1
2

A single changing hypernetwork to represent (social-)ecological dynamics

Gaucherel C.[1]*, Cosme M. [1], Noûs C. [2] & Pommereau F.[3]

[1] UMR AMAP – INRAE, University of Montpellier, CNRS, CIRAD, IRD, Montpellier, France
[2] Cogitamus Laboratory, France
[2] IBISC – University of Evry-Val d'Essonne, Unviersity of Paris-Saclay, Evry, France

*Corresponding author
Correspondence: cedric.gaucherel@inrae.fr

14
15

ABSTRACT

To understand and manage (social-)ecological systems, we need an intuitive and rigorous way to represent them. Recent ecological studies propose to represent interaction networks into modular graphs, multiplexes and higher-order interactions. Along these lines, we argue here that non-pairwise interactions are common in ecology and environmental sciences, necessitating renewed concepts and tools for handling them. In addition, such interaction networks often change sharply, due to appearing and disappearing species and components. We illustrate in a simple example that any ecosystem can be represented by a single hypergraph, here called the ecosystem hypernetwork. Moreover, we highlight that any ecosystem hypernetwork exhibits a changing topology summarizing its long term dynamics (e.g. species extinction/invasion, pollutant or human arrival). Qualitative and discrete-event models developed in computer science appear suitable for modeling hypergraph (topological) dynamics. Hypernetworks thus also provide a conceptual foundation for theoretical as well as more applied studies in ecology (at large), as they form the qualitative backbone of ever-changing ecosystems.

Keywords: Ecosystems, interaction networks, hypergraphs, Petri nets, discrete-event models.

Introduction

Ecosystems are under threat and we need to think carefully about their representations for many
reasons. A sound representation helps to conceptualize the study, understanding and ultimately, the
management of ecosystems. Ecological and epistemological works continually propose new
representations of these complex objects (Gignoux et al. 2011, Schwartz and Jax 2011), such as class
diagrams and interaction networks (Strogatz 2001, Loreau et al. 2003, Proulx et al. 2005), to name but two.
Such representations should firstly be useful and provide relevant insights. Secondly, even in case of well-
understood ecosystems, tracking their changes and managing them efficiently, necessitates dealing with a
reliable and intuitive representation. While redefining the ecosystem is outside the scope of this paper, we
at least acknowledge that all ecosystems are made up of biotic, abiotic and often anthropogenic
components. These components continuously interact through various (bio-ecological, social-economic,
physicochemical) processes (e.g., Frontier et al. 2008, Gignoux et al. 2011, Gaucherel 2018). Defined as
such, the ecosystem may be, and often is, represented as a set of (material) variables and (abstract)
processes (Fontaine et al. 2011, Pilosof et al. 2017, Landi et al. 2018). While a set-based representation
does not assign an order to ecosystem content, matrix or graph representations more explicitly specify
their complex, directed (oriented) interconnections. Such representations often exhibit specific structures,
such as multiple interactions between variables (Kéfi et al. 2017, Gaucherel and Pommereau 2019) or
nested (Bastolla et al. 2009) and modular sets (Dicks et al. 2002) of variables.

Species networks and social networks provide two textbook cases of network representations. It is
common to represent species communities in the form of a graph in which the main variables (species or
populations) are represented by nodes and their ecological interactions by edges. Species networks
generally consist in sub-systems of an ecosystem, including e.g., food webs, plant-pollinator networks,
host-parasite networks and competition networks. Some of these graphical models have also been
combined with a dynamic model to examine variations in species abundances and process flows (fluxes)
on the graph (Proulx et al. 2005, Delmas et al. 2018, Landi et al. 2018). The use of multilayer, multiplex and
multilevel networks has recently been proposed to merge distinct sub-systems in the same network (Kéfi
et al. 2016, Pilosof et al. 2017). These multigraphs are representations combining several initially separate
graphs involving the same variables into a multilayered graph.

Latterly, it has been proposed to study *hypergraphs* for their ability to grasp non-dyadic (i.e., non-
pairwise) interactions between species, also called *Higher-Order Interactions*, as many ecological
interactions appear to be mediated by a third (or more) species (Werner and Peacor 2003, Golubski et al.
2016, Delmas et al. 2018). Such multi-component interactions thus play an important role in system
dynamics. Hypergraphs have rarely been used in biology and ecology (but see Billick and Case 1994, Klamt
et al. 2009 , Valverde et al. 2020). Several definitions have been proposed for the higher-order interactions
they capture: a pairwise interaction modulated by (i.e., in the context of) a third species, a non-additive
effect of two species on the per-capita growth of another, or the failure of a null pairwise model (Sanchez
2019). However, the hypergraphs and higher-order interactions found in the literature are mainly confined
to interactions made up of three components and are static. Here, we venture to add that i) higher-order
interactions not only concern species interactions, ii) are not at all restricted to three-way interactions and,
ultimately, iii) often change sharply in the long term. In parallel, social networks are also much used and
studied. In the same vein as species networks, anthropogenic nets can be plotted in graph form, with which
humans or human groups shown as nodes and each human-related process (e.g., communication, trading,
financial exchange) as an edge connecting the nodes (e.g., Buldyrev et al. 2010, Brummitt et al. 2012).
Depending on the question being addressed, such human-related graphs are sometimes purposefully

