Zinc stimulation of phytoplankton in a low carbon dioxide, coastal Antarctic environment

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Abstract

Zinc (Zn) is a key micronutrient used by phytoplankton for carbon (C) acquisition, yet there have been few observations of its influence on natural oceanic phytoplankton populations. In this study, we observed Zn limitation of growth in the natural phytoplankton community of Terra Nova Bay, Antarctica, due to low (~220 μatm) pCO_2 conditions, in addition to primary iron (Fe) limitation. Shipboard incubation experiments amended with Zn and Fe resulted in significantly higher chlorophyll *a* content and dissolved inorganic carbon drawdown compared to Fe addition alone. Zn and Fe response proteins detected in incubation and environmental biomass provided independent verification of algal co-stress for these micronutrients. These observations of Zn limitation under low pCO_2 conditions demonstrate Zn can influence coastal primary productivity. Yet, as surface ocean pCO_2 rises with continued anthropogenic emissions, the occurrence of Zn/C co-limitation will become rarer, impacting the biogeochemical cycling of Zn and other trace metal micronutrients.

Significance Statement

Zinc has tremendous nutritional importance in life, and due to its central role in carbon acquisition in major phytoplankton species, Zn is directly connected to ocean carbon cycling. As a result, it has been hypothesized that co-limitation by zinc and CO_2 could influence primary productivity, but a connection between Zn and CO_2 has not yet been observed in field experiments. The present study demonstrates that Zn can be a limiting nutrient in natural waters, particularly in coastal environments where CO_2 is diminished by bloom events, and explores the implications for Zn-C co-limitation in the context of climate change— the oceans may have

decreased susceptibility to Zn-C co-limitation in the future due to the rapid input of

anthropogenic CO_2 to the atmosphere.

Introduction

Primary productivity in the oceans is a key component of the global carbon cycle and is

largely controlled by the availability of nitrogen (N), phosphorus (P), and iron (Fe). Yet there is

increasing evidence that other micronutrients such as zinc (Zn), cobalt (Co), and vitamin B_{12} can

also influence phytoplankton productivity, often as secondary limiting nutrients after N, P, or Fe

are added[1]. Zn can be particularly scarce in the photic zone[2,3] where total dissolved Zn (dZn_T)

can be as low as 0.1 nM in seawater due to biological uptake and complexation by organic

ligands[4,5] which further lower Zn bioavailability[6–8]. Marine eukaryotic algae and copiotrophic

bacteria possess a large metabolic demand for Zn that is on par with that of Fe[6,9] and polar

phytoplankton in particular demonstrate an enhanced Zn demand based on genomic and

laboratory analyses[10]. Despite the scarcity of bioavailable Zn in the surface ocean and its high

cellular demand, there have been relatively few experimental studies examining the ability of Zn

addition to stimulate natural phytoplankton communities **(Supplementary Table 1)**, with

findings that include negative results[11–13], slight Zn stimulatory results[14], Zn stimulation within

Fe and Si uptake experiments[15], Zn primary and secondary limitation in the North Pacific in an

unreplicated experiment[3], and secondary Zn limitation after primary Si limitation in the Costa

Rica Dome[16].

In contrast to the limited field evidence, laboratory studies have unequivocally

demonstrated that marine phytoplankton can easily be Zn-limited in culture due to their large Zn

requirement and that this effect is exacerbated at low pCO_2[17,18] due to the use of Zn as a required

catalytic cofactor within carbonic anhydrase (CA) metalloenzymes[18]. CAs catalyze the reversible

dehydration of HCO_3^- to CO_2, the substrate required by the carbon fixing enzyme RUBISCO. As

HCO_3^- constitutes about 90% of the dissolved inorganic carbon (DIC) pool in the surface ocean,

sufficient CA activity works to prevent carbon stress in marine phytoplankton by ensuring

adequate CO_2 supply to RUBISCO. In some CA classes, the Zn cofactor can be substituted for

by Co[19,20], and in some species, alternate CAs can be produced that use Cd or manganese

(Mn)[21,22] as the cofactor, conferring a biochemical substitution flexibility with respect to metal

availability known as cambialism[7]. While low pCO_2 events are frequently observed in coastal

environments due to high biological productivity[23–25], a connection between Zn and CO_2 has not

yet been observed in field experiments, and it has remained unclear under what conditions Zn

can act as a limiting oceanic micronutrient. However, a connection between Cd, which can

metabolically substitute for Zn when Zn is depleted[26], and CO_2 has been observed, with

enhanced Cd uptake and increased particulate Cd in productive coastal Californian waters

experiencing low pCO_2 conditions[27,28].

As low CO_2 conditions often result during algal blooms in coastal environments, it has

been hypothesized that the combination of high biomass and resulting low CO_2 in these

environments may cause severe Zn depletion that may limit algal growth rates[18]. In this study,

we assessed these environments for evidence of Zn limiting conditions in the field by performing

Zn addition experiments with the natural phytoplankton assemblage in Terra Nova Bay (TNB),

Antarctica. As TNB is located at the western edge of the Ross Sea and hosts large seasonal

blooms dominated by diatoms[29] and the prymnesiophyte *Phaeocystis antarctica*[30,31] (*Phaeocystis*

hereon) that result in low CO_2 [29,32], TNB was an ideal site to test for Zn/C co-limitation in the

field.

Results and discussion

Biogeochemical characterization of Terra Nova Bay

We begin by biogeochemically characterizing the study site within Terra Nova Bay

(TNB), Antarctica, which was temporally sampled for over one month (January 9 – February 18,

2018) during the 2017-2018 CICLOPS expedition (**Fig. 1a; Supplementary Table 2**) to

concurrently characterize the natural phytoplankton bloom progression and biogeochemical

changes in the water column. Low (~200 μatm) pCO_2 surface waters throughout TNB (**Fig. 1b**)

resulted in a potentially Zn/C co-limiting environment for native phytoplankton.

TNB hosted a large phytoplankton bloom as indicated by high (> 3000 ng L^{-1})

chlorophyll *a* concentrations in January that waned into February (**Fig. 1c**), that is characteristic

of Antarctic polynya environments. This phytoplankton community initially consisted of a mixed

assemblage of both diatoms as indicated by fucoxanthin (fuco, **Fig. 1d**) and *Phaeocystis* as

verified by shipboard microscopy and as indicated by 19'-hexanoyloxyfucoxanthin (19'-hex,

Fig. 1e). Surface fucoxanthin concentrations >200 ng L^{-1} were observed at the late TNB stations

(**Fig. 1d**) while 19'-hex decreased to ~20 ng L^{-1} (**Fig. 1e**), indicating that the stations sampled in

late February were dominated by diatoms rather than *Phaeocystis*. This was consistent with

historical observations of phytoplankton succession patterns in TNB[30,31,33]. Pronounced and

progressively deepening total dissolved Zn (dZn$_T$) depletion was observed in our temporal, upper

water column sampling study of the TNB region spanning over a month (January-February 2018;

Fig. 1f). dZn was depleted down to an average of 0.82 ± 0.47 nM at 10 m over all TNB stations

with Zn-depleted areas of the water column progressively deepening to 50 m and then to ~100 m

throughout January and February (**Fig. 1f**) indicative of strong Zn uptake and export from the

euphotic zone. In the Ross Sea, dissolved iron (dFe) has previously been demonstrated to be the

primary limiting nutrient for phytoplankton growth[12,34,35]. Consistent with this, across all TNB

stations, dFe_T in the upper 50 m remained below 1 nM (**Supplementary Figure 1a**) as observed

previously in this region[36].

Temporal dynamics of the TNB bloom (**Supplementary Figure 12**) further showed that

total Zn uptake (ρZn, measured concurrently using a new stable isotope tracer method[37]) was

highest in the shallow euphotic zone in early January and waned into February (**Fig. 1g**),

following trends seen in chlorophyll *a* (chl *a,* **Fig. 1c**) and the phytoplankton pigments

fucoxanthin (**Fig. 1d**) and 19'-hex (**Fig. 1e**). This ρZn trend was consistent with laboratory

studies demonstrating the substantial Zn requirements of both diatoms and *Phaeocystis*

antarctica[20,38]. Within TNB, the negative linear relationship (n = 15, R^2 = 0.63) between surface

seawater (5 m underway system) pCO_2 partial pressure and ρZn (10 m) (**Fig. 1h**), as well as the

positive linear relationship (n=9, R^2= 0.65) between chlorophyll *a* concentrations and ρZn, both

measured at 10 m (**Fig. 1i**), demonstrated that the combination of high biomass and low CO_2 lead

to substantial increases in Zn uptake. A positive linear trend (n = 27, R^2 = 0.87) comparing

surface seawater pCO_2 and DIC concentrations confirmed this biological influence on carbon

dynamics, with pCO_2 responding to DIC rather than alkalinity or salinity (**Supplementary**

Figure 2).

Biogeochemical characterization of the incubation study site

We next describe the biogeochemistry of the water column at one TNB station (station

27, referred to as the "experimental site" herein) as this particular site was chosen for the

multifactor shipboard incubation experiment (**Fig. 1a,b**; red star). This experimental site

harbored a coastal bloom and was biologically and chemically characterized as having high *in*

situ chlorophyll *a* levels (maximum of 3259 ng L^{-1} at 30 m; **Supplementary Figure 3a**), and a

decrease in total dissolved inorganic carbon (DIC_T; 2186 μmol kg^{-1} at 5 m compared to the deep

water (200-1065 m) average of 2224 ± 2.1 µmol kg^{-1}, **Supplementary Figure 3b**). Within the

water column, dZn demonstrated a pronounced decrease from 5.1 nM at 50 m to 0.9 nM at 10 m,

representing an 82% decrease (and a 76% decrease comparing the minimum dZn_T value at 10 m

to the average deepwater (210 – 1000 m) concentration of 3.9 nM $\pm$ 0.4; **Supplementary Figure**

3c), consistent with prior observations of surface dZn depletion in this region[36]. While Zn

speciation was not measured in this study, it is likely that Zn organic complexation further

depleted total bioavailable Zn. A concurrent study of Co speciation observed complete

complexation of dissolved Co in the upper water column in TNB, consistent with this idea[39].

