

1 Easy, fast and reproducible Stochastic Cellular 2 Automata with chouca

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26 **ABSTRACT**

27 Stochastic cellular automata (SCA) are models that describe spatial dynamics using a grid of
28 cells that switch between discrete states over time. They are widely used to understand how
29 small-scale processes scale up to affect ecological dynamics at larger spatial scales, and have
30 been applied to a wide diversity of theoretical and applied problems in all systems, such as
31 arid ecosystems, coral reefs, forests, bacteria, or urban growth.

32
33 Despite their wide applications, SCA implementations are often ad-hoc, lacking performance,
34 guarantees of correctness and poorly reproducible. *De novo* implementation of SCA for each
35 specific system and application also represents a major barrier for many practitioners. To
36 provide a unifying, well-tested technical basis to this class of models and facilitate their
37 implementation, we built *chouca*, an R package that translates intuitive SCA model definitions
38 into compiled code, and runs simulations in an efficient way.

39
40 *chouca* supports a wide set of SCA as well as deterministic cellular automata, with
41 performance typically two to three orders of magnitude above typical implementations in
42 interpreted languages (e.g. R, Python), all while maintaining an intuitive interface in the R
43 environment. Exact and mean-field simulations can be run, and both numerical and graphical
44 results can be easily exported.

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Besides providing better reproducibility and accessibility, a fast engine for SCA unlocks novel, computationally intensive statistical approaches, such as simulation-based inference of ecological interactions from field data, which represents by itself an important avenue for research. By providing an easy and efficient entry point to SCAs, *chouca* lowers the bar to the use of this class of models for ecologists, managers and general practitioners, providing a leveled-off reproducible platform while opening novel methodological approaches.

Keywords: spatial ecology, spatial modelling, cellular automaton, R package

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Introduction

56 The analysis of spatial patterns has proven essential to understand ecological system dynamics, and
57 various modelling approaches have helped ground empirical patterns into ecological theory. Among such
58 approaches, models based on Stochastic Cellular Automata (hereafter SCA), also called Probabilistic or
59 Random Cellular Automata, or Locally-interacting Markov Chains, have been a particularly useful, heuristic
60 and widely used approach (Wolfram, 1984; Louis & Nardi, 2018). Cellular automata are based on a grid of
61 cells that switch over time between a finite number of states. Most often, SCA are considered to be
62 distributed on a rectangular grid, though other geometries can exist (van Baalen, M., 2000). A famous
63 deterministic cellular automaton (CA) is Conway's game of life, which is defined by two discrete states
64 ("dead" and "alive") and a set of deterministic rules to make cells switch between them (Gardner, 1970;
65 Bays, 2010). Stochastic cellular automata follow the same principles, but state transitions occur with a
66 given probability instead of being based on deterministic rules. The probabilities of a cell switching from
67 one state to another is assumed to depend on model parameters, the global state of the system (the
68 proportion of cells in each state), and the local neighborhood of the focal cell. In all cases, the system future
69 dynamics is probabilistically defined by its current state, *i.e.* dynamics are memoryless.

70 The use of SCA in ecology is widespread, as they have been used to describe the dynamics of a large
71 array of ecosystems, including mussel beds (Guichard et al., 2003), arid ecosystems (Kéfi et al., 2007),
72 forests (Heinonen & Pukkala, 2007), rocky shores (Wootton, 2001), coral reefs (Muthukrishnan et al., 2016;
73 Génin et al., 2024) and plant communities (Lanzer & Pillar, 2002). SCA are often used to understand how
74 local processes can scale up to affect landscape-wide properties, such as the persistence or extinction of a
75 given species, the type of spatial patterns arising at the scale of a landscape (Pascual et al., 2002), or the
76 spread of fire or epidemics. The latter cases is a classical use-case of SCA for applied ecology, where data
77 on local processes affecting forest stands can be coupled with GIS data to provide guidance on forest fire
78 sensitivity (Yassemi et al., 2008).

79 The popularity of SCA is probably rooted in their relatively light need for formal mathematics compared
80 to other approaches modelling spatial processes (*e.g.* partial derivative equations): only the probabilities
81 of transitions between states need to be defined. The drawback of this simplicity is that the numerical
82 simulation of SCA is typically computationally-intensive. Current approaches to do so efficiently rely on
83 approximations, either assuming no spatial structure (mean field approximation), or using approximations
84 to obtain analytical results on spatial structure, such as spatial moment equations (Lion, 2016; of which
85 pair approximation is a specific case; Matsuda et al., 1992; Iwasa, 2000). However, these approximations
86 are inappropriate when long-range correlations occur within a landscape (Iwasa, 2000), or when the full
87 simulated landscape is needed as a model output, for example to compute spatial metrics (Génin et al.,
88 2018). In many cases, the explicit numerical simulation must be run, which is often done on grids that may
89 not be large enough to fully capture emergent spatial patterns (van de Koppel et al., 2011; Majumder et
90 al., 2021). On top of these performance issues, most ecological studies based on SCA come with their own
91 implementation. This opens the possibility of errors in code and often makes it difficult to reproduce model
92 simulations. We aim at alleviating these issues with *chouca*, which provides a unifying and well-tested
93 technical basis to SCA. Our goal is to improve the performance and accessibility of this class of models, and
94 ultimately allow ecologists to spend more time exploring the behavior of their models, rather than on their
95 implementation.