biased by more social, more economic or political processes. Rarely, these interaction networks are merged
into the same larger graph for study, such as in decision systems (Tixier et al. 2013, Battiston et al. 2020,
Felipe-Lucia et al. 2021). As humans in a society, we understand that social networks are highly dynamic,
frequently losing or gaining new components (individuals or groups) and interactions, such as in ecosystem
services (Scholes et al. 2013, Mao et al. 2021). In principle, there is nothing to prevent the merging of
interaction networks involving human groups with networks involving species once they interact (Felipe-
Lucia et al. 2021), and so on with any other (e.g., physicochemical) variable in the same social-ecological
system.
89
In this paper, we will build on the representations above to put forward three interlinked propositions:
i) any ecosystem may be efficiently represented by a single graph, here called the *ecosystem network*,
whatever the interactions it involves; ii) this single graph would gain from being a hypergraph (called
hypernetwork), made up of non-dyadic interactions; and iii) the single hypernetwork changes in the long
term, hence the need to formally document these changes to improve our understanding of ecosystem
dynamics. In other words, we aim to show that the whole (social-)ecological system (i.e., with all the
processes and components it involves) can be represented by a single comprehensive graph which may be
more accurately defined as a changing hypergraph. In addition, many ecological interactions are non-
dyadic, i.e., they involve more than two variables and, reciprocally, most variables are simultaneously
involved in many interactions (Billick and Case 1994, Valverde et al. 2020). In this paper, we will not present
new models and new results, rather than exemplify the proposed concepts with our past basic (Gaucherel
and Pommereau 2019, Pommereau et al. 2022a) and applied studies (Mao et al. 2021, Cosme et al. 2022).
As a simple illustration, we will model a termite population that is eating, producing mushrooms and
reproducing through various processes (Turner 2009). We will show the numerous advantages of
representing the ecosystem as a changing hypernetwork, and will discuss the conceptual, explanatory and
practical insights it provides.

Methods and Results

A single graph for the ecosystem
To summarize any ecosystem on the basis of its interactions and interaction networks implies an
understanding, first and foremost, of what an interaction is. An interaction can be defined in several ways,
although most ecologists would accept that it is an abstraction of "real relationships" between species or
between any other ecosystem components (Delmas et al. 2018, Landi et al. 2018). For example, a
"predation interaction" takes the form of the catching and eating by (at least) one predator individual of
(at least) one prey individual (Nakazawa 2020). Hence, what is usually called predation is an abstract
concept, a kind of immaterial process, with a duration in time and aggregating many events. When
ecologists study interaction networks to understand or predict their dynamics, they mostly explore
quantitative variations in variables (abundances) through various flows (matter and energy exchanges)
connecting them (Lobry and Sari 2015, Nakazawa 2020) (Fig. 1a).
118
Everything that has been said about species may pertain to any other component of a social-ecological
system, depending on the question being addressed. For example, the economic interaction stating that
"a producer sells their productions to a consumer" is composed of a number of occurrences of isolated and
observable sales between two social units (Strogatz 2001, Gross and Sayama 2009). Similarly, the social
interaction "a scientist communicates their results" is composed of a number of utterances or sentences
between the scientist and the audience. The fact that the scientist may simultaneously be communicator,
consumer and predator eating a salad for lunch, strongly militates in favor of merging the respective social,

economic and ecological networks in which they are involved (Felipe-Lucia et al. 2021). The reverse is also true: many processes require several components to be applied. To "communicate their results", the scientist may use a computer, the salad they have eaten and potentially studied, and the audience. Hence, as a first step, a single graph can be built in each social-ecological system, depending on the question being addressed (Fig. 1b).

Figure 1

Figure 1 - Tentative representations of corresponding quantitative (a) and qualitative (b) predator/prey dynamics. The total biomasses of predator (in black) and prey (blue) populations fluctuate until they collapse (a). In the long term, the trophic network may be represented as a qualitative graph, potentially involving other species (in grey), shifting sharply from one discrete state to the next, lacking the corresponding prey and predator nodes.