Consistent with high macronutrient abundance in this region (as observed previously and in the

present study; **Supplemental Figure 1b,c,d**), surface macronutrient concentrations (nitrate and

nitrite, phosphate, and silicic acid) were partially depleted at the experimental site with 64%,

46%, and 29% decreases in N+N, P, and Si, respectively, comparing 10 m and average deep

water (200 – 1000 m) values (**Supplementary Figure 3d**). Observations of rapid Zn uptake at

the experimental site (**Supplementary Figure 3e**) can explain the dZn_T drawdown described

above—Zn uptake rates of 46 pmol L^{-1} d^{-1} (at 10 m) are of the appropriate scale to induce the

multi-nanomolar surface water depletion during the austral spring and summer season.

Evidence for Zn stimulation of phytoplankton in shipboard incubations

A multifactor incubation experiment was conducted using surface waters collected by

trace metal clean fish sampler (7 m) fed into a shipboard cleanroom at the experimental site to

examine controls on primary productivity, with triplicate treatments of Zn amended (+Zn; 2 nM

as $ZnCl_2$), Fe amended (+Fe; 1 nM as $FeCl_2$), and Fe and Zn amended (+Fe+Zn) incubations, in

addition to unamended controls. Addition of Fe alone (+Fe) resulted in significantly higher chl *a*

content compared to controls (p = 9.5e-5) after six days (T6) (**Fig. 2a**), demonstrating primary Fe

limitation as observed previously in the Ross Sea[31,34]. However, addition of Zn alone (+Zn) also

resulted in significantly higher chl a content compared to the controls (p = 0.011), implying that

a subset of the incubated phytoplankton population benefitted from the addition of Zn alone,

without additional Fe, and was thus experiencing primary Zn limitation (**Fig. 2a**). The combined

addition of +Fe+Zn resulted in the highest average chl a content among all treatments at T6, with

4.5 ± 0.21 µg L^{-1} compared to 3.9 ± 0.35 µg L^{-1} and 2.8 ± 0.15 µg L^{-1} achieved by +Fe and +Zn

alone, respectively, demonstrating additive co-limitation between Zn and Fe[40]. The Zn stock

solution was analyzed to confirm these results were not caused by inadvertent iron contamination

(see Methods). Moreover, significant differences between controls and metal treatments (and low

variance within treatments) indicated source seawater was also not metal contaminated.

Significant differences in seawater chemistry were also observed within these incubations over

time, with larger decreases in DIC$_T$ in all metal treatments compared to the control (-12.7 µmol

kg^{-1} for +Fe (p = 5.3e-6), -8.2 µmol kg^{-1}for +Zn (p = 5e-5), and -18.5 µmol kg^{-1} for +Fe+Zn (p =

2.2e-16); **Fig. 2b**). The decrease in DIC$_T$ observed with +Fe+Zn was significantly larger than

that achieved with +Fe alone (p = 4.4e-3; **Fig. 2b**). Statistically significant differences in

measured parameters among treatments are summarized in **Supplementary Table 3.**

The application of algal Fe- and Zn-stress proteins provided an independent line of

evidence corroborating the results described above. This included the detection of the recently

described Zn/Co responsive protein ZCRP-A (a putative Zn chaperone)[41] as a biomarker of Zn

stress as well as the iron starvation-induced proteins ISIP1A, ISIP2A and ISIP3 (ISIPs) as

biomarkers of Fe stress (**Supplementary Table 4**). The ISIPs represent a group of unrelated

proteins that are upregulated under Fe limitation in various algal species. ISIP1 proteins are

responsible for endocytosis of siderophore-bound iron, ISIP2 proteins are involved in Fe^{3+}

uptake, and ISIP3 has been suggested to act as an Fe storage protein[42,43]. Within the T6

incubation biomass, there was an increased abundance of ISIPs in the control and +Zn treatment

and a corresponding decrease in ISIP protein abundance within the +Fe and +Fe+Zn treatments,

consistent with primary Fe limitation and the expected response to Fe addition (**Fig. 2c**). ISIPs

were taxonomically assigned to diatoms, *Phaeocystis*, and dinoflagellates (**Supplementary**

Figure 4a). Similarly, we detected the presence of ZCRP-A in the T6 control, indicating the

predilection for Zn limitation. The strongest expression of ZCRP-A was detected in the +Fe

treatment (**Fig. 2c**) indicative of Fe addition driving the community towards increased Zn stress.

Notably, ZCRP-A was still detected in the +Fe+Zn treatment (**Fig. 2c**), implying that the added

Zn was unable to completely satiate Zn demand as phytoplankton biomass increased (as

indicated by the increase in chl *a* at T6, **Fig. 2a**), despite added Zn (2 nM) being double that of

added Fe (1 nM). Sequence analysis of the contigs identified as ZCRP-A homologs in these

incubations revealed that all contigs contained one or more canonical conserved motifs found in

COG0523 family proteins such as ZCRP-A (**Supplementary Figure 5**). Coupled with evidence

from prior laboratory studies[41], this provides further support for the role of ZCRP-A in

responding to Zn scarcity. The substantial drawdown of surface seawater pCO_2 to 221 µatm at

the incubation site (**Fig. 2d**) (a ~45% decrease compared to offshore waters in the Ross Sea

measured during the same time frame; **Fig. 1b**), is consistent with an increased Zn demand

resulting from low pCO_2 conditions resulting in the detection of Zn stress proteins.

ZCRP-A proteins were taxonomically assigned to chlorophytes, prasinophytes,

dinoflagellates, and *Phaeocystis* (**Supplementary Figure 4a**), with the detection of *Phaeocystis*

ZCRP-A only in the +Fe treatment. The detection of ZCRP-A attributed to *Phaeocystis*, but the

nondetection of ZCRP-A attributed to diatoms, implies that diatoms present were outcompeting

Phaeocystis for Zn. In addition to these metal stress-responsive proteins, variations in other

functional proteins were observed among treatments. For example, RUBISCO and RNA

polymerase were most abundant under +Zn+Fe conditions, and the latter was co-located within

the same hierarchical cluster as ZCRP-A, consistent with their shared connection to cellular

carbon metabolism (**Fig. 2c).** Our observations of Fe and Zn biomarkers shifting in abundance in

response to their respective metal treatment provides independent evidence for Zn/Fe co-

limitation and additional verification that neither incubation treatment (+Zn nor +Fe) was

contaminated with the other metal throughout the experiment.

A third independent line of evidence for the nutritional influence of Zn on TNB

phytoplankton was obtained from *in situ* cellular stoichiometry. Particulate Zn:P ratios (Zn:P)

analyzed from biomass collected at the surface of this experimental station were consistent with

ratios from Zn limited culture studies. Particulate Zn:C ratios reported previously in Zn-limiting

culture studies of the diatom *Thalassiosira pseudonana*[18] were converted to Zn:P ratios using the

Redfield ratio[44] (**Supplementary Table 6**). We then compared these ratios and associated

growth rates with Zn:P measured within biomass collected at 10, 25, 50 and 100 m at the

experimental site. At all of these surface depths, Zn:P measured at the experimental site was ~

2E-4 mol:mol, which, in comparison to cultured diatom Zn:P ratios, fell within in the range of

severely Zn-limited growth rates (**Supplemental Figure 6**), again demonstrating the propensity

for Zn limited growth in this region and corroborating the incubation results.

Further consistent with the observed Zn stimulation of biomass in the incubations, the

largest decreases in macronutrient (P and N+N) concentrations in these incubations at T6 were

observed in the +Fe+Zn treatment (**Supplementary Figure 7a,b**), as was the largest increase in

particulate organic carbon (POC; **Supplementary Figure 7c**). POC collected from the +Zn and

+Fe+Zn incubations was characterized by larger C:N atomic ratios (5.9 and 6.2, respectively)

compared to the +Fe and T6 control (5.2 and 5.3; **Supplementary Figure 7d**) consistent with the

stimulation of carbon uptake upon Zn addition alleviating carbon stress. Significantly higher

bacterial abundances in both +Fe (p = 9.1e-4) and +Fe+Zn (p = 6.3e-4) treatments relative to the

T6 control (**Supplementary Figure 7e**) indicated the alleviation of bacterial Fe limitation,

consistent with prior reports[45–47].

Taxonomic characterization of incubation results

The measurement of phytoplankton pigments over time within the shipboard incubations

revealed a diverse taxonomic response to metal amendments. Measured pigments included

fucoxanthin (fuco), 19'-hexanoyloxyfucoxanthin (19'-hex), prasinoxanthin (prasino), chlorophyll

b (chl *b*), and chlorophyll c3 (chl c3). Fuco is produced by both diatoms and by *Phaeocystis*

under certain conditions, while 19'-hex and chl c3 are indicative of *Phaeocystis* in this region[48].

Fuco:19'-hex ratios significantly increased in the +Fe (p = 4.2e-4) and +Fe+Zn treatments (p =

2.7e-3) (**Supplementary Figure 8a**) due to no significant change in fuco (**Supplementary**

Figure 8b) and decreased 19'-hex (**Supplementary Figure 8c**) relative to the T6 control.