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Supported models

97 The R package *chouca* works with 2-dimensional rectangular grids of cells (a "landscape"). Each cell can
98 be in one of a finite set S of n elementary states $S_1 \dots S_n$. Probabilities of transition are assumed to depend
99 only on (i) the proportion of neighbors in each state, captured by the vector $\mathbf{q} = (q_1, \dots, q_n)$, (ii) the
100 proportion of cells in a given state in the whole landscape, $\mathbf{p} = (p_1, \dots, p_n)$, and (iii) a set of constant model
101 parameters θ .

102 *chouca* has been primarily designed for modeling the dynamics of sessile organisms over space, which
103 do not move and reproduce through the dispersal of propagules. This includes a wide range of organisms,
104 ranging from forests, herbaceous plants or coral reefs, but the supported models may also be useful to

105 describe other ecological situations. For the moment, the types of models that can be implemented with
 106 chouca currently exclude cellular automata in which an intermediate distance of interaction is considered
 107 (e.g. through a dispersion kernel; Muthukrishnan et al., 2016), or those in which a preferential direction
 108 exists (e.g. modeling water redistribution on a slope; Mayor et al., 2013), though these limitations are
 109 planned to be lifted in later versions. More importantly, SCA that are based on the dynamics of pairs of
 110 cells, for instance in which two cells swap their respective state, cannot be implemented directly here, as
 111 state transitions are defined at the level of a cell. This is typical when representing the movements of an
 112 organism in a landscape (Pascual et al., 2002), and we provide links to alternative software at the end of
 113 the article for these models.

114 In the package, the probabilities of transition of a cell from state i to j , $P(S_i \rightarrow S_j)$, can be any function
 115 of \mathbf{p} , \mathbf{q} , and $\boldsymbol{\theta}$ – however, it will work best and without approximation if it has the following form:

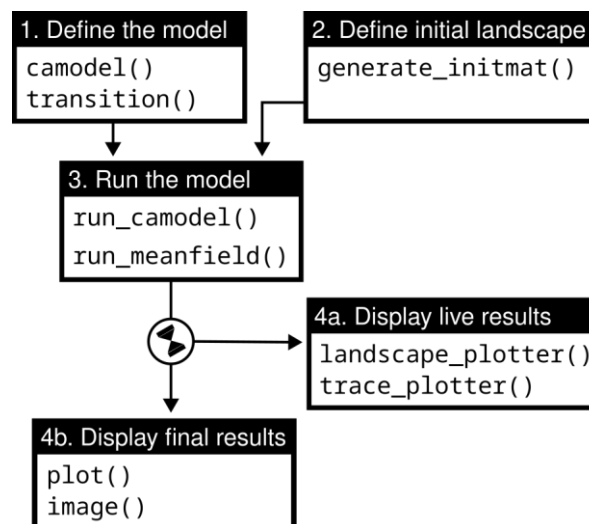
116 (1)
$$P(S_i \rightarrow S_j) = a_0 + g_1(q_1) + \dots + g_n(q_n) + \zeta(\mathbf{q}, \mathbf{q}) + \zeta(\mathbf{q}, \mathbf{p}) + \zeta(\mathbf{p}, \mathbf{p})$$

117 where, for any transition $S_i \rightarrow S_j$, a_0 is a constant, g_s is any univariate function of q_s , and $\zeta(\mathbf{x}, \mathbf{y})$ is the
 118 sum, defined for two vectors $\mathbf{x} = (x_1, \dots, x_n)$ and $\mathbf{y} = (y_1, \dots, y_n)$ as

119 (2)
$$\zeta(\mathbf{x}, \mathbf{y}) = a_1 x_1^{\alpha_1} y_1^{\beta_1} + a_2 x_1^{\alpha_2} y_2^{\beta_2} + a_3 x_2^{\alpha_3} y_1^{\beta_3} + \dots + a_K x_n^{\alpha_K} y_n^{\beta_K}$$

120 in which the $(a_i)_{i \in [1:K]}$, $(\alpha_i)_{i \in [1:K]}$ and $(\beta_i)_{i \in [1:K]}$ are constants and K is the total number of terms of
 121 the sum. In practice, this functional form is flexible enough to approximate the probabilities of transition
 122 of many ecological models.

123 Implementing and working with an SCA in chouca typically consists in four steps, in which the user 1)
 124 defines the model states and the transitions between them, (2) creates an initial landscape (grid of cells),
 125 (3) runs the model, and (4) displays or extracts the results (Figure 1). We detail in this paper this workflow
 126 (Figure 1) – documentation is available throughout the package, individually for each function, or as a
 127 whole in an R “vignette”, accessible with the command `vignette(“chouca-package”)`.
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130 **Figure 1** - Main tasks (boxes) of the chouca package, and their associated sets of functions

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Example applications

A simple model of mussel bed

132 To illustrate how a stochastic cellular automaton can be defined with chouca, we use the model of
 133 Guichard et al. (2003), which describes the dynamics of mussels colonizing rocks exposed to waves. This
 134 model has three cell states (i) “disturbed”, (ii) “empty” and (iii) “occupied” (by mussels). During a single
 135

136 time-step, a disturbed cell becomes an empty cell with probability 1. Such transition can be defined in
137 chouca by using a call to the R function `transition()`:

```
138 transition(from = "disturbed", to = "empty", ~ 1)
```

139 This statement declares a transition from a “disturbed” state to an “empty” state, with the last
140 argument being a symbolic expression starting with “~”, that describes how to compute the probability,
141 here being simply equal to the constant “1”.