In previous studies, we suggested calling this comprehensive graph the *ecosystem network* (EN)
(Gaucherel and Pommereau 2019, Cosme et al. 2022), as it represents flows and is not simply an abstract
mathematical graph. This EN usually concerns a social-ecological system interaction network with no
limitations on the natures of its components and processes, insofar as the question being addressed
requires all these components. Hereafter, we illustrate the following concepts on the basis of a termite
colony (Table 1). The same graph reprensenting the system may be displayed with several algorithms for
highlighting distinct features (Fig. 2a-b), such as ecosystem boundaries or multiscale (nested) structures.
All the graphs examined so far in various disciplines are indeed "sub-graphs" or modules of the same single
"super-graph" of the whole social-ecological system. Often, this is the same node of the EN, representing
a material component, which is potentially involved in highly distinct (trophic, physicochemical and/or
human-related) interaction networks (Fontaine et al. 2011, Gaucherel 2019). The reasons that we usually
avoid merging all these modular graphs are: specific monodisciplinary questions being addressed; data,
knowledge or skills being unavailable; to avoid time-consuming analyses often with limited capabilities; or
for all these reasons combined. As seen recently in multilayer, multiplex and multilevel analyses, individual
graphs of the same system may be merged into a single larger graph (Fontaine et al. 2011, Pilosof et al.
2017). This observation strongly suggests that each node is connected to nodes of distinct natures, through
interactions of various natures too (Kéfi et al. 2016, Gaucherel et al. 2020). In contrast to these studies,
though, we do not make any hypothesis here about the shape of the EN. It may be layered, modular, nested
or have another specific structure: we do not need to assume its shape, but instead allow the graph
topology to emerge from the purported, observed and known interactions (Mao et al. 2021, Cosme et al.
2022).

a)

b)

Figure 2

Furthermore, it follows from this observation that every node, as a component, is a super-graph (made
up of finer sub-nodes) in itself. Indeed, it is always possible, and sometimes relevant, to subdivide a
population node into individuals or groups (e.g., population stages, age-wise, sex-wise), and to emphasize
their various actions, at various levels of organization. All these representations need to be useful for the

question being addressed, though, as it is often pointless to subdivide a component into a new subgraph
to achieve a more detailed representation of the system. This remains valid for a specific interaction which
it is pointless to subdivide into several (pairwise) processes. Whenever it is relevant for the question being
addressed to zoom in ("unfold") and/or zoom out ("fold") on part of the single graph under examination,
it is possible to split nodes into sub-nodes connected to other nodes already present with distinct processes
(Gaucherel 2019). Such a procedure obviously requires caution as it quickly encompasses distinct levels of
organization (or scales) too.
178
From graph to hypergraph
In real systems, most variables are connected to many others, such as the aforementioned scientist
who interacts with other scientists for the purpose of communication, with salads as food and with
producers for shopping, through the processes of talking, eating and buying. Conversely, most processes
involve many components (variables) simultaneously, such as the communication process requiring the
scientist, potentially an organism under examination (the salad), and the audience. This observation
remains valid for any subgraph and discipline such as, for example, a physicochemical graph of chemical
reactions requiring reactants and catalyzers (Fontana and Buss 1994, Cumming et al. 2014, Fages et al.
2018). For these reasons, most interactions are called non-dyadic (i.e., not reduced to pairwise
interactions), a point that has been made recently for ecological networks too (Werner and Peacor 2003,
Golubski et al. 2016, Delmas et al. 2018). In ecology and elsewhere, it may be that such non-dyadic
interactions are the norm rather than the exception. This leads to complex ENs with a large number of
complex interactions, which require appropriate concepts for disentangling them. Computer science has
developed conceptual and technical tools to handle such non-dyadic interactions and compute their
dynamic consequences. Known as *hypergraphs*, a number of associated tools have been developed to
handle them (e.g., Reisig 2013, Battiston et al. 2020).
195
A *hypergraph* is a generalization of a graph in which edges, called *hyperedges*, may have several
incoming and outgoing nodes. In ecology, any variable may be connected through many processes and,
above all, any process may connect many variables, as in the case of the termite colony (Fig. 3a). In our
studies, we use hypergraphs by dictating that nodes are the material components of social-ecological
systems and that edges are the abstract processes involving these components (Mao et al. 2021, Cosme et
al. 2022). Hence, any hypergraph is a means of representing the state of the ecosystem via the composition
of its components, while still considering their associated interactions. In addition, it is possible to define
the processes involved on the basis of equations or rules that represent the components' interactions. Such
process definitions, be they dyadic or not, naturally direct the associated EN hyperedges in most
environmental studies. Any hypergraph may be represented by a Petri net, among other convenient tools
developed in computer science (Fig. 3b). Petri nets were developed in the 1930s for modeling chemical
reactions requiring precisely such connections and have since been used in many other fields (Pommereau
2010, Reisig 2013, Pommereau et al. 2022b). In such a representation, variables are known as *places*
(round-shaped nodes) and interact through processes called *transitions* (square-shaped nodes), shown as
arcs. Any hyperedge may be represented by the set of arcs (or *hyperarcs*) of a Petri net, connecting the
nodes as the pre-condition and other nodes as the post-condition of the focal transition (Fig. 3b). Petri net
arcs may be weighted, and places may be multi-valued, holding varied numbers of tokens, thus
constraining their flows in the Petri net. Such bipartite graphs are equivalent to hypergraphs as they allow
an arbitrary number of nodes to be connected to an arbitrary number of edges and vice versa (e.g., the
termite colony EN, Fig. 3c).

