1 **Prior heat stress increases pathogen susceptibility in the model cnidarian *Exaptasia***
2 ***diaphana***
3
**Abstract**

Anthropogenic climate change has significantly altered terrestrial and marine ecosystems
globally, often in the form of climate-related events such as thermal anomalies and disease
outbreaks. Although the isolated effects of these stressors have been well documented, a growing
body of literature suggests that stressors often interact, resulting in complex effects on
ecosystems, including coral reefs where sequential associations between heat stress and disease
have had profound impacts. Here we used the model cnidarian *Exaiptasia diaphana* to
investigate mechanisms linking prior heat stress to increased disease susceptibility. We examined
anemone pathogen susceptibility and physiology (symbiosis, immunity, and energetics)
following recovery from heat stress. We observed significantly increased pathogen susceptibility
in anemones previously exposed to heat stress. Notably, prior heat stress reduced anemone
energetic reserves (carbohydrate concentration), and activity of multiple immune components.
Minimal effects of prior heat stress on symbiont density were observed. Together, results suggest
changes in energetic availability might have the strongest effect on pathogen susceptibility and
immunity following heat stress. The results presented here provide critical insight regarding the
interplay between heat stress recovery and pathogen susceptibility in cnidarians and are an
important first step towards understanding temporal associations between these stressors.

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23 **Keywords**
24 **cnidarian, heat stress, symbiosis, disease, multiple stressors, immunity**
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## Introduction

Anthropogenic pressure has led to environmental and ecological changes in terrestrial and aquatic ecosystems across the world, often in the form of climate-related events.[1,2] Marine ecosystems, in particular, have been significantly affected by multiple climate-induced stressors, such as ocean warming and disease outbreaks[3-5]. Both of these stressors directly affect organismal physiology through alteration of metabolic function, and indirectly affect a variety of ecological processes, such as oxygen availability, nutrient cycling, and ocean circulation[3,6]. As such, ocean warming, specifically marine heatwaves, are a leading cause of mass mortality in marine organisms[7-10]. Similarly, increases in the frequency and severity of disease outbreaks pose significant threats to marine ecosystems. Marine taxa are affected by a large diversity of pathogens ranging from viruses and bacteria to fungi and metazoans[11], many of which have triggered significant mortality events in a variety of marine organisms[12-15].

Although the isolated effects of ocean warming, disease outbreaks, and other climate-associated stressors have been well documented within marine ecosystems[3,5], a growing body of literature suggests that these stressors often coincide. Depending on the circumstances and stressors in question, interactions may range from cumulative to synergistic, or even antagonistic in nature.[16,17] Associations between heat stress and disease outbreaks are particularly prevalent and have often contributed to extensive marine mortality.[18] Consequently, studies have attempted to improve understanding of associations between heat stress and disease in marine ecosystems, but most have focused only on simultaneous synergy between the two stressors.[19-21] However, frequent increases in disease outbreaks following marine heatwaves suggest that these two stressors may more commonly accumulate in a sequential, rather than simultaneous, manner.[22-25] Despite these observations, the mechanisms driving sequential temporal patterns between heat stress and disease are poorly understood and necessitate further investigation.

Coral reef ecosystems present a unique opportunity to investigate sequential associations between heat stress and disease. These ecosystems are commonly faced with a variety of climate stressors[26], which have resulted in unprecedented declines in coral cover over the past several decades.[27,28] Increasing sea surface temperature (SST), in particular, has exacerbated the frequency and severity of coral decline. Two major drivers of these declines are heat-induced bleaching—loss of endosymbiotic algae[29,30]—and disease[31]. Although isolated heat-induced bleaching[32-34] and disease are common[35-38], these events have also been frequently observed in

close spatial and temporal associations.[39-43] In many cases, coral disease follows bleaching[25],
with extensive disease-related mortality often reported in the weeks to months after heat
stress.[24,44] Despite these observations, nearly all experimental studies have focused on the
synergistic effects of simultaneous heat stress and disease, specifically documenting increases in
disease susceptibility when these stressors occur at the same time.[45,46] Synchronous associations
between heat stress and increased pathogen susceptibility may be linked to heat induced
suppression of immunity,[47-51] though other studies have instead demonstrated minimal or
positive effects of heat stress on coral immunity.[52-54] Furthermore, it is unclear if similar
mechanisms underly sequential associations between heat stress and disease outbreaks.

Sequential associations between heat stress and disease outbreaks may be mediated by
several mechanisms. First, heat stress may induce changes in host energetic reserves, which lead
to immunosuppression. The loss of symbionts which typically occurs as a result of heat stress
may severely limit the energetic resources available to the host.[55-57] According to resource
allocation theory, organisms must distribute a finite amount of resources amongst diverse
biological processes—growth, reproduction, metabolism, immunity, etc.[58-60] This may involve
trade-offs where resources are shifted between competing needs in order to maximize fitness.[58-60]
Consequently, reduction in the total amount of resources may alter the proportions in which
resources are distributed, potentially resulting in the suppression of one or more biological
processes. As such, immune suppression is one of the most commonly documented energetic
trade-offs associated with resource limitation.[61,62] Heat-stress induced limitations in energetics
may exacerbate resource tradeoffs and result in changes to resource allocation, including
allocation to host immune responses. Alternatively, immune-energetic trade-offs may also occur
in the absence of resource limitation in response to biotic or abiotic pressures: organisms may
dynamically shift resources between processes in response to pressing needs caused by
environmental stress or other changes in biological condition.[63]

Alternatively, dynamic changes in symbiosis during heat stress and associated recovery may
contribute to observed links between heat stress and disease outbreaks. Suppression of host
immunity by symbionts has been hypothesized as an important mechanism required for the
successful establishment and maintenance of symbiotic relationships.[64] Previous laboratory
studies have confirmed negative associations between symbiosis and immunity across several
cnidarians.[65-68] However, the dynamics and consequences of immune-symbiosis interplay during

the breakdown and re-establishment of symbiosis remain unclear. One possibility that the rapid

repopulation of endosymbionts following heat-induced dysbiosis exacerbates symbiotic

immunosuppression, potentially explaining, in part, naturally observed patterns of disease after

heat-induced bleaching events.

Here we describe the results of a study aimed at assessing the sequential associations

between heat stress and disease outbreaks in cnidarians, with a specific focus on the roles of

changes in energetics, symbiont density, and host immunity in these processes. In order to

comprehensively address this question, we leveraged the powerful emergent cnidarian model

species, *Exaiptasia diaphana* (common name Aiptasia). While tropical scleractinian corals are

exceptionally important from an ecological standpoint, their precarious species statuses and

inherent experimental limitations complicate the laboratory investigation of complex biological

phenomenon in these species. In contrast with tropical corals, Aiptasia is easily reared,

maintained, and destructively sampled in large amounts without many of the experimental and

ethical challenges of working with tropical corals.[69] Aiptasia also maintains symbiotic

relationships with endosymbiotic algae of the family Symbiodiniaceae, like tropical corals.[70,71]

Finally, Aiptasia is a particularly handy system due to the existence of established, genetically

identical clonal lines, allowing for control of genetic variation in experimentation.[69] We

leveraged the *E. diaphana* model system to investigate the mechanisms linking heat stress and

disease susceptibility in cnidarians. Specifically, we examined host immunity, symbiont density,

and energetic reserves following heat-stress, as well as the effects of prior heat stress on

pathogen susceptibility, in two genetically distinct clonal lines of *E. diaphana*. We document a

significant association between prior heat stress and subsequent increased pathogen susceptibility

and highlight the potential physiological mechanisms of this association. The results described

are an important first step towards understanding temporal links between two of the most

significant climate-associated stressors faced by cnidarians.

**Methods**

*Experimental Design*

Anemones were obtained from laboratory-maintained clonal lines of symbiotic *E. diaphana*

(VWB9 and H2; V. Weis).  Anemones from each clonal line were randomly assigned to one of

four possible treatment combinations:  ambient temperature and placebo, heat stress and placebo,

ambient temperature and infection, or heat stress and infection (Table 1). Individuals were
transferred accordingly into sterile, lidded 6-well culture plates filled with 10 mL of filter-sterile
artificial sea water (FASW). Each well contained a single anemone and plates were distributed
such that each treatment combination had 30 plates. Plates assigned to ambient temperature
treatment were maintained in a single temperature-controlled incubator (Percival Scientific;
Model AL-41L4) at 27°C for the duration of the experiment. Plates assigned to heat stress
treatment were randomly divided between two separate temperature-controlled incubators
(VWR; Model 2005, VWR; Model 3734). Prior to the start of the experiment, all individuals
were acclimated to well plates for four days at 27°C in their respective incubators. Throughout
the entire experiment, anemones in all three incubators were maintained under the same light
conditions (12hr:12hr light: dark cycles with approximately 40-60 µmol·m$^{-2}$ PAR) and received
bi-weekly feeding (*Artemia nauplii)* and subsequent water changes until the point of infection.
Plate locations were randomized within and between incubators at the time of water changes, as
appropriate.