142 The model assumes that the establishment of new individuals always occurs next to existing mussels.
143 In other words, for a focal cell in the “bare” state, its probability of switching to the “mussel” state is not
144 constant, but given by αq_{mussel} , where α is a productivity rate, and q_{mussel} is the proportion of cell
145 neighbors in the mussel state. Such transition is defined by the following call in R:

```
146 transition(from = "empty", to = "mussel", ~ alpha * q["mussel"])
```

147 Here, `q["mussel"]` is used to refer to the proportion of cell neighbors in the “mussel” state, a
148 continuous number between 0 and 1. Similarly, `p["mussel"]` could have been used to refer to the global
149 proportion of cells in the landscape in the “mussel” state.

150 Mussels can be perturbed by incoming waves, which dislodge them and turn them into “disturbed”
151 cells. In this model, the probability of a mussel cell to become disturbed is the sum of a baseline term δ ,
152 and an additional term d , which is non-zero only if the mussel cell has one or more disturbed neighbors:

```
153 transition(from = "mussel", to = "disturbed", ~ delta + d * ( q["disturbed"] > 0 ))
```

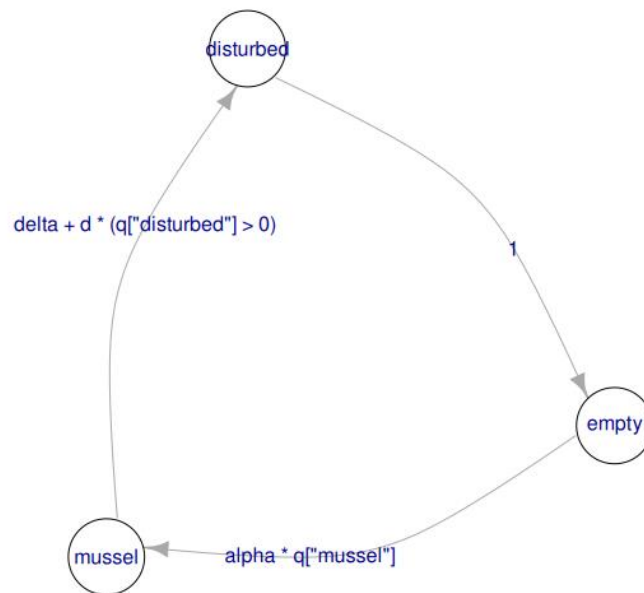
154 The original model considers that cells are neighbors when they share an edge (current options include
155 a 4-way or von-Neumann neighborhood, the other option being a Moore or 8-way neighborhood), and
156 uses a toric space for simulations, meaning that the up/leftmost cells of the grid are neighbors of the
157 bottom/rightmost cells. Putting everything together, this model can be defined using the following syntax:

```
158 musselbed_mod <- camodel(  
159   transition(from = "disturbed", to = "empty", ~ 1),  
160   transition(from = "empty", to = "mussel", ~ alpha * q["mussel"]),  
161   transition(from = "mussel", to = "disturbed", ~ delta + d * ( q["disturbed"] > 0 )),  
162   parms = list(alpha = 1, delta = 0.2, d = 1), # parameters  
163   wrap = TRUE, # toric space  
164   neighbors = 4 # 4-way neighborhood  
165 )
```

166 The call to the `camodel()` function above estimates the constants in the functional form described in
167 equation 1 to match the parameters of the model. If this process yields large residual error, for example
168 because the transition probabilities do not correspond to the functional form in equation 1, a warning is
169 produced. The structure of the model can be displayed on the R console using `print()`, or as a graph using
170 the generic function `plot()` (Figure 2).

```
171 > print(musselbed_mod)  
172 Stochastic Cellular Automaton  
173 States: disturbed empty mussel  
174  
175 Transition: disturbed -> empty  
176 ~ 1  
177 Transition: empty -> mussel  
178 ~ alpha * q["mussel"]  
179 Transition: mussel -> disturbed  
180 ~ delta + d * (q["disturbed"] > 0)  
181  
182 Neighborhood: 4x4  
183 Wrap: TRUE  
184 Max error: 0 (OK)
```

185 Max rel error: 0 (OK)



186

187 **Figure 2** Structure of the mussel bed model displayed as a graph, in which nodes are states, and
188 arrows represent the possible transitions, with the expression used to compute their probabilities

189 Once the model is created, an initial landscape can be defined with a given number of rows and columns
190 using `generate_initmat()`, which creates a landscape with randomly-distributed cell states in space, but
191 following the specified initial covers:

```
192 init_landscape <- generate_initmat(musselbed_mod,  
193                                 pvec = c(disturbed = 0.1, empty = 0.1, mussel = 0.8),  
194                                 nrow = 64, ncol = 64)
```