Figure 3

Simultaneously, it becomes relevant to analyze the EN structure itself and to highlight the graph
topology for ecological insights (Fig. 2b-3b). Here, no shape is assumed for the EN; instead, the graph
topology is allowed to emerge from the chosen variables and known interactions, bringing interacting
variables closer and moving non-interacting variables further out. With an appropriate graph layout (Fig.

2b), algorithms and graph analyses (Kamada and Kawai 1989), it is possible to produce representations that
help in understanding node properties (e.g., whether they influence or are influenced by other nodes) and
many other properties (e.g., betweenness, connectivity index) as is already routinely the case in ecology
(e.g., Urban and Keitt 2001, Foltête and Giraudoux 2012). For example, a relevant *boundary* of the
ecosystem under examination may be computed by excluding nodes that influence others but are not
influenced in turn (e.g., the sun and other so-called *control variables*), and including all the others (e.g., a
species simultaneously eating and being eaten) (Fig. 2b).

Qualitative ecosystem dynamics
Graphs and hypergraphs are by definition discrete. However, the interaction network they represent
might not be (Dicks et al. 2002, Landi et al. 2018, Valverde et al. 2020), as their nodes may represent
continuous variables and their edges continuous flows between them (Fig. 1a). In this paper, we propose
to approximate the interaction network of any ecosystem with a qualitative hypergraph (still generically
denoted as 'EN' hereafter). This hypergraph carries information about the potential qualitative dynamics
of components, namely their appearance and disappearance (Fig. 1b). This description is relevant for long-
term dynamics, as an observer who waits long enough will no longer measure flows and abundances, but
appearance and disappearance of system components (and processes). In this spirit, we define a system
component as a *Boolean* variable, i.e. having two discrete values, describing the functional
presence/absence of the component in the system. A component that is functionally present is considered
to pass a predefined threshold, present in most non-linear processes (Thomas 1973), and to be able to
influence (or be influenced by and possibly, no more influence) other components (Fig. 1b). Conversely, a
process (potentially generating an *event*) is defined as a discrete change in the values of certain variables
(the consequence) triggered by other variables (the condition). According to Petri net formalism
(Pommereau 2010, Pommereau et al. 2022b), these events are represented by *rules* written as "condition
>> consequence" (Table 1). The change in one (or more) variable(s) is called a *transition*, which creates a
new system *state*.

Table 1 - List of the eight termite colony variables with their acronyms and descriptions (columns 1
and 2), and the ten termite colony processes with their descriptions (columns 3 and 4). Simplified
from the initial publications (Gaucherel et al. 2017, Gaucherel and Pommereau 2019, Gaucherel et
al. 2020).

inhabitants:	Descriptions	rules:	Descriptions
Rp+:	Reproductives	R1: Rp+ >> Ec+	Reproductives lay eggs
Wk-:	Workers	R2: Rp+, Ec+ >> Wk+	Eggs grow into workers
Sd-:	Soldiers	R3: Wk+ >> Wd+, Te+, Fg+, Ec+	Workers collect wood resource and maintain termitomyces and their gardens
Te-:	Termitomyces	R4: Wk+, Wd+ >> Sd+, Rp+	Workers and resources supply soldiers and maintain reproductives
(spatial) structures:		R5: Wk+, Te+ >> Wd-	Workers and termitomyces consume wood resources
Ec-:	Egg chambers	R6: Wd- >> Wk-, Te-	Without wood resources, workers and termitomyces die
Fg-:	Fungal gardens	R7: Wk- >> Fg-, Te-, Sd-	Without workers, fungal gardens (and thus termitomyces) and soldiers disappear
Resources and competitors:		R8: Wk-, Rp- >> Ec-	Without workers and reproductives, eggs cannot be laid
Wd-:	Wood	R9: Ac+, Sd- >> Wk-, Rp-	When competitors attack the termite colony without soldiers, workers and reproductives die

Ac+:	Ant competitors

The sequences of states and transitions are the system *dynamics*. The dynamics of a qualitative EN would enable the sharp changes in the long-term dynamics of the ecosystem to be understood. In a way, the single qualitative EN constitutes the *skeleton*, a kind of backbone of the (social-)ecological system under study, intentionally neglects subtle variations and flows (Gaucherel and Pommereau 2019, Cosme et al. 2022), i.e. the component dynamics, and enables the computing of qualitative ecosystem dynamics, the network dynamics (Gaucherel et al. 2017, Gaucherel 2019) (Appendix 1). An EN ecosystem interaction network hosts the tangled dynamics of species abundances and other components and of flows between them. Many studies have developed powerful models to simulate and understand such dynamics in networks with frozen topology (Thébault and Fontaine 2010, Kéfi et al. 2015). Using discrete-event models, it becomes possible to model changing networks and their discrete dynamics (i.e., state and transitions). A system state is defined by the values of all the variables. In the graph, a system state corresponds to a given token distribution (called the *marking* in Petri nets) among the nodes (Fig 4a). A state transition occurs when the marking changes, that is to say, when one or more tokens are produced or consumed in the graph. Starting from an initial state (blue hexagone in Fig 4a), the graph enables the computation of all reachable system states and transitions (following green and pink states). These states and transitions are usually represented as a State-Transition Graph (STG, Fig. 4a) (Berthomieu et al. 2004, Reisig 2013).