*Experimental Heat Stress & Pathogen Challenge*

Heat stressed anemones were exposed to a temperature profile which consisted of a seven-
day ramping period wherein temperatures were raised 1°C per day to a final temperature of 34°C
which was maintained for five days before an immediate return to ambient temperatures (27°C)
on day six. The chosen protocol was developed based on previous studies that used 34°C as a
target temperature[45,72,73] and pilot data which demonstrated this regime induced bleaching
(measured as decline in chlorophyll fluorescence) with limited mortality**.** This approach was
employed to mimic a sublethal bleaching event from which individuals recover so as to test our
hypotheses regarding the impact of reestablishment of symbiosis on host immunity. During the
experiment, bleaching was confirmed by visual assessment of anemone color and was found to
be consistent with the expected timeline. Following heat stress, anemones were immediately
returned to 27°C and allowed to recover for two weeks. At the conclusion of this recovery
period, surviving anemones from both temperature treatments were inoculated with either a
pathogen or a placebo (FASW) by plate, according to *a priori* assignments. Pathogen exposure
was conducted using the known coral pathogen *V. coralliilyticus* BAA 450[74] at a predetermined
dosage of 1 x 10$^8$ CFU·mL$^{-1}$.[45] Final bacterial concentrations were confirmed with serial

dilutions and plating. Following exposure, a subset of anemones was monitored for survivorship

(**Table 2**): mortality data was collected every two hours for the first 24 hours, and then daily for

the remainder of the experiment (96 hours post-infection) at which point 89% of exposed

anemones had died. Individuals were considered dead when they exhibited more than 75% tissue

degradation or had 75% tissue degradation for more than 24 hours.

*Sample Collection and Processing*

Twelve hours after inoculation, a second subset of randomly selected anemones was sampled

for the assessment of symbiont density, immune activity, and energetic reserves (**Table 2**).

Anemones were anesthetized in 10 mL of $MgCl_2$ (0.37 M) prior to sampling. First, we excised a

single tentacle for estimates of symbiont density described below. Subsequent tentacle

processing was modified from existing protocols.[75] Following tentacle excision, the remaining

anemone was halved using a sterile razor blade; one half was flash frozen for the immune and

energetic assays described herein. Frozen tissue samples were homogenized with 3 mm glass

beads in 1.5 mL of 100 mM Tris + 0.05 M DTT buffer (pH 7.8) for 1 minute at 5 $m \cdot s^{-1}$ (Tim

Bateman, *pers. comm.*) using a bead mill (Fisherbrand). For each sample, a 200 µL aliquot of

homogenate was transferred into a separate microcentrifuge tube and stored at -20°C for melanin

assays. The remaining homogenate for each sample was then transferred into a new 2 mL

microcentrifuge tube and centrifuged at 2250xg at 4°C for five minutes. Resulting supernatants

were aliquoted accordingly for each set of assays (protein, lipid, carbohydrates, ash-free dry

weight) and stored at -80°C (protein fraction) or -20°C (all other fractions) prior to downstream

analyses.

*Symbiont Density Estimation*

Mounted tentacles were imaged using brightfield and fluorescence microscopy on the

Cytation1 Cell Imaging Multimode Reader (Biotek). Approximate tentacle area was determined

via brightfield microscopy. Symbiont autofluorescence was measured using a chlorophyll a

fluorescent filter (Agilent; EX 445/45 nm, EM 685/40 nm). Brightfield and fluorescent images of

each tentacle were combined to standardize fluorescence by tentacle area. The average pixel

intensity of the chlorophyll a fluorescence within each standardized tentacle area was used as a

proxy for symbiont density.

*Immune Assays*

Cnidarian immunity was measured using a suite of standardized immune assays originally

developed for scleractinian corals[76,77] and adapted for use in *E. diaphana*. We measured

responses related to three major innate immune pathways: antioxidant, phenoloxidase, and

antibacterial pathways. All immune assays were run in triplicate on separate 96-well plates and

measured using the Cytation1 Cell Imaging Multimode Reader and Gen5 Software (Biotek).

Negative controls for all assays were 100mM Tris + 0.05 mM DTT buffer (homogenization

buffer). All enzymatic immune assays were standardized by protein concentration using

established protocols for Red660 colorimetric assays[76,77], while melanin concentration was

standardized by desiccated tissue weight.

Catalase (CAT)

Cnidarians produce antioxidant enzymes as a mechanism to protect against reactive

oxygen species generated by both hosts and pathogens during infection and associated

immune responses.[78-80]   Catalase activity was measured using established protocols.[76,77]

Briefly, 5 µL of protein sample were combined with 45 uL of 50 mM PBS (7.0 pH) and 75

µL of 25 mM $H_2O_2$ in a UV 96-well plate. Immediately following addition of $H_2O_2$,

absorbance at 240 nm was read every 45 seconds for 15 minutes. Catalase activity was

determined from the steepest point of the reaction curve. Total hydrogen peroxide scavenged

during this period was calculated using a standard curve of $H_2O_2$. Final catalase activity is

reported as $H_2O_2$ scavenged per minute per mg of protein.

Total Phenoloxidase Activity (TPO)

The phenoloxidase pathway is an enzymatic cascade resulting in the synthesis of melanin

– an effector molecule involved in wound healing and pathogen encapsulation.[78-80] Here we

quantify the phenoloxidase cascade using both total phenoloxidase activity (enzymatic

activity of prophenoloxidase + phenoloxidase) and melanin concentration (see next section).

Total phenoloxidase  activity was estimated using modifications to existing protocols.[77]

Briefly, 20 µL of protein sample were combined with 20 µL of 50 mM PBS (7.0 pH) and 25

µL of 0.1 mg·mL$^{-1}$ trypsin. Samples were then incubated on ice for 30 minutes to allow for

cleavage of prophenoloxidase (PPO) into phenoloxidase (PO). We then measured total

potential phenoloxidase activity (i.e. activity of any existing phenoloxidase in the sample +

activity of prophenoloxidase activated during trypsin incubation). Following incubation, 30 µL of 10mM L-DOPA were added to each well to initiate melanin synthesis. Absorbance at 490 nm was read every 10 minutes for 4 hours. Total phenoloxidase activity was calculated from the steepest portion of the reaction curve and standardized to sample protein concentration.

Melanin (MEL)

Melanin tissue concentration was estimated using reserved tissue samples following established protocols.[76,77] Briefly, samples were desiccated in a speed vac (Eppendorf Concentrator Plus) for a minimum of 12 hours. Following desiccation, samples were weighed to obtain total dry tissue weight. Then, 1.0 mm glass beads (Sigma Aldrich) and 400 µL of 10 M NaOH were added to each sample. Samples were vortexed for 20 seconds and stored in the dark for 48 hours. During this incubation, samples were vortexed briefly at 36 and 48 hours. At the end of this period, samples were centrifuged at 183 x $g$ for 10 minutes at room temperature. A melanin standard processed identically to samples, was used to create a standard curve. To measure melanin concentration, 40 µL of the resulting supernatant from samples or melanin standard dilutions were transferred to half volume UV 96-well plates. Absorbance was measured at 490 nm and converted to melanin concentration using the standard curve. All melanin values were standardized by desiccated tissue weight.

Antibacterial Activity (AB)

Finally, we quantified antibacterial activity as a metric of the production of antimicrobial peptides and metabolites that can neutralize a wide variety of pathogenic bacteria.[78-80] Antibacterial activity here was quantified by measuring bacterial growth rates of *Vibrio coralliilyticus* in the presence of protein extracts generated from host tissue. Antibacterial activity against *Vibrio coralliilyticus* BAA 450 was calculated using established protocols.[76] Prior to the assay, samples were diluted to a standardized protein concentration. To initiate the reaction, 60 µL of sample (or buffer control) were combined with 140 µL of *V. coralliilyticus* diluted to an OD600 of 0.2. Absorbance was read at 600 nm every 10 minutes for 6 hours to generate bacterial growth curves. Percent inhibition of bacterial growth was calculated by comparing sample growth rate to growth rate of buffer controls from the steepest part of the curve.

*Energetic Assays*

Total lipids and total carbohydrates assays were used to determine approximate host

energetic budget. Lipids and carbohydrates are commonly used proxies for energetic budgets in

multiple cnidarian species[81]. Total lipid and total carbohydrate assays were run in triplicate and

duplicate, respectively, and measured in separate 96-well plates using the Cytation1 Cell

Imaging Multimode Reader and Gen5 Software (Biotek). Tris + DTT buffer was used as a

negative control for both assays.

Ash Free Dry Weight

Sample ash-free dry weight was calculated to standardize both lipid and carbohydrate

assays using modifications from previous methods.[82] All samples were weighed using an

analytical scale. Briefly, a recorded volume of sample was added to tared aluminum boats

and dried at 60°C in a drying oven for 48 hours. Samples were removed from the drying

oven and weighed before combusting at 500°C for 4 hours in a furnace box. Following

combustion, samples were weighed one final time. Ash-free dry weight was calculated as the

average difference between the dry weights and ash weights of each sample and standardized

to 125 µL volume.

Lipids

Lipid concentrations were estimated using a total lipid assay following established

protocols.[83] Samples were dried for 5 hours at room temperature in a speed vac and then

combined with 500 µL 2:1 chloroform: methanol solution and 100µL of 0.05 M NaCl. Tubes

were vortexed for 20 seconds before 2-hour incubation on a shaking incubator at 200 RPM,

with brief vortexing approximately every 30 minutes. At the end of this period, samples were

centrifuged at 1650 x *g* for 10 minutes at room temperature. One hundred microliters from

the bottom organic layer of each sample were transferred in triplicate to a 96-well PCR plate.