Phaeocystis contributions to total fuco concentrations are typically minimal at the low Fe levels

of the Ross Sea, though *Phaeocystis* can revert to making fuco rather than 19'-hex when released

from Fe limitation[48], as was evident in these incubations by decreased 19'-hex:chl c3 ratios with

+Fe and +Fe+Zn (**Supplementary Figure 8d**). *Phaeocystis* therefore likely contributed to total

fuco by responding to Fe addition. Notably, significant decreases in both fuco:chl *a* and 19'-

hex:chl *a* (**Supplementary Figure 8e,f**) in all treatments compared to the T6 control indicated

that other phytoplankton groups contributed to chl *a* (**Fig. 2a**) without contributing fuco nor

19'hex. Increases in chl *b* (**Supplementary Figure 8g**) and prasinoxanthin (**Supplementary**

Figure 8h) suggested that small green algae such as chlorophytes and prasinophytes also

responded to +Fe and +Zn independently, consistent with the detection of chlorophyte ZCRP-A

in these incubations (**Supplementary Figure 4a**). We used metaproteomic data from T6

incubation biomass to supplement the pigment data and to further describe the major

phytoplankton taxa present in the incubations at T6. Taxonomically assigned protein spectral

counts were summed by genus, and indicated that *Phaeocystis*, diatoms, dinoflagellates, and

cryptophytes were the most abundant phytoplankton taxa as represented in the metaproteomic

data, with cryptophytes, dinoflagellates, and *Phaeocystis* responding to metal treatments,

consistent with the pigment results (**Supplementary Figure 8i**). Photosynthetic efficiency of

photosystem II (Fv/Fm) significantly increased with +Fe (p = 0.011) and with +Fe+Zn (p =

0.0036) at T4 (96 hours) compared to T4 controls, (note that there was no difference among

treatments at T6) but did not significantly increase with +Zn alone, implying Fv/Fm may not be

useful as a diagnostic for Zn stress and that caution should be used in interpreting its signals

universally (**Supplementary Figure 8j**).

Selective zooplankton grazing on small diatoms and solitary *Phaeocystis* cells may also

have played a role in affecting phytoplankton biomass and the observed pigment:chl *a* ratios. For

instance, higher ratios of phaeophytin:total phaeopigments were observed in +Fe and +Zn

amended incubations (**Supplementary Figure 9**) which may reflect grazing on solitary

Phaeocystis cells, as high phaeophytin:total phaeopigments ratios were previously observed in

Phaeocystis dominated waters of the Ross Sea[33].

 Detection of Zn- and Fe-stress protein biomarkers in the water column

Metaproteomic and metatranscriptomic analyses of biomass within the water column at

the experimental site provided additional confirmation of the incubation results by demonstrating

the occurrence of phytoplankton micronutrient Zn- and Fe- stress-response proteins within the

water column, without influence from incubation conditions. In addition to ISIPs and ZCRP-A,

we additionally detected ZCRP-B (a putative membrane-tethered Zn-binding protein)[41], Zrt/Irt-

like (ZIP) Zn transporters, which are known to be used for Zn^{2+} and other divalent metal cation

uptake in marine phytoplankton[42,49,50], and θ (theta) and δ (delta) carbonic anhydrases (CAs). θ-

CAs with Zn^{2+} coordination sites have been documented in diatoms[51], including the polar diatom

Chaetoceros neogracile RS19[52], but there are no studies to date investigating enzyme activity or

efficiency with Co^{2+} or Cd^{2+}. In contrast, δ-CA (i.e, *Thalassiosira weissflogii* TWCA1) is known

to function with either Co^{2+} or Zn^{2+} as a cofactor[53] conferring metabolic flexibility when Zn^{2+} is

scarce.

Both proteins and transcripts of Fe and Zn stress biomarkers were observed throughout

the water column at the experimental site. RUBISCO, ZCRP-A, and ISIP protein spectral counts

within the 3μm size fraction were most abundant at the surface and decreased with depth (**Fig.**

3a,b,d), consistent with the depletion of trace metals in the photic zone due to high-biomass

bloom conditions (**Fig. 1c**). Throughout the water column, ISIPs were predominantly attributed

to *Phaeocystis* and to the diatom genera *Fragilariopsis, Chaetoceros,* and *Pseudo-nitzschia* (**Fig.**

3i, Supplementary Figure 10b). ZCRP-A was detected in both 3 and 51 μm filter pore-size

fractions (**Fig. 3b**) and was predominantly attributed to *Phaeocystis* (**Supplementary Figure**

10c). ZCRP-A transcripts in the euphotic zone corresponded mainly to *Phaeocystis* and to the

diatom genera *Chaetoceros* (**Fig. 3i**). We note that ZCRP-B, a protein also found to be

upregulated in marine diatoms under low Zn/Co and characterized as a putative membrane-

tethered Zn/Co protein ligand[41] was most abundant in the 0.2 μm fraction below 100 m (**Fig. 3c**).

As ZCRP-B shares ~30% similarity to the bacterial ABC-type nickel transporter component

NikA, spectral counts within the bacterial 0.2 μm fraction most likely reflect true bacterial NikA.

BLAST analysis of all ZCRP-B contigs confirmed that all ZCRP-B hits (across all size fractions)

corresponded to bacteria (**Fig. 3i**).

ZIPs were almost solely found in the 51 μm fraction, likely due to the capture of

abundant *Phaeocystis* colonies and chain-forming diatoms (**Fig. 3e; Supplementary Figure**

10d). The assignment of the majority of ISIPs and ZCRP-A proteins to *Phaeocystis* in the upper

water column provides additional evidence that *Phaeocystis* was likely Zn/Fe co-limited at the

study site, consistent with incubation results (**Fig. 2c, Supplementary Figure 4a**). The enhanced

abundance of diatom θ-CA and δ-CA within the water column (**Fig. 3f,g**), as well as transcripts

for the diatom Cd carbonic anhydrase CDCA, which can replace Zn^{2+} with Cd^{2+} as the catalytic

cofactor[21], in the 3 μm and 51 μm fractions at 200 m (**Fig. 3h**) was indicative of a sinking, prior

diatom bloom event[54]. θ-CA and δ-CA were predominantly taxonomically assigned to the diatom

genera *Pseudo-nitzschia* and *Chaetoceros*, respectively, while CDCA transcripts belonged to the

diatom genera *Chaetoceros* and *Corethron* (**Fig. 3i**). The presence of θ-CA, but lack of δ-CA,

assigned to *Chaetoceros* is consistent with proteomic analysis of the polar diatom *Chaetoceros*

neogracile RS19 grown in culture under Zn limiting conditions[52].

Analysis of Zn/CO₂ growth limitation thresholds

In this study, we observed multiple independent lines of evidence that phytoplankton

within incubation experiments and the field environment were experiencing Zn and Fe

nutritional stress. Coupled with observations of greatly reduced C and Zn availability at the

experimental site, which had nearly half the pCO_2 of atmosphere-equilibrated waters (221 μatm

compared to ~420 μatm; **Fig. 1b, Fig. 2d**), and the greater demand for Zn as a catalytic cofactor

within CAs that are required for CO_2 acquisition, TNB was likely experiencing Zn-C

biochemically dependent co-limitation[7], in addition to the expected regional iron limitation. The

low pCO_2 waters of the experimental site incubations naturally provided this low CO_2 backdrop,

allowing the stimulatory effects of Zn addition described above to be observed. This observation

of Zn stimulation within low pCO_2 environments has not been observed previously, and these

results imply that other highly productive coastal environments with low pCO_2 are also

susceptible to Zn-C limitation, potentially limiting the extent of coastal phytoplankton blooms

and carbon export.

Using the available quantitative constraints on Zn and CO_2 co-limitation thresholds

available from the literature (see Methods), we estimated that the threshold for Zn-CO_2 limitation

in algae occurs at 259 μatm pCO_2. We then compared this laboratory-determined Zn/C limitation

threshold to both the *in situ* 221 μatm pCO_2 at our field study site, and to the historical, global

trend in surface ocean pCO_2 (**Fig. 4**). It is evident that average, global surface ocean pCO_2 levels

are rapidly rising above both the laboratory-estimated 259 μatm pCO_2 Zn/C limitation threshold

and our field observation value of 221 μatm[55] (**Fig. 4a,b**) implying that over time, larger amounts

of the surface ocean will cease to experience Zn/C limitation. In high-productivity regions,

organisms may still experience Zn limitation (as shown in this study in TNB), but the fraction of

the ocean that experiences these conditions will continue to shrink, alleviating Zn limitation and

potentially increasing overall biological production. Though a fraction of the modern-day surface

ocean is currently at ≤ 250 ppm pCO_2 (this fraction is predominantly comprised of polar regions,

Supplementary Figure 11a), this represents a large decrease in oceanic extent compared to only

100 years ago (**Supplementary Figure 11b**). The continued rise in global pCO_2 levels as seen

over the last ~100 years is potentially driving a shift away from Zn/C co-limitation, reducing the

advantages for algal species with more efficient carbon acquisition capabilities, and also

triggering broader changes in the biogeochemical cycling of Zn and other bioavailable trace

metals within the oceans. Conversely, during interglacial periods, global atmospheric CO_2 was

similar to or below that of this study, implying zinc-carbon co-limiting conditions for

phytoplankton would have been prevalent throughout the oceans[56].

This study provides experimental field evidence using multiple independent lines of

evidence (chlorophyll, DIC, Zn and Fe biomarker proteins, cellular stoichiometry), combined

with contextual biogeochemical (pCO_2, dZn, and Zn uptake) and metaproteomic environmental

data to demonstrate the presence of Zn limitation in Terra Nova Bay of the Ross Sea. Given the

large Zn requirements of diatoms and other phytoplankton, it is perhaps surprising that more

frequent positive observations of Zn nutritional effects have not been seen. We believe this is due

to the combined challenges of avoiding Zn contamination and the need to initiate experiments in

environments experiencing diminished pCO_2. Low pCO_2 conditions are well known to occur

throughout coastal environments globally and could be further examined for Zn effects. While

there are elaborate biochemical capabilities available to many marine algae for dealing with Zn

scarcity[41], the geographic extent of Zn/C co-limitation will further decrease in the coming

decades with rising anthropogenic CO_2 emissions. Despite this, the biochemical demand for Zn

in marine organisms remains substantial, with cellular demand rivaling that of Fe. The multitude

of metabolic functions requiring Zn, including but not limited to carbonic anhydrase activity,

implies the need for further exploration of Zn influences in a changing ocean environment.