In this STG, any cycle (i.e. a set of mutually reachable states) or fixed state (a state with no outgoing transition) becomes now identifiable and informative (Gaucherel and Pommereau 2019, Gaucherel et al. 2020) (colors in Fig. 4a). Hence, it is possible to automatically compute the discrete-event dynamics of the system from the hypergraph structure (Fig 4b). This is precisely what we modeled in several studies from a chosen EN of temperate and tropical social-ecological systems (Mao et al. 2021, Cosme et al. 2022). By using specific tools, it is also possible to formalize specific questions about the dynamics displayed in the STG, without explicitly computing them, a relevant advantage when it becomes too large to be fully computed (Fages et al. 2004, Largouët et al. 2012, Thomas et al. 2022). In addition to an intelligible visualization of ecosystem structure and dynamics, we now have the benefit of the whole toolbox associated with such representations (Fig. 4b), as with the Petri nets that computer science has developed (Berthomieu et al. 2004, Rodríguez and Schwoon 2013, Pommereau et al. 2022b). The demonstration of the relevance and application of Petri nets and related concepts is outside the scope of this conceptual paper, but interested readers may refer to recent papers (Gaucherel and Pommereau 2019, Mao et al. 2021, Cosme et al. 2022). Such models automatically produce STGs comparable to those already developed in many empirical ecological studies (Stringham et al. 2003, Bestelmeyer et al. 2017). Looking ahead, several computing tools are even able to automatically detect various kinds of topological structures and regularities/invariants directly within the EN (Rodríguez and Schwoon 2013, Thomas et al. 2022). For example, in theory, individuals or populations that may exhibit cycles or that may collapse could be detected by such invariants directly within the EN representation (Fig. 4b). This view supports an innovative conceptual representation of the ecosystem.

a)

b)

Figure 4

A changing (and spatial) hypergraph

303
Disregarding for a moment variable abundances and process flows, this EN is in no way frozen. In the
long term, each component and each process potentially disappears and (re)appears in line with the
presence/absence of its material components (Fig. 1b). They can appear (e.g., species invasions or
pollutants) and/or disappear (species extinctions or migrations out of the system being studied). The aim
is to model network dynamics, i.e., a changing topology (the neighboring relationships), not to be confused
with the dynamics of components in a static network (flows, "carried" by the frozen graph). In the long

term, the nodes and edges of the graphical representation themselves become fragile and ephemeral, depending on the other connected variables, what previous studies have called *ecosystem development* (Gaucherel et al. 2017, Gaucherel 2018). Conversely, in short-term dynamics with a similar level of detail, the EN retains the same structure (Fig. 1 or any state in Fig 4a). This assumption of a fixed topology is still widely used in most ecological studies today (e.g., Proulx et al. 2005, Golubski et al. 2016, Kéfi et al. 2016).

In the long term, the (qualitative) EN topology is continuously changing, and these structural (or network) dynamics may be largely disconnected from short-term component dynamics. Any sharp change in the system will change the network topology, which is modeled by the token flow. Modeling such a changing topology is highly difficult, however, because it implies anticipating any disappearance and (above all) appearance of components. Practically, it is easier to initially assume the *maximal* EN topology by including all possible components, and then mimicking their possible disappearance by removing their associated nodes and/or tokens. For example, in our past studies, we model structural changes in the EN representing the appearance and disappearance of components by changing the Boolean values of their associated variables and nodes (Fig. 1b). This is known as the EDEN (for Ecological and Social Discrete-Event Network) framework (Gaucherel and Pommereau 2019, Pommereau et al. 2022a). Such changes are generated by certain rules representing system processes and their related transition events. The STG precisely computes the successive topologies of the EN, considering the present and absent components and processes of the termite colony (nodes in Fig. 4a). Hence, any STG state displays a distinct EN topology, mimicked by the variables changing states (Fig. 3b). This makes the EN an ever-changing hypergraph whose topology needs to be reconstructed for an understanding of the long-term dynamics (Appendix 2).

In parallel, providing another perspective, all the points discussed previously may be included in spatially implicit or explicit ecosystem models. Indeed, space itself is easily represented as a graph: in the case of a spatially explicit ecosystem, the localities involved are potentially represented as nodes or (unfolded as) a set of nodes themselves, and may be either individualized or merged into the same EN. Space has long been modeled in this way, as a graph with changing topologies (Burel and Baudry 2003, Gaucherel et al. 2012). Conceptually, there is nothing to prevent the merging of the spatial representation in the graph with the more functional EN we have described here: whenever various localities need to be described in the same extensive overall super-graph (Felipe-Lucia et al. 2021), their own EN (but not their corresponding spatial node) may be merged with neighboring ENs too. Hence, each ecosystem may be represented as a changing single hypergraph, including its spatial structure.

