Soil health increases primary productivity across Europe

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Abstract

A healthy soil is at the core of sustainable management and policy, but its importance for plant productivity across environmental gradients and land-use types remains poorly understood. To address this gap, we conducted a pan-European field study including 588 sites from 27 countries to investigate the link between soil health and primary productivity across three major land-use types: woodlands, grasslands, and croplands. We found that mean soil health (a composite index based on soil properties, biodiversity, and plant disease control) in woodlands was 31.4% higher than in grasslands, and 76.1% higher than in croplands. Soil health was positively linked to cropland and grassland productivity at the continental scale. Woodland productivity was best explained by climate. Among microbial diversity indicators, we observed a positive association between the richness of Acidobacteria, Firmicutes, and Proteobacteria and primary productivity. Among microbial functional groups, we found that nitrogen-fixing bacteria and mycorrhizal fungi positively related to primary productivity in croplands and grasslands, while plant pathogens showed a negative relationship. Together, our results point to the importance of soil biodiversity and soil health for maintaining primary productivity across contrasting land-use types.

Keywords:

Soil health, primary productivity, soil biodiversity, large-scale assessment, mycorrhizal fungi, pathogens, soil organic carbon

Introduction

Plant productivity is fundamental for supporting key ecosystem services such as food provision and climate regulation (Barrios, 2007). A range of studies have shown that plant diversity in ecosystems promotes plant productivity and yield (Dietrich et al., 2023; Fraser et al., 2015; Isbell et al., 2011, 2015; Le Provost et al., 2023; Tilman et al., 1997). However, much less is known on how soil health including soil biodiversity relates to primary productivity, particularly across contrasting land-uses at large scales. Soil health refers to the capacity of soils to continuously support plant productivity and ecosystem functions (Lehmann et al., 2020; Qiao et al., 2022). Soil health assessments have traditionally focused on soil properties including microbial biomass, available nutrients, and organic carbon (Bünemann et al., 2018; Lehmann et al., 2020). However, the broader implications of integrating soil biodiversity into these assessments, especially when conducted across diverse land-uses and large spatial scales, remain underexplored yet potentially significant for ecosystem management.

Despite the growing interest in soil health and how it contributes to human well-being (Bender et al., 2016; Wall et al., 2015), current scientific evidence that demonstrates the importance of soil health at large scale is missing (Lehmann et al., 2020; Rinot et al., 2019). A recent study demonstrated that soil health is positively linked to plant yield in croplands in China (Li et al., 2021). Yet, the relative contribution of soil health to explain primary productivity across contrasting ecosystems (i.e., woodlands, grasslands, croplands) remains unknown. The mechanisms by which soil health drives productivity are many-fold: soil organic carbon, for example, positively contributes to primary productivity as it increases soil water retention and nutrient availability (Rawls et al., 2003). On the other hand, the presence and richness of plant pathogens in the soil can lead to diseases, reducing plant growth and yield (Maron et al., 2011). The link between soil health and plant productivity between different land-use types might differ due to variations in biological, physical, and chemical soil properties. Mature woodlands, for example, are stable ecosystems and thought to display high levels of soil health and are dominated by trees with deep root systems that can access nutrients from deeper soil layers and can store nutrients in their biomass for decades (Fernández-Martínez et al., 2014). Additionally, the slow decomposition of leaf litter in woodlands contributes to soil organic matter and nutrient availability over time, which may lead to a gradual increase in primary productivity as the forest matures (Krishna & Mohan, 2017). Consequently, as it is already high, soil health may be less important for productivity in woodlands and other factors (e.g., water availability, temperature) may have a bigger impact on woodland productivity. On the other hand, grasslands and croplands are subjected to more disturbance and are often

characterized by dense and short root networks of grasses and forbs that grow and die
seasonally, leading to a rapid turnover of nutrients and organic matter (Gill & Jackson, 2000;
Wang et al., 2019).

Recently, it was observed that soil microbial diversity at the European scale is highest in
croplands compared to other ecosystems, suggesting that diversity and plant productivity may
be decoupled in these human-made ecosystems (Karimi et al., 2019; Köninger et al., 2023;
Labouyrie et al., 2023; Szoboszlay et al., 2017). We also know that certain microbial groups are
particularly beneficial for primary productivity, including nitrogen-fixing (N-fixing) bacteria and
mycorrhizal fungi (Bauer et al., 2012; van der Heijden et al., 2015). However, the question of
whether productivity is similarly driven by biological components of soil health (i.e., microbial
diversity) across contrasting land-use types remains unexplored, as a systematic large-scale
comparison among different land-uses is lacking and large-scale monitoring efforts are scarce
(Guerra et al., 2021; Heintz-Buschart et al., 2020).

Here, we tested whether soil health is positively linked to primary productivity across Europe.
In addition, we tested which soil variables best explained primary productivity. Our analysis is
based on 588 sites from 27 countries, and we assessed three major land-use types in Europe:
woodlands (n = 185), grasslands (n = 126), and croplands (n = 277). As a measure of soil health,
we used a composite index built from seven soil properties including abiotic and biotic factors
(i.e., soil organic carbon, phosphorus content, water infiltration potential, microbial biomass,
richness of N-fixing bacteria, richness of mycorrhizal -arbuscular and ectomycorrhizal- fungi,
and plant disease control - see methods). As a measure of primary productivity, we used the
normalized difference vegetation index (NDVI) derived from remote sensing data obtained
from each of the 588 sampling locations. Specifically, we tested whether the relationship
between soil health and productivity differs among contrasting land-use types. We
hypothesized that less disturbed ecosystems (i.e., woodlands) would show healthier soils
compared to grasslands and croplands. We also hypothesized that primary productivity in
grasslands and croplands would be more immediately impacted by soil health compared to
woodlands.

Methods

Field survey and soil sampling

This study was carried out within the EU Statistical Office's Land Use and Coverage Area frame Survey (LUCAS), the largest pan-European scheme for assessing soil characteristics in relation to land cover and use (Orgiazzi et al., 2018). As part of the 2018 LUCAS Soil survey, 881 fresh soil samples were collected across Europe and stored at -20°C before further processing (Orgiazzi et al., 2022). The sampling followed a composite strategy: the final sample collected at each location comprised five topsoil (0-20 cm) subsamples that are physically mixed to form a single composite sample. The first subsample was taken at the precise GIS (geographical information system) coordinates of the pre-established LUCAS point, and the remaining four subsamples were taken 2 m from the central one following the cardinal directions (North, East, South, and West). For this study, we focused on 277 cropland, 126 grassland, and 185 woodland sites across 27 countries (Figure 1). Croplands included both permanent (e.g., fruit trees, olive groves, vineyards) and non-permanent (e.g., cereals, legumes) sites. Grasslands included sites covered by communities of grasses, grass-like plants, or forbs. Woodlands included both broadleaved and coniferous forests. Sampling lasted from May to December 2018. The selection was based on availability of all critical variables (soil biodiversity, edaphic factors, and primary productivity).