Materials and Methods

Study area and sample collection

Sample collection occurred during the CICLOPS (Cobalamin and Iron Co-limitation of Phytoplankton Species) expedition (NBP18-01) aboard the RVIB *Nathaniel B. Palmer*, December 11, 2017 – March 3, 2018 in the Amundsen Sea and Ross Sea of the Southern Ocean (**Fig. 1a**). Station metadata is provided in **Supplementary Table 2.** All stations were assumed to be representative of TNB during this temporal study (as evident in the total dissolved metal, macronutrient, and chlorophyll *a* datasets). Water samples for dissolved trace metal analyses were collected using trace metal sampling protocols described previously[57]. A trace metal clean rosette suspended on a Kevlar line and equipped with twelve 8L X-Niskin bottles (Ocean Test Equipment) was used to collect seawater at depths ranging from 10 – 600 m. Niskin bottles were transported to a positive-pressure trace metal clean shipboard van for filtration upon surfacing. Total fluorescence on the vertical profiles was measured using an ECO chlorophyll fluorometer (Wet Labs) equipped to the rosette. The rosette also included instrumentation for measuring conductivity and temperature (Sea-Bird Electronics).

Preparation of plasticware

Polyethylene and polycarbonate sampling and incubation bottles were rigorously cleaned to remove trace metal contaminants before use. Bottles were rinsed with 18.2 Ω Milli-Q water (Millipore), soaked for 72h in <1% Citranox detergent, rotated, soaked for an additional 72h, and then rinsed five times with Milli-Q water. Bottles were then filled with 10% HCl (Baker instar-analyzed) by volume and soaked for a minimum of one week, rotated, and soaked for another

week. Bottles were then rinsed five times with dilute acid (HCl, pH 2) and stored double-bagged

in plastic zip bags. All cleaning work was conducted in a Class 100 clean room.

Underway seawater pCO$_2$ partial pressure

Surface water pCO$_2$ measurements were conducted aboard the RVIB *Nathaniel B.*

Palmer using an underway method consisting of an air-water equilibrator and IR CO$_2$ analyzer

developed and operated by the Lamont-Doherty Earth Observatory (LDEO) group[58].

TDIC and POC measurements and calculation of pCO$_2$

Total alkalinity (TA) and dissolved inorganic carbon (DIC) were measured on CTD and

incubations samples in near real-time aboard the NBP. Dissolved inorganic carbon (DIC) and

total alkalinity (TA) samples were collected following the protocols of Dickson et. al 2007[59].

DIC analyses were conducted within ~4 h of collection. We acidified 1.25 mL of sample using an

automated custom-built injection and bubble stripping system coupled to an infrared gas analyzer

(LICOR LI7000) and integrated the infrared absorption signal versus time for each stripped gas

sample to yield a total mass of CO$_2$. Each sample was analyzed in triplicate or greater. Since

microbubbles regularly formed as samples warmed between sample acquisition and DIC

analysis, every integration curve was visually inspected and those curves that exhibited evidence

for bubbles were rejected. Certified reference materials (Dickson CRM batch 169) were analyzed

between every 3 to 4 unknowns. The estimated precision based upon unknowns (>860 samples

run in triplicate) and CRM replicates (n = 738) was ± 2.0 µmol kg^{-1} (±1 SD). As described by

DeJong et al. 2015[60], analyses for TA on filtered samples were completed within ~12 h of

collection by using a potentiometric titrator (Metrohm 855 Robotic Titrosampler). The estimated

precision based on replicate analyses of CRMs (n = 195) was ± 2.6 µmol kg^{-1} (±1 SD). DIC/TA

pairs were used to calculate pCO_2 values using in situ temperature and salinity. Parameters were

input to the MATLAB version of CO2SYS[61], and pCO_2 was calculated using the dissociation

constants determined by the Dickson and Millero refit of Mehrbach's data (option 4).

We acknowledge offsets in seawater pCO_2 calculated from TA-DIC pairs and those

measured using the ship's underway system. In general, values inferred from TA-DIC pairs are

higher than underway measurements, sometimes by up to 70 ppm. Errors in acid dissociation

constants due to temperature effects[62] cannot explain this offset, as using these updated constants

only exacerbates the offsets rather than reconciling them. While these offsets are beyond the

scope of the current manuscript, they hint at oceanographic effects on the carbonate system in

very cold seas (e.g. large temperature gradients and their effect on the equilibration of the

carbonate system). Nevertheless, both calculated and underway pCO_2 measurements are

significantly below atmospheric CO_2 values, suggesting carbon limitation of the microbial

community in Terra Nova Bay.

Analysis of historical atmospheric and surface ocean pCO_2 trends

Decadal surface ocean pCO_2 reconstructions from Jiang et al. 2023[55] were downloaded,

binned by decade, and plotted using the 'violinplot' in MATLAB. Atmospheric pCO_2 data was

assembled from the running Mauna Loa record[63], and from measurements made on Antarctic firn

ice[64].

Calculation of Zn-and pCO_2 co-limitation of phytoplankton thresholds

There are few experimental measurements of Zn- and pCO_2-co-limitation, either in the

lab or *in situ*. This study represents one of the first demonstrations of Zn- and pCO_2 co-limitation

at a pCO_2 of ~220 ppm (Station 27, **Fig. 2d**). In the literature, several models exist to interpret

co-limitation[7,65]. For this study we chose to use the biochemically dependent co-limitation model

for growth rate (μ):

$$\mu = V_{max} \frac{[CO_{2(aq)}]}{\varphi K_s + [CO_{2(aq)}]}, \quad (1)$$

where V_{max} is the maximum growth rate, $[CO_{2(aq)}]$ is the aqueous CO_2 concentration of the

growth medium in micromoles per kilogram of seawater, K_s is the half-saturation constant in

micromoles per kilogram of seawater, and φ is a Zn-dependent growth term:

$$\varphi = \frac{[dZn] + K_{s,Zn}}{[dZn]} \quad (2)$$

Here, the dissolved Zn concentration in the growth medium [dZn] is modified by a Zn-dependent

saturation constant ($K_{s,Zn}$). Few studies have enough experimental data to robustly establish a

kinetic relationship between [dZn] and $[CO_{2(aq)}]$, so we compiled several estimates for these

terms from the literature. Diatom growth rates under pCO_2 limitation were taken from Riebesell

et al. (1993) [66]. Reported pH and temperature measurements for each treatment, a total alkalinity

of 2300 μmol kg^{-1}, and a salinity of 35 were used to calculate aqueous CO_2 concentrations using

CO2SYSv3.2.1.[67]. Reported V_{max} and K_s for *D. brightwellii, T. punctigera,* and *R. alata* were

1.46, 1.30, and 0.93 d^{-1}, and 1.4, 1.2, and 2.1 μmol kg^{-1}, respectively. Values for coccolithophore

growth ($K_s = 0.97$ μmol kg^{-1}, $V_{max} = 4.7$ d^{-1}) were taken from Krumhardt et al. (2017) [68]. A value

for $K_{s,Zn}$ of 300 pmol L^{-1} was taken from Buitenhuis et al. (2003). This value is for the

coccolithophore *E. huxleyii* generated under varying CO_2 and Zn conditions, and Zn response

growth curves under single CO_2 conditions (ambient) are similar to other diatoms like *T.*

pseudonana[69]. The value of 300 pmol L^{-1} appears high, but is tied to the functional form of the

biochemically co-limitation equation[7,65]. Based on the same dataset and different models for co-

limitation, Buitenhuis et al. (2003) arrived at $K_{s,Zn}$ values of ranging from 38 pmol L^{-1} to 300

pmol L^{-1}. They calculated Zn-limitation alone, at CO_2-replete conditions, of 19 pmol L^{-1}. Thus,

the chosen value of $K_{s,Zn}$ is not a reflection of high Zn demand but determined by the

functionality of biochemical co-limitation by Zn and C.

To calculate φ, a surface ocean Zn concentration of 50 pmol L^{-1} was assumed [2,70]. While

these concentrations reflect total dissolved Zn, the relationship between bioavailable free Zn and

dZn, especially in the field, remains unclear. Eq. 1 was then used to calculate effective CO_2

concentrations (and thus pCO_2 values) at which growth is halved, or in other words, $\mu = 0.5V_{max}$.

We note that this calculation is distinct from the CO_2 half-saturation constant because of the co-

limitation by Zn. The median pCO_2 threshold for 50% growth from the three diatom species was

278 ppm. Including coccolithophores decreases the median to 259 ppm. These values are slightly

higher than the *in situ* evidence for Zn limitation at 220 ppm presented in the present study. Our

results cannot be considered as an upper bound for Zn-CO_2 limitation, but serve as evidence for

growth limitation under those specific environmental conditions.

Analyses of total dissolved metals using isotope dilution

The analysis of total dissolved metals has been described previously [71]. Briefly, seawater

collected shipboard by pressure-filtering X-Niskin bottles through an acid-washed 142 mm, 0.2

μM polyethersulfone Supor membrane filter (Pall) within 3 hours of rosette recovery using high

purity (99.999%) N_2 gas and stored at 4ºC. All sample collection occurred shipboard within an

on-deck trace metal clean van. Samples were acidified to pH 1.7 with high purity HCl (Optima)

within 7 months of collection and were stored acidified at room temperature for over 1 year prior

to analysis.

Quantification of total dissolved Fe, Mn, Ni, Cu, Zn, and Cd was performed using isotope

dilution. Acidified seawater samples were spiked with a stable isotope spike solution artificially

enriched in ^{57}Fe, ^{61}Ni, ^{65}Cu, ^{67}Zn, and ^{110}Cd (Oak Ridge National Laboratory). Concentrations

and spike ratios were verified by ICP-MS using a multi-element standard curve (SPEX

CertiPrep). Preconcentration of spiked seawater samples for total dissolved metal analysis was

performed using the automated solid phase extraction system seaFAST-pico (Elemental

Scientific) in offline concentration mode with an initial volume of 15mL and elution volume of

500μL [72,73]. Following preconcentration, multielemental quantitative analysis was performed

using an iCAP-Q inductively coupled plasma-mass spectrometer (ICP-MS) (Thermo Scientific).

Concentrations of Mn, Fe, Ni, Cu, Zn and Cd were determined using a six-point external

standard curve of a multi-element standard (SPEX CertiPrep), diluted to range from 1-10 ppb in

5% nitric acid. An indium standard (SPEX CertiPrep) was similarly added to these standard

stocks, diluted to range 1-10 ppb. Instrument injection blanks consisted of 5% nitric acid in

Milli-Q. Standard curve R^2 values were ≥0.98 for all metals monitored. Method accuracy and

precision were assessed using the 2009 GEOTRACES coastal surface seawater (GSC) standard

(n = 8; **Supplementary Table 5**), which produced values consistent with consensus results.

