Easy, fast and reproducible Stochastic Cellular Automata with chouca

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ABSTRACT

Stochastic cellular automata (SCA) are models that describe spatial dynamics using a grid of cells that switch between discrete states over time, depending only on the current state (Markovian processes). They are widely used to understand how small-scale processes scale up to affect ecological dynamics at larger spatial scales, and have been applied to a wide diversity of theoretical and applied problems in all systems, such as arid ecosystems, coral reefs, forests, bacteria, or urban growth.

Despite their wide applications, SCA implementations found in the ecological literature are often ad-hoc, lacking performance and guarantees of correctness. More importantly, *de novo* implementation of SCA for each specific system and application represents a major barrier for many practitioners. To provide a unifying, well-tested technical basis to this class of models and facilitate their implementation, we built *chouca*, an R package that translates intuitive SCA model definitions into compiled code, and runs simulations in a reproducible and efficient way.

chouca supports a wide set of SCA as well as deterministic cellular automata, with performance typically two to three orders of magnitude above *ad hoc* implementations found in the literature, all while maintaining an intuitive interface in the R environment. Exact and

45 mean-field simulations can be run, and both numerical and graphical results can be easily
46 exported.
47
Besides providing better reproducibility and accessibility, a fast engine for SCA unlocks novel,
computationally intensive statistical approaches, such as simulation-based inference of
ecological interactions from field data, which represents by itself an important avenue for
research. By providing an easy and efficient entry point to SCAs, *chouca* lowers the bar to the
use of this class of models for ecologists, managers and general practitioners, providing a
leveled-off reproducible platform while opening novel methodological approaches.
54
55 *Keywords:* spatial ecology, spatial modelling, cellular automaton, R package
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Introduction

The analysis of spatial patterns has proven essential to understand ecological system dynamics, and various modelling approaches have helped ground empirical patterns into ecological theory. Among such approaches, models based on Stochastic Cellular Automata (hereafter SCA), also called Probabilistic or Random Cellular Automata, or Locally-interacting Markov Chains, have been a particularly useful, heuristic and widely used approach (Wolfram, 1984; Louis & Nardi, 2018). Cellular automata are based on a grid of cells that switch over time between a finite number of states. Most often, SCA are considered to be distributed on a rectangular grid, though other geometries can exist (van Baalen, M., 2000). A famous deterministic cellular automaton (CA) is Conway's game of life, which is defined by two discrete states ("dead" and "alive") and a set of deterministic rules to make cells switch between them. Stochastic cellular automata follow the same principles, but state transitions occur with a given probability instead of being based on deterministic rules. The probabilities of a cell switching from one state to another is assumed to depend on model parameters, the global state of the system (the proportion of cells in each state), and the local neighborhood of the focal cell. In all cases, the system future dynamics is probabilistically defined by its current state, *i.e.* dynamics are memoryless.

The use of SCA in ecology is widespread, as they have been used to describe the dynamics of a large array of ecosystems, including mussel beds (Guichard et al., 2003), arid ecosystems (Kéfi et al., 2007), forests (Heinonen & Pukkala, 2007), rocky shores (Wootton, 2001), coral reefs (Muthukrishnan et al., 2016; Génin et al., 2024) and plant communities (Lanzer & Pillar, 2002). SCA are often used to understand how local processes can scale up to affect landscape-wide properties, such as the persistence or extinction of a given species, the type of spatial patterns arising at the scale of a landscape (Pascual et al., 2002), or the spread of fire or epidemics. The latter cases is a classical use-case of SCA for applied ecology, where data on local processes affecting forest stands can be coupled with GIS data to provide guidance on forest fire sensitivity (Yassemi et al., 2008).

The popularity of SCA is probably rooted in their relatively light need for formal mathematics compared to other approaches modelling spatial processes (*e.g.* partial derivative equations): only the probabilities of transitions between states need to be defined. The drawback of this simplicity is that the numerical simulation of SCA is typically compute-intensive. Current approaches to do so efficiently rely on approximations, either assuming no spatial structure (mean field approximation) or simplifying the spatial dynamics to pairs of neighboring cells ("pair approximation"; Matsuda et al., 1992; Iwasa, 2000). However, these approximations are inappropriate when long-range correlations occur within a landscape (Iwasa, 2000), or when the full simulated landscape is needed as a model output, for example to compute spatial metrics (Génin et al., 2018). In many cases, the explicit numerical simulation must be run, which is often done on small grids to reduce computation time. This is an issue as some theoretical results involving spatial structure or system stability have been shown to depend on grid size (van de Koppel et al., 2011; Majumder et al., 2021). On top of these performance issues, most ecological studies based on SCA come with their own, ad-hoc implementation. This opens the possibility of errors in code and often makes it difficult to reproduce model simulations. We aim at alleviating these issues with *chouca*, which provides a unifying and well-tested technical basis to SCA. Our goal is to improve the performance and accessibility of this class of models, and ultimately allow ecologists to spend more time exploring the behavior of their models, rather than on their implementation.