Soil microbial diversity

Soil DNA was extracted from each of the 588 composite soil samples using the Qiagen DNeasy PowerSoil HTP 96 Kit Q12955-4. Three 0.2 g aliquots per sample were extracted and pooled. Negative and positive controls were included to determine the presence of any external contamination. DNA extracts were quantified and checked for quality using a Qubit™ 1X dsDNA HS Assay Kit and a Qubit™ 3 Fluorometer (Invitrogen). Primer sets for barcode amplification were 515F (GTGYCAGCMGCCGCGGTAA) and 926R (GGCCGYCAATTYMTTTRAGTTT) for the bacterial 16S region (Caporaso et al., 2011; Parada et al., 2016), and ITS9mun (GTACACACCGCCCGTCG) and ITS4ngsUni (CGCCTSCSCTTANTDATATGC) for the fungal ITS region (Tedersoo & Anslan, 2019). Sequencing was performed by Illumina MiSeq platform with 2 x 300 paired-end mode for bacterial amplicons and PacBio Sequel II platform for fungal amplicons. Additional information including DNA amplification and library preparation is available in Labouyrie et al., 2023.

The Illumina and PacBio amplicon data (for bacteria and fungi, respectively) were
demultiplexed using LotuS2 (Özkurt et al., 2022), and paired-end reads were assembled setting
a minimum overlap of 10 bp using FLASH 1.2.10 (Magoč & Salzberg, 2011). Reads failing to
merge were removed from the final dataset. For bacteria, operational taxonomic units (OTUs)
were generated with the *usearch* option (version 10.0.024) available in UPARSE (Edgar, 2013).
Merged reads were truncated up to the 16S primer sequences (i.e., 515F and 926R) and
filtered for the presence of primer sequences (allowing a maximum of 2 mismatches). Merged
reads were further quality-filtered with usearch by removing sequences with more than one
expected error, and duplicated sequences were collapsed with *fqtrim* version 0.9.7 (Pertea,
2015). Obtained OTUs were annotated with the taxonomy data available from the Ribosomal
Database Project (Cole et al., 2014). To avoid sequencing artifacts, only OTUs with at least 50
read counts and present in at least 10 samples were kept. For fungi, 98%-OTUs were obtained
as described in Tedersoo et al., 2021. Fungal taxonomic annotation was performed using
BLAST+ version 2.11.0 (Camacho et al., 2009) against the UNITE 9.1 database (Nilsson et al.,
2019). Bacterial and fungal datasets were further rarefied to comparable sequencing depths
(bacteria: 40,109 reads; fungi: 502 reads). Rarefied datasets were used to obtain values of total
richness (i.e., OTU number), community composition (extracted values from the first
dimension of NMDS based on Bray-Curtis dissimilarities), and the richness of major microbial
groups (i.e., phyla). Rarefied datasets were also used to obtain the richness of major functional
groups as retrieved from the FAPROTAX (bacteria) and FungalTraits (fungi) databases (Louca et
al., 2016; Põlme et al., 2020) (Table S1). Briefly, we retrieved the number of OTUs classified as
bacterial chemoheterotrophs, nitrogen-fixing bacteria, fungal saprotrophs, fungal plant
pathogens, and mycorrhizal fungi, as described in Liu et al., 2022. Further details on
bacterial/fungal sequence processing, filtering, and taxonomic/functional annotation are
available in Labouyrie et al., 2023.

Edaphic and climatic factors

The following measurements were taken for each sampling site to ensure a comprehensive
understanding of climatic and edaphic factors (Ballabio et al., 2019): Total nitrogen and
phosphorus content were assessed to determine nitrogen and phosphorus availability in the
soil. Soil organic carbon was quantified to estimate the amount of carbon stored in the soil
organic matter. Soil pH and microbial biomass were used as measures of acidity and activity of
the living component of the soil, respectively. Soil bulk density was used as a surrogate of
water infiltration potential. Soil texture was characterized by measuring silt, sand, and clay

relative content. Additionally, to capture the broader climatic context of each site, mean
annual precipitation and mean annual temperature (i.e., averaged monthly values over a long
period of time; 1970-2000) were retrieved from the WorldClim database (www.worldclim.org).
These parameters collectively provided a detailed profile of the environmental and soil
conditions at each sampling site (Table S1). Specific details on the methods employed for these
measurements are available in Labouyrie et al., 2023, Orgiazzi et al., 2018 and Smith et al.,
2021.

Primary productivity

The Net Primary Productivity (NPP) was derived from images taken by the NASA Terra and
Aqua satellites using their MODerate resolution Imaging Spectroradiometers (MODIS). The
NPP is calculated as the sum of the MODIS eight-days Net Photosynthesis (PSN) products. PSN
is in turn estimated from the Fraction of Absorbed Photosynthetically Active Radiation
(FAPAR,), which is the fraction of the incoming solar radiation (in the photosynthetically active
spectrum) absorbed by plants. FAPAR is closely related to NDVI (normalized difference
vegetation index), that has been used as a proxy for plant biomass in several studies relating
soil biodiversity and primary productivity (Delgado-Baquerizo et al., 2016; Fan et al., 2023; Liu
et al., 2022). However, the derivation of PSN from FAPAR, provides a more reliable estimate of
the NPP as it accounts for the ground conditions in which the photosynthesis occurs and for
the plant maintenance respiration. Thus, this computation requires additional information
such as the maximum radiation conversion efficiency, the ground temperature, and vapour
pressure. The calculation is performed following the MOD17 PSN/NPP algorithm (Running et
al., 1999). For this study the NPP was derived from MODIS products for year 2018 (to match
with the LUCAS survey) at a resolution of 500 m (as described in Ballabio et al., 2016, 2019).
The NPP values were derived for the months before, during, and after sampling at each LUCAS
location and subsequently averaged.

Soil health

In this study, we used seven explanatory variables (selection based on Lehmann et al., 2020),
namely soil organic carbon, phosphorus content, water infiltration potential, microbial
biomass, richness of N-fixing bacteria, richness of mycorrhizal fungi, and plant disease control)
to build a soil health index (Table S2). These variables were selected as they collectively
represent a broad range of soil characteristics that are critical for understanding soil health-
productivity relationships. Soil organic carbon is crucial for nutrient cycling and soil structure;

phosphorus content is a key nutrient for plant growth; water infiltration potential influences
water availability and root development; microbial biomass indicates the biological activity
essential for nutrient transformation; richness of N-fixing bacteria and mycorrhizal fungi
reflects soil capacity for nitrogen fixation and nutrient uptake, respectively; and plant disease
control highlights the role of soil in supporting plant health. The soil health index was
calculated using z-score transformation: each variable was standardized to have a mean of 0
and a standard deviation of 1. The soil health index for each site was then computed by
summing its respective z-scores (i.e., equally weighting). This approach allowed us to integrate
multiple soil health indicators into a single composite index, facilitating a comprehensive
assessment.