A serial dilution of corn oil standard dissolved in chloroform was also plated for the creation

of a standard curve. Next, 50 µL of methanol were added to each well. The plate was

incubated in a 70°C water bath for at least 15 minutes to evaporate the solvent, after which

150 µL of 18 M sulfuric acid ($H_2SO_4$) was added to each well. The plate was then covered

and incubated at 90°C for 20 minutes and cooled to 4°C. Following this incubation, the plate

was vortexed and 75 µL of each sample and standard were transferred into a new 96-well

microplate. Following baseline absorbance measurement at 540 nm, 34.5 µL of mg·mL$^{-1}$

vanillin in 17% phosphoric acid was then added to each well. After a 5-minute dark

incubation period, absorbance values were read again at 540 nm and subtracted from the

initial reads. Average differences in absorbance values were converted to lipid concentration

using the standard curve and standardized by initial sample volume and ash-free dry weight.

<u>Carbohydrates</u>

Carbohydrate concentrations were estimated following established protocols.[84] Fifty

microliters of either sample or glucose standard were plated in duplicate on a 96-well

microplate. Next, 150 µL of 18 M $H_2SO_4$ was added to each well, followed immediately by

30 µL of 5% phenol. Plates were incubated (uncovered) in a hot water bath for 5 minutes at

80°C and allowed to cool for 15-20 minutes. Absorbance values were measured at 485 nm

and 750 nm. Total carbohydrates were calculated by converting the average difference

between 750 nm and 485 nm absorbance values to glucose concentrations using a standard

curve. Carbohydrate values were standardized to tissue weight using ash-free dry weight.

*Statistical Analyses*

All statistical analyses were performed using R Version 4.4.0.[85] Prior to hypothesis testing

and modeling sample outliers were identified using the Rosner test from the EnvStats package

(v2.8.1[86]) and removed. Shapiro-Wilke and Bartlett tests were used to confirm the assumptions

of normal distribution and equal variances, respectively. Some assays failed to meet assumptions

necessary for parametric testing, even following transformation, necessitating the use of non-

parametric analyses.

<u>Mortality</u>

Survivorship data from our first subset of anemones were used to assess the effects of

pathogen exposure, heat stress and clonal line on mortality. First, we tested for a significant

effect of pathogen exposure on mortality using a Kaplan-Meier analysis (Survival ~ pathogen

challenge; n = 184; survival package v3.5.5[87]). Mortality data within pathogen exposed

groups was then statistically evaluated using standard approaches for survival analyses,

specifically a Cox proportional hazard's model, which both allows for censorship of

individuals (i.e. those that did not die) and incorporates data regarding rate of mortality.[88,89]

We investigated the interactive effects of heat stress and clonal line on mortality within the
subset of pathogen-exposed anemones only (Survival ~ temperature * clonal line; n = 92).

<u>Univariate hypothesis testing</u>

A second subset of anemones reserved for assessment of symbiont density, energetic
reserves, and immune activity, was used to investigate hypotheses related to mechanisms of
differential mortality. Given that there were no interactions between temperature treatment
and clonal line in our cox proportional hazards modeling, we did not test for interactions in
these physiological metrics. To isolate the effects of temperature from the effects of pathogen
challenge, we separated out placebo samples and tested for differences in symbiont density
(mean chlorophyll fluorescence intensity) or energetic reserves (total lipids and total
carbohydrates) between temperature treatments (ambient vs. heat) and clonal lines within the
placebo groups. The impacts of heat stress or clonal line on symbiont density, lipids, and
carbohydrates were assessed independently with separate Welch's two-sample t-tests
(symbiont density and lipids) or non-parametric Wilcoxon Rank Sum tests (carbohydrates).

<u>Redundancy analysis</u>

To investigate the impact of our independent variables (temperature, pathogen exposure,
and clonal line) and covariates (symbiont density, total lipids, and total carbohydrates) on
host immune activity, we ran a redundancy analysis (RDA) using the vegan package
(v2.6.6).[90] Before starting with this multivariate approach, we tested for and confirmed no
collinearity among our variables. Outliers in predictor variables were removed and all
continuous variables were centered and scaled before separation into predictor (i.e., symbiont
density, lipids, carbohydrates) and response (i.e., catalase, total phenoloxidase, and
antibacterial activity, and melanin concentration) data frames. An RDA was used to identify
which variables best predicted variation in immunity across our samples. We started by
running an RDA with all possible predictor variables (i.e. clonal line, temperature treatment,
pathogen exposure, symbiont density, lipids, and carbohydrates). From there, we used a
forward selection to identify which predictor variables explained significant variation in our
dataset. After identifying significant predictor variables, we ran a final RDA using the best-
fit model determined by forward selection (~ Treatment + Carbohydrates + Lipids).

 Linear modeling

Finally, we used linear modeling to assess variation in individual immune parameters

with respect to the statistically relevant factors determined via RDA (treatment,

carbohydrate, lipids). For normally distributed immune parameters (catalase, total

phenoloxidase, antibacterial activity) we used GLMs. Global models with all possible

additive and interactive variables were analyzed for catalase, total phenoloxidase, and

antibacterial activity independently. All models were dredged and ranked via AICc using the

MuMIn package (v1.47.5)[91], where applicable. Models with delta AICc values less than or

equal to 2 were averaged to determine the best-fit models for each immune parameter. For

melanin, which was right skewed with multiple zero measurements, we adopted a two-tiered

strategy. First, we used a probit model (i.e. a generalized linear model with binomial error

structure) to predict the presence of melanin among samples using heat treatment,

carbohydrate and lipid concentration as predictors. We then removed samples in which

melanin was not detected and used multiple regression to predict melanin concentrations

employing the same predictors.

**Results**

*Mortality*

Anemone mortality was significantly greater in anemones exposed to the pathogen (Kaplan-

Meier; $p < 0.001$; **Supplemental Fig. 1**). Of those anemones exposed to pathogen, 89% died

during the 96-hour period, compared to 0% mortality within the placebo group. While there were

no significant interactive effects of temperature treatment and clonal line on anemone mortality

within pathogen exposed anemones controls (hazard ratio = 1.32; 95% CI [0.52 – 3.34]; $p =$

0.562), mortality rate was significantly greater in the heat stressed anemones relative to the

ambient controls (hazard ratio = 2.29; 95% CI [1.11 – 4.75]; $p = 0.0255$) and among H2

anemones relative to VWB9 anemones (hazard ratio = 0.197; 95% CI [0.10-0.37]; $p < 0.001$;

**Fig. 1; Supplemental Fig. 2**). When considering just those anemones exposed to pathogens,

82% of ambient and 97% of heat stressed individuals died within the 96-hour period.

Furthermore, 100% of H2 and 82% of VWB9 anemones died.

*Symbiont Density & Energetic Reserves*

Symbiont density, estimated by mean chlorophyll a fluorescence pixel intensity, was

significantly higher in heat stressed anemones compared to ambient controls (Welch's two

sample t-test; t = -2.52, df = 40.181, $p$ = 0.0157. 95% CI [-5617 – -620]; **Fig. 2a**). Symbiont

density also varied across clonal lines (Welch's two sample t test; t = 2.65, df = 38.3, $p$ = 0.0118;

95% CI [768 – 5766]); H2 had significantly greater symbiont density than VWB9 anemones

**(Fig. 2a).** For energetic reserves, total carbohydrates varied between treatment groups.

Carbohydrates were significantly lower in heat treated anemones compared to anemones that

remained at ambient temperature (Wilcoxon Rank Sum test; W = 392, $p$ < 0.01; **Fig. 2b**), but not

significantly different between clonal lines (Wilcoxon Rank Sum test; W = 220, p = 0.358; **Fig.**

**2b**). Total lipids did not significantly differ between either temperature treatment (Welch's two

sample t-test; t = -0.444, df = 40.12, $p$ = 0.659, 95% CI [-0.143 – 0.0917], **Fig 2c**) or clonal lines

(Welch's two sample t-test; t = -1.27, df = 40.9, $p$ = 0.213, 95% CI [-0.185 – 0.0424], **Fig 2c**).

*Immune Responses*

The RDA explained significant variation along the first axis of the ordination, RDA1, which

explained 18.12% of the total variation **(Fig. 3)**. Temperature treatment (ANOVA, $p$ = 0.003),

lipid concentration (ANOVA, $p$ = 0.003), and carbohydrate concentration (ANOVA, $p$ = 0.034),

all explained significant variation in immune responses. Temperature treatment explained

significant variation along both the RDA1 and RDA2 axes (**Fig. 3** boxplots; t-test, p < 0.0001).

However, carbohydrate concentration explained variation mostly along the first RDA axis (**Fig. 3**

regression bottom-left; Pearson correlation, p < 0.0001), while lipid concentration explained

variation primarily on the second axis (**Fig. 3** regression top-right; Pearson correlation, p <

0.0001). The immune variables used to construct the ordination show strong association with the

temperature treatments with all arrows representing the immune variables pointing towards the

group of samples from the ambient treatment.

Linear modeling of the effects of each of our significant predictors on individual immune

metrics revealed metric-specific effects (**Table 3**). Three metrics of immunity (CAT, TPO, and

MEL) were significantly lower in heat treated anemones relative to ambient temperatures (CAT

GLM, $p$ = 0.0129; TPO GLM, p = 0.0212; MEL zero-included binomial, $p$ < 0.0430; **Fig. 4).**

Additionally multiple metrics of immunity were positively associated with energetic

measurements. Both CAT and MEL were positively associated with carbohydrate concentration

(CAT GLM, $p$ < 0.001; MEL zero-excluded GLM, $p$ < 0.001; **Fig. 5**), and MEL was also

positively associated with lipid concentration (zero-excluded GLM, $p < 0.001$; **Fig. 5**).