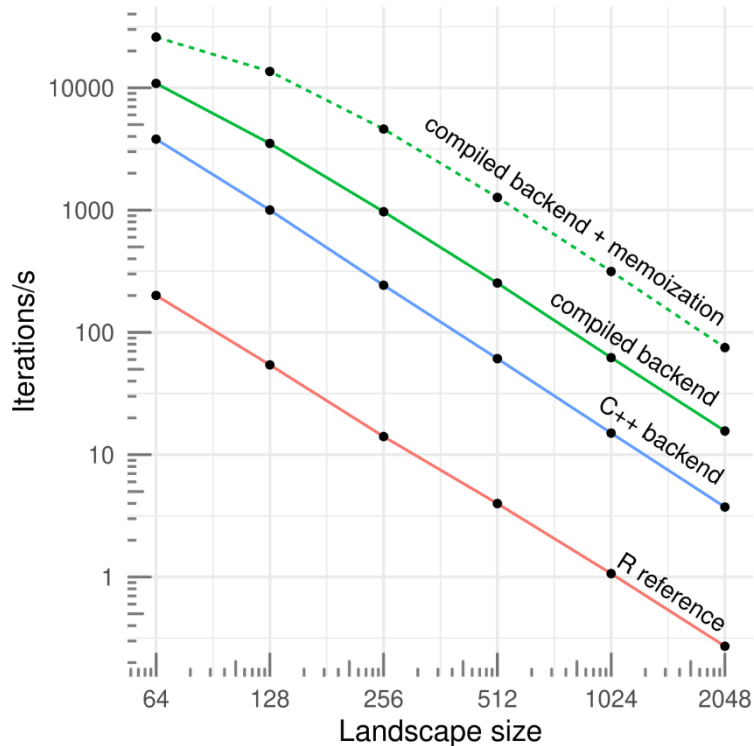
195 The model can then be simulated for a given number of time steps, below from 0 to 50, using
196 `run_camodel()`. Standard methods such as `plot()` or `image()` can be used to display the global covers
197 after the model has run, or the resulting landscapes (respectively):

```
198 output <- run_camodel(musselbed_mod, init_landscape, times = seq(0, 50))  
199 plot(output)  
200 image(output)
```

201 By default, `chouca` uses a C++ backend based on Rcpp (Eddelbuettel 2013), which has reasonable
202 performance. This can be improved by compiling the model code just before the simulation is run, and
203 using memoization so that transition probabilities are computed only once for cells with the same
204 neighborhood configuration. This typically improve performance by two orders of magnitude over a typical
205 implementation (Figure 3; Schneider et al., 2016), which can be further increased by parallelizing
206 computations over multiple cores, though this is a less efficient approach. Enabling these options can be
207 done by passing control arguments as an R list object:

```
208 control_args <- list(engine = "compiled",  
209                    cores = 4,  
210                    precompute_probas = TRUE)  
211 output <- run_camodel(musselbed_mod, init_landscape, niter = 256, control = control_args)
```

212 This “control” list defines how the simulation is run, which data to save from the simulation, or the
213 textual output to print while the simulation is running (a complete list of options is in the help page
214 ?run_camode1). The user can also supply custom functions that will be run as the simulation is running.
215 This can be useful, for example, to display landscapes, covers or compute statistics on the 2D landscape as
216 the simulation is running, which we illustrate in the following sections.
217



218

219 **Figure 3** Simulation speeds for a set of three simple ecological models (2-3 states and 3-4 transitions)
220 according to the grid size, for a pure-R implementation (R reference; Schneider et al., 2016), and
221 three backends provided by chouca (blue and green lines). Single-core performance on a 2020
222 desktop computer.

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224 Graphical explorations

225 Because an SCA describes dynamics over landscapes, they are particularly well-adapted to pattern-
226 oriented modelling (Grimm et al., 1996), in which a model is defined and revised based on a qualitative or
227 quantitative comparison with empirical patterns. Likelihood-based approaches are increasingly popular to
228 compare models with data (Hartig et al., 2011 and section below), but the qualitative comparison and
229 visual exploration of model dynamics remains an essential phase for spatial models. To make this modeling
230 step more accessible, we made it easy to investigate visually the behavior of models, and illustrate here
231 this approach with an epidemiological example.

232 Keeling (2000) uses an SCA-based approach to investigate the spread of a parasite over space, with an
233 application to forests. This very simple model uses three states, “host”, “parasitized”, “empty”, and can be
234 defined as follows in chouca:

```
235 mod <- camodel(transition(from = "empty", to = "host",  
236 ~ 1 - ( 1 - g )^( 4 * q["host"] ) ),  
237 transition(from = "host", to = "parasitized",  
238 ~ 1 - ( 1 - T )^( 4 * q["parasitized"] ) ),  
239 transition(from = "parasitized", to = "empty",  
240 ~ 1),  
241 parms = list(g = 0.05, T = 0.5),  
242 wrap = TRUE,  
243 neighbors = 4)
```

244 where g is the growth rate of the host, and T the transmissibility of the parasite (this model assumes
245 that infection is always fatal, so the transition from “parasitized” to “empty” is equal to 1).