Statistics

Statistical analyses were performed using R software, version 4.2.1. Only soil samples for
which all explanatory variables (i.e., soil biodiversity and climatic, edaphic, and spatial factors)
were available were used for this study, so none of the values in any explanatory variables was
imputed or calculated from the median value of the respective dataset. A total of 588 sites
were used for this study, including 277 croplands, 126 grasslands, and 185 woodlands (Figure
1). To assess the most important predictors of primary productivity within the three land-use
systems, we used Random Forest (RF) (Cutler et al., 2007) followed by structural equation
modeling (Eisenhauer et al., 2015).

First, we selected a wide range of explanatory variables including biodiversity indicators (i.e.,
OTU richness, diversity, community composition), together with climatic and edaphic factors.
The total number of explanatory variables was 34 (Table S1). We then investigated the
correlation among all variables to ensure that there was no strong collinearity and reduced the
variables that were strongly correlated (i.e., Spearman rank correlation coefficient higher than
+0.70, or lower than -0.70). Correlations were run using the *corr* function within the *corrplot*
package (Wei & Simko, 2017). This approach reduced the 34 explanatory variables to 24
(croplands), 23 (grasslands), and 19 (woodlands) (Table S1). The variables removed included
alpha and beta diversity indices (strong correlation with specific microbial groups, e.g.,
Proteobacteria), total nitrogen (strong correlation with soil organic carbon), richness of
chemoheterotrophic bacteria (strong correlation with Actinobacteria and total bacteria), and
silt and sand content (strong correlation with clay content).

To further clarify the relative contribution of climate, edaphic factors, and soil biodiversity to
explain patterns in primary productivity, we conducted RF analyses. The RF analyses were
performed independently for the three land-use types using 10,000 permutations. RF analyses
were performed using the *rfPermute* package (Archer, 2016). Predictor importance was
estimated by the percent increase in Mean Squared Error (MSE) when each predictor was
permuted. This reflects the contribution of each predictor to model accuracy; a higher increase
in MSE indicates greater importance for accurate predictions. Finally, we employed Structural
Equation Modelling (SEM) to evaluate the direct and indirect relationships between climate,
edaphic factors, soil biodiversity, and primary productivity. We adapted the procedure
described in Delgado-Baquerizo et al., 2016 using the R package *piecewiseR* for SEM (Lefcheck,
2016).

Results and discussion

Here we assessed the link between soil health and primary productivity, examining 588 sites
across three land-use types (i.e., croplands, grasslands, and woodlands) in Europe (Figure 1).
We found the highest soil health values in woodlands compared to grasslands and croplands.
According to our initial hypothesis, soil health in woodlands was ~31% higher than in
grasslands, and ~76% higher than in croplands (Figure 2). Soil health was positively linked to
primary productivity when all sites were taken together (Figure 2). In line with our
expectations, soil health was positively linked to primary productivity in grasslands and
croplands, but not in woodlands (Figure 2). Our soil health index encapsulated a range of
abiotic and biotic soil factors, thus providing a comprehensive assessment of soil quality. This
approach is advantageous as it represents a holistic measure that might serve as a decision-
making tool for farmers and land managers. In our study, utilizing a comprehensive soil health
index resulted in a stronger correlation with primary productivity than any single variable
alone. Specifically, the correlation between soil health and productivity was 3.17 times higher
than the average of individual variables for croplands, 2.64 times higher for grasslands, and
1.38 times higher for woodlands (Table S3).

Next, we assessed which individual variables best explained primary productivity. Overall, our
results demonstrate that a combination of climatic and edaphic factors, together with soil
biodiversity, explains patterns in primary productivity across contrasting land-use types at the
European scale: our structural equation models explained 40% of the variability in primary
productivity in grasslands, 35% in woodlands, and 31% in croplands (Figures 3 and 4). Our
results demonstrate that primary productivity in cropland and grassland is best explained by
various factors of soil health, including edaphic properties (e.g., organic carbon) and soil
microbial diversity. In woodlands, our results show that temperature is a major regulator of
primary productivity, although specific components of soil health (e.g., phosphorus content,
organic matter, and the richness of Actinobacteria and Proteobacteria) also do play a role.
Moreover, we observed that various components of microbial diversity act as relevant
predictors of primary productivity, including the richness of Acidobacteria, Firmicutes,
Proteobacteria, and mycorrhizal fungi (Figure 3, Figure 4).

Overall, our results indicate a positive relationship between soil health and cropland
productivity (Figure 2). Random Forest analyses further indicated that pH was the main
predictor of primary productivity in croplands. This confirms previous evidence on the role of
pH in driving crop productivity (Couto, 2018). Accordingly, we observed the lowest cropland

productivity in highly alkaline (pH ≈ 8) soils (Figure S1). Other edaphic factors showed positive

effects on cropland productivity, namely soil organic carbon and phosphorus content (Figure 3,

Figure S1). Among climatic factors, we observed a negative effect of increasing temperature on

cropland productivity, while the effect of increasing precipitation was positive (Figure 3, Figure

S1). Our results further indicate that microbial community composition was the best microbial

predictor of cropland productivity (Figure S2). In line with this, random forest analyses also

indicated that specific microbial groups predicted cropland productivity: that was the case for

Acidobacteria, Firmicutes, and Proteobacteria. The richness of these bacterial groups was

positively correlated with productivity, while a negative effect was observed for Actinobacteria

(Figure 3, Figure S3). Structural equation models (SEMs) further demonstrated that edaphic

factors (e.g., soil texture, soil organic carbon, pH, phosphorus content) affected cropland

productivity also indirectly by leading to changes on the diversity of soil microorganisms,

particularly the bacterial taxa Acidobacteria and Firmicutes, and the fungal phylum

Chytridiomycota (Figure 4). Structural equation models confirmed a positive direct effect of

increased richness of Acidobacteria and Firmicutes on cropland productivity, while the

opposite was observed for Chytridiomycota (i.e., increased OTU richness led to lower

productivity). Acidobacteria are abundant soil-borne microorganisms known to be involved in

nutrient mineralization and organic matter degradation (Banerjee et al., 2016; Kielak et al.,

2016; Sikorski et al., 2022). Our results also align with a recent study where Acidobacteria

positively correlated with cropland productivity in a four-decade fertilization experiment (Fan

et al., 2021). Similarly, members of the phylum Firmicutes, known for their widespread

presence in soils and for promoting plant growth locally, appear to also enhance cropland

productivity on a continental scale (Fierer, 2017; Hashmi et al., 2020). Finally, SEMs indicated

that the richness of Chytridiomycota was negatively related to cropland productivity (Figure 4).