Discussion and Conclusion

The so-called ecosystem hypernetwork (EN, i.e., the social-ecological interaction hypergraph) fits our need for efficient representation of any complex ecosystem, and offers several additional advantages. It summarizes the internal structure of the system, a kind of synthetic skeleton of the components and processes involved. In this study, we highlighted three points, illustrated with specific Petri net models. Firstly, it is possible to collect all the ecosystem's interactions into a single EN, regardless of their nature, number and how they intertwine. This proposed representation still shows intentional limitations: for example, in this representation, components never interact directly; here, interactions are always connected to a process. Similarly, no process can impact directly another process without being connected to a component. This single EN is possibly not unique, as ecosystem models are driven by a research question, thus leading to various representations of the same system.

Secondly, the EN is modelled as a hypergraph, which represents non-dyadic interactions more
explicitly. Events and components defining the system can be represented as distinct nodes of a bipartite
graph, as in Petri nets or, in our case, a qualitative hypernetwork (Fig. 3). Thirdly, the EN representing any
ecosystem is dynamic, with a changing topology. We provided an example of a framework (EDEN)
computing qualitative ecosystem dynamics based on transitions producing/consuming tokens in the
hypergraph. This enables a State-Transition Graph to be computed, collecting all the ecosystem states
reachable from the initial state, according to predefined events (Fig. 4). This provides a coherent and
effective overview of the dynamics of any ecosystem.

Appendices

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Data, scripts, code, and supplementary information availability

Scripts and code are available online: No data were used in this study; the method sources are archived at https://github.com/fpom/ecco

Conflict of interest disclosure

The authors declare that they comply with the PCI rule of having no financial conflicts of interest in relation to the content of the article.

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References

Bastolla, U., M. A. Fortuna, A. Pascual-Garcia, A. Ferrera, B. Luque, and J. Bascompte. 2009. The architecture of mutualistic networks minimizes competition and increases biodiversity. Nature **458**:1018-1020.

Battiston, F., G. Cencetti, I. Iacopini, V. Latora, M. Lucas, A. Patania, J.-G. Young, and G. Petri. 2020. Networks beyond pairwise interactions: Structure and dynamics. Physics Reports **874** 1-92.

Berthomieu, B., P.-O. Ribet, and F. Vernadat. 2004. The tool TINA -- Construction of Abstract State Spaces for Petri Nets and Time Petri Nets. International Journal of Production Research **42**.

Bestelmeyer, B. T., A. Ash, J. R. Brown, B. Densambuu, M. Fernández-Giménez, J. Johanson, M. Levi, D. Lopez, R. Peinetti, L. Rumpff, and P. Shaver. 2017. State and Transition Models: Theory, Applications, and Challenges. Pages 303-345 Rangeland systems. Springer.

Billick, I., and T. J. Case. 1994. Higher Order Interactions in Ecological Communities: What Are They and How Can They be Detected? Ecology **75**:1530-1543.

Brummitt, C. D., R. M. D'Souza, and E. A. Leicht. 2012. Suppressing cascades of load in interdependent networks. Proceedings of the National Academy of Sciences **109**:E680-E689.

Buldyrev, S. V., R. Parshani, G. Paul, H. E. Stanley, and S. Havlin. 2010. Catastrophic cascade of failures in interdependent networks. Nature **464**:1025-1028.

Burel, F., and J. Baudry. 2003. Landscape ecology: concepts, methods, and applications. Science Publishers, Enfield.