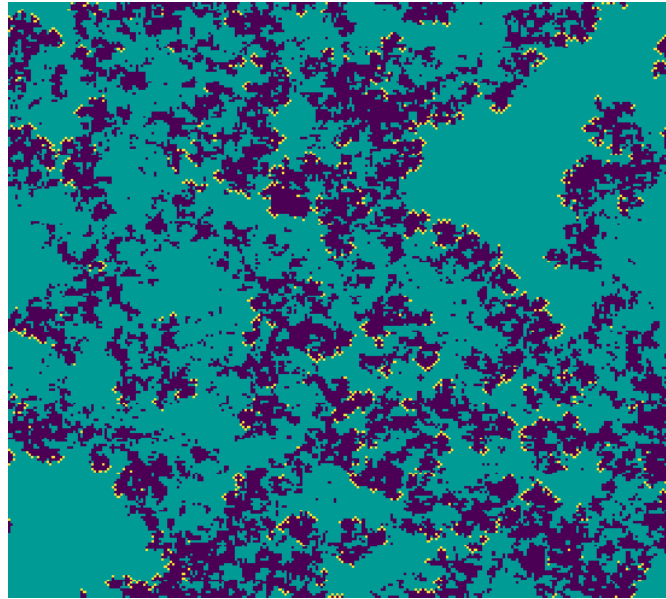
246 This model can produce interesting “epidemiological fronts”, which stem from the way the parasite
247 spreads to its neighboring hosts, killing them and leaving behind empty, bare patches that are later
248 recolonized by the host, albeit at a slower rate. This phenomenon of fronts may be difficult to quantify
249 formally, but can be easily visualized from model outputs. This can be done in chouca by setting the
250 `run_camodel()` function to display the results as the simulation is run. To do so, we adjust the control list
251 to include a function that displays the landscapes:

```
252 options <- list(custom_output_fun = landscape_plotter(mod),  
253 custom_output_every = 1)
```

254 then run the model on a 256×256 grid seeded with 10% of cells in the “parasitized” state:

```
255 initmm <- generate_initmat(mod, c(host = 0.9, parasitized = 0.1, empty = 0),  
256 nrow = 256, ncol = 256)  
257 output <- run_camodel(mod, initmm, times = seq(0, 1024), control = options)
```

258 The above lines of code will run the model and display the changing landscape, which allows
259 investigating the spreading patterns of the parasite. Once this visualization step is done (Figure 4, and
260 animated version in SM1), the options set to visualize the landscape can be removed to reduce simulation
261 times, for example to investigate the model behavior along a range of parameter values.



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Figure 4 Landscape patterns of the host-parasitized-empty model (respectively green, yellow and blue) as displayed on screen.

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Inference of local interactions from landscape-scale patterns

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Because SCA are defined on grids, a natural application is to compare their output to empirical raster data, such as remote-sensing images, to infer local-scale ecological interactions from landscape-wide spatial patterns. Arid systems provide a good illustration of this approach: in those systems, interactions between plants are often a balance of negative effects, through competition for nutrients or water, with positive effects, for example with an increased survival of seedlings below the canopy of taller plants (Valiente-Banuet & Ezcurra, 1991). These effects can be strong enough for new plants to mostly establish below existing plants, which results in their aggregation into patches, and has important consequences for the resilience of those systems to changes in aridity (Kéfi et al., 2007). The sizes and numbers of those patches can be readily quantified from remote-sensing images, and such patterns can be used to infer whether facilitation occurs between plants (Xu et al., 2015; Chen et al., 2022). This is traditionally done by summarizing the spatial structure into spatial statistics, such as spatial autocorrelation (Sankaran et al., 2017) or type of patch size distribution (Siteur et al., 2023; Pichon et al., 2024) and linking the observed changes in those metrics to theoretical results (Scanlon et al., 2007; Kéfi et al., 2024). However, such qualitative comparison logically results in a qualitative and corroborative result, *i.e.* is there or is there not facilitation between plants, rather than a more informative quantitative result, *i.e.* how strong facilitation is between plants. A quantitative inference must be based on an approach that links quantitatively some aspects of the spatial patterns to the strength of facilitation. This can be done by defining a model that links the strength of facilitation with the expected patterns, and finding model parameters that maximize agreement between model output and data. This is expensive computationally, but this limitation is alleviated by a fast SCA engine such as *chouca*, as we show below.

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We define here a model of an arid ecosystem with two states, “bare” and “vegetated”. A bare cell can become a vegetated cell with the probability p_{plant} . A vegetated cell can become bare (plants die) with the probability $d(1 - \beta q_{plant})$, where d is a constant mortality rate, that is reduced by the coefficient β . β captures the local effect of plants, either facilitation when they increase the survival of plants near existing vegetation ($\beta > 0$), or competition when survival is decreased ($\beta < 0$). This model is implemented in *chouca* using:

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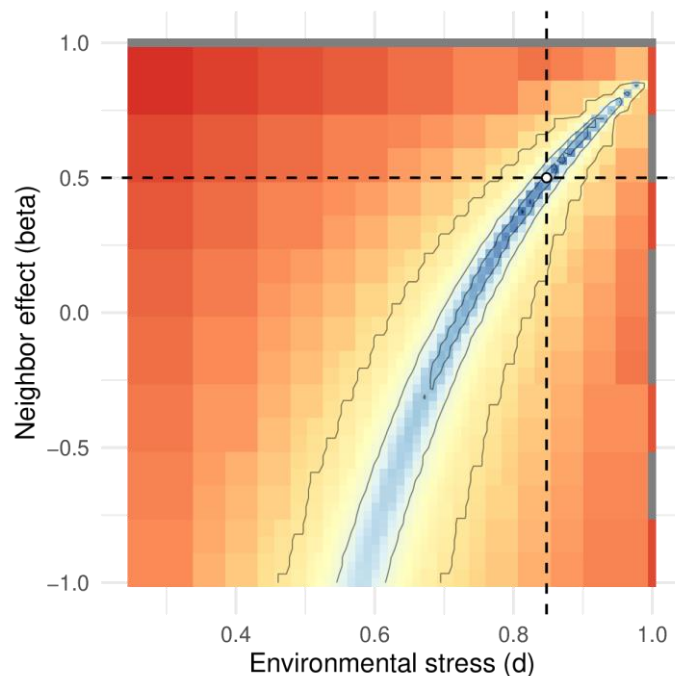
```
facilitation_mod <- camodel(  
  transition(from = "bare", to = "plant", ~ p["plant"] ),  
  transition(from = "plant", to = "bare", ~ d * ( 1 - beta * q["plant"] ) ),  
  parms = list(beta = 0.5, d = 0.85),  
  wrap = TRUE,  
  neighbors = 4
```

298)

299 We ran the model till equilibrium on a 1024×1024 grid, to simulate a landscape that would be obtained
300 from empirical data (e.g. a remote sensing image) using $(d, \beta) = (0.85, 0.5)$. From this landscape used as
301 observed data, we computed the distribution of pairs, which summarizes all the possible pairs of
302 neighboring cells present in the landscape $\mathbf{n}_{obs} = (n_{p,0}, n_{p,p}, n_{0,0})$. We then define the likelihood
303 $P(\mathbf{n}_{obs} | d, \beta)$ assuming these observed number of pairs follow a multinomial distribution of size N_p (the
304 sum of all pairs, a fixed number given the grid size and neighborhood type), and probabilities $\boldsymbol{\mu} =$
305 $(\mu_{p,0}, \mu_{p,p}, \mu_{0,0})$. $\boldsymbol{\mu}$ defines the relative probabilities of observing each type of pair in the grid, which
306 depend on the particular values of d and β , and can be estimated by simulating the landscape with these
307 parameter values. This estimate of the likelihood formally quantifies the agreement between model and
308 data, and allows exploring the parameter space in terms of d and β to find the most likely parameter
309 combinations given the observed patterns.

310 We find that this approach can recover the parameter values used for d and β , with only one global
311 maximum in the likelihood function (Figure 5; code provided in SM2), showing that the distribution of pairs
312 is a sufficient statistic to recover the model parameters. Applying such an approach to empirical data would
313 require further testing of model assumptions, for example, to investigate whether facilitation occurs on
314 the recruitment of new plants instead of on the mortality of adult plants – this can be done by simply
315 changing the model definition above. Because this approach is likelihood-based, model support can be
316 compared using traditional statistics, such as AIC, and a Bayesian approach can be used to estimate credible
317 intervals on parameter values, or use informative priors grounded in knowledge about the system (Hartig
318 et al., 2011).

319 Using this type of simulation-based inference is very expensive computationally, as this simple test
320 requires running around a thousand simulations. This would require long-running computations with usual
321 SCA implementations, but takes just under five minutes with chouca on a 2018 laptop computer (8 cores).
322 This way of calibrating models has seldom been used in spatial ecology – a fast SCA implementation is
323 essential to make it more accessible.



324

325 **Figure 5** Likelihood surface as a function of the two model parameters d and β . Blue values indicate
326 parameter combinations with high likelihood. The white dot and lines indicate the estimated
327 parameter values.

328

Conclusion

329 chouca is an easy-to-use package to model, simulate, and visualize SCA in a reproducible way, that
330 enables an interactive design and revision of models as well as novel methodological approaches. chouca
331 does not support all types of cellular automata, and does not replace more generic modelling frameworks
332 such as NetLogo (Wilensky, 1999) or simecol (Petzoldt & Rinke, 2007), but allows very efficient
333 simulation for the types of SCA it supports. The package focuses on processes defined at the level of a cell,
334 while a large set of SCA define processes at the level of pair of cells, a model representation other software
335 packages can provide (e.g. CellLab-CTS; Tucker et al., 2016). It is important to note that because chouca
336 splits the definition of an SCA model from its simulation phase, it may use different backends for simulation.
337 This opens the possibility of future improvements to already-implemented models without modification of
338 existing code.

339 Because of the relatively low bar of entry of SCA compared to other forms of spatial modelling, we
340 hope this work will contribute to make more accessible the testing of hypotheses linking spatial processes
341 to patterns in ecology, as well exploring system-level consequences of specific local management decisions.
342 chouca is available and maintained on CRAN, and welcomes comments, feedback and bug reports on its
343 home page at <https://github.com/alexgenin/chouca>.

344

Data, script and code availability

345 All code used for this work is freely-available at Zenodo under a CC-BY license (Génin, 2024), along with
346 chouca version v0.1.99 used at the time of writing.