While this fungal group is best known for parasitizing vertebrates, some species are

responsible for plant disease in Europe (van de Vossenberg et al., 2022). In line with this, we

observed a negative correlation between the richness of fungal plant pathogens and cropland

productivity (Figure S3).

The grasslands included in this study were, on average, slightly more acidic than croplands and

showed higher microbial biomass and higher organic carbon content, but lower phosphorus

content compared (see full details in Table S4). Bacterial richness was lower in grasslands

compared to croplands, but fungal richness was slightly higher (Table S4). As for croplands, our

results indicate a positive relationship between soil health and grassland productivity (Figure

2). Random forest analyses retained four factors as predictors of primary productivity in

grasslands, namely pH, soil organic carbon, and the richness of Proteobacteria and mycorrhizal

fungi (Figure 3). Similar to croplands, pH showed a hump-shaped to negative response, with

productivity peaking at pH ≈ 6. The relationship between grassland productivity and soil

organic carbon, Proteobacteria, and mycorrhizal fungi was positive. Grasslands store one third

of the terrestrial carbon stock, and most of it is stored belowground as soil organic carbon

(Conant, 2012). Recent studies show that plant diversity increases soil organic carbon storage

in grasslands by elevating carbon inputs to belowground biomass and promoting microbial

necromass contribution to soil organic carbon storage (Lange et al., 2015; Xu et al., 2020). Our

results add to this and show that soil organic carbon is positively linked to plant productivity

for grasslands at the continental scale, indicating that soil organic carbon favors carbon

storage, but also has positive effects on productivity. Although we did not differentiate

between necromass and living biomass in our study, we also found positive correlations

between bacterial richness/ microbial biomass and grassland productivity (Figure S1, Figure

S2). Structural equation models further supported the idea that grassland productivity was

determined by pH and soil organic carbon (Figure 4). Moreover, SEMs indicated a direct

positive effect of Proteobacteria richness on grassland productivity, and a negative direct

effect of Chytridiomycota.

Finally, woodlands showed the lowest pH and phosphorus content compared to croplands and

grasslands, but the highest organic carbon content. Woodlands also showed the lowest values

of bacterial and fungal richness (see full details in Table S4). Random forest modelling of the

185 woodland sites included in this study revealed that the best predictors of woodland

productivity were monthly temperature, clay content, soil organic carbon, bulk density,

richness of Firmicutes, microbial biomass, and monthly precipitation (Figure 3). Only bulk

density showed a negative effect on woodland productivity, all other predictors were

positively associated to it. SEMs revealed that woodland productivity was shaped by a complex

interplay between climatic factors (i.e., temperature), edaphic factors (i.e., clay content, soil

organic carbon, and phosphorus content) and soil biodiversity (richness of Actinobacteria and

Proteobacteria). Specifically, SEMs depicted a strong positive effect of temperature on

woodland productivity, although factors associated to soil diversity (e.g., richness of

Actinobacteria and Proteobacteria) also shaped woodland productivity (Figure 4). Recent

studies suggest that fungal communities are major drivers of ecosystem functioning in

woodlands (Anthony et al., 2022; Osburn et al., 2023). Our SEM adds to this that fungal

communities (particularly plant-beneficial members within Mortierellomycota, Basidiomycota,

and mycorrhizal fungi) in woodlands are shaped by a combination of climatic factors (i.e.,

precipitation) and edaphic properties. Importantly, we also show that future studies
addressing the impacts of the soil microbiome on woodland productivity need to include
bacteria. This is because major bacterial divisions (i.e., Actinobacteria and Proteobacteria)
were closely related to primary productivity in the woodlands included in our study.

In our study, primary productivity was significantly lower in croplands, compared to grasslands
and woodlands, following previous research (Krause et al., 2022) (Figure S4). In line with our
hypothesis, primary productivity relied more strongly on soil health in grasslands and
croplands than in woodlands. Woodland productivity was largely governed by temperature, in
line with previous findings (Collalti et al., 2020; Morin et al., 2018). Moreover, we observed the
highest soil health values in woodlands. We argue that above a tipping point of soil health,
other factors (i.e., climate), might become more important when it comes to explaining
primary productivity. Finally, agricultural crops often have a short lifespan, rely on a fast
delivery of nutrients and might be more dependent on soil health compared to less dynamic
and more stable ecosystems like woodlands. In line with this, a recent study by Walder et al.,
2023 assessed soil health and crop productivity in 60 fields managed either conventionally,
under no-till, or organically. That study found that the management type with the lowest soil
health, conventionally managed croplands, relied most on soil health when it comes to
explaining plant productivity. Similarly, a recent greenhouse experiment showed that soil
health is positively linked to nitrogen use efficiency (Toda et al., 2023). This indicates that soil
health is especially important as driver of plant productivity and other ecosystem processes in
disturbed agricultural soils. Among the different soil properties here assessed, we found strong
correlations between total nitrogen and phosphorus and plant productivity across land-use
types. We highlight that future studies should focus on their inorganic forms, as these are
more readily available to plants in soil. This aspect was not addressed in our study due to the
high number of samples involved.

Our study being based on soil health indices, we acknowledge that the diversity of soil systems
makes it challenging to capture soil complexity into a single metric. Also, the assessment of soil
health depends on the number of soil variables available, which is especially difficult for large
scale assessments such as in this study. Here, we observed significant relationships between
soil health and primary productivity, pointing to the importance of soil health as driver of plant
productivity. However, these data are based on correlations, and further experimental studies
need to complement and validate these findings empirically testing causality at specific
locations (e.g., testing whether an increase in soil health indeed promotes productivity). In line

with this, previous studies have successfully used both laboratory-based approaches and
manipulative field studies to support findings from observational approaches (Delgado-
Baquerizo et al., 2017; Sünnemann et al., 2023). Finally, we highlight the need to complement
our NDVI-based vegetation analysis with actual plant biomass measurements, although we
recognize the considerable technical challenges associated with conducting such detailed
biomass studies at large scales.

In summary, our results provide compelling evidence that primary productivity is linked to soil
health at the continental scale, particularly for grasslands and croplands. Among the soil health
indicators assessed here, we highlight the positive effects of soil organic carbon and the
richness of Acidobacteria, Firmicutes, and Proteobacteria on primary productivity. In contrast
the richness of Actinobacteria and major plant pathogens including Chytridiomycota was
negatively related primary productivity. Our study points to the importance of soil health as a
driver of primary productivity and it highlights the need to include microorganisms in soil
monitoring schemes to gain a comprehensive understanding of ecosystem functioning in
terrestrial ecosystems.