Macronutrient, pigment, and Fv/Fm analyses

Seawater for macronutrient (silicate, phosphate, nitrate, and nitrite) analyses were filtered

through 0.2 μm pore-size Supor membrane filters and frozen at sea in acid-washed 60-mL high-

density polyethylene bottles until analysis. Macronutrient analyses were conducted by nutrient

autoanalyzer (Technicon Autoanalyzer II) by Joe Jennings at Oregon State University. The

chemotaxonomic distribution of phytoplankton pigments was determined using HPLC as

described previously[74]. Photosynthetic efficiency of photosystem II (Fv/Fm) was measured using

a Phyto PAM phytoplankton analyzer (Walz, Effeltrich, Germany) as described previously[75].

Bacterial abundance

One ml samples for heterotrophic prokaryotes abundance (HPA) analysis were fixed for

10 min with a mix of paraformaldehyde and glutaraldehyde (1% and 0.05% final concentration,

respectively), frozen in liquid nitrogen and stored at $-80°C$ until analysis. After thawing,

samples were stained with SYBR Green (Invitrogen Milan, Italy) using 10^{-3} dilution of stock

solution for 15 min at room temperature. Cell concentrations were assessed using a FACSVerse

flow cytometer (BD BioSciences Inc., Franklin Lakes, USA) equipped with a 488 nm Ar laser

and standard set of optical filters. FCS Express software was used for analyzing the data and HP

were discriminated from other particles on the basis of scatter and green fluorescence from

SYBR Green [76].

Zn uptake rates using ^{67}Zn

^{67}Zn stable isotope uptake experiments were performed to quantify the movement of

dissolved Zn to the particulate phase (in units of pmol $L^{-1} d^{-1}$). Briefly, raw (unfiltered) seawater

was collected using the trace metal rosette over a depth range of $10 - 600$ m into 250mL trace

metal clean polycarbonate bottles. Bottles were spiked with ^{67}Zn such that the total added

(spiked) concentration of Zn was 2 nM. Immediately after spiking, incubation bottles were

sealed, inverted to mix, and transferred to flow-through on-deck incubators for 24hr. Biomass

was collected after 24hr by vacuum filtering each incubation sample at 34.5 kPa (5 psi) onto an

acid-cleaned 3μm pore-size acrylic copolymer Versapore filter (Pall) mounted on an acid-

cleaned plastic filtration rig. Sample filters were retrieved from storage at -80°C, removed from

cryovials using plastic acid-washed forceps, and transferred into trace metal clean 15 mL PFA

vials with 4 mL of 5% HNO_3 (Optima) containing a 1 ppb Indium (In) internal standard. Filters

were digested for ~3.5h at 140°C using a HotBlock® heating block (Environmental Express,

USA) before they were removed and discarded. After evaporating the remaining solution to just

dryness, the residue was resuspended in 2 mL of 5% HNO_3 (Optima) by light vortexing. Process

blank filters were digested and processed as sample filters were. This experiment was also

carried out using ^{110}Cd as a tracer of Cd uptake in separate incubation bottles (data not shown

here). Digests were analyzed in duplicate by ICP-MS using a Thermo ICAP-Q plasma mass

spectrometer calibrated to a multi-element standard curve (Spex Certiprep) over a range of 1 –

20 ppb. Natural Cd and Zn isotope abundances of the standards were assumed to calculate

concentrations of ^{110}Cd, ^{111}Cd, ^{114}Cd, ^{67}Zn, ^{66}Zn, and ^{68}Zn. Total Zn uptake (pmol L^{-1} d^{-1}) was

calculated using particulate ^{67}Zn and total water column dZn measurements as described

previously[77]. The particulate metal measurements captured contributions from the active

transport of metal into cells, nonspecific metal adsorption to cell surfaces, metal adsorption to

non-living particulate organic matter, and metal adsorption to particulate inorganic matter,

though we expect active transport into cells to dominate the measured particulate isotopic signal

due to the high abundance of actively growing autotrophic cells in the photic zone observed in

Southern Ocean during austral summer. Particulate Zn:P measurements presented in Figure 3d,e

were calculated using particulate Zn measured on Cd-spiked filters and thus do not include any

pZn contribution from Zn tracer addition. Particulate phosphorus concentrations were measured

by ICP-MS simultaneously and were calibrated to a standard curve ranging from 100 to 3200

ppb using a 1 ppm certified P stock (Alfa Aesar Specpure). All SPEX and P standard curves had

R^2 values > 0.99. The Zn stock solution used in the incubation experiments was similarly

analyzed by ICP-MS to confirm that the stock was not Fe contaminated— this analyses showed

that less than 2.3 pM (which was near the instrument blank level for this analysis) of iron was

added for every 2 nM of zinc, far less than needed to stimulate phytoplankton to the extent

observed in our experiments.

Shipboard incubation experiments

Incubation experiments were conducted at station 27 (-74.9870°N, 165.8898°E). Raw

surface seawater was pumped directly into a cleanroom container van, collected into acid-

cleaned 50L carboys, and dispensed into acid-washed 1L polycarbonate bottles using a trace

metal sampling system with acid-washed polypropylene tubing and a teflon diaphragm pump.

Incubation bottles were first rinsed with seawater then filled. Seawater was collected at 16:05

UTC. Triplicate incubation bottles were amended with +Fe (1 nM), +Zn (2 nM) and +Fe+Zn,

sealed, and placed into a flow-through on-deck incubator with light screens that shaded the

incubator to 20% percent ambient surface irradiance. Incubations were sampled at 0, 48, 96, and

144 hours (corresponding to T0, T2, T4, and T6 timepoints) for analysis by filtering onto GFF

filters for chlorophyll (all time points, biological triplicates), pigment analyses (T6, biological

triplicates), and proteomic analyses (T6, pooled biological triplicates). Chlorophyll was extracted

immediately, otherwise samples were frozen at -80°C until further analyses, with pigment and

protein samples kept in -80°C freezers, liquid nitrogen dewars, or dry ice coolers at all times

during transport back to the laboratories. All amendments and sampling were conducted in a

positive-pressure, clean room van with laminar flow hoods and plastic sheeting to minimize

trace-metal contamination.

Metaproteomic analysis

Water column metaproteomic biomass was collected onto 0.2, 3, and 51 µm pore-size

filters using in-situ battery operated McLane pumps. Incubation metaproteomic samples were

collected on GFF filters as described above. Filters were frozen at -80°C and transported as

described above until laboratory extractions. Frozen filters were thawed and extracted as

described previously [78]. Briefly, filters were extracted using SDS detergent to solubilize both

membrane and soluble proteins and subjected to a 10-min heating incubation at 95 °C. Protein

material was subsequently purified, alkylated, reduced and digested with trypsin (mass

spectrometry grade; Promega Corporation) at a trypsin:protein ratio of 1:20 and peptides were

purified using C18 tips (Pierce C18 Tips, 100-µl bed; Thermo Fisher Scientific). Purified

peptides were diluted to 0.1 µg µl^{-1} and 20 µl (2 µg) was injected onto a Dionex UltiMate 3000

RSLCnano LC system (Thermo Fisher Scientific) with an additional RSLCnano pump run in

online two-dimensional active modulation mode coupled to a Thermo Fusion Orbitrap mass

spectrometer.

A translated metatranscriptome (see below) was used as a reference protein database and

peptide spectra matches were performed using the SEQUEST algorithm within Proteome

Discoverer v.2.1 (Thermo Fisher Scientific) with a fragment tolerance of 0.6 Da and parent

tolerance of 10 ppm. Identification criteria consisted of a peptide threshold of 95% (false

discovery rate (FDR) = 0.1%) and protein threshold of 99% (1 peptide minimum, FDR = 0.8%)

in Scaffold v.5 (Proteome Software) resulting in 5,387 proteins identified in the incubation

experiment and 27,924 proteins identified in the water column. To avoid double-counting mass

spectra, exclusive spectral counts were used for the downstream proteomic analysis. Exclusive

spectral counts were normalized using the normalized spectral abundance factor (NSAF)

calculation [79] to allow for a comparison of protein abundance across samples while remaining

consistent with the metatranscriptomic procedure, see Cohen et al. 2021 for details. Counts

associated with redundant ORFs (sharing identical taxonomic and functional assignments) were

summed together.

Relative abundances of *Phaeocystis*, cryptophytes, dinoflagellates, and diatoms ("major

groups") within the T6 incubation biomass were estimated by summing exclusive total protein

spectral counts across all detected proteins that were assigned to the same group (as noted by

best hit species; see metatranscriptomic methods below). Protein hits with LPI scores < 0.5, low

assigned species identity (< 50%), or no assigned species were excluded.

Metatranscriptomic analysis

RNA sequencing was performed using the Illumina HiSeq platform. Transcriptomic

assemblies were generated for biomass collected using McLane pumps filtered through 0.2, 3,

and 51 μm pore-size filters. In order to enrich metatranscriptomic libraries derived from 0.2 μm

filters in prokaryotic transcripts and libraries derived from 3 μm and 51 μm filters in eukaryotic

transcripts, 0.2 μm libraries were generated from total rRNA-depleted mRNA and 3 μm and 51

μm libraries were generated from polyA mRNA. Total RNA was extracted from 0.2 μm, 3 μm,

and 51μm filters using Macherey-Nagel a NucleoMag RNA kit (Macherey-Nagel GmbH &

Co.KG). Cleared lysate was loaded into a 96 deep-well plate and put on an epMotion 5075 TMX

liquid handler to complete the RNA extraction following the Machery-Nagel standard protocol.