347

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356

357

Conflict of interest disclosure

358 The authors declare they have no conflict of interest relating to the content of this article.

359

References

- 360 Bays C (2010) Introduction to Cellular Automata and Conway's Game of Life. In: *Game of Life Cellular*
361 *Automata* (ed Adamatzky A), pp. 1–7. Springer London, London. [https://doi.org/10.1007/978-1-84996-](https://doi.org/10.1007/978-1-84996-217-9_1)
362 [217-9_1](https://doi.org/10.1007/978-1-84996-217-9_1)
- 363 Chen BJW, Teng SN, Zheng G, Cui L, Li S, Staal A, Eitel JUH, Crowther TW, Berdugo M, Mo L, Ma H, Bialic-
364 Murphy L, Zohner CM, Maynard DS, Averill C, Zhang J, He Q, Evers JB, Anten NPR, Yizhaq H, Stavi I, Argaman
365 E, Basson U, Xu Z, Zhang M, Niu K, Liu Q, Xu C (2022) Inferring plant–plant interactions using remote
366 sensing. *Journal of Ecology*, **110**, 2268–2287. <https://doi.org/10.1111/1365-2745.13980>
- 367 Gardner M (1970) The fantastic combinations of John Conway's new solitaire game 'life.' *Scientific*
368 *American*, 120–123.
- 369 Génin A (2024) Scripts used in the publication "Easy, fast and reproducible Stochastic Cellular Automata
370 with chouca" Génin et al. 2024. <https://doi.org/10.5281/ZENODO.11567662>

- 371 Génin A, Majumder S, Sankaran S, Danet A, Guttal V, Schneider FD, Kéfi S (2018) Monitoring ecosystem
372 degradation using spatial data and the R package spatialwarnings. *Methods in Ecology and Evolution*, **9**,
373 2067–2075. <https://doi.org/10.1111/2041-210X.13058>
- 374 Génin A, Navarrete SA, Garcia-Mayor A, Wieters EA (2024) Emergent spatial patterns can indicate
375 upcoming regime shifts in a realistic model of coral community. *The American Naturalist*.
- 376 Grimm V, Frank K, Jeltsch F, Brandl R, Uchmański J, Wissel C (1996) Pattern-oriented modelling in
377 population ecology. *Science of The Total Environment*, **183**, 151–166. [https://doi.org/10.1016/0048-](https://doi.org/10.1016/0048-9697(95)04966-5)
378 [9697\(95\)04966-5](https://doi.org/10.1016/0048-9697(95)04966-5)
- 379 Guichard F, Halpin PM, Allison GW, Lubchenco J, Menge BA (2003) Mussel disturbance dynamics:
380 signatures of oceanographic forcing from local interactions. *The American Naturalist*, **161**, 889–904.
- 381 Hartig F, Calabrese JM, Reineking B, Wiegand T, Huth A (2011) Statistical inference for stochastic
382 simulation models - theory and application: Inference for stochastic simulation models. *Ecology Letters*,
383 **14**, 816–827. <https://doi.org/10.1111/j.1461-0248.2011.01640.x>
- 384 Heinonen T, Pukkala T (2007) The use of cellular automaton approach in forest planning. , **37**.
- 385 Iwasa Y (2000) Lattice models and pair approximation in ecology. In: *The geometry of ecological*
386 *interactions: simplifying spatial complexity* Cambridge Studies in Adaptive Dynamics., pp. 227–251.
387 Cambridge University Press.
- 388 Keeling MJ (2000) Evolutionary Dynamics in Spatial Host–Parasite Systems. In: *The Geometry of*
389 *Ecological Interactions* (eds Dieckmann U, Law R, Metz JAJ), pp. 271–291. Cambridge University Press.
390 <https://doi.org/10.1017/CBO9780511525537.018>
- 391 Kéfi S, Génin A, Garcia-Mayor A, Guirado E, Cabral JS, Berdugo M, Guerber J, Solé R, Maestre FT (2024)
392 Self-organization as a mechanism of resilience in dryland ecosystems. *Proceedings of the National Academy*
393 *of Sciences*, **121**, e2305153121. <https://doi.org/10.1073/pnas.2305153121>
- 394 Kéfi S, Rietkerk M, Alados CL, Pueyo Y, Papanastasis VP, ElAich A, de Ruiter PC (2007) Spatial vegetation
395 patterns and imminent desertification in Mediterranean arid ecosystems. *Nature*, **449**, 213–217.
396 <https://doi.org/10.1038/nature06111>
- 397 van de Koppel J, Gupta R, Vuik C (2011) Scaling-up spatially-explicit ecological models using graphics
398 processors. *Ecological Modelling*, **222**, 3011–3019. <https://doi.org/10.1016/j.ecolmodel.2011.06.004>
- 399 Lanzer ATS, Pillar VD (2002) Probabilistic cellular automaton: model and application to vegetation
400 dynamics. *Community Ecology*, **3**, 159–167. <https://doi.org/10.1556/ComEc.3.2002.2.3>
- 401 Lion S (2016) Moment equations in spatial evolutionary ecology. *Journal of Theoretical Biology*, **405**,
402 46–57. <https://doi.org/10.1016/j.jtbi.2015.10.014>
- 403 Louis P-Y, Nardi FR (Eds.) (2018) *Probabilistic Cellular Automata: Theory, Applications and Future*
404 *Perspectives*. Springer International Publishing, Cham. <https://doi.org/10.1007/978-3-319-65558-1>
- 405 Majumder S, Das A, Kushal A, Sankaran S, Guttal V (2021) Finite-size effects, demographic noise, and
406 ecosystem dynamics. *The European Physical Journal Special Topics*, **230**, 3389–3401.
407 <https://doi.org/10.1140/epjs/s11734-021-00184-z>
- 408 Matsuda H, Ogita N, Sasaki A, Sato K (1992) Statistical Mechanics of Population: The Lattice Lotka-
409 Volterra Model. *Progress of Theoretical Physics*, **88**, 1035–1049. <https://doi.org/10.1143/ptp/88.6.1035>
- 410 Mayor ÁG, Kéfi S, Bautista S, Rodríguez F, Cartení F, Rietkerk M (2013) Feedbacks between vegetation
411 pattern and resource loss dramatically decrease ecosystem resilience and restoration potential in a simple
412 dryland model. *Landscape Ecology*, **28**, 931–942. <https://doi.org/10.1007/s10980-013-9870-4>
- 413 Muthukrishnan R, Lloyd-Smith JO, Fong P (2016) Mechanisms of resilience: empirically quantified
414 positive feedbacks produce alternate stable states dynamics in a model of a tropical reef (B Silliman, Ed.).
415 *Journal of Ecology*, **104**, 1662–1672. <https://doi.org/10.1111/1365-2745.12631>
- 416 Pascual M, Roy M, Guichard F, Flierl G (2002) Cluster size distributions: signatures of self-organization
417 in spatial ecologies. *Philosophical Transactions of the Royal Society B: Biological Sciences*, **357**, 657–666.
418 <https://doi.org/10.1098/rstb.2001.0983>
- 419 Petzoldt T, Rinke K (2007) **simecol**: An Object-Oriented Framework for Ecological Modeling in R.
420 *Journal of Statistical Software*, **22**. <https://doi.org/10.18637/jss.v022.i09>
- 421 Pichon B, Donnet S, Gounand I, Kéfi S (2024) *Estimating distances to tipping points from dryland*
422 *ecosystem images*. Ecology. <https://doi.org/10.1101/2024.02.20.581244>
- 423 Sankaran S, Majumder S, Viswanathan A, Guttal V (2017) Patchiness and scale-free correlations:
424 characterising criticality in ecosystems. *bioRxiv*, 233429.