Figure 1. Study site locations. Geographic distribution across the European continent of the

588 locations used in this study. Three major land-use types were covered in the survey:

cropland (n = 277, yellow), grassland (n = 126, green), and woodland (n = 185, purple).

402   **Figure 2. Relationship between soil health and primary productivity.** Soil health index

403   across land-use types (woodlands, grasslands, and croplands), and relationship between the

404   soil health index and primary productivity. Explained variation is expressed as R^2 (0-1), and

405   significance is indicated by asterisks (***; p-value < 0.001, **; p-value < 0.010, *; p-value <

406   0.050). Non-significant (p-value > 0.05) regressions are denoted with "n.s.".

Figure 3. Relative importance of the main predictors of primary productivity. Relative importance of main predictors (top-10) for primary productivity based on random forest analysis in croplands (A, n = 277), grasslands (B, n = 126) and woodlands (C, n = 185). Dark colors indicate positive predictor effects on primary productivity, and light colors indicate negative effects. Variance explained by models is indicated as percentage (%). Asterisks indicate a significant effect based on the averaged model coefficients (***; p-value < 0.001, **; p-value < 0.010, *; p-value < 0.050).

Figure 4. Structural equation modeling (SEM) describing the relationships between climate, edaphic factors, soil biodiversity, and primary productivity. Direct and indirect effects of climate, edaphic factors, and soil biodiversity on primary productivity across different land-use types. Variables included in each category (i.e., climatic factors, edaphic factors, and soil biodiversity) are indicated on the right panels. The proportion of explained

variance (R^2) appears alongside primary productivity in each model. Goodness-of-fit statistics

are also included. Dashed lines indicate that none of the variables included was significant (p-

value > 0.05). d.f.: degrees of freedom, AIC: Akaike information criterion. Variables included

within "Edaphic factors" are pH, clay content (C), soil organic carbon (SOC), phosphorus

content (P), soil bulk density (BD), and microbial biomass (MB). Variables included within

"Climatic factors" are monthly air temperature (T) and monthly precipitation (P). Variables

included within "Soil biodiversity" are Acidobacteria (Aci), Actinobacteria (Act),

Mortierellomycota (Mor), Verrucomicrobia (Ver), Firmicutes (Fir), Nitrospirae (Nit),

Chytridiomycota (Chy), Mucoromycota (Muc), Ascomycota (Asc), Basidiomycota (Bas),

Proteobateria (Pro), N-fixing bacteria (Nfix), Mycorrhiza (Myc), and plant pathogens (Pat).

"PP" stands for primary productivity.

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References

Anthony, M. A., Crowther, T. W., van der Linde, S., Suz, L. M., Bidartondo, M. I., Cox, F., Schaub, M., Rautio, P., Ferretti, M., Vesterdal, L., De Vos, B., Dettwiler, M., Eickenscheidt, N., Schmitz, A., Meesenburg, H., Andreae, H., Jacob, F., Dietrich, H.-P., Waldner, P., … Averill, C. (2022). Forest tree growth is linked to mycorrhizal fungal composition and function across Europe. *The ISME Journal*, *16*(5), 1327–1336. https://doi.org/10.1038/s41396-021-01159-7

Archer, E. (2016). rfPermute: estimate permutation p-values for random forest importance metrics. *R Package Version*, *1*(2).

Ballabio, C., Lugato, E., Fernández-Ugalde, O., Orgiazzi, A., Jones, A., Borrelli, P., Montanarella, L., & Panagos, P. (2019). Mapping LUCAS topsoil chemical properties at European scale using Gaussian process regression. *Geoderma*, *355*, 113912. https://doi.org/https://doi.org/10.1016/j.geoderma.2019.113912

Ballabio, C., Panagos, P., & Monatanarella, L. (2016). Mapping topsoil physical properties at European scale using the LUCAS database. *Geoderma*, *261*, 110–123. https://doi.org/https://doi.org/10.1016/j.geoderma.2015.07.006

Banerjee, S., Kirkby, C. A., Schmutter, D., Bissett, A., Kirkegaard, J. A., & Richardson, A. E. (2016). Network analysis reveals functional redundancy and keystone taxa amongst bacterial and fungal communities during organic matter decomposition in an arable soil. *Soil Biology and Biochemistry*, *97*, 188–198. https://doi.org/https://doi.org/10.1016/j.soilbio.2016.03.017

Barrios, E. (2007). Soil biota, ecosystem services and land productivity. *Ecological Economics*, *64*(2), 269–285. https://doi.org/https://doi.org/10.1016/j.ecolecon.2007.03.004

Bauer, J. T., Kleczewski, N. M., Bever, J. D., Clay, K., & Reynolds, H. L. (2012). Nitrogen-fixing bacteria, arbuscular mycorrhizal fungi, and the productivity and structure of prairie grassland communities. *Oecologia*, *170*, 1089–1098.

Bender, S. F., Wagg, C., & van der Heijden, M. G. A. (2016). An Underground Revolution: Biodiversity and Soil Ecological Engineering for Agricultural Sustainability. *Trends in Ecology & Evolution*, *31*(6), 440–452. https://doi.org/https://doi.org/10.1016/j.tree.2016.02.016

Bünemann, E. K., Bongiorno, G., Bai, Z., Creamer, R. E., De Deyn, G., De Goede, R., Fleskens, L., Geissen, V., Kuyper, T. W., & Mäder, P. (2018). Soil quality–A critical review. *Soil Biology and Biochemistry*, *120*, 105–125.

Camacho, C., Coulouris, G., Avagyan, V., Ma, N., Papadopoulos, J., Bealer, K., & Madden, T. L. (2009). BLAST+: architecture and applications. *BMC Bioinformatics*, *10*, 1–9.

Caporaso, J. G., Lauber, C. L., Walters, W. A., Berg-Lyons, D., Lozupone, C. A., Turnbaugh, P. J., Fierer, N., & Knight, R. (2011). Global patterns of 16S rRNA diversity at a depth of millions of sequences per sample. *Proceedings of the National Academy of Sciences*, *108*(supplement_1), 4516–4522.

Cole, J. R., Wang, Q., Fish, J. A., Chai, B., McGarrell, D. M., Sun, Y., Brown, C. T., Porras-Alfaro, A., Kuske, C. R., & Tiedje, J. M. (2014). Ribosomal Database Project: data and tools for high throughput rRNA analysis. *Nucleic Acids Research*, *42*(D1), D633–D642.

