophila nigrosparsa*. *Global Ecol Conservation* **18**, e00638
(2019). https://doi.org:10.1016/j.gecco.2019.e00638

3   Rezende, E. L., Bozinovic, F., Szilágyi, A. & Santos, M. Predicting temperature mortality
and selection in natural *Drosophila* populations. *Science* **369**, 1242-1245 (2020).
https://doi.org:10.1126/science.aba9287 PMID - 32883867

4   Oliver, E. C. *et al.* Marine heatwaves off eastern Tasmania: Trends, interannual
variability, and predictability. *Progress in oceanography* **161**, 116-130 (2018).

5   Munday, P. L., Jones, G. P., Pratchett, M. S. & Williams, A. J. Climate change and the
future for coral reef fishes. *Fish and Fisheries* **9**, 261-285 (2008).

6   Terblanche, J. S., Deere, J. A., Clusella-Trullas, S., Janion, C. & Chown, S. L. Critical
thermal limits depend on methodological context. *Proc Royal Soc B Biological Sci* **274**,
2935-2943 (2007).

7   Walsh, B. S. *et al.* The impact of climate change on fertility. *Trends in Ecology &*
Evolution **34**, 249-259 (2019).

8   Inouye, D. W. Climate change and phenology. *Wiley Interdiscip Rev Clim Change* **13**
(2022). https://doi.org:10.1002/wcc.764

9   Amasino, R. Vernalization, Competence, and the Epigenetic Memory of Winter. *Plant*
Cell **16**, 2553-2559 (2004). https://doi.org:10.1105/tpc.104.161070 PMID - 15466409

10   Kim, D.-H. Current understanding of flowering pathways in plants: focusing on the
vernalization pathway in *Arabidopsis* and several vegetable crop plants. *Hortic Environ*
Biotechnology **61**, 209-227 (2020). https://doi.org:10.1007/s13580-019-00218-5

11   Rogers-Bennett, L., Dondanville, R. F., Moore, J. D. & Vilchis, L. I. Response of Red
Abalone Reproduction to Warm Water, Starvation, and Disease Stressors: Implications of
Ocean Warming. *J Shellfish Res* **29**, 599-611 (2010).
https://doi.org:10.2983/035.029.0308

12   Breckels, R. D. & Neff, B. D. The effects of elevated temperature on the sexual traits,
immunology and survivorship of a tropical ectotherm. *J Exp Biol* **216**, 2658-2664 (2013).
https://doi.org:10.1242/jeb.084962 PMID - 23531818

13   Paxton, C. W., Baria, M. V. B., Weis, V. M. & Harii, S. Effect of elevated temperature
on fecundity and reproductive timing in the coral *Acropora digitifera*. *Zygote* **24**, 511-516
(2016). https://doi.org:10.1017/s0967199415000477 PMID - 26349407

14   Hurley, L. L., McDiarmid, C. S., Friesen, C. R., Griffith, S. C. & Rowe, M. Experimental
heatwaves negatively impact sperm quality in the zebra finch. *Proc Royal Soc B*

Biological Sci **285**, 20172547 (2018). https://doi.org:10.1098/rspb.2017.2547 PMID -
29343605

15 Parratt, S. R. *et al.* Temperatures that sterilize males better match global species
distributions than lethal temperatures. *Nat Clim Change* **11**, 481-484 (2021).
https://doi.org:10.1038/s41558-021-01047-0

16 Okamoto, D. K., Schmitt, R. J., Holbrook, S. J. & Reed, D. C. Fluctuations in food
supply drive recruitment variation in a marine fish. *Proc Royal Soc B Biological Sci* **279**,
4542-4550 (2012). https://doi.org:10.1098/rspb.2012.1862 PMID - 23015631

17 Stoner, A. W. & Ray-Culp, M. Evidence for Allee effects in an over-harvested marine
gastropod: density-dependent mating and egg production. *Mar Ecol Prog Ser* **202**, 297-
302 (2000). https://doi.org:10.3354/meps202297

18 Gaines, S. D. & Bertness, M. D. Dispersal of juveniles and variable recruitment in sessile
marine species. *Nature* **360**, 579-580 (1992). https://doi.org:10.1038/360579a0

19 Kelly, M. S. Environmental parameters controlling gametogenesis in the echinoid
Psammechinus miliaris. *Journal of Experimental Marine Biology and Ecology* **266**, 67-
80 (2001).