**Discussion**

As anthropogenic climate change worsens, marine ecosystems are faced with a multitude of

stressors that may form complex associations with each other.[3,5,17] However, empirical study of

interactive stressors, particularly those occurring asynchronously, has been limited, leaving many

questions regarding the nature and mechanisms of associations amongst stressors.[92] Cnidarians,

which face a diversity of climate related stressors[21,80,93,94], provide a relevant model to investigate

the effects of multiple stressors (both interactive and sequential) on marine species. Here we

conducted one of the first experimental studies investigating the consequences of prior heat

stress on subsequent pathogen susceptibility. Our results support a long-held theory[95] that

previous heat stress increases later pathogen susceptibility and provide important new insight

regarding these patterns. Specifically, we highlight changes in host cnidarian immunity

following heat stress which may be driven by shifting resource allocation and likely contribute to

observed patterns of differential pathogen susceptibility. These results are an essential first step

towards understanding the cellular mechanisms which contribute to associations between

sequential stressors.

*Mortality differs across temperature treatments and clonal lines*

Susceptibility to pathogen exposure was significantly higher among anemones exposed to

heat stress. This result is in agreement with documented increases in disease and disease-related

mortality following coral bleaching in the wild.[24,40,44,96] Notably, while previous laboratory

experiments have documented impacts of simultaneous heat stress and pathogen

exposure[45,47,50,51,97] ours is the first to document the prolonged consequences of heat stress on

pathogen susceptibility, more accurately representing observed environmental patterns in nature.

Future studies combining our approach with emergent experimental disease techniques[98,99] will

allow for more broadscale applications to complex cnidarian and marine disease systems.

We also observed significant genotypic differences in mortality. Anemones from the VWB9

clonal line were less susceptible to pathogen-induced mortality than H2. Both clonal lines

originate from similar geographic regions and host the same species of algal endosymbiont

(*Brevolium minutum*)[100], suggesting some degree of fine scale genetic, epigenetic, or plastic

variation contribute to observed physiological differences between the lines. While intraspecific

variation in cnidarian disease resistance has been linked to genetic variation in the wild[101,102],

there are a lack of studies investigating this variation in laboratory settings.[103] Similarly, most

experimental studies investigating pathogen dynamics in Aiptasia use only a single clonal

line.[45,104-106] Our findings suggest that inclusion of genetic variation in cnidarian studies,

including those involving *E. diaphana*, may provide increased insight regarding the roles of

genetic variability in contributing to stress resistance or susceptibility.

*Mechanisms of differential mortality: roles of symbiont density and energetic reserves*

Beyond documenting differences in mortality as a result of environmental and genetic

factors, we employed an integrative approach to investigate the underlying mechanisms driving

these differences. Given the well-known contributions of symbiont density and energetic

reserves to cnidarian host immunity and pathogen susceptibility[65,107,108], as well as the

documented effects of heat stress on these metrics[57,109-113], we sought to examine differences in

symbiont density and energetic reserves between temperature treatments. Contrary to expected

results, symbiont density was higher in previously heat stressed individuals following two weeks

of recovery, potentially due to *E. diaphana*'s capacity for exceptionally rapid recovery from

dysbiosis.[75] Thus, the two-week recovery between heat stress and pathogen challenge was likely

ample time for anemones to recover from any dysbiosis and even overcompensate. Elevated

symbiont densities in previously heat stressed individuals may be the result of unregulated

symbiont growth during recovery. As negative associations have frequently been observed

between symbiont density and immunity in other cnidarians[64-66], the unchecked growth of

symbionts during recovery may have contributed to differential mortality. Consistent with this

hypothesis, H2 anemones, which were more susceptible to pathogens, also had consistently

higher symbiont densities. In contrast to symbiont density, carbohydrate concentration was

reduced in those anemones which had undergone previous heat stress. Thus, while symbiont

density appears to recover quickly, carbohydrate concentration appears to recover more slowly,

potentially contributing to increased pathogen susceptibility in previously heat stressed

individuals. Time series and finer-scale approaches leveraging metabolomic or carbon-flux

modeling are needed to determine the prolonged effects of heat stress on host energetic budget,

and subsequent effects on host pathogen susceptibility.

*Mechanisms of differential mortality: roles of host immunity*

Next, we focused on the potential roles of shifts in immunological activity which might

contribute to observed patterns of mortality. When considering immunity using multivariate
statistics, we saw significant effects of temperature, carbohydrate and lipid concentration on
immune response. Particularly, we saw indication of immune suppression in anemones that had
undergone prior heat stress. Examination of the effects of heat stress on individual immune
parameters revealed strong suppression of catalase activity, total phenoloxidase activity, and
melanin concentration. While numerous studies have evidenced the short term effects of heat
stress (days – weeks) on cnidarian host immunity[48,114-116] the long-term effects of heat stress
(months – years) on immune activity are not well documented.[117,118] Many studies have noted
heat-induced activation of a variety of immune parameters, including antioxidant (e.g., catalase)
and phenoloxidase enzymes[48,114,117], though some studies report minimal or negative impacts of
heat on immunity.[52-54] Positive associations between heat stress and immune responses during
simultaneous stressors may be driven by overlap between response mechanisms; some heat stress
biomarkers (ie HSP70) are known to activate immune pathways such as prophenoloxidase.[119,120]
In contrast to the synergy which often occurs between heat stress and immune activation during
simultaneous stressors, we document pronounced suppression of immunity (catalase,
phenoloxidase, melanin) after two weeks of recovery from heat stress. Our approach suggests
significant immune suppression occurs during, or as a result of, recovery from heat stress, though
the exact timing is unclear.  Frequent, repeated sampling during heat stress and recovery will
improve understanding of the prolonged effects of increasing SSTs on cnidarian immunity.

*Investigating the role of symbiotic immunosuppression*

While immune suppression following heat stress may be the result a number of mechanisms,
one of the most commonly proposed mechanisms involves dynamic changes in symbiosis. While
our observation of higher symbiont density in groups more susceptible to pathogen challenge
(i.e. previously heat stressed and H2 anemones) is consistent with this hypothesis, we did not
document any significant associations between symbiont density and our measured metrics of
immunity. Two potential explanations exist: 1) we measured a limited suite of immune metrics;
symbiont-induced immune suppression may have effects on unmeasured metrics, or 2) indirect
or time lagged effects of changing symbiont densities may be influencing observed patterns of
differential mortality. Specifically, rapid and excessive symbiont proliferation (evidenced by
higher symbiont densities in heat treated anemones) may have had significant prolonged effects
on host immunity that are not detected at the time of sampling (i.e. time lagged effects). Future

studies incorporating more immune parameters and time-series approaches to sampling may be
used to test these hypotheses.

*Investigating the role of energetic immunosuppression and the importance of resource*
*allocation*

A second potential hypothesis to explain observed associations between heat stress and
immunity is that changes in resource allocation as a result of resource limitations during
dysbiosis may result in immune suppression. The strongest support for this hypothesis comes
from our analyses involving carbohydrate concentrations, which were significantly reduced as a
result of prior heat stress. Additionally, carbohydrate concentration was strongly positively
associated with both catalase activity and melanin concentration, both of which were more
active/abundant in samples from the ambient group. Combined, these data suggest that
reductions in carbohydrate concentrations due to prior heat stress have direct effects on
immunity; as energetic reserves decrease in response to prior heat-stress certain immune
responses appear suppressed. These findings are consistent with synthesis of previous findings.
Heat stress both causes reallocation of host energetic reserves to stress responses, and often
impacts overall energetic reserves through the reduction in symbiont density which provide
nutritional resources to hosts.[113,121-126] This loss of energetic resources likely results in a decrease
in the total energy available to allocate to immune defenses, consistent with resource allocation
theory.[61,62,127]

Associations between lipid and melanin concentrations further support the notion that
changes in energy affect host immunity. Melanin concentration was also significantly positively
correlated to lipid concentration, which was not affected by prior heat stress. Reduced effects of
prior heat stress on lipid compared to carbohydrate are consistent with previous findings; stored
carbohydrates, rather than lipids, are preferentially used as an energy source by hosts during
bleaching recovery in other cnidarians.[113] The association between melanin and lipids, which
were relatively unaffected by prior heat stress, suggests that melanin production potentially
draws resources from both lipids and carbohydrate sources. The use of both lipids and
carbohydrates to fuel melanin production could account for comparatively reduced impacts of
heat on this metric compared to other immune parameters like catalase. Together, our results
suggest that changes in total energetic resources directly impact host immunity in the aftermath
of heat stress.

*Conclusion & Future Directions*

Multiple stressors are a common consequence of anthropogenic climate change and can have

pervasive effects on a variety of marine ecosystems.[17] As such, an improved understanding of

organismal responses to climate-induced stressors is imperative to the conservation of these

ecosystems. Here we detail one of the first studies to link prior heat stress and increased disease

susceptibility in a marine species. Our study establishes the ability to reproduce heat stress and

disease associations under laboratory conditions using a tractable model system, allowing for

more nuanced and robust investigations of ecologically relevant phenomena. Using this

experimental approach, we were able to demonstrate prolonged effects of prior heat stress on

host immunity, which may contribute to observed increased pathogen-induced mortality

following recovery from heat stress. Furthermore, our results implicate changes in resource

availability as a potential mechanism of heat-induced immunosuppression. Further studies

incorporating time series approaches will be necessary to fully disentangle the observed patterns

and associated mechanisms and assess their broad applicability to marine systems. Specifically,

future work should aim to sample across multiple timepoints from the onset of bleaching through

acute and long-term recovery to determine the effects of dynamic changes in symbiont density

and energetic reserves on host immunity and pathogen susceptibility. This, coupled with more

nuanced approaches such as fine-scale metabolomics and transcriptomics, should provide a

clearer picture of the nuances of changing resource allocation, symbiont density, and immunity

which link prior heat stress to increased pathogen susceptibility in corals. In summary, our

results provide an important starting point for understanding the links between heat stress and

disease that will aid in the conservation of marine species, including cnidarians, in the face of a

changing climate. Most importantly, they generate testable hypotheses that will be invaluable in

better understanding the effects of multiple stressors on diverse species and ecosystem functions.