Collalti, A., Ibrom, A., Stockmarr, A., Cescatti, A., Alkama, R., Fernández-Martínez, M.,
Matteucci, G., Sitch, S., Friedlingstein, P., Ciais, P., Goll, D. S., Nabel, J. E. M. S.,
Pongratz, J., Arneth, A., Haverd, V., & Prentice, I. C. (2020). Forest production
efficiency increases with growth temperature. *Nature Communications*, *11*(1),
5322. https://doi.org/10.1038/s41467-020-19187-w
Conant, R. T. (2012). Grassland soil organic carbon stocks: status, opportunities,
vulnerability. *Recarbonization of the Biosphere: Ecosystems and the Global Carbon*
Cycle, 275–302.
Couto, W. (2018). Soil pH and plant productivity. In *Handbook of agricultural*
productivity (pp. 71–84). CRC Press.
Cutler, D. R., Edwards Jr, T. C., Beard, K. H., Cutler, A., Hess, K. T., Gibson, J., & Lawler,
J. J. (2007). Random forests for classification in ecology. *Ecology*, *88*(11), 2783–
2792.
Delgado-Baquerizo, M., Maestre, F. T., Reich, P. B., Jeffries, T. C., Gaitan, J. J., Encinar,
D., Berdugo, M., Campbell, C. D., & Singh, B. K. (2016). Microbial diversity drives
multifunctionality in terrestrial ecosystems. *Nature Communications*, *7*(1), 10541.
https://doi.org/10.1038/ncomms10541
Delgado-Baquerizo, M., Trivedi, P., Trivedi, C., Eldridge, D. J., Reich, P. B., Jeffries, T. C.,
& Singh, B. K. (2017). Microbial richness and composition independently drive soil
multifunctionality. *Functional Ecology*, *31*(12), 2330–2343.
https://doi.org/https://doi.org/10.1111/1365-2435.12924
Dietrich, P., Eisenhauer, N., & Roscher, C. (2023). Linking plant diversity–productivity
relationships to plant functional traits of dominant species and changes in soil
properties in 15-year-old experimental grasslands. *Ecology and Evolution*, *13*(3),
e9883. https://doi.org/https://doi.org/10.1002/ece3.9883
Edgar, R. C. (2013). UPARSE: highly accurate OTU sequences from microbial amplicon
reads. *Nature Methods*, *10*(10), 996–998.
Eisenhauer, N., Bowker, M. A., Grace, J. B., & Powell, J. R. (2015). From patterns to
causal understanding: structural equation modeling (SEM) in soil ecology.
Pedobiologia, *58*(2–3), 65–72.
Fan, K., Chu, H., Eldridge, D. J., Gaitan, J. J., Liu, Y.-R., Sokoya, B., Wang, J.-T., Hu, H.-W.,
He, J.-Z., Sun, W., Cui, H., Alfaro, F. D., Abades, S., Bastida, F., Díaz-López, M.,
Bamigboye, A. R., Berdugo, M., Blanco-Pastor, J. L., Grebenc, T., … Delgado-
Baquerizo, M. (2023). Soil biodiversity supports the delivery of multiple
ecosystem functions in urban greenspaces. *Nature Ecology & Evolution*, *7*(1), 113–
126. https://doi.org/10.1038/s41559-022-01935-4
Fan, K., Delgado-Baquerizo, M., Guo, X., Wang, D., Zhu, Y., & Chu, H. (2021).
Biodiversity of key-stone phylotypes determines crop production in a 4-decade
fertilization experiment. *The ISME Journal*, *15*(2), 550–561.
https://doi.org/10.1038/s41396-020-00796-8
Fernández-Martínez, M., Vicca, S., Janssens, I. A., Sardans, J., Luyssaert, S., Campioli,
M., Chapin III, F. S., Ciais, P., Malhi, Y., Obersteiner, M., Papale, D., Piao, S. L.,
Reichstein, M., Rodà, F., & Peñuelas, J. (2014). Nutrient availability as the key
regulator of global forest carbon balance. *Nature Climate Change*, *4*(6), 471–476.
https://doi.org/10.1038/nclimate2177

Fierer, N. (2017). Embracing the unknown: disentangling the complexities of the soil
microbiome. *Nature Reviews Microbiology*, *15*(10), 579–590.
https://doi.org/10.1038/nrmicro.2017.87
Fraser, L. H., Pither, J., Jentsch, A., Sternberg, M., Zobel, M., Askarizadeh, D., Bartha, S.,
Beierkuhnlein, C., Bennett, J. A., Bittel, A., Boldgiv, B., Boldrini, I. I., Bork, E.,
Brown, L., Cabido, M., Cahill, J., Carlyle, C. N., Campetella, G., Chelli, S., … Zupo, T.
(2015). Worldwide evidence of a unimodal relationship between productivity and
plant species richness. *Science*, *349*(6245), 302–305.
https://doi.org/10.1126/science.aab3916
Gill, R. A., & Jackson, R. B. (2000). Global patterns of root turnover for terrestrial
ecosystems. *The New Phytologist*, *147*(1), 13–31.
Guerra, C. A., Bardgett, R. D., Caon, L., Crowther, T. W., Delgado-Baquerizo, M.,
Montanarella, L., Navarro, L. M., Orgiazzi, A., Singh, B. K., & Tedersoo, L. (2021).
Tracking, targeting, and conserving soil biodiversity. *Science*, *371*(6526), 239–241.
Hashmi, I., Bindschedler, S., & Junier, P. (2020). Firmicutes. In N. Amaresan, M. Senthil
Kumar, K. Annapurna, K. Kumar, & A. Sankaranarayanan (Eds.), *Beneficial
Microbes in Agro-Ecology* (pp. 363–396). Academic Press.
https://doi.org/https://doi.org/10.1016/B978-0-12-823414-3.00018-6
Heintz-Buschart, A., Guerra, C., Djukic, I., Cesarz, S., Chatzinotas, A., Patoine, G.,
Sikorski, J., Buscot, F., Küsel, K., & Wegner, C.-E. (2020). Microbial diversity-
ecosystem function relationships across environmental gradients. *Research Ideas
and Outcomes*, *6*, e52217.
Isbell, F., Calcagno, V., Hector, A., Connolly, J., Harpole, W. S., Reich, P. B., Scherer-
Lorenzen, M., Schmid, B., Tilman, D., van Ruijven, J., Weigelt, A., Wilsey, B. J.,
Zavaleta, E. S., & Loreau, M. (2011). High plant diversity is needed to maintain
ecosystem services. *Nature*, *477*(7363), 199–202.
https://doi.org/10.1038/nature10282
Isbell, F., Craven, D., Connolly, J., Loreau, M., Schmid, B., Beierkuhnlein, C., Bezemer, T.
M., Bonin, C., Bruelheide, H., de Luca, E., Ebeling, A., Griffin, J. N., Guo, Q.,
Hautier, Y., Hector, A., Jentsch, A., Kreyling, J., Lanta, V., Manning, P., …
Eisenhauer, N. (2015). Biodiversity increases the resistance of ecosystem
productivity to climate extremes. *Nature*, *526*(7574), 574–577.
https://doi.org/10.1038/nature15374
Karimi, B., Dequiedt, S., Terrat, S., Jolivet, C., Arrouays, D., Wincker, P., Cruaud, C.,
Bispo, A., Chemidlin Prévost-Bouré, N., & Ranjard, L. (2019). Biogeography of Soil
Bacterial Networks along a Gradient of Cropping Intensity. *Scientific Reports*, *9*(1),
3812. https://doi.org/10.1038/s41598-019-40422-y
Kielak, A. M., Cipriano, M. A. P., & Kuramae, E. E. (2016). Acidobacteria strains from
subdivision 1 act as plant growth-promoting bacteria. *Archives of Microbiology*,
198(10), 987–993. https://doi.org/10.1007/s00203-016-1260-2
Köninger, J., Ballabio, C., Panagos, P., Jones, A., Schmid, M. W., Orgiazzi, A., & Briones,
M. J. I. (2023). Ecosystem type drives soil eukaryotic diversity and composition in
Europe. *Global Change Biology*.
Krause, A., Papastefanou, P., Gregor, K., Layritz, L. S., Zang, C. S., Buras, A., Li, X., Xiao,
J., & Rammig, A. (2022). Quantifying the impacts of land cover change on gross
primary productivity globally. *Scientific Reports*, *12*(1), 18398.
https://doi.org/10.1038/s41598-022-23120-0