20 Rogers-Bennett, L., Klamt, R. & Catton, C. A. Survivors of climate driven abalone mass
mortality exhibit declines in health and reproduction following kelp forest collapse.
Frontiers in Marine Science **8**, 725134 (2021).

21 Okamoto, D. K., Schroeter, S. C. & Reed, D. C. Effects of ocean climate on
spatiotemporal variation in sea urchin settlement and recruitment. *Limnol Oceanogr* **65**,
2076-2091 (2020). https://doi.org:10.1002/lno.11440

22 Ebert, T. A., Schroeter, S. C., Dixon, J. D. & Kalvass, P. Settlement patterns of red and
purple sea urchins (*Strongylocentrotus franciscanus* and *S. purpuratus*) in California,
USA. *Mar Ecol Prog Ser*, 41-52 (1994).

23 Farmanfarmaian, A. & Giese, A. C. Thermal tolerance and acclimation in the western
purple sea urchin, Strongylocentrotus purpuratus. *Physiological Zoology* **36**, 237-243
(1963).

24 Azad, A. K., Pearce, C. M. & McKinley, R. S. Influence of stocking density and
temperature on early development and survival of the purple sea urchin,
Strongylocentrotus purpuratus (S timpson, 1857). *Aquaculture Research* **43**, 1577-1591
(2012).

25 Azad, A. K., Pearce, C. M. & McKinley, R. S. Influence of microalgal species and
dietary rations on larval development and survival of the purple sea urchin,
Strongylocentrotus purpuratus (Stimpson, 1857). *Aquaculture* **322**, 210-217 (2011).

26 Strathmann, M. F. *Reproduction and development of marine invertebrates of the northern*
Pacific coast: data and methods for the study of eggs, embryos, and larvae. (University
of Washington Press, 2017).

27   Basch, L. V. & Tegner, M. J. Reproductive responses of purple sea urchin
626   (*Strongylocentrotus purpuratus*) populations to environmental conditions across a coastal
627   depth gradient. *Bulletin of Marine Science* **81**, 255-282 (2007).

628   28   Pearse, J. S. in *Advances in invertebrate reproduction*   (ed S. Adams W. H. Clark Jr)
629   (1981).

630   29   Cochran, R. C. & Engelmann, F. Environmental regulation of the annual reproductive
631   season of Strongylocentrotus purpuratus (Stimpson). *The Biological Bulletin* **148**, 393-
632   401 (1975).

633   30   Gronau, Q. F., Singmann, H. & Wagenmakers, E.-J. bridgesampling: An R package for
634   estimating normalizing constants. *Journal of Statistical Software* **92**, 1-29 (2020).
635   https://doi.org:doi:10.18637/jss.v092.i10

636   31   Byrne, M. Annual reproductive cycles of the commercial sea urchin *Paracentrotus*
637   *lividus* from an exposed intertidal and a sheltered subtidal habitat on the west coast of
638   Ireland. *Marine Biology* **104**, 275-289 (1990).

639   32   Huey, R. B. & Kingsolver, J. G. Climate warming, resource availability, and the
640   metabolic meltdown of ectotherms. *The American Naturalist* **194**, E140-E150 (2019).

641   33   Spindel, N. B. *Impacts of abiotic and biotic stressors on the energetics and nutritional*
642   *composition of temperate sea urchins* PhD thesis, Florida State University, ("In Prep").

643   34   Cushing, D. in *Advances in marine biology* Vol. 26   249-293 (Elsevier, 1990).

644   35   Kapsenberg, L., Okamoto, D. K., Dutton, J. M. & Hofmann, G. E. Sensitivity of sea
645   urchin fertilization to pH varies across a natural pH mosaic. *Ecology and Evolution* **7**,
646   1737-1750 (2017).