**Data and Code Availability**

All raw data and R code used in described analyses is supplied as supplementary files.

**Conflict of Interest**

The authors declare they have no competing interests.

## References Cited

Rosenzweig, C. *et al.* Attributing physical and biological impacts to anthropogenic climate change. *Nature* **453**, 353-357, doi:10.1038/nature06937 (2008).

Hansen, G. & Stone, D. Assessing the observed impact of anthropogenic climate change. *Nature Climate Change* **6**, 532-537, doi:10.1038/nclimate2896 (2015).

Doney, S. C. *et al.* Climate Change Impacts on Marine Ecosystems. *Annual Review of Marine Science* **4**, 11-37, doi:10.1146/annurev-marine-041911-111611 (2011).

Scavia, D. *et al.* Climate change impacts on U.S. Coastal and Marine Ecosystems. *Estuaries* **25**, 149-164, doi:10.1007/BF02691304 (2002).

Hoegh-Guldberg, O. & Bruno, J. F. The impact of climate change on the world's marine ecosystems. *Science* **328**, 1523-1528, doi:10.1126/science.1189930 (2010).

Smith, K. E. *et al.* Biological Impacts of Marine Heatwaves. *Annual Review of Marine Science* **15**, 119-145, doi:10.1146/annurev-marine-032122-121437 (2023).

Genin, A., Levy, L., Sharon, G., Raitsos, D. E. & Diamant, A. Rapid onsets of warming events trigger mass mortality of coral reef fish. *Proc Natl Acad Sci U S A* **117**, 25378-25385, doi:10.1073/pnas.2009748117 (2020).

Wild, S. *et al.* Long-term decline in survival and reproduction of dolphins following a marine heatwave. *Curr Biol* **29**, R239-R240, doi:10.1016/j.cub.2019.02.047 (2019).

Rivetti, I., Fraschetti, S., Lionello, P., Zambianchi, E. & Boero, F. Global warming and mass mortalities of benthic invertebrates in the Mediterranean Sea. *PLOS ONE* **9**, e115655, doi:10.1371/journal.pone.0115655 (2014).

Smale, D. A. Impacts of ocean warming on kelp forest ecosystems. *New Phytol* **225**, 1447-1454, doi:10.1111/nph.16107 (2020).

Bateman, K. S., Feist, S. W., Bignell, J. P., Bass, D. & Stentiford, G. D. in *Marine Disease Ecology*   0 (Oxford University Press, 2020).

Sanderson, C. E. & Alexander, K. A. Unchartered waters: Climate change likely to intensify infectious disease outbreaks causing mass mortality events in marine mammals. *Glob Chang Biol* **26**, 4284-4301, doi:10.1111/gcb.15163 (2020).

Patterson, K. R. Modelling the impact of disease-induced mortality in an exploited population: the outbreak of the fungal parasite ( Ichthyophonus)(hoferi) in the North Sea herring (Clupea harengus). *Canadian Journal of Fisheries and Aquatic Sciences* **53**, 2870-2887, doi:10.1139/f96-234 (1996).

Hamilton, S. L. *et al.* Disease-driven mass mortality event leads to widespread extirpation and variable recovery potential of a marine predator across the eastern Pacific. *Proc Biol Sci* **288**, 20211195, doi:10.1098/rspb.2021.1195 (2021).

Short, F. T., Muehlstein, L. K. & Porter, D. EELGRASS WASTING DISEASE: CAUSE AND RECURRENCE OF A MARINE EPIDEMIC. *The Biological Bulletin* **173**, 557-562, doi:10.2307/1541701 (1987).

Adams, S. M. Assessing cause and effect of multiple stressors on marine systems. *Marine Pollution Bulletin* **51**, 649-657, doi:https://doi.org/10.1016/j.marpolbul.2004.11.040 (2005).

Crain, C. M., Kroeker, K. & Halpern, B. S. Interactive and cumulative effects of multiple human stressors in marine systems. *Ecol Lett* **11**, 1304-1315, doi:10.1111/j.1461-0248.2008.01253.x (2008).

Harvell, C. D. *et al.* Climate Warming and Disease Risks for Terrestrial and Marine Biota. *Science* **296**, 2158-2162, doi:10.1126/science.1063699 (2002).

19   Harvell, C. D. *et al.* Disease epidemic and a marine heat wave are associated with the
continental-scale collapse of a pivotal predator (Pycnopodia helianthoides). *Science*
*Advances* **5**, eaau7042, doi:10.1126/sciadv.aau7042.
20   Olsen, Y. S. & Duarte, C. M. Combined effect of warming and infection by Labyrinthula
sp. on the Mediterranean seagrass Cymodocea nodosa. *Marine Ecology Progress Series*
**532**, 101-109, doi:10.3354/meps11343 (2015).
21   Ban, S. S., Graham, N. A. & Connolly, S. R. Evidence for multiple stressor interactions
and effects on coral reefs. *Glob Chang Biol* **20**, 681-697, doi:10.1111/gcb.12453 (2014).
22   Groner, M. L. *et al.* Warming sea surface temperatures fuel summer epidemics of
eelgrass wasting disease. *Marine Ecology Progress Series* **679**, 47-58 (2021).
23   Oliver, E. C. J. *et al.* The unprecedented 2015/16 Tasman Sea marine heatwave. *Nat*
*Commun* **8**, 16101, doi:10.1038/ncomms16101 (2017).
24   Miller, J. *et al.* Coral disease following massive bleaching in 2005 causes 60% decline in
coral cover on reefs in the US Virgin Islands. *Coral Reefs* **28**, 925, doi:10.1007/s00338-
009-0531-7 (2009).
25   Croquer, A. & Weil, E. Changes in Caribbean coral disease prevalence after the 2005
bleaching event. *Dis Aquat Organ* **87**, 33-43, doi:10.3354/dao02164 (2009).
26   Smith, S. V. & Buddemeier, R. W. Global Change and Coral Reef Ecosystems. *Annu.*
*Rev. Ecol. Syst.* **23**, 89-118 (1992).
27   Gardner, T. A., Côté, I. M., Gill, J. A., Grant, A. & Watkinson, A. R. Long-Term Region-
Wide Declines in Caribbean Corals. *Science* **301**, 958-960, doi:10.1126/science.1086050
(2003).
28   Bruno, J. F. & Selig, E. R. Regional Decline of Coral Cover in the Indo-Pacific: Timing,
Extent, and Subregional Comparisons. *PLOS ONE* **2**, e711,
doi:10.1371/journal.pone.0000711 (2007).
29   Brown, B. E. Coral bleaching: causes and consequences. *Coral Reefs* **16**, S129-S138,
doi:10.1007/s003380050249 (1997).
30   Douglas, A. E. Coral bleaching—how and why? *Marine Pollution Bulletin* **46**, 385-392,
doi:10.1016/S0025-326X(03)00037-7 (2003).
31   Harvell, D. *et al.* Coral disease, environmental drivers, and the balance between coral and
microbial associates. *Oceanography* **20**, 172-195 (2007).
32   Johnston, M. A. *et al.* Coral bleaching and recovery from 2016 to 2017 at East and West
Flower Garden Banks, Gulf of Mexico. *Coral Reefs* **38**, 787-799, doi:10.1007/s00338-
019-01788-7 (2019).
33   Winter, A., Appeldoorn, R., Bruckner, A., Williams Jr, E. & Goenaga, C. Sea surface
temperatures and coral reef bleaching off La Parguera, Puerto Rico (northeastern
Caribbean Sea). *Coral Reefs* **17**, 377-382 (1998).
34   McClanahan, T. R., Ateweberhan, M., Muhando, C. A., Maina, J. & Mohammed, M. S.
Effects of Climate and Seawater Temperature Variation on Coral Bleaching and
Mortality. *Ecological Monographs* **77**, 503-525, doi:10.1890/06-1182.1 (2007).
35   Richardson, L. L., Goldberg, W. M., Carlton, R. G. & Halas, J. C. Coral disease outbreak
in the Florida Keys: plague type II. *Revista de Biología tropical* **46**, 187-198 (1998).
36   Bruckner, A. W., Bruckner, R. J. & Williams Jr, E. H. Spread of a black-band disease
epizootic through the coral reef system in St. Ann's Bay, Jamaica. *Bulletin of Marine*
*Science* **61**, 919-928 (1997).
37   Dahlgren, C., Pizarro, V., Sherman, K., Greene, W. & Oliver, J. Spatial and temporal