Krishna, M. P., & Mohan, M. (2017). Litter decomposition in forest ecosystems: a
review. *Energy, Ecology and Environment*, *2*(4), 236–249.
https://doi.org/10.1007/s40974-017-0064-9
Labouyrie, M., Ballabio, C., Romero, F., Panagos, P., Jones, A., Schmid, M. W.,
Mikryukov, V., Dulya, O., Tedersoo, L., & Bahram, M. (2023). Patterns in soil
microbial diversity across Europe. *Nature Communications*, *14*(1), 3311.
Lange, M., Eisenhauer, N., Sierra, C. A., Bessler, H., Engels, C., Griffiths, R. I., Mellado-
Vázquez, P. G., Malik, A. A., Roy, J., Scheu, S., Steinbeiss, S., Thomson, B. C.,
Trumbore, S. E., & Gleixner, G. (2015). Plant diversity increases soil microbial
activity and soil carbon storage. *Nature Communications*, *6*(1), 6707.
https://doi.org/10.1038/ncomms7707
Le Provost, G., Schenk, N. V, Penone, C., Thiele, J., Westphal, C., Allan, E., Ayasse, M.,
Blüthgen, N., Boeddinghaus, R. S., Boesing, A. L., Bolliger, R., Busch, V., Fischer,
M., Gossner, M. M., Hölzel, N., Jung, K., Kandeler, E., Klaus, V. H., Kleinebecker, T.,
… Manning, P. (2023). The supply of multiple ecosystem services requires
biodiversity across spatial scales. *Nature Ecology & Evolution*, *7*(2), 236–249.
https://doi.org/10.1038/s41559-022-01918-5
Lefcheck, J. S. (2016). piecewiseSEM: Piecewise structural equation modelling in r for
ecology, evolution, and systematics. *Methods in Ecology and Evolution*, *7*(5), 573–
579.
Lehmann, J., Bossio, D. A., Kögel-Knabner, I., & Rillig, M. C. (2020). The concept and
future prospects of soil health. *Nature Reviews Earth & Environment*, *1*(10), 544–
553.
Li, P., Li, Y., Xu, L., Zhang, H., Shen, X., Xu, H., Jiao, J., Li, H., & Hu, F. (2021). Crop yield-
soil quality balance in double cropping in China's upland by organic amendments:
A meta-analysis. *Geoderma*, *403*, 115197.
https://doi.org/https://doi.org/10.1016/j.geoderma.2021.115197
Liu, S., García-Palacios, P., Tedersoo, L., Guirado, E., van der Heijden, M. G. A., Wagg,
C., Chen, D., Wang, Q., Wang, J., Singh, B. K., & Delgado-Baquerizo, M. (2022).
Phylotype diversity within soil fungal functional groups drives ecosystem stability.
Nature Ecology & Evolution, *6*(7), 900–909. https://doi.org/10.1038/s41559-022-
01756-5
Louca, S., Parfrey, L. W., & Doebeli, M. (2016). Decoupling function and taxonomy in
the global ocean microbiome. *Science*, *353*(6305), 1272–1277.
Magoč, T., & Salzberg, S. L. (2011). FLASH: fast length adjustment of short reads to
improve genome assemblies. *Bioinformatics*, *27*(21), 2957–2963.
Maron, J. L., Marler, M., Klironomos, J. N., & Cleveland, C. C. (2011). Soil fungal
pathogens and the relationship between plant diversity and productivity. *Ecology*
Letters, *14*(1), 36–41.
Morin, X., Fahse, L., Jactel, H., Scherer-Lorenzen, M., García-Valdés, R., & Bugmann, H.
(2018). Long-term response of forest productivity to climate change is mostly
driven by change in tree species composition. *Scientific Reports*, *8*(1), 5627.
https://doi.org/10.1038/s41598-018-23763-y
Nilsson, R. H., Larsson, K.-H., Taylor, A. F. S., Bengtsson-Palme, J., Jeppesen, T. S.,
Schigel, D., Kennedy, P., Picard, K., Glöckner, F. O., & Tedersoo, L. (2019). The
UNITE database for molecular identification of fungi: handling dark taxa and
parallel taxonomic classifications. *Nucleic Acids Research*, *47*(D1), D259–D264.