The R package *chouca* works with 2-dimensional rectangular grids of cells (a "landscape"). Each cell can be in one of a finite set S of n elementary states $S_1 \ldots S_n$. Probabilities of transition are assumed to depend only on (i) the proportion of neighbors in each state, captured by the vector $\boldsymbol{q} = (q_1, \ldots, q_n)$, (ii) the proportion of cells in a given state in the whole landscape, $\boldsymbol{p} = (p_1, \ldots, p_n)$, and (iii) a set of constant model parameters $\boldsymbol{\theta}$. It is important to note that this excludes cellular automata in which an intermediate distance of interaction is considered (*e.g.* through a dispersion kernel; Muthukrishnan et al., 2016), or those in which a preferential direction exists (*e.g.* modeling water redistribution on a slope; Mayor et al., 2013). Other types of SCA not fitting these constraints are those in which two cells swap their respective state, *e.g.* when modelling the movements of a predator in a landscape (Pascual et al., 2002). The probability of transition from state i to j, $P(S_i \rightarrow S_j)$, can be any function of $\boldsymbol{p}$, $\boldsymbol{q}$, and $\boldsymbol{\theta}$ – however, it will work best and without approximation if it has the following form:

(1) $\qquad P(S_i \rightarrow S_j) = a_0 + g_1(q_1) + \cdots + g_n(q_n) + \zeta(\boldsymbol{q}, \boldsymbol{q}) + \zeta(\boldsymbol{q}, \boldsymbol{p}) + \zeta(\boldsymbol{p}, \boldsymbol{p})$

where, for any transition $S_i \rightarrow S_j$, a_0 is a constant, g_s is any univariate function of q_s, and $\zeta(\boldsymbol{x}, \boldsymbol{y})$ is the
sum, defined for two vectors $\boldsymbol{x} = (x_1, \ldots, x_n)$ and $\boldsymbol{y} = (y_1, \ldots, y_n)$ as

(2) $\qquad \zeta(\boldsymbol{x}, \boldsymbol{y}) = a_1 x_1^{\alpha_1} y_1^{\beta_1} + a_2 x_1^{\alpha_2} y_2^{\beta_2} + a_3 x_2^{\alpha_3} y_1^{\beta_3} + \cdots + a_K x_n^{\alpha_K} y_n^{\beta_K}$

in which the $(a_i)_{i \in [1:K]}$, $(\alpha_i)_{i \in [1:K]}$ and $(\beta_i)_{i \in [1:K]}$ are constants and K is the total number of terms of
the sum. In practice, this functional form is flexible enough to approximate the probabilities of transition
of many ecological models.
Implementing and working with an SCA in chouca typically consists in four steps, in which the user 1)
defines the model states and the transitions between them, (2) creates an initial landscape (grid of cells),
(3) runs the model, and (4) displays or extracts the results (Figure 1). We detail in this paper this workflow
(Figure 1) – documentation is available throughout the package, individually for each function, or as a
whole in an R "vignette", accessible with the command vignette("chouca-package").
121

123   **Figure 1** - Main tasks (boxes) of the chouca package, and their associated sets of functions

124 **Using chouca: a simple model**
125
To illustrate how a stochastic cellular automaton can be defined with chouca, we use the model of
Guichard et al. (2003), which describes the dynamics of mussels colonizing rocks exposed to waves. This
model has three cell states (i) "disturbed", (ii) "empty" and (iii) "occupied" (by mussels). During a single
time-step, a disturbed cell becomes an empty cell with probability 1. Such transition can be defined in
chouca by using a call to the R function transition():

```
transition(from = "disturbed", to = "empty", ~ 1)
```