patterns of stony coral tissue loss disease outbreaks in the Bahamas. *Frontiers in Marine*
*Science* **8**, 682114 (2021).
38    Gil-Agudelo, D., Smith, G. & Weil, E. The white band disease type II pathogen in Puerto
Rico. *Revista de biologia tropical* **54**, 59-67 (2006).
39    Brandt, M. E. & McManus, J. W. Disease incidence is related to bleaching extent in reef-
building corals. *Ecology* **90**, 2859-2867, doi:10.1890/08-0445.1 (2009).
40    Harvell, D., Kim, K., Quirolo, C., Weir, J. & Smith, G. Coral bleaching and disease:
contributors to 1998 mass mortality in Briareum asbestinum (Octocorallia, Gorgonacea).
*Hydrobiologia* **460**, 97-104, doi:10.1023/A:1013169331913 (2001).
41    Howells, E. J., Vaughan, G. O., Work, T. M., Burt, J. A. & Abrego, D. Annual outbreaks
of coral disease coincide with extreme seasonal warming. *Coral Reefs* **39**, 771-781,
doi:10.1007/s00338-020-01946-2 (2020).
42    Burke, S. *et al.* The impact of rising temperatures on the prevalence of coral diseases and
its predictability: A global meta-analysis. *Ecology Letters* **26**, 1466-1481,
doi:https://doi.org/10.1111/ele.14266 (2023).
43    Jones, N. P., Kabay, L., Semon Lunz, K. & Gilliam, D. S. Temperature stress and disease
drives the extirpation of the threatened pillar coral, Dendrogyra cylindrus, in southeast
Florida. *Scientific Reports* **11**, 14113, doi:10.1038/s41598-021-93111-0 (2021).
44    Miller, J., Waara, R., Muller, E. & Rogers, C. Coral bleaching and disease combine to
cause extensive mortality on reefs in US Virgin Islands. *Coral Reefs* **25**, 418-418,
doi:10.1007/s00338-006-0125-6 (2006).
45    Zaragoza, W. J. *et al.* Outcomes of infections of sea anemone Aiptasia pallida with
Vibrio spp. pathogenic to corals. *Microb Ecol* **68**, 388-396, doi:10.1007/s00248-014-
0397-2 (2014).
46    Cervino, J. M. *et al.* Relationship of Vibrio species infection and elevated temperatures to
yellow blotch/band disease in Caribbean corals. *Appl Environ Microbiol* **70**, 6855-6864
(2004).
47    Vidal-Dupiol, J. *et al.* Thermal stress triggers broad Pocillopora damicornis
transcriptomic remodeling, while Vibrio coralliilyticus infection induces a more targeted
immuno-suppression response. *PLoS One* **9**, e107672, doi:10.1371/journal.pone.0107672
(2014).
48    Palmer, C. V. Warmer Water Affects Immunity of a Tolerant Reef Coral. *Frontiers in*
*Marine Science* **5** (2018).
49    Zhou, Z. *et al.* Suppression of NF-kappaB signal pathway by NLRC3-like protein in
stony coral Acropora aculeus under heat stress. *Fish Shellfish Immunol* **67**, 322-330,
doi:10.1016/j.fsi.2017.06.027 (2017).
50    Bisanti, L. *et al.* How does warmer sea water change the sensitivity of a Mediterranean
thermophilic coral after immune-stimulation? *Coral Reefs* **43**, 137-150,
doi:10.1007/s00338-023-02454-9 (2024).
51    Takagi, T., Yoshioka, Y., Zayasu, Y., Satoh, N. & Shinzato, C. Transcriptome Analyses
of Immune System Behaviors in Primary Polyp of Coral Acropora digitifera Exposed to
the Bacterial Pathogen Vibrio coralliilyticus under Thermal Loading. *Mar Biotechnol*
*(NY)* **22**, 748-759, doi:10.1007/s10126-020-09984-1 (2020).
52    van de Water, J. *et al.* Antimicrobial and stress responses to increased temperature and
bacterial pathogen challenge in the holobiont of a reef-building coral. *Mol Ecol* **27**, 1065-
1080, doi:10.1111/mec.14489 (2018).

53    Eliachar, S. *et al.* Heat stress increases immune cell function in Hexacorallia. *Front*
*Immunol* **13**, 1016097, doi:10.3389/fimmu.2022.1016097 (2022).
54    van de Water, J. A. J. M. *et al.* Elevated seawater temperatures have a limited impact on
the coral immune response following physical damage. *Hydrobiologia* **759**, 201-214,
doi:10.1007/s10750-015-2243-z (2015).
55    Tremblay, P., Naumann, M. S., Sikorski, S., Grover, R. & Ferrier-Pagès, C. Experimental
assessment of organic carbon fluxes in the scleractinian coral Stylophora pistillata during
a thermal and photo stress event. *Marine Ecology Progress Series* **453**, 63-77,
doi:10.3354/meps09640 (2012).
56    Grottoli, A. G., Rodrigues, L. J. & Palardy, J. E. Heterotrophic plasticity and resilience in
bleached corals. *Nature* **440**, 1186-1189, doi:10.1038/nature04565 (2006).
57    Rodrigues, L. J. & Grottoli, A. G. Energy reserves and metabolism as indicators of coral
recovery from bleaching. *Limnology and Oceanography* **52**, 1874-1882,
doi:10.4319/lo.2007.52.5.1874 (2007).
58    Ng'oma, E., Perinchery, A. M. & King, E. G. How to get the most bang for your buck:
the evolution and physiology of nutrition-dependent resource allocation strategies.
*Proceedings of the Royal Society B: Biological Sciences* **284**, 20170445,
doi:10.1098/rspb.2017.0445 (2017).
59    Zera, A. J. & Harshman, L. G. The Physiology of Life History Trade-Offs in Animals.
*Annual Review of Ecology and Systematics* **32**, 95-126,
doi:10.1146/annurev.ecolsys.32.081501.114006 (2001).
60    Garland, T., Downs, C. J. & Ives, A. R. Trade-Offs (and Constraints) in Organismal
Biology. *Physiological and Biochemical Zoology* **95**, 82-112, doi:10.1086/717897
(2022).
61    Rauw, W. M. Immune response from a resource allocation perspective. *Front Genet* **3**,
267, doi:10.3389/fgene.2012.00267 (2012).
62    Houston, A. I., McNamara, J. M., Barta, Z. & Klasing, K. C. The effect of energy
reserves and food availability on optimal immune defence. *Proc Biol Sci* **274**, 2835-2842,
doi:10.1098/rspb.2007.0934 (2007).
63    McNamara, J. M. & Buchanan, K. L. Stress, resource allocation, and mortality.
*Behavioral Ecology* **16**, 1008-1017, doi:10.1093/beheco/ari087 (2005).
64    Mansfield, K. M. & Gilmore, T. D. Innate immunity and cnidarian-Symbiodiniaceae
mutualism. *Dev Comp Immunol* **90**, 199-209, doi:10.1016/j.dci.2018.09.020 (2019).
65    Fuess, L. E., Palacio-Castro, A. M., Butler, C. C., Baker, A. C. & Mydlarz, L. D.
Increased Algal Symbiont Density Reduces Host Immunity in a Threatened Caribbean
Coral Species, Orbicella faveolata. *Frontiers in Ecology and Evolution* **8**,
doi:10.3389/fevo.2020.572942 (2020).
66    Berthelier, J. *et al.* Implication of the host TGFβ pathway in the onset of symbiosis
between larvae of the coral Fungia scutaria and the dinoflagellate Symbiodinium sp.
(clade C1f). *Coral Reefs* **36**, 1263-1268, doi:10.1007/s00338-017-1621-6 (2017).
67    Detournay, O., Schnitzler, C. E., Poole, A. & Weis, V. M. Regulation of cnidarian-
dinoflagellate mutualisms: Evidence that activation of a host TGFbeta innate immune
pathway promotes tolerance of the symbiont. *Dev Comp Immunol* **38**, 525-537,
doi:10.1016/j.dci.2012.08.008 (2012).
68    Jacobovitz, M. R. *et al.* Dinoflagellate symbionts escape vomocytosis by host cell
immune suppression. *Nat Microbiol* **6**, 769-782, doi:10.1038/s41564-021-00897-w

(2021).
69  Weis, V. M., Davy, S. K., Hoegh-Guldberg, O., Rodriguez-Lanetty, M. & Pringle, J. R.
Cell biology in model systems as the key to understanding corals. *Trends Ecol Evol* **23**,
369-376, doi:10.1016/j.tree.2008.03.004 (2008).
70  LaJeunesse, T. C. *et al.* Systematic Revision of Symbiodiniaceae Highlights the
Antiquity and Diversity of Coral Endosymbionts. *Curr Biol* **28**, 2570-2580.e2576,
doi:10.1016/j.cub.2018.07.008 (2018).
71  Voolstra, C. R. A journey into the wild of the cnidarian model system Aiptasia and its
symbionts. *Mol Ecol* **22**, 4366-4368, doi:10.1111/mec.12464 (2013).
72  Bellis, E. S. & Denver, D. R. Natural variation in responses to acute heat and cold stress
in a sea anemone model system for coral bleaching. *The Biological Bulletin* **233**, 168-181
(2017).
73  Cleves, P. A., Krediet, C. J., Lehnert, E. M., Onishi, M. & Pringle, J. R. Insights into
coral bleaching under heat stress from analysis of gene expression in a sea anemone
model system. *Proceedings of the National Academy of Sciences* **117**, 28906-28917
(2020).
74  Ben-Haim, Y. *et al.* Vibrio coralliilyticus sp. nov., a temperature-dependent pathogen of
the coral Pocillopora damicornis. *International journal of systematic and evolutionary*
*microbiology* **53**, 309-315 (2003).
75  Tortorelli, G., Belderok, R., Davy, S. K., McFadden, G. I. & van Oppen, M. J. H. Host
Genotypic Effect on Algal Symbiosis Establishment in the Coral Model, the Anemone
Exaiptasia diaphana, From the Great Barrier Reef. *Frontiers in Marine Science* **6** (2020).
76  Pinzón C, J. H., Beach-Letendre, J., Weil, E. & Mydlarz, L. D. Relationship between
Phylogeny and Immunity Suggests Older Caribbean Coral Lineages Are More Resistant
to Disease. *PLOS ONE* **9**, e104787, doi:10.1371/journal.pone.0104787 (2014).
77  Mydlarz, L. D. & Palmer, C. V. The presence of multiple phenoloxidases in Caribbean
reef-building corals. *Comparative Biochemistry and Physiology Part A: Molecular &*
*Integrative Physiology* **159**, 372-378, doi:https://doi.org/10.1016/j.cbpa.2011.03.029
(2011).
78  Parisi, M. G., Parrinello, D., Stabili, L. & Cammarata, M. Cnidarian Immunity and the
Repertoire of Defense Mechanisms in Anthozoans. *Biology (Basel)* **9**, 283,
doi:10.3390/biology9090283 (2020).
79  Mydlarz, L. D., Fuess, L., Mann, W., Pinzón, J. H. & Gochfeld, D. J. Cnidarian
immunity: from genomes to phenomes. *The cnidaria, past, present and future: the world*
*of medusa and her sisters*, 441-466 (2016).
80  Palmer, C. V. & Traylor-Knowles, N. G. in *Advances in Comparative Immunology*  (ed
Edwin L. Cooper)  51-93 (Springer International Publishing, 2018).
81  Lesser, M. P. Using energetic budgets to assess the effects of environmental stress on
corals: are we measuring the right things? *Coral Reefs* **32**, 25-33, doi:10.1007/s00338-
012-0993-x (2013).
82  Fitt, W. K., McFarland, F. K., Warner, M. E. & Chilcoat, G. C. Seasonal patterns of
tissue biomass and densities of symbiotic dinoflagellates in reef corals and relation to
coral bleaching. *Limnology and Oceanography* **45**, 677-685,
doi:https://doi.org/10.4319/lo.2000.45.3.0677 (2000).
83  Cheng, Y.-S., Zheng, Y. & VanderGheynst, J. S. Rapid quantitative analysis of lipids
using a colorimetric method in a microplate format. *Lipids* **46**, 95-103,