Orgiazzi, A., Ballabio, C., Panagos, P., Jones, A., & Fernández-Ugalde, O. (2018). LUCAS
Soil, the largest expandable soil dataset for Europe: a review. *European Journal of*
Soil Science, *69*(1), 140–153. https://doi.org/https://doi.org/10.1111/ejss.12499
Orgiazzi, A., Panagos, P., Fernández-Ugalde, O., Wojda, P., Labouyrie, M., Ballabio, C.,
Franco, A., Pistocchi, A., Montanarella, L., & Jones, A. (2022). LUCAS Soil
Biodiversity and LUCAS Soil Pesticides, new tools for research and policy
development. *European Journal of Soil Science*, *73*(5), e13299.
https://doi.org/https://doi.org/10.1111/ejss.13299
Osburn, E. D., Yang, G., Rillig, M. C., & Strickland, M. S. (2023). Evaluating the role of
bacterial diversity in supporting soil ecosystem functions under anthropogenic
stress. *ISME Communications*, *3*(1), 66.
Özkurt, E., Fritscher, J., Soranzo, N., Ng, D. Y. K., Davey, R. P., Bahram, M., &
Hildebrand, F. (2022). LotuS2: an ultrafast and highly accurate tool for amplicon
sequencing analysis. *Microbiome*, *10*(1), 1–14.
Parada, A. E., Needham, D. M., & Fuhrman, J. A. (2016). Every base matters: assessing
small subunit rRNA primers for marine microbiomes with mock communities,
time series and global field samples. *Environmental Microbiology*, *18*(5), 1403–
1414.
Pertea, G. (2015). fqtrim: v0. 9.4 release. *Zenodo: Geneva, Switzerland*.
Põlme, S., Abarenkov, K., Henrik Nilsson, R., Lindahl, B. D., Clemmensen, K. E.,
Kauserud, H., Nguyen, N., Kjøller, R., Bates, S. T., & Baldrian, P. (2020).
FungalTraits: a user-friendly traits database of fungi and fungus-like
stramenopiles. *Fungal Diversity*, *105*, 1–16.
Qiao, L., Wang, X., Smith, P., Fan, J., Lu, Y., Emmett, B., Li, R., Dorling, S., Chen, H., &
Liu, S. (2022). Soil quality both increases crop production and improves resilience
to climate change. *Nature Climate Change*, *12*(6), 574–580.
Rawls, W. J., Pachepsky, Y. A., Ritchie, J. C., Sobecki, T. M., & Bloodworth, H. (2003).
Effect of soil organic carbon on soil water retention. *Geoderma*, *116*(1), 61–76.
https://doi.org/https://doi.org/10.1016/S0016-7061(03)00094-6
Rinot, O., Levy, G. J., Steinberger, Y., Svoray, T., & Eshel, G. (2019). Soil health
assessment: A critical review of current methodologies and a proposed new
approach. *Science of the Total Environment*, *648*, 1484–1491.
Running, S. W., Nemani, R., Glassy, J. M., & Thornton, P. E. (1999). *MODIS daily*
photosynthesis (PSN) and annual net primary production (NPP) product (MOD17)
Algorithm Theoretical Basis Document.
Sikorski, J., Baumgartner, V., Birkhofer, K., Boeddinghaus, R. S., Bunk, B., Fischer, M.,
Fösel, B. U., Friedrich, M. W., Göker, M., & Hölzel, N. (2022). The evolution of
ecological diversity in Acidobacteria. *Frontiers in Microbiology*, *13*, 78.
Smith, L. C., Orgiazzi, A., Eisenhauer, N., Cesarz, S., Lochner, A., Jones, A., Bastida, F.,
Patoine, G., Reitz, T., & Buscot, F. (2021). Large-scale drivers of relationships
between soil microbial properties and organic carbon across Europe. *Global*
Ecology and Biogeography, *30*(10), 2070–2083.
Sünnemann, M., Beugnon, R., Breitkreuz, C., Buscot, F., Cesarz, S., Jones, A., Lehmann,
A., Lochner, A., Orgiazzi, A., Reitz, T., Rillig, M. C., Schädler, M., Smith, L. C.,
Zeuner, A., Guerra, C. A., & Eisenhauer, N. (2023). Climate change and cropland
management compromise soil integrity and multifunctionality. *Communications*
Earth & Environment, *4*(1), 394. https://doi.org/10.1038/s43247-023-01047-2

Szoboszlay, M., Dohrmann, A. B., Poeplau, C., Don, A., & Tebbe, C. C. (2017). Impact of
land-use change and soil organic carbon quality on microbial diversity in soils
across Europe. *FEMS Microbiology Ecology*, *93*(12), fix146.
https://doi.org/10.1093/femsec/fix146

Tedersoo, L., & Anslan, S. (2019). Towards PacBio-based pan-eukaryote metabarcoding
using full-length ITS sequences. *Environmental Microbiology Reports*, *11*(5), 659–
668.

Tedersoo, L., Mikryukov, V., Anslan, S., Bahram, M., Khalid, A. N., Corrales, A., Agan, A.,
Vasco-Palacios, A.-M., Saitta, A., & Antonelli, A. (2021). The Global Soil
Mycobiome consortium dataset for boosting fungal diversity research. *Fungal*
Diversity, *111*, 573–588.

Tilman, D., Lehman, C. L., & Thomson, K. T. (1997). Plant diversity and ecosystem
productivity: theoretical considerations. *Proceedings of the National Academy of*
Sciences, *94*(5), 1857–1861.

Toda, M., Walder, F., & van der Heijden, M. G. A. (2023). Organic management and soil
health promote nutrient use efficiency. *Journal of Sustainable Agriculture and*
Environment.

van de Vossenberg, Bart. T. L. H., Prodhomme, C., Vossen, J. H., & van der Lee, T. A. J.
(2022). Synchytrium endobioticum, the potato wart disease pathogen. *Molecular*
Plant Pathology, *23*(4), 461–474.
https://doi.org/https://doi.org/10.1111/mpp.13183

van der Heijden, M. G. A., Martin, F. M., Selosse, M.-A., & Sanders, I. R. (2015).
Mycorrhizal ecology and evolution: the past, the present, and the future. *New*
Phytologist, *205*(4), 1406–1423.
https://doi.org/https://doi.org/10.1111/nph.13288

Walder, F., Büchi, L., Wagg, C., Colombi, T., Banerjee, S., Hirte, J., Mayer, J., Six, J.,
Keller, T., Charles, R., & van der Heijden, M. G. A. (2023). Synergism between
production and soil health through crop diversification, organic amendments and
crop protection in wheat-based systems. *Journal of Applied Ecology*, *60*(10),
2091–2104. https://doi.org/https://doi.org/10.1111/1365-2664.14484

Wall, D. H., Nielsen, U. N., & Six, J. (2015). Soil biodiversity and human health. *Nature*,
528(7580), 69–76. https://doi.org/10.1038/nature15744

Wang, J., Sun, J., Yu, Z., Li, Y., Tian, D., Wang, B., Li, Z., & Niu, S. (2019). Vegetation type
controls root turnover in global grasslands. *Global Ecology and Biogeography*,
28(4), 442–455. https://doi.org/https://doi.org/10.1111/geb.12866

Wei, T., & Simko, V. (2017). *R package "corrplot": Visualization of a Correlation Matrix*
(Version 0.84). Vienna.

Xu, S., Eisenhauer, N., Ferlian, O., Zhang, J., Zhou, G., Lu, X., Liu, C., & Zhang, D. (2020).
Species richness promotes ecosystem carbon storage: evidence from biodiversity-
ecosystem functioning experiments. *Proceedings of the Royal Society B: Biological*
Sciences, *287*(1939), 20202063. https://doi.org/10.1098/rspb.2020.2063

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