For 3 μm and 51 μm samples with total RNA greater than 1 ug, 800 ng of total RNA was used

for preparing poly A libraries with an Illumina Stranded mRNA Prep Ligation kit (Illumina),

following the manufacturer's protocol. For the 3 μm and 51 μm samples with total RNA less

than 1 ug, 20 ng of total RNA was used as input for the SMART-Seq v4 Ultra Low Input RNA

kit (Takara Bio USA. Inc), which converts poly(A) RNA to full-length cDNA using a modified

oligo (dT) primer with simultaneous cDNA amplification. The resulting double-stranded cDNA

was then fragmented using a Covaris E210 system with the target size of 300bp. Libraries were

prepared from fragmented double-stranded cDNA using an Illumina Stranded mRNA Prep

Ligation kit (Illumina). For RNA obtained from 0.2 μm filters, ribosomal RNA was removed

using a riboPOOL Seawater Kit (Galen Laboratory Supplies, North Haven, Connecticut, USA).

The riboPOOL Seawater Kit is a customized mixture of Removal Solutions: Pan-Prokaryote

riboPOOL, Pan-Plant riboPOOL and Pan-Mammal in a ratio of 6:1:1. The rRNA-depleted total

RNA was used for cDNA synthesis by Ovation RNA-Seq System V2 (TECAN, Redwood City,

USA). Double stranded cDNA was then prepared for the libraries using an Illumina Stranded

mRNA Prep Ligation kit (Illumina). Ampure XP beads (Beckman Coulter) were used for final

library purification. Library quality was analyzed on a 2200 TapeStation System with an Agilent

High Sensitivity DNA 1000 ScreenTape System (Agilent Technologies, Santa Clara, CA, USA).

Resulting libraries were subjected to paired-end Illumina sequencing via NovaSeq S4.

The input paired-end fastq sequences are trimmed of sequencing adapters, primers and

low quality bases by using either BLASTN (NCBI, v2.2.25) [80] or trimmomatic, v0.36 [81]. The

trimmed paired and unpaired sequences were then depleted of rRNA sequences with riboPicker

v0.4.3 [82]. The command-line program clc_assembler, v5.2.1 (Qiagen) was used to assemble

processed sequences into contigs and ORFs were identified by FragGeneScan, v1.31 [83]. The

trimmed sequences were mapped to the predicted ORFs using the command-line program

clc_mapper, v5.2.1 (Qiagen) to generate mapped raw read counts for each ORF. The raw counts

were normalized initially to RPKM values, to account for variations in inter-sample sequencing

depth and the ORF sequence length [84]. The RPKM values were subsequently converted to TPM

(transcripts per million) units for estimation of the relative RNA abundance among samples [85].

The ORFs were annotated for putative function by several programs in parallel using BLASTP

against PhyloDB, hidden Markov models PFAM and TIGRFAM by HMMER, v3.3.2 [86], KEGG

Ortholog HMM by kofamscan, v1.3.0 [87], and transmembrane HMM by TMHMM [88]. Additional

annotations were generated by similarity searches using BLASTP to transporter (PhyloDB),

organelle (PhyloDB) and KOG [89] databases. The ORFs are assigned to the best taxonomic

species/group as determined by LPI (Lineage Probability Index) analysis [90]. The final list of

curated ORFs was generated by removing ORFs with low mapping coverage (<50 reads total

over all samples) and with no BLAST hits and no known domains.

Statistical analysis and data visualization

ANOVA and Dunnett tests were performed using MATLAB 2019a. Statistics are

summarized in Supplementary Table 3. The linear regressions shown in Fig. 1h,i are a two-way

(type-II) linear regressions, which allows for error in both the x and y parameters and does not

assume dependence between the x and y axes. The two-way regression MATLAB function

(lsqfitma.m) is available in python at https://github.com/rebecca-chmiel/GP15. Figures were

made using matplotlib (version 3.5.0), Ocean Data View (version 5.5.2), Excel (2019), and

RStudio (version 1.3.1093).

Data availability

CICLOPS (NBP18-01) total dissolved Zn, Zn uptake rates, macronutrients, and pigment

datasets are available through the NSF's Biological and Chemical Oceanography Data

Management Office (BCO-DMO) repository under project number 774945 (https://www.bco-

dmo.org/project/774945). The mass spectrometry global proteomics data for CICLOPS bottle

incubations and water column analyses have been deposited with the ProteomeXchange

Consortium through the PRIDE repository under the project name "Zinc-iron co-limitation of

natural marine phytoplankton assemblages in coastal Antarctica" with project accession number

PXD037056 (https://www.ebi.ac.uk/pride/archive/projects/PXD037056). This data is accessible

for review by using the following login information: username reviewer_pxd037056@ebi.ac.uk,

password: lFdOUoEb. The translated transcriptome used for spectrum to peptide matching has

been deposited in the National Center for Biotechnology Information sequence read archive

under BioProject accession no. PRJNA890306

(https://www.ncbi.nlm.nih.gov/bioproject/?term=PRJNA890306) and RNA-Seq BioSample

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(https://www.ncbi.nlm.nih.gov/biosample/?term=SAMN31286421).

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Author contributions statement

All authors contributed to data acquisition and analysis. RMK, NLS, LEL, RJC, DR, DMM,

MRM, FB, RBD, GRD and MAS implemented the shipboard incubation study. RMK, MMB,

RJC, DR, TJH, and AVS contributed to the formal analysis. NLS, LEL, FB, OM, RF, and GRD

contributed to pigment datasets and interpretations. CB provided bacterial abundance data. RBD

contributed DIC and *in situ* pCO_2 data. AVS contributed analyses and discussion regarding

historical pCO_2 data. AEA contributed the metatranscriptome reference database used for

proteomic analyses. RMK and MAS wrote the original draft. RMK, MAS, GRD, AVS, TJH, and

RBD contributed to review and editing. All authors approved the final submitted manuscript.

Competing interests statement

The authors declare no competing interests.

References

1. Moore, C. M. *et al.* Processes and patterns of oceanic nutrient limitation. *Nat. Geosci.* **6**,

701–710 (2013).

2. Bruland, K. W. Oceanographic distributions of cadmium, zinc, nickel, and copper in the

North Pacific. *Earth Planet. Sci. Lett.* **47**, 176–198 (1980).

3. Jakuba, R. W., Saito, M. A., Moffett, J. W. & Xu, Y. Dissolved zinc in the subarctic North

Pacific and Bering Sea: Its distribution, speciation, and importance to primary producers.

Glob. Biogeochem. Cycles (2012) doi:10.1029/2010GB004004.

4. Bruland, K. W. Complexation of zinc by natural organic ligands in the central North Pacific.

Limnol. Oceanogr. **34**, 269–285 (1989).

5. Baars, O. & Croot, P. L. The speciation of dissolved zinc in the Atlantic sector of the

Southern Ocean. *Deep Sea Res. Part II Top. Stud. Oceanogr.* **58**, 2720–2732 (2011).

6. Sunda, W. G. & Huntsman, S. A. Effect of Zn, Mn, and Fe on Cd accumulation in

phytoplankton: Implications for oceanic Cd cycling. *Limnol. Oceanogr.* **45**, 1501–1516

(2000).

7. Saito, M. A., Goepfert, T. J. & Ritt, J. T. Some thoughts on the concept of colimitation:

Three definitions and the importance of bioavailability. *Limnol. Oceanogr.* **53**, 276–290

(2008).

8. Lhospice, S. *et al.* Pseudomonas aeruginosa zinc uptake in chelating environment is

primarily mediated by the metallophore pseudopaline. *Sci. Rep.* **7**, 17132 (2017).

9. Mazzotta, M. G. *et al.* Characterization of the metalloproteome of Pseudoalteromonas (BB2-

AT2): biogeochemical underpinnings for zinc, manganese, cobalt, and nickel cycling in a

ubiquitous marine heterotroph. *Metallomics* **13**, (2021).

10. Ye, N. *et al.* The role of zinc in the adaptive evolution of polar phytoplankton. *Nat. Ecol.*

Evol. **6**, 965–978 (2022).

11. Scharek, R., Van Leeuwe, M. A. & De Baar, H. J. W. Responses of Southern Ocean

phytoplankton to the addition of trace metals. *Deep-Sea Res. Part II Top. Stud. Oceanogr.*

(1997) doi:10.1016/S0967-0645(96)00074-4.

12. Coale, K. H., Wang, X., Tanner, S. J. & Johnson, K. S. Phytoplankton growth and biological

response to iron and zinc addition in the Ross Sea and Antarctic Circumpolar Current along

170°W. *Deep Sea Res. Part II Top. Stud. Oceanogr.* **50**, 635–653 (2003).

13. Ellwood, M. J. Zinc and cadmium speciation in subantarctic waters east of New Zealand.

Mar. Chem. (2004) doi:10.1016/j.marchem.2004.01.005.

14. Crawford, D. W. *et al.* Influence of zinc and iron enrichments on phytoplankton growth in

the northeastern subarctic Pacific. *Limnol. Oceanogr.* **48**, 1583–1600 (2003).

15. Franck, V., Bruland, K., Hutchins, D. & Brzezinski, M. Iron and zinc effects on silicic acid

and nitrate uptake kinetics in three high-nutrient, low-chlorophyll (HNLC) regions. *Mar.*

Ecol. Prog. Ser. **252**, 15–33 (2003).

16. Dreux Chappell, P. *et al.* Preferential depletion of zinc within Costa Rica upwelling dome

creates conditions for zinc co-limitation of primary production. *J. Plankton Res.* **38**, 244–255

(2016).

17. Morel, F. M. M. *et al.* Zinc and carbon co-limitation of marine phytoplankton. *Nature* **369**,

740–742 (1994).

18. Sunda, W. G. & Huntsman, S. A. Effect of CO 2 supply and demand on zinc uptake and

growth limitation in a coastal diatom. *Limnol. Oceanogr.* **50**, 1181–1192 (2005).

19. Price, N. M. & Morel, F. M. M. Cadmium and cobalt substitution for zinc in a marine

diatom. *Nature* **344**, 658–660 (1990).

20. Kellogg, R. M. *et al.* Efficient zinc/cobalt inter-replacement in northeast Pacific diatoms and

relationship to high surface dissolved Co : Zn ratios. *Limnol. Oceanogr.* **65**, 2557–2582

(2020).