774    doi:10.1007/s11745-010-3494-0 (2011).

775 84  Masuko, T. *et al.* Carbohydrate analysis by a phenol-sulfuric acid method in microplate
776    format. *Anal Biochem* **339**, 69-72, doi:10.1016/j.ab.2004.12.001 (2005).

777 85  R: A language and environment for statistical computing (R Foundation for Statistical
778    Computing, Vienna, Austria, 2021).

779 86  Millard, S. P. *EnvStats: An R Package for Environmental Statistics*. (Springer, 2013).

780 87  A Package for Survival Analysis in R (2024).

781 88  Johnson, L. L. & Shih, J. H. in *Principles and Practice of Clinical Research* 273-282
782    (2007).

783 89  Christensen, E. Multivariate survival analysis using Cox's regression model. *Hepatology*
784    **7**, 1346-1358, doi:10.1002/hep.1840070628 (1987).

785 90  Oksanen, J. *et al.* vegan: Community Ecology Package. (2024).

786 91  MuMIn: Multi-Model Inference (2023).

787 92  Pirotta, E. *et al.* Understanding the combined effects of multiple stressors: A new
788    perspective on a longstanding challenge. *Sci Total Environ* **821**, 153322,
789    doi:10.1016/j.scitotenv.2022.153322 (2022).

790 93  Maor-Landaw, K. & Levy, O. in *The Cnidaria, Past, Present and Future: The world of*
791    *Medusa and her sisters* (eds Stefano Goffredo & Zvy Dubinsky) 523-543 (Springer
792    International Publishing, 2016).

793 94  Weil, E., Weil-Allen, A. & Weil, A. Coral and Cnidarian Welfare in a Changing Sea. *The*
794    *Welfare of Invertebrate Animals*, 123-145 (2019).

795 95  Muller, E. M., Bartels, E. & Baums, I. B. Bleaching causes loss of disease resistance
796    within the threatened coral species Acropora cervicornis. *eLife* **7**, e35066,
797    doi:10.7554/eLife.35066 (2018).

798 96  Precht, W. F., Gintert, B. E., Robbart, M. L., Fura, R. & van Woesik, R. Unprecedented
799    Disease-Related Coral Mortality in Southeastern Florida. *Scientific Reports* **6**, 31374,
800    doi:10.1038/srep31374 (2016).

801 97  Palmer, C. V. *et al.* Patterns of coral ecological immunology: variation in the responses
802    of Caribbean corals to elevated temperature and a pathogen elicitor. *J Exp Biol* **214**,
803    4240-4249, doi:10.1242/jeb.061267 (2011).

804 98  Traylor-Knowles, N. *et al.* Gene Expression Response to Stony Coral Tissue Loss
805    Disease Transmission in M. cavernosa and O. faveolata From Florida. *Frontiers in*
806    *Marine Science* **8**, doi:10.3389/fmars.2021.681563 (2021).

807 99  Huntley, N. *et al.* Experimental transmission of Stony Coral Tissue Loss Disease results
808    in differential microbial responses within coral mucus and tissue. *ISME Communications*
809    **2**, 46, doi:10.1038/s43705-022-00126-3 (2022).

810 100  Thornhill, D. J., Xiang, Y., Pettay, D. T., Zhong, M. & Santos, S. R. Population genetic
811    data of a model symbiotic cnidarian system reveal remarkable symbiotic specificity and
812    vectored introductions across ocean basins. *Mol Ecol* **22**, 4499-4515,
813    doi:10.1111/mec.12416 (2013).

814 101  Vollmer, S. V. & Kline, D. I. Natural Disease Resistance in Threatened Staghorn Corals.
815    *PLOS ONE* **3**, e3718, doi:10.1371/journal.pone.0003718 (2008).

816 102  Miller, M. W. *et al.* Genotypic variation in disease susceptibility among cultured stocks
817    of elkhorn and staghorn corals. *PeerJ* **7**, e6751, doi:10.7717/peerj.6751 (2019).

818 103  Wright, R. M. *et al.* Intraspecific differences in molecular stress responses and coral
819    pathobiome contribute to mortality under bacterial challenge in Acropora millepora |

Scientific Reports. *Scientific Reports* **7** (2017).

104  Brown, T. & Rodriguez-Lanetty, M. Defending against pathogens – immunological priming and its molecular basis in a sea anemone, cnidarian. *Scientific Reports* **5**, 17425, doi:10.1038/srep17425 (2015).

105  Seneca, F., Davtian, D., Boyer, L. & Czerucka, D. Gene expression kinetics of Exaiptasia pallida innate immune response to Vibrio parahaemolyticus infection. *BMC Genomics* **21**, 768, doi:10.1186/s12864-020-07140-6 (2020).

106  Valadez-Ingersoll, M. *et al.* Starvation differentially affects gene expression, immunity and pathogen susceptibility across symbiotic states in a model cnidarian. *Proceedings of the Royal Society B* **291**, 20231685 (2024).

107  Changsut, I., Womack, H. R., Shickle, A., Sharp, K. H. & Fuess, L. E. Variation in symbiont density is linked to changes in constitutive immunity in the facultatively symbiotic coral, *Astrangia poculata*. *Biology Letters* **18**, 20220273, doi:doi:10.1098/rsbl.2022.0273 (2022).

108  Valadez-Ingersoll, M. *et al.* Nutrient deprivation differentially affects gene expression, immunity, and pathogen susceptibility across symbiotic states in a model cnidarian. *bioRxiv*, 2023.2007. 2030.551141 (2023).

109  Jokiel, P. L. & Coles, S. L. Response of Hawaiian and other Indo-Pacific reef corals to elevated temperature. *Coral Reefs* **8**, 155-162, doi:10.1007/bf00265006 (1990).

110  Douglas, A. E. Coral bleaching--how and why? *Mar Pollut Bull* **46**, 385-392, doi:10.1016/S0025-326X(03)00037-7 (2003).

111  Imbs, A. B. & Yakovleva, I. M. Dynamics of lipid and fatty acid composition of shallow-water corals under thermal stress: an experimental approach. *Coral Reefs* **31**, 41-53, doi:10.1007/s00338-011-0817-4 (2011).

112  Ermolenko, E. V. & Sikorskaya, T. V. Lipidome of the reef-building coral Acropora cerealis: Changes under thermal stress. *Biochemical Systematics and Ecology* **97**, doi:10.1016/j.bse.2021.104276 (2021).

113  Kochman, N. a.-R., Grover, R., Rottier, C., Ferrier-Pages, C. & Fine, M. The reef building coral Stylophora pistillata uses stored carbohydrates to maintain ATP levels under thermal stress. *Coral Reefs* **40**, 1473-1485, doi:10.1007/s00338-021-02174-y (2021).

114  Krueger, T. *et al.* Differential coral bleaching—Contrasting the activity and response of enzymatic antioxidants in symbiotic partners under thermal stress. *Comparative Biochemistry and Physiology Part A: Molecular & Integrative Physiology* **190**, 15-25, doi:https://doi.org/10.1016/j.cbpa.2015.08.012 (2015).

115  Palmer, C. V. *et al.* Patterns of coral ecological immunology: variation in the responses of Caribbean corals to elevated temperature and a pathogen elicitor. *Journal of Experimental Biology* **214**, 4240-4249, doi:10.1242/jeb.061267 (2011).

116  Wall, C. B. *et al.* The effects of environmental history and thermal stress on coral physiology and immunity. *Marine Biology* **165**, 56, doi:10.1007/s00227-018-3317-z (2018).

117  Wall, C. B. *et al.* Shifting baselines: Physiological legacies contribute to the response of reef corals to frequent heatwaves. *Functional Ecology* **35**, 1366-1378, doi:10.1111/1365-2435.13795 (2021).