Cosme, M., C. Hély, F. Pommereau, P. Pasquariello, C. Tiberi, A. C. Treydte, and C. Gaucherel.
2022. Qualitative Modeling for Bridging Expert-Knowledge and Social-Ecological Dynamics
of an East African Savanna. Land **11**:https://doi.org/10.3390/land11010042.
Cumming, G. S., A. Buerkert, E. M. Hoffmann, E. Schlecht, S. v. Cramon-Taubadel, and T.
Tscharntke. 2014. Implications of agricultural transitions and urbanization for ecosystem
services. Nature **515** 50-57.
Delmas, E., M. Besson, M.-H. Brice, L. A. Burkle, G. V. Dalla Riva, M.-J. Fortin, D. Gravel, P. R.
Guimarães Jr., D. H. Hembry, E. A. Newman, J. M. Olesen, M. M. Pires, J. D. Yeakel, and T.
Poisot. 2018. Analysing ecological networks of species interactions. Biological Reviews
94:16-36.
Dicks, L. V., S. A. Corbet, and R. F. Pywell. 2002. Compartmentalization in plant–insect flower
visitor webs. Journal of Animal Ecology **71**:32-43.
Fages, F., T. Martinez, D. Rosenblueth, and S. Soliman. 2018. Influence Networks compared with
Reaction Networks: Semantics, Expressivity and Attractors. IEEE/ACM Transactions on
Computational Biology and Bioinformatics, Institute of Electrical and Electronics Engineers
99:1-14.
Fages, F., S. Soliman, and N. Chabrier-Rivier. 2004. Modelling and querying interaction networks in
the biochemical abstract machine BIOCHAM. Journal of Biological Physics and Chemistry
4:64-73.
Felipe-Lucia, M. R., A. M. Guerrero, S. M. Alexander, J. Ashander, J. A. Baggio, M. L. Barnes, Ö.
Bodin, A. Bonn, M.-J. Fortin, R. S. Friedman, J. A. Gephart, K. J. Helmstedt, A. A. Keyes, K.
Kroetz, F. Massol, M. J. O. Pocock, J. Sayles, R. M. Thompson, S. A. Wood, and L. E. Dee.
2021. Conceptualizing ecosystem services using social–ecological networks. Trends in
Ecology & Evolution **2944**:12.
Foltête, J. C., and P. Giraudoux. 2012. A graph-based approach to investigating the influence of the
landscape on population spread processes. Ecological Indicators **18**:684-692.
Fontaine, C., P. R. Guimaraes Jr, S. Kéfi, N. Loeuille, J. Memmott, W. H. van der Putten, F. J. F. van
Veen, and E. Thébault. 2011. The ecological and evolutionary implications of merging
different types of networks. Ecology Letters **14**:1170-1181.
Fontana, W., and L. W. Buss. 1994. What would be conserved if "the tape were played twice"?
Proceedings of the National Academy of Sciences (PNAS) **91**:757-761.
Frontier, S., D. Pichod-Viale, A. Lepêtre, D. Davoult, and C. Luczak. 2008. Ecosystèmes. Structure,
Fonctionnement, Evolution. 4ème édition edition. Dunod, Paris.
Gaucherel, C. 2018. Physical concepts and ecosystem ecology: a revival? Journal of Ecosystem and
Ecography **8**:DOI: 10.4172/2157-7625.1000259.
Gaucherel, C. 2019. The Languages of Nature. When nature writes to itself. Lulu editions, Paris,
France.
Gaucherel, C., F. Boudon, T. Houet, M. Castets, and C. Godin. 2012. Understanding Patchy
Landscape Dynamics: Towards a Landscape Language. PLoS One **7**:e46064.
Gaucherel, C., and F. Pommereau. 2019. Using discrete systems to exhaustively characterize the
dynamics of an integrated ecosystem. Methods in Ecology and Evolution **00**:1–13.
Gaucherel, C., F. Pommereau, and C. Hély. 2020. Understanding ecosystem complexity via
application of a process-based state space rather than a potential. Complexity **Article ID**
7163920,.
Gaucherel, C., H. Théro, A. Puiseux, and V. Bonhomme. 2017. Understand ecosystem regime shifts
by modelling ecosystem development using Boolean networks. Ecological Complexity
31:104-114.
Gignoux, J., I. D. Davies, S. R. Flint, and J.-D. Zucker. 2011. The ecosystem in practice: interest and
problems of an old definition for constructing ecological models. Ecosystems **14** 1039-1054.
Golubski, A. J., E. E. Westlund, J. Vandermeer, and M. Pascual. 2016. Ecological Networks over the
Edge:Hypergraph Trait-Mediated Indirect Interaction (TMII) Structure trends in Ecology &
Evolution **31**:344-354.
Gross, T., and H. Sayama. 2009. Adaptive Networks: Theory, Models and Applications. Springer.
Kamada, T., and S. Kawai. 1989. An algorithm for drawing general undirected graphs. Information
Processing Letters **31**:7-15.

Kéfi, S., E. L. Berlow, E. A. Wieters, L. N. Joppa, S. A. Wood, U. Brose, and S. A. Navarrete. 2015.
Network structure beyond food webs: mapping non-trophic and trophic interactions on
Chilean rocky shores. Ecology **96**:291-303.

Kéfi, S., V. Miele, E. Wieters, S. Navarrete, and E. Berlow. 2016. How Structured Is the Entangled
Bank? The Surprisingly Simple Organization of Multiplex Ecological Networks Leads to
Increased Persistence and Resilience. PLoS Biology **14**:e1002527.

Kéfi, S., E. Thébault, A. Eklöf, M. Lurgi, A. J. Davis, M. Kondoh, and J. A. Krumins. 2017. Toward
Multiplex Ecological Networks: Accounting for Multiple Interaction Types to Understand
Community Structure and Dynamics.*in* J. C. Moore, P. C. d. Ruiter, K. S. McCann, and V.
Wolters, editors. Adaptive Food Webs. Stability and Transitions of Real and Model
Ecosystems. Cambridge University Press, Published online.