647   36   Tegner, M. & Dayton, P. Sea urchins, El Niños, and the long term stability of Southern
648   California kelp forest communities. *Marine ecology progress series. Oldendorf* **77**, 49-63
649   (1991).

650   37   Seward, L. C. The relationship between green sea urchin spawning, spring phytoplankton
651   blooms, and the winter-spring hydrography at selected sites in Maine. (2002).

652   38   Zhadan, P. M., Vaschenko, M. A. & Almyashova, T. N. Spawning failure in the sea
653   urchin Strongylocentrotus intermedius in the northwestern Sea of Japan: Potential
654   environmental causes. *Journal of Experimental Marine Biology and Ecology* **465**, 11-23
655   (2015).

656   39   Lester, S. E., Tobin, E. D. & Behrens, M. D. Disease dynamics and the potential role of
657   thermal stress in the sea urchin, *Strongylocentrotus purpuratus*. *Canadian Journal of*
658   *Fisheries and Aquatic Sciences* **64**, 314-323 (2007).

659   40   Behrens, M. D. & Lafferty, K. D. Effects of marine reserves and urchin disease on
660   southern Californian rocky reef communities. *Mar Ecol Prog Ser* **279**, 129-139 (2004).

661   41   Lafferty, K. D. Fishing for lobsters indirectly increases epidemics in sea urchins.
662   *Ecological Applications* **14**, 1566-1573 (2004).

42 Arafeh-Dalmau, N. *et al.* Extreme marine heatwaves alter kelp forest community near its
equatorward distribution limit. *Frontiers in Marine Science* **6**, 499 (2019).

43 Filbee-Dexter, K. *et al.* Marine heatwaves and the collapse of marginal North Atlantic
kelp forests. *Scientific Reports* **10**, 13388 (2020).

44 McPherson, M. L. *et al.* Large-scale shift in the structure of a kelp forest ecosystem co-
occurs with an epizootic and marine heatwave. *Communications biology* **4**, 298 (2021).

45 Michaud, K. M., Reed, D. C. & Miller, R. J. The Blob marine heatwave transforms
California kelp forest ecosystems. *Communications Biology* **5**, 1143 (2022).

46 Smale, D. A. Impacts of ocean warming on kelp forest ecosystems. *New Phytologist* **225**,
1447-1454 (2020).

47 Thomsen, M. S. *et al.* Local extinction of bull kelp (*Durvillaea* spp.) due to a marine
heatwave. *Frontiers in Marine Science* **6**, 84 (2019).

48 Wernberg, T. in *Ecosystem collapse and climate change*   325-343 (Springer, 2021).

49 Smith, K. E. *et al.* Socioeconomic impacts of marine heatwaves: Global issues and
opportunities. *Science* **374**, eabj3593 (2021).

50 Walsh, B. S. *et al.* Female fruit flies cannot protect stored sperm from high temperature
damage. *J Therm Biol* **105**, 103209 (2022).

51 Carter, M. L. *et al.*   (UC San Diego Library Digital Collections, 2022).

52 Reynolds, R. W. *et al.* Daily High-Resolution-Blended Analyses for Sea Surface
Temperature. *J Climate* **20**, 5473-5496 (2007). https://doi.org:10.1175/2007jcli1824.1

53 Spindel, N. B. *Ecophysiology of Ectothermic Ecosystem Engineers: Bioenergetic Effects*
of Climate and Food on Dominant Consumers and Their Consequences for Coastal
Ecosystems PhD thesis, The Florida State University, (2023).

54 Kass, R. E. & Raftery, A. E. Bayes factors. *Journal of the american statistical*
association **90**, 773-795 (1995).

55 Meng, X.-L. & Wong, W. H. Simulating ratios of normalizing constants via a simple
identity: a theoretical exploration. *Statistica Sinica*, 831-860 (1996).

56 Bürkner, P.-C. brms: An R package for Bayesian multilevel models using Stan. *Journal*
of statistical software **80**, 1-28 (2017).

57 R Core Team. R: A Language and Environment for Statistical Computing (R Foundation
for Statistical Computing, Vienna, Austria., 2023).