118  Mydlarz, L. D., Couch, C. S., Weil, E., Smith, G. & Harvell, C. D. Immune defenses of healthy, bleached and diseased Montastraea faveolata during a natural bleaching event.

*Diseases of Aquatic Organisms* **87**, 67-78, doi:10.3354/dao02088 (2009).

119  Baruah, K., Norouzitallab, P., Linayati, L., Sorgeloos, P. & Bossier, P. Reactive oxygen
species generated by a heat shock protein (Hsp) inducing product contributes to Hsp70
production and Hsp70-mediated protective immunity in Artemia franciscana against
pathogenic vibrios. *Dev Comp Immunol* **46**, 470-479, doi:10.1016/j.dci.2014.06.004
(2014).

120  Baruah, K., Ranjan, J., Sorgeloos, P., Macrae, T. H. & Bossier, P. Priming the
prophenoloxidase system of Artemia franciscana by heat shock proteins protects against
Vibrio campbellii challenge. *Fish Shellfish Immunol* **31**, 134-141,
doi:10.1016/j.fsi.2011.04.008 (2011).

121  Anthony, K. R. N., Connolly, S. R. & Hoegh-Guldberg, O. Bleaching, energetics, and
coral mortality risk: Effects of temperature, light, and sediment regime. *Limnology and*
*Oceanography* **52**, 716-726, doi:10.4319/lo.2007.52.2.0716 (2007).

122  Ferrier-Pagès, C., Rottier, C., Beraud, E. & Levy, O. Experimental assessment of the
feeding effort of three scleractinian coral species during a thermal stress: Effect on the
rates of photosynthesis. *Journal of Experimental Marine Biology and Ecology* **390**, 118-
124, doi:10.1016/j.jembe.2010.05.007 (2010).

123  Tremblay, P., Gori, A., Maguer, J. F., Hoogenboom, M. & Ferrier-Pages, C.
Heterotrophy promotes the re-establishment of photosynthate translocation in a symbiotic
coral after heat stress. *Sci Rep* **6**, 38112, doi:10.1038/srep38112 (2016).

124  Anthony, K. R. N., Hoogenboom, M. O., Maynard, J. A., Grottoli, A. G. & Middlebrook,
R. Energetics approach to predicting mortality risk from environmental stress: a case
study of coral bleaching. *Functional Ecology* **23**, 539-550, doi:10.1111/j.1365-
2435.2008.01531.x (2009).

125  Kenkel, C. D., Meyer, E. & Matz, M. V. Gene expression under chronic heat stress in
populations of the mustard hill coral (Porites astreoides) from different thermal
environments. *Mol Ecol* **22**, 4322-4334, doi:10.1111/mec.12390 (2013).

126  Paradis, B. T., Henry, R. P. & Chadwick, N. E. Compound effects of thermal stress and
tissue abrasion on photosynthesis and respiration in the reef-building coral Acropora
cervicornis (Lamarck, 1816). *Journal of Experimental Marine Biology and Ecology* **521**,
doi:10.1016/j.jembe.2019.151222 (2019).

127  Brunner, F. S., Schmid-Hempel, P. & Barribeau, S. M. Protein-poor diet reduces host-
specific immune gene expression in Bombus terrestris. *Proc Biol Sci* **281**,
doi:10.1098/rspb.2014.0128 (2014).

**Tables and Figures**

**Table 1:** Initial anemone distribution for H2 and VWB9 conal lines across each of four
treatment combinations.

| | VWB9 | | H2 | |
|---|---|---|---|---|
| | Ambient | Heat | Ambient | Heat |
| Placebo | 86 | 86 | 86 | 86 |
| Pathogen | 86 | 86 | 86 | 86 |

906
907
908
909
910 **Table 2**: Final sample sizes for physiological and survivorship analyses.

| Survivorship | VWB9 | | H2 | |
|---|---|---|---|---|
| | Ambient | Heat | Ambient | Heat |
| Placebo | 30 | 26 | 19 | 21 |
| Infection | 30 | 25 | 20 | 13 |
| Physiology | VWB9 | | H2 | |
| | Ambient | Heat | Ambient | Heat |
| Placebo | 14 | 13 | 12 | 11 |
| Pathogen | 17 | 12 | 11 | 10 |

924 **Table 3:** Best-fit linear models for each immune parameter when including temperature
925 treatment, clonal line, immune challenge, and lipids as predictors. † indicate models that did not
926 undergo model dredging and averaging. Asterisks (*) represent significant p-value ($\alpha = 0.05$).
927

| Catalase | | | | | |
|---|---|---|---|---|---|
| **Predictors** | **Estimates** | **SE** | **SEadj** | **z** | **p value** |
| (Intercept) | 15.68 | 0.901 | 0.913 | 17.18 | **< 0.001 \*\*\*** |
| Treatment | -3.12 | 1.24 | 1.25 | 2.49 | **0.0129\*** |
| Carbs | 141.9 | 36.76 | 37.01 | 3.80 | **< 0.001 \*\*\*** |
| Lipid | 0.431 | 1.47 | 1.48 | 0.290 | 0.772 |
| Lipids : Treatment | 0.775 | 2.10 | 2.11 | 0.367 | 0.714 |
| Carbs : Treatment | 13.62 | 45.99 | 45.42 | 0.293 | 0.769 |
| **Antibacterial Activity** | | | | | |
| **Predictors** | **Estimates** | **SE** | **SEadj** | **z** | **p value** |
| (Intercept) | 12.58 | 5.50 | 5.55 | 2.27 | **0.0235\*** |
| Treatment | -0.642 | 1.699 | 1.717 | 0.374 | 0.708 |
| Carbs | 79.64 | 121.4 | 122.4 | 0.651 | 0.515 |
| Lipids | -8.11 | 9.17 | 9.26 | 0.876 | 0.381 |
| **Total Phenoloxidase** | | | | | |
| **Predictors** | **Estimates** | **SE** | **SEadj** | **z** | **p value** |
| (Intercept) | 0.0164 | 0.0016 | 0.00117 | 13.98 | **< 0.001 \*\*\*** |
| Treatment | -0.00323 | 0.00138 | 0.00140 | 2.304 | **0.0212\*** |
| Carbs | 0.0210 | 0.0514 | 0.0519 | 0.405 | 0.686 |

| Melanin† | | | |
|---|---|---|---|
| **Predictors** | $X^2$ | **DF** | **p value** |
| Treatment | 4.097 | 1 | **0.0430\*** |
| Carbs | 0.3801 | 1 | 0.5376 |
| Lipids | 1.677 | 1 | 0.1960 |

| 0s removed Melanin† | | | | |
|---|---|---|---|---|
| **Predictors** | **Estimates** | **SE** | **t** | **p value** |
| (Intercept) | -1.98E-03 | 8.98E-04 | -2.20 | **0.0321\*** |
| Treatment | -1.81E-04 | 4.62E-04 | -0.392 | 0.697 |
| Carbs | 0.0706 | 0.0215 | 3.28 | **< 0.001 \*\*\*** |
| Lipids | 6.64E-03 | 1.38E-3 | 4.68 | **< 0.001 \*\*\*** |

**Figure 1:** (a-b) Kaplan-Meier survivorship curves for pathogen exposed anemones split based on comparisons across (a) temperature treatment and (b) clonal line. (c) Cox proportional hazards models of pathogen exposed anemones compared across temperature treatment and clonal line. Vertical line at 1 indicates the reference to which the second group is compared. The distance from 1 indicates the probability of mortality (i.e. heat stress is about twice as likely to die from immune challenge as heat). Both treatment (p = 0.0255) and genotype (p < 0.001) had significant impacts on survival. Hazard ratios are displayed for both predictors and the interaction term with 95% confidence intervals.

**Figure 2:** Box plots displaying **a)** symbiont density, **b)** lipid concentration, and **c)** carbohydrate concentration by temperature treatment and clonal line. Raw data values represented by overlaid points; boxes and points are colored based on clonal line. Significance (*p values*) for treatment and clonal line is reported.

978　**Figure 3.** RDA plot (center) showing significant predictors (black labels) for explaining variation in immunity. Data points represent individual samples, and ordination was constructed using distances generated from immune parameters. Color indicates temperature treatment, shape indicates clonal line, and open shape/filled shape indicates immune challenge. Black labels represent predictor variables and white labels represent response variables. Arrows represent strength and directionality of each variable in multi-variate space. Boxplots on the top and far left indicate the distribution of heat and ambient points along the first and second axes to demonstrate significant variation explained by temperature treatment in each direction. Regressions on the bottom and far right show the correlation between carbohydrates and lipids on the first and second axes respectively to demonstrate significant explanation of variation by those variables on each axis. MEL: melanin; CAT: catalase; TPO: total phenoloxidase; AB: antibacterial activity.

**Figure 4:** Box plot displaying differences in **a)** catalase activity, **b)** total phenoloxidase activity, and **c)** melanin concentration as a result of temperature treatment, divided by clonal line. Raw data values represented by overlaid points; boxes and points are colored based on clonal line. Significance (*p values*) for treatment is reported.

**Figure 5:** Best-fit linear models of the association between individual immune assays and energetic metrics: **a)** catalase activity and carbohydrate concentration, **b)** melanin concentration and carbohydrate concentration, and **c)** melanin concentration and lipid concentration Points represent individual samples color coded by clonal line. Trendlines represent linear models of association between immune and energetic metrics (independent of clonal lines). Significance (*p values*) for associations are reported.