21. Lane, T. W. & Morel, F. M. A biological function for cadmium in marine diatoms. *Proc.*

Natl. Acad. Sci. U. S. A. **97**, 4627–31 (2000).

22. Jensen, E. L., Clement, R., Kosta, A., Maberly, S. C. & Gontero, B. A new widespread

subclass of carbonic anhydrase in marine phytoplankton. *ISME J.* **13**, 2094–2106 (2019).

23. Salisbury, J. E. *et al.* Seasonal observations of surface waters in two Gulf of Maine estuary-

plume systems: Relationships between watershed attributes, optical measurements and

surface pCO2. *Estuar. Coast. Shelf Sci.* **77**, 245–252 (2008).

24. Chen, S. *et al.* A machine learning approach to estimate surface ocean pCO2 from satellite

measurements. *Remote Sens. Environ.* **228**, 203–226 (2019).

25. Cai, W.-J. *et al.* Controls on surface water carbonate chemistry along North American ocean

margins. *Nat. Commun.* **11**, 2691 (2020).

26. Morel, F. M. M., Lam, P. J. & Saito, M. A. Trace metal substitution in marine

phytoplankton. *Annu. Rev. Earth Planet. Sci.* **48**, 491–517 (2020).

27. Cullen, J. T., Lane, T. W., Morel, F. M. M. & Sheerell, R. M. Modulation of cadmium

uptake in phytoplankton by seawater CO2 concentration. *Nature* (1999) doi:10.1038/46007.

28. Cullen, J. T. & Sherrell, R. M. Effects of dissolved carbon dioxide, zinc, and manganese on

the cadmium to phosphorus ratio in natural phytoplankton assemblages. *Limnol. Oceanogr.*

50, 1193–1204 (2005).

29. Arrigo, K. R., Lowry, K. E. & van Dijken, G. L. Annual changes in sea ice and

phytoplankton in polynyas of the Amundsen Sea, Antarctica. *Deep Sea Res. Part II Top.*

Stud. Oceanogr. **71–76**, 5–15 (2012).

30. Smith, W. O. *et al.* Interannual variations in nutrients, net community production, and

biogeochemical cycles in the Ross Sea. *Deep-Sea Res. Part II Top. Stud. Oceanogr.* (2006)

doi:10.1016/j.dsr2.2006.02.014.

31. Mangoni, O. *et al.* Phaeocystis antarctica unusual summer bloom in stratified antarctic

coastal waters (Terra Nova Bay, Ross Sea). *Mar. Environ. Res.* (2019)

doi:10.1016/j.marenvres.2019.05.012.

32. DeJong, H. B. & Dunbar, R. B. Air-Sea CO2 Exchange in the Ross Sea, Antarctica. *J.*

Geophys. Res. Oceans **122**, 8167–8181 (2017).

33. DiTullio, G. R. & Smith, W. O. Spatial patterns in phytoplankton biomass and pigment

distributions in the Ross Sea. *J. Geophys. Res. Oceans* **101**, 18467–18477 (1996).

34. Martin, J. H., Fitzwater, S. E. & Gordon, R. M. Iron deficiency limits phytoplankton growth

in Antarctic waters. *Glob. Biogeochem. Cycles* **4**, 5–12 (1990).

35. Sedwick, P. N. *et al.* Early season depletion of dissolved iron in the Ross Sea polynya:

Implications for iron dynamics on the Antarctic continental shelf. *J. Geophys. Res.* **116**,

C12019 (2011).

36. Fitzwater, S. E., Johnson, K. S., Gordon, R. M., Coale, K. H. & Smith, W. O. Trace metal

concentrations in the Ross Sea and their relationship with nutrients and phytoplankton

growth. *Deep-Sea Res. Part II Top. Stud. Oceanogr.* **47**, 3159–3179 (2000).

37. Kellogg, R. M. Assessing the potential for Zn limitation of marine primary production:

proteomic characterization of the low Zn stress response in marine diatoms. (Massachusetts

Institute of Technology; the Woods Hole Oceanographic Institution, 2022).

38. Saito, M. A. & Goepfert, T. J. Zinc-cobalt colimitation of *Phaeocystis antarctica*. *Limnol.*

Oceanogr. **53**, 266–275 (2008).

39. Chmiel, R. J. *et al. Low Cobalt Inventories in the Amundsen and Ross Seas Driven by High*

Demand for Labile Cobalt Uptake Among Native Phytoplankton Communities.

https://egusphere.copernicus.org/preprints/2023/egusphere-2023-402/ (2023)

doi:10.5194/egusphere-2023-402.

40. Sperfeld, E., Raubenheimer, D. & Wacker, A. Bridging factorial and gradient concepts of

resource co-limitation: towards a general framework applied to consumers. *Ecol. Lett.* **19**,

201–215 (2016).

41. Kellogg, R. M. *et al.* Adaptive responses of marine diatoms to zinc scarcity and ecological

implications. *Nat. Commun. 2022 131* **13**, 1–13 (2022).

42. Allen, A. E. *et al.* Whole-cell response of the pennate diatom Phaeodactylum tricornutum to

iron starvation. *Proc. Natl. Acad. Sci.* **105**, 10438–10443 (2008).

43. Behnke, J. & LaRoche, J. Iron uptake proteins in algae and the role of Iron Starvation-

Induced Proteins (ISIPs). *Eur. J. Phycol.* **55**, 339–360 (2020).

44. Redfield, A. C. The Biological Control of Chemical Factors in the Environment. *Am. Sci.*

(1958) doi:10.2307/27828530.

45. Sun, Y., Debeljak, P. & Obernosterer, I. Microbial iron and carbon metabolism as revealed

by taxonomy-specific functional diversity in the Southern Ocean. *ISME J.* **15**, 2933–2946

(2021).

46. Obernosterer, I., Fourquez, M. & Blain, S. Fe and C co-limitation of heterotrophic bacteria in

the naturally fertilized region off the Kerguelen Islands. *Biogeosciences* **12**, 1983–1992

(2015).

47. Fourquez, M. *et al.* Microbial Competition in the Subpolar Southern Ocean: An Fe–C Co-

limitation Experiment. *Front. Mar. Sci.* **6**, 776 (2020).

48. DiTullio, G. R., Garcia, N., Riseman, S. F. & Sedwick, P. N. Effects of iron concentration on

pigment composition in Phaeocystis antarctica grown at low irradiance. *Biogeochemistry* **83**,

71–81 (2007).

49. Milner, M. J., Seamon, J., Craft, E. & Kochian, L. V. Transport properties of members of the

ZIP family in plants and their role in Zn and Mn homeostasis. *J. Exp. Bot.* **64**, 369–381

(2013).

50. Bender, S. J. *et al.* Colony formation in Phaeocystis antarctica: connecting molecular

mechanisms with iron biogeochemistry. *Biogeosciences* **15**, 4923–4942 (2018).

51. Jensen, E. L., Maberly, S. C. & Gontero, B. Insights on the Functions and Ecophysiological

Relevance of the Diverse Carbonic Anhydrases in Microalgae. *Int. J. Mol. Sci.* **21**, 2922

(2020).

52. Kellogg, R. M. *et al.* Lack of a Zn/Co substitution ability in the polar diatom *Chaetoceros*

neogracile RS19. *Limnol. Oceanogr.* **67**, 2265–2280 (2022).

53. Lane, T. W. & Morel, F. M. M. Regulation of Carbonic Anhydrase Expression by Zinc,

Cobalt, and Carbon Dioxide in the Marine Diatom Thalassiosira weissflogii. *Plant Physiol.*

123, 345–352 (2000).

54. Subhas, A. V., Adkins, J. F., Dong, S., Rollins, N. E. & Berelson, W. M. The carbonic

anhydrase activity of sinking and suspended particles in the North Pacific Ocean. *Limnol.*

Oceanogr. **65**, 637–651 (2019).

55. Jiang, L. *et al.* Global Surface Ocean Acidification Indicators From 1750 to 2100. *J. Adv.*

Model. Earth Syst. **15**, e2022MS003563 (2023).

56. Ahn, J. & Brook, E. J. Atmospheric CO_2 and Climate on Millennial Time Scales During the

Last Glacial Period. *Science* **322**, 83–85 (2008).

57. Cutter, G. A. & Bruland, K. W. Rapid and noncontaminating sampling system for trace

elements in global ocean surveys. *Limnol. Oceanogr. Methods* (2012)

doi:10.4319/lom.2012.10.425.

58. Takahashi, T., Sutherland, S. C. & Kozyr, A. Global Ocean Surface Water Partial Pressure of

CO2 Database (LDEO Database Version 2019): Measurements Performed During 1957-

2019 (NCEI Accession 0160492). *NOAA Natl. Cent. Environ. Inf.* (2020).

59. Dickson, A. G., Sabine, C. L. & Christian, J. R. *Guide to Best Practices for Ocean CO2*

Measurement. https://repository.oceanbestpractices.org/handle/11329/249 (2007)

doi:10.25607/OBP-1342.

60. DeJong, H. B., Dunbar, R. B., Mucciarone, D. & Koweek, D. A. Carbonate saturation state

of surface waters in the Ross Sea and Southern Ocean: controls and implications for the

onset of aragonite undersaturation. *Biogeosciences* **12**, 6881–6896 (2015).

61. Van Heuven, S., Pierrot, D., Rae, J. W. B., Lewis, E. & Wallace, D. W. R. MATLAB

Program Developed for CO2 System Calculations. ORNL/CDIAC-105b. Carbon Dioxide

Information Analysis Center (CDIAC)

https://doi.org/10.3334/CDIAC/OTG.CO2SYS_MATLAB_V1.1 (2011).

62. Sulpis, O., Lauvset, S. K. & Hagens, M. Current estimates of K1* and K2* appear

inconsistent with measured CO2 system parameters in cold oceanic regions. *Ocean Sci.* **16**,

847–862 (2020).

63. Keeling, C. D. *et al.* Atmospheric carbon dioxide variations at Mauna Loa Observatory,

Hawaii. *Tellus Dyn. Meteorol. Oceanogr.* **28**, 538 (1976).