- 425 Scanlon TM, Caylor KK, Levin SA, Rodriguez-Iturbe I (2007) Positive feedbacks promote power-law
426 clustering of Kalahari vegetation. *Nature*, **449**, 209–212. <https://doi.org/10.1038/nature06060>
- 427 Schneider FD, Danet A, Génin A, Guttal V, Kéfi S, Majumder S, Sankaran S (2016) R-package caspr.
428 Siteur K, Liu Q-X, Rottschäfer V, van der Heide T, Rietkerk M, Doelman A, Boström C, van de Koppel J
429 (2023) Phase-separation physics underlies new theory for the resilience of patchy ecosystems. *Proceedings*
430 *of the National Academy of Sciences*, **120**, e2202683120. <https://doi.org/10.1073/pnas.2202683120>
- 431 Tucker GE, Hobbey DEJ, Hutton E, Gasparini NM, Istanbuluoglu E, Adams JM, Nudurupati SS (2016)
432 CellLab-CTS 2015: continuous-time stochastic cellular automaton modeling using Landlab. *Geoscientific*
433 *Model Development*, **9**, 823–839. <https://doi.org/10.5194/gmd-9-823-2016>
- 434 Valiente-Banuet A, Ezcurra E (1991) Shade as a Cause of the Association Between the Cactus
435 *Neobuxbaumia Tetetzo* and the Nurse Plant *Mimosa Luisana* in the Tehuacan Valley, Mexico. *The Journal*
436 *of Ecology*, **79**, 961. <https://doi.org/10.2307/2261091>
- 437 van Baalen, M. (2000) Pair approximations for different spatial geometries. In: *The geometry of*
438 *ecological interactions: simplifying spatial complexity* Cambridge Studies in Adaptive Dynamics., pp. 359–
439 387.
- 440 Wilensky U (1999) NetLogo.
- 441 Witten TA, Sander LM (1983) Diffusion-limited aggregation. *Physical Review B*, **27**, 5686–5697.
442 <https://doi.org/10.1103/PhysRevB.27.5686>
- 443 Wolfram S (1984) Cellular automata as models of complexity. *Nature*, **311**, 419–424.
- 444 Wootton JT (2001) Local interactions predict large-scale pattern in empirically derived cellular
445 automata. *Nature*, **413**, 841–844. <https://doi.org/10.1038/35101595>
- 446 Xu C, Holmgren M, Van Nes EH, Maestre FT, Soliveres S, Berdugo M, Kéfi S, Marquet PA, Abades S,
447 Scheffer M (2015) Can we infer plant facilitation from remote sensing? a test across global drylands.
448 *Ecological Applications*, **25**, 1456–1462. <https://doi.org/10.1890/14-2358.1>
- 449 Yassemi S, Dragicevic S, Schmidt M (2008) Design and implementation of an integrated GIS-based
450 cellular automata model to characterize forest fire behaviour. *Ecological Modelling*, 71–84.

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