Klamt, S., U.-U. Haus, and F. Theis. 2009 Hypergraphs and Cellular Networks. PLoS Computational
Biology **5**:e1000385.

Landi, P., H. O. Minoarivelo, Å. Brännström, C. Hui, and U. Dieckmann. 2018. Complexity and
stability of ecological networks: a review of the theory. Population Ecology **60**:319-345.

Largouët, C., M.-O. Cordier, Y.-M. Bozec, Y. Zhao, and G. Fontenelle. 2012. Use of timed automata
and model-checking to explore scenarios on ecosystem models. Environmental Modelling &
Software **30** 123-138.

Lobry, C., and T. Sari. 2015. Migrations in the Rosenzweig-MacArthur model and the "atto-fox"
problem. ARIMA Journal **20**:95-125.

Loreau, M., N. Mouquet, and R. D. Holt. 2003. Meta-ecosystems: a theoretical framework for a
spatial ecosystem ecology. Ecology Letters **6**:673-679.

Mao, Z., J. Centanni, F. Pommereau, A. Stokes, and C. Gaucherel. 2021. Maintaining biodiversity
promotes the multifunctionality of social-ecological systems: Holistic modelling of a
mountain system. Ecosystem services **47**:101220.

Nakazawa, T. 2020. Species interaction: Revisiting its terminology and concept. Ecological Research
35:1106-1113.

Pilosof, S., M. A. Porter, M. Pascual, and S. Kéfi. 2017. The multilayer nature of ecological networks.
Nature Ecology and Evolution **1**:1-9.

Pommereau, F. 2010. Algebras of coloured Petri nets. Lambert Academic Publishing (LAP).

Pommereau, F., C. Thomas, and C. Gaucherel. 2022a. EDEN framework for interactive analysis of
ecosystems models.*in* PNSE'22: International Workshop on Petri Nets and Software
Engineering, Bergen, NW.

Pommereau, F., C. Thomas, and C. Gaucherel. 2022b. Petri Nets Semantics of Reaction Rules: A
Language for Ecosystems Modelling. Pages 175-194 *in* Application and Theory of Petri Nets
and Concurrency (PTRE NETS), Bergen, Norway.

Proulx, S. R., D. E. L. Promislow, and P. C. Phillips. 2005. Network thinking in ecology and
evolution. Trends in Ecology and Evolution **20**:345-353.

Reisig, W. 2013. Understanding Petri Nets. Springer Berlin Heidelberg, Berlin, Heidelberg.

Rodríguez, C., and S. Schwoon. 2013. Cunf: A tool for unfolding and verifying Petri nets with read
arcs. Springer.

Sanchez, A. 2019. Defining Higher-Order Interactions in Synthetic Ecology: Lessons from Physics
and Quantitative Genetics Cell Systems **9**:519-520.

Scholes, R., B. Reyers, R. Biggs, M. Spierenburg, and A. Duriappah. 2013. Multi-scale and cross-
scale assessments of social-ecological systems and their ecosystem services. Current Opinion
in Environmental Sustainability **5**:16-25.

Schwartz, A., and K. Jax. 2011. Ecology revisited. Reflecting on Concepts, Advancing Science.
Springer.

Stringham, T. K., W. C. Krueger, and P. L. Shaver. 2003. State and transition modeling: An
ecological process approach. Journal of Range Management **56**:106 -113.

Strogatz, S. H. 2001. Exploring complex networks. Nature **410**:268-276.

Thébault, E., and C. Fontaine. 2010. Stability of Ecological Communities and the Architecture of
Mutualistic and Trophic Networks. Science **329**:853 -856.

Thomas, C., M. Cosme, C. Gaucherel, and F. Pommereau. 2022. Model-checking ecological state-
transition graphs. PLoS Computational Biology **18**:e1009657.

Thomas, R. 1973. Boolean formalization of genetic control circuits. Journal of Theoretical Biology
42:563-585.
Tixier, P., N. Peyrard, J.-N. Aubertot, S. Gaba, J. Radoszycki, G. Caron-Lormier, F. Vinatier, G.
Mollot, and R. Sabbadin. 2013. Modelling Interaction Networks for Enhanced Ecosystem
Services in Agroecosystems. Pages 437-480 Advances in Ecological Research. Elsevier Ltd
Turner, J. S. 2009. The extended organism: the physiology of animal-built structures. Harvard
University Press.
Urban, D., and T. Keitt. 2001. Landscape Connectivity: A Graph-Theoretic Perspective. Ecology
82:1205-1218.
Valverde, S., B. Vidiella, R. Montañez, A. Fraile, S. Sacristán, and F. García-Arenal. 2020.
Coexistence of nestedness and modularity in host–pathogen infection networks. Nature
Ecology & Evolution:https://doi.org/10.1038/s41559-41020-41130-41559.
Werner, E. E., and S. D. Peacor. 2003. A review of trait-mediated indirect interactions in ecological
communities. Ecology **84**:1083–1100.
523