64. Etheridge, D. M. *et al.* Natural and anthropogenic changes in atmospheric CO_2 over the last

1000 years from air in Antarctic ice and firn. *J. Geophys. Res. Atmospheres* **101**, 4115–4128

(1996).

65. Buitenhuis, E. T., Timmermans, K. R. & de Baar, H. J. W. Zinc-bicarbonate colimitation of

Emiliania huxleyi. *Limnol. Oceanogr.* **48**, 1575–1582 (2003).

66. Riebesell, U., Wolf-Gladrow, D. A. & Smetacek, V. Carbon dioxide limitation of marine

phytoplankton growth rates. *Nature* **361**, 249–251 (1993).

67. Sharp, J. D. *et al.* CO2SYSv3 for MATLAB. Zenodo

https://doi.org/10.5281/ZENODO.3950562 (2023).

68. Krumhardt, K. M., Lovenduski, N. S., Iglesias-Rodriguez, M. D. & Kleypas, J. A.

Coccolithophore growth and calcification in a changing ocean. *Prog. Oceanogr.* **159**, 276–

295 (2017).

69. Sunda, W. G. & Huntsman, S. A. Cobalt and zinc interreplacement in marine phytoplankton:

Biological and geochemical implications. *Limnol. Oceanogr.* **40**, 1404–1417 (1995).

70. Wyatt, N. J. *et al.* Biogeochemical cycling of dissolved zinc along the GEOTRACES South

Atlantic transect GA10 at 40°S. *Glob. Biogeochem. Cycles* **28**, 44–56 (2014).

71. Cohen, N. R. *et al.* Hydrothermal trace metal release and microbial metabolism in the

northeastern Lau Basin of the South Pacific Ocean. *Biogeosciences* **18**, 5397–5422 (2021).

72. Rapp, I., Schlosser, C., Rusiecka, D., Gledhill, M. & Achterberg, E. P. Automated

preconcentration of Fe, Zn, Cu, Ni, Cd, Pb, Co, and Mn in seawater with analysis using high-

resolution sector field inductively-coupled plasma mass spectrometry. *Anal. Chim. Acta*

(2017) doi:10.1016/j.aca.2017.05.008.

73. Wuttig, K. *et al.* Critical evaluation of a seaFAST system for the analysis of trace metals in

marine samples. *Talanta* (2019) doi:10.1016/j.talanta.2019.01.047.

74. DiTullio, G. R., Geesey, M. E., Leventer, A. & Lizotte, M. P. Algal pigment ratios in the

Ross Sea: Implications for Chemtax analysis of Southern Ocean data. in 35–51 (2003).

doi:10.1029/078ARS03.

75. Schanke, N. L. *et al.* Biogeochemical and ecological variability during the late summer–early

autumn transition at an ice-floe drift station in the Central Arctic Ocean. *Limnol. Oceanogr.*

(2021) doi:10.1002/lno.11676.

76. Balestra, C., Alonso-Sáez, L., Gasol, J. M. & Casotti, R. Group-specific effects on coastal

bacterioplankton of polyunsaturated aldehydes produced by diatoms. *Aquat. Microb. Ecol.*

(2011) doi:10.3354/ame01486.

77. Cox, A. D., Noble, A. E. & Saito, M. A. Cadmium enriched stable isotope uptake and

addition experiments with natural phytoplankton assemblages in the Costa Rica Upwelling

Dome. *Mar. Chem.* **166**, 70–81 (2014).

78. Hughes, C. S. *et al.* Ultrasensitive proteome analysis using paramagnetic bead technology.

Mol. Syst. Biol. **10**, 757 (2014).

79. Zhang, Y., Wen, Z., Washburn, M. P. & Florens, L. Refinements to Label Free Proteome

Quantitation: How to Deal with Peptides Shared by Multiple Proteins. *Anal. Chem.* **82**,

2272–2281 (2010).

80. Altschul, S. F., Gish, W., Miller, W., Myers, E. W. & Lipman, D. J. Basic local alignment

search tool. *J. Mol. Biol.* (1990) doi:10.1016/S0022-2836(05)80360-2.

81. Bolger, A. M., Lohse, M. & Usadel, B. Trimmomatic: a flexible trimmer for Illumina

sequence data. *Bioinformatics* **30**, 2114–2120 (2014).

82. Schmieder, R., Lim, Y. W. & Edwards, R. Identification and removal of ribosomal RNA

sequences from metatranscriptomes. *Bioinformatics* **28**, 433–435 (2012).

83. Rho, M., Tang, H. & Ye, Y. FragGeneScan: predicting genes in short and error-prone reads.

Nucleic Acids Res. **38**, e191–e191 (2010).

84. Mortazavi, A., Williams, B. A., McCue, K., Schaeffer, L. & Wold, B. Mapping and

quantifying mammalian transcriptomes by RNA-Seq. *Nat. Methods* **5**, 621–628 (2008).

85. Li, B. & Dewey, C. N. RSEM: accurate transcript quantification from RNA-Seq data with or

without a reference genome. *BMC Bioinformatics* **12**, 323 (2011).

86. Eddy, S. R. Accelerated Profile HMM Searches. *PLoS Comput. Biol.* **7**, e1002195 (2011).

87. Aramaki, T. *et al.* KofamKOALA: KEGG Ortholog assignment based on profile HMM and

adaptive score threshold. *Bioinformatics* **36**, 2251–2252 (2020).

88. Krogh, A., Larsson, B., von Heijne, G. & Sonnhammer, E. L. L. Predicting transmembrane

protein topology with a hidden markov model: application to complete genomes11Edited by

F. Cohen. *J. Mol. Biol.* **305**, 567–580 (2001).

89. Tatusov, R. L. *et al.* The COG database: an updated version includes eukaryotes. *BMC*

Bioinformatics **4**, 41 (2003).

90. Podell, S. & Gaasterland, T. DarkHorse: a method for genome-wide prediction of horizontal

gene transfer. *Genome Biol.* **8**, R16 (2007).

Figures

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Figure 1. **Temporal biogeochemistry of Terra Nova Bay, Jan-Feb 2018.** (**a**) Sampling
locations over the Ross Sea shelf in Terra Nova Bay, Antarctica. (**b**) Location of station 27 and
surrounding seawater pCO_2 measured over a three-day transit northwards represented in color
scale. Locations sampled on January 11, 2018 and on January 14, 2018 are indicated. (**c**)
chlorophyll *a* (chl *a*), (**d**) fucoxanthin (fuco), (**e**) 19' hexanoyloxyfucoxanthin (19'-hex), (**f**) total
dissolved Zn (dZn_T), and (**g**) total Zn uptake rates for all Terra Nova Bay stations over time
measured in the upper 250m. Station numbers are given at the top of (c) and (f). Station data is
presented in order of sampling date, from the earliest (Stn 22, early January) to the latest (Stn 79,
late February). (**h**) negative linear relationship between seawater CO_2 partial pressure (pCO_2) at
5 m depth and Zn uptake rates (ρZn) measured at surface ($\leq 10m$) depths for 15 stations of the
CICLOPS cruise (n = 15, R^2 = 0.63). (**i**) positive linear relationship between chlorophyll *a* at 10
m depth and Zn uptake rates (ρZn) measured at 10 m depth for 9 stations of the CICLOPS cruise
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sample due to icebreaking duties for the McMurdo Station resupply ship. Stations indicated in
(a) are those where the trace metal rosette (TMR) was deployed; pigment data was supplemented
with additional stations using a CTD.

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Figure 2. Evidence for Zn co-limitation with Fe in bottle incubations. (**a**) Chlorophyll *a* and
(**b**) total salinity-normalized DIC in each treatment at T6. Significant differences among groups
were found using one-way ANOVA and post-hoc Dunnett test (*** $p < 0.001$, ** $p < 0.01$, * $p <$
0.05). Error bars are the standard deviation of biological triplicates (n=3). Individual data points
are overlaid (white circles). (**c**) Heatmap of log2 exclusive protein spectral counts showing
relative protein abundance at the KOG annotation level. The dendrogram shows similarity in
spectral abundance among samples based on Euclidean distance and hierarchical clustering. Each
row represents a KOG annotation. Spectral counts associated with identical KOG descriptions
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Ribulose-1,5-biphosphate carboxylase/oxygenase (RUBISCO), zinc/cobalt responsive protein A
(ZCRP-A), and iron starvation induced proteins (ISIP1, ISIP2A, ISIP3) are shown. (**d**), pCO_2
measured over time within TNB during a 3-day transit. The vertical red line denotes the pCO_2
level at the time of initial seawater collection at station 27.

Figure 3. Metatranscriptomic and metaproteomic detection of Zn- and Fe-related proteins of interest at station 27. Depth profiles of summed NSAF-normalized protein spectral counts of (**a**) RUBISCO, (**b**) ZCRP-A, (**c**) ZCRP-B, (**d**) iron starvation induced proteins (ISIPs), (**e**) ZIP, (**f**) Delta CA, and (**g**) Theta CA for each size fraction (0.2, 3 and 51µm). (**h**) TPM-normalized transcript read counts of CDCA. (**i**) stacked pie charts depicting relative community composition for proteins of interest for euphotic (< 200 m) and mesopelagic (≥ 200 m) depths. The outer rings show community composition based on NSAF-normalized protein spectral counts while the inner rings are TPM-normalized transcript read counts. Proteins and transcripts in (i) were summed across all size fractions. ISIPs are the combined spectral counts of ISIP1A, ISIP2A and ISIP3.

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Figure 4. **The partial pressure of CO_2 (pCO₂) and associated phytoplankton responses from this study and the literature**. **(a)** Pre-industrial and decadal surface ocean pCO₂ reconstructions plotted as violin plots, with a running black line through the median values. The atmospheric curve is a composite of ice core data (dashed yellow line[64]) and the Mauna Loa record (solid yellow line[63]). Thresholds for zinc-limited growth are plotted as the median of previous laboratory results (259 µatm, dark green line; see Methods), and estimated from the *in situ* results of this study (220 µatm, light green line). **(b)** Data in (a) plotted as a histogram comparing preindustrial and modern (2010) pCO₂ values, with the same Zn growth limitation thresholds indicated.

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Main Text Figures

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