132   This statement declares a transition from a "disturbed" state to an "empty" state, with the last
133 argument being a symbolic expression starting with "~", that describes how to compute the probability,
134 here being simply equal to the constant "1".
135   The model assumes that the establishment of new individuals always occurs next to existing mussels.
136 In other words, for a focal cell in the "bare" state, its probability of switching to the "mussel" state is not
137 constant, but given by αq_{mussel}, where α is a productivity rate, and q_{mussel} is the proportion of cell
138 neighbors in the mussel state. Such transition is defined by the following call in R:

139   ```
transition(from = "empty", to = "mussel", ~ alpha * q["mussel"])
```

Here, `q["mussel"]` is used to refer to the proportion of cell neighbors in the "mussel" state, a
continuous number between 0 and 1. Similarly, `p["mussel"]` could have been used to refer to the global
proportion of cells in the landscape in the "mussel" state.

Mussels can be perturbed by incoming waves, which dislodge them and turn them into "disturbed"
cells. In this model, the probability of a mussel cell to become disturbed is the sum of a baseline term $\delta$,
and an additional term $d$, which is non-zero only if the mussel cell has one or more disturbed neighbors:

`transition(from = "mussel", to = "disturbed", ~ delta + d * ( q["disturbed"] > 0 ))`

The original model considers that cells are neighbors when they share an edge (4-way or von-Neumann
neighborhood, the other option being a Moore or 8-way neighborhood), and uses a toric space for
simulations, meaning that the up/leftmost cells of the grid are neighbors of the bottom/rightmost cells.
Putting everything together, this model can be defined using the following syntax:

```
151     musselbed_mod <- camodel(
152       transition(from = "disturbed", to = "empty",  ~ 1),
153       transition(from = "empty", to = "mussel",      ~ alpha * q["mussel"]),
154       transition(from = "mussel", to = "disturbed", ~ delta + d * ( q["disturbed"] > 0 )),
155       parms = list(alpha = 1, delta = 0.2, d = 1), # parameters
156       wrap = TRUE, # toric space
157       neighbors = 4 # 4-way neighborhood
158     )
```

159   The call to the `camodel()` function above estimates the constants in the functional form described in
160   equation 1 to match the parameters of the model. If this process yields large residual error, for example
161   because the transition probabilities do not correspond to the functional form in equation 1, a warning is
162   produced. The structure of the model can be displayed on the R console using `print()`, or as a graph using
163   the generic function `plot()` (Figure 2).

```
164     > print(musselbed_mod)
165     Stochastic Cellular Automaton
166     States: disturbed empty mussel
167
168     Transition: disturbed -> empty
169        ~   1
170     Transition: empty -> mussel
171        ~   alpha * q["mussel"]
172     Transition: mussel -> disturbed
173        ~   delta + d * (q["disturbed"] > 0)
174
175     Neighborhood: 4x4
176     Wrap: TRUE
177     Max error: 0 (OK)
178     Max rel error: 0 (OK)
```

Figure 2 Structure of the mussel bed model displayed as a graph, in which nodes are states, and
arrows represent the possible transitions, with the expression used to compute their probabilities

Once the model is created, an initial landscape can be defined with a given number of rows and columns
using `generate_initmat()`, which creates a landscape with randomly-distributed cell states in space, but
following the specified initial covers:

```
init_landscape <- generate_initmat(musselbed_mod,
                                   pvec = c(disturbed = 0.1, empty = 0.1, mussel = 0.8),
                                   nrow = 64, ncol = 64)
```

The model can then be simulated for a given number of time steps, below from 0 to 50, using
`run_camodel()`. Standard methods such as `plot()` or `image()` can be used to display the global covers
after the model has run, or the resulting landscapes (respectively):

```
output <- run_camodel(musselbed_mod, init_landscape, times = seq(0, 50))
plot(output)
image(output)
```

By default, chouca uses a C++ backend based on Rcpp (Eddelbuettel 2013), which has reasonable
performance. This can be improved on by compiling the model code just before the simulation is run, and
using memoization so that transition probabilities are computed only once for cells with the same
neighborhood configuration. This allows reaching typical simulation speeds above 10,000 iterations.s^{-1} on
a typical 128 x 128 grid (Figure 3), which can be further increased by parallelizing computations over
multiple cores, though this is a less efficient approach. Enabling these options can be done by passing
control arguments as an R list object:

```
control_args <- list(engine = "compiled",
                     cores = 4,
                     precompute_probas = TRUE)
output <- run_camodel(musselbed_mod, init_landscape, niter = 256, control = control_args)
```

This "control" list defines how the simulation is run, which data to save from the simulation, or the
textual output to print while the simulation is running (a complete list of options is in the help page

207   `?run_camodel`). The user can also supply custom functions that will be run as the simulation is running.
208   This can be useful, for example, to display landscapes, covers or compute statistics on the 2D landscape as
209   the simulation is running, which we illustrate in the following sections.
210

212   **Figure 3** Simulation speeds for a set of three simple ecological models (2-3 states and 3-4 transitions)
213   according the the grid size, for a pure-R implementation (R reference; Schneider et al. 2016), and
214   three backends provided by chouca (blue and green lines). Single-core performance on a 2020
215   desktop computer.

Example applications

Graphical explorations

Because an SCA describes dynamics over landscapes, they are particularly well-adapted to pattern-oriented modelling (Grimm et al., 1996), in which a model is defined and revised based on a qualitative or quantitative comparison with empirical patterns. Likelihood-based approaches are increasingly popular to compare models with data (Hartig et al., 2011 and section below), but the qualitative comparison and visual exploration of model dynamics remains an essential phase for spatial models. To make this modeling step more accessible, we made it easy to investigate visually the behavior of models, and illustrate here this approach with an epidemiological example.

(Keeling, 2000) uses an SCA-based approach to investigate the spread of a parasite over space, with an application to forests. This model uses three states, "host", "parasitized", "empty", and can be defined as follows in chouca:

```
mod <- camodel(transition(from = "empty", to = "host",
                          ~ 1 - ( 1 - g )^( 4 * q["host"] ) ),
               transition(from = "host", to = "parasitized",
                          ~ 1 - ( 1 – T )^( 4 * q["parasitized"] )),
               transition(from = "parasitized", to = "empty",
                          ~ 1),
               parms = list(g = 0.05, T = 0.5),
               wrap = TRUE,
               neighbors = 4)
```

where g is the growth rate of the host, and T the transmissibility of the parasite (this model assumes that infection is always fatal, so the transition from "parasitized" to "empty" is equal to 1).

This model can produce interesting "epidemiological fronts", which stem from the way the parasite spreads to its neighboring hosts, killing them and leaving behind empty, bare patches that are later recolonized by the host, albeit at a slower rate. This phenomenon of fronts may be difficult to quantify formally, but can be easily visualized from model outputs. This can be done in chouca by setting the run_camodel() function to display the results as the simulation is run. To do so, we adjust the control list to include a function that displays the landscapes:

```
options <- list(custom_output_fun = landscape_plotter(mod),
                custom_output_every = 1)
```

then run the model on a 256×256 grid seeded with 10% of cells in the "parasitized" state:

```
initmm <- generate_initmat(mod, c(host = 0.9, parasitized = 0.1, empty = 0),
                           nrow = 256, ncol = 256)
output <- run_camodel(mod, initmm, times = seq(0, 1024), control = options)
```

The above lines of code will run the model and display the changing landscape, which allows investigating the spreading patterns of the parasite. Once this visualization step is done (Figure 4, and animated version in SM1), the options set to visualize the landscape can be removed to reduce simulation times, for example to investigate model behavior along a range of parameter values.

257 **Figure 4** Landscape patterns of the host-parasitized-empty model (respectively green, yellow and
258 blue) as displayed on screen

Inference of local interactions from landscape-scale patterns

259
Because SCA are defined on grids, a natural application is to compare their output to empirical raster
data, such as remote-sensing images, to infer local-scale ecological interactions from landscape-wide
spatial patterns. Arid systems provide a good illustration of this approach: in those systems, plants often
facilitate each other, which results in their aggregation into patches, and has important consequences for
the resilience of those systems to changes in aridity (Kéfi et al., 2007). The sizes and numbers of those
patches can be readily quantified from remote-sensing images, and such patterns can be used to infer
whether facilitation occurs between plants (Xu et al., 2015; Chen et al., 2022). This is traditionally done by
summarizing the spatial structure into spatial statistics, such as spatial autocorrelation (Sankaran et al.,
2017) or type of patch size distribution (Siteur et al., 2023; Pichon et al., 2024) and linking the observed
changes in those metrics to theoretical results (Scanlon et al., 2007; Kéfi et al., 2024). However, such
qualitative comparison logically results in a qualitative and corroborative result, *i.e.* is there or is there not
facilitation between plants, rather than a more informative quantitative result, *i.e.* how strong facilitation
is between plants. A quantitative inference must be based on an approach that links quantitatively some
aspects of the spatial patterns to the strength of facilitation. This can be done by defining a model that
links the strength of facilitation with the expected patterns, and finding model parameters that maximize
agreement between model output and data. This is expensive computationally, but this limitation is
alleviated by a fast SCA engine such as chouca, as we show below.

We define here a model of an arid ecosystem with two states, "bare" and "vegetated". A bare cell can
become a vegetated cell with the probability p_{plant}. A vegetated cell can become bare (plants die) with the
probability $d(1 - \beta q_{plant})$, where d is a constant mortality rate, that is reduced by the coefficient β. β
captures the local effect of plants, either facilitation when they increase the survival of plants near existing
vegetation ($\beta > 0$), or competition when survival is decreased ($\beta < 0$). This model is implemented in
chouca using:

```
283    facilitation_mod <- camodel(
284      transition(from = "bare", to = "plant", ~ p["plant"] ),
285      transition(from = "plant", to = "bare", ~ d * ( 1 – beta * q["plant"]) ),
286      parms = list(beta = 0.5, d = 0.85),
287      wrap = TRUE,
288      neighbors = 4
289    )
```

We ran the model till equilibrium on a 1024×1024 grid, to simulate a landscape that would be obtained
from empirical data (e.g. a remote sensing image) using $(d, \beta) = (0.85, 0.5)$. From this landscape used as
observed data, we computed the distribution of pairs, which summarizes all the possible pairs of
neighboring cells present in the landscape $\boldsymbol{n_{obs}} = (n_{p,0}, n_{p,p}, n_{0,0})$. We the define the likelihood
$P(\boldsymbol{n_{obs}} \mid d, \beta)$ assuming these observed number of pairs follow a multinomial distribution of size N_p (the
sum of all pairs, a fixed number given the grid size and neighborhood type), and probabilities $\boldsymbol{\mu} =$
$(\mu_{p,0}, \mu_{p,p}, \mu_{0,0})$. $\boldsymbol{\mu}$ defines the relative probabilities of observing each type of pair in the grid, which
depend on the particular values of d and β, and can be estimated by simulating the landscape with these
parameter values. This estimate of the likelihood formally quantifies the agreement between model and
data, and allows exploring the parameter space in terms of d and β to find the most likely parameter
combinations given the observed patterns.
We find that this approach can recover the parameter values used for d and β, with only one global
maximum in the likelihood function (Figure 5; code provided in SM2), showing that the distribution of pairs
is a sufficient statistic to recover the model parameters. Applying such an approach to empirical data would
require further testing of model assumptions, for example, to investigate whether facilitation occurs on
the recruitment of new plants instead of on the mortality of adult plants – this can be done by simply
changing the model definition above. Because this approach is likelihood-based, model support can be
compared using traditional statistics, such as AIC, and a Bayesian approach can be used to estimate credible
intervals on parameter values, or use informative priors grounded in knowledge about the system (Hartig
et al., 2011).
Using this type of simulation-based inference is very expensive computationally, as this simple test
requires running around a thousand simulations. This would require long-running computations with usual
SCA implementations, but takes just under five minutes with chouca on a 2018 laptop computer (8 cores).
This way of calibrating models has seldom been used in spatial ecology – a fast SCA implementation is
essential to make it more accessible.

Figure 5 Likelihood surface as a function of the two model parameters d and β. Blue values indicate
parameter combinations with high likelihood. The white dot and lines indicate the estimated
parameter values.

Conclusion

`chouca` is an easy-to-use package to model, simulate, and visualize SCA in a reproducible way, that enables an interactive design and revision of models as well as novel methodological approaches. chouca does not support all types of cellular automata, and does not replace more generic modelling frameworks such as `NetLogo` (Wilensky, 1999) or `simecol` (Petzoldt & Rinke, 2007), but allows efficient simulations for the type of SCA it supports. It is important to note that because chouca splits the definition of an SCA model from its simulation phase, it may use different backends for simulation. This opens the possibility of future improvements to already-implemented models without modification of existing code.

Because of the relatively low bar of entry of SCA compared to other forms of spatial modelling, we hope this work will contribute to make more accessible the testing of hypotheses linking spatial processes to patterns in ecology, as well exploring system-level consequences of specific local management decisions. `chouca` is available on CRAN, and welcomes comments, feedback and bug reports on its home page at https://github.com/alexgenin/chouca.

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