**Title**

Low salinity significantly alters the bacterial community of the marine
foundation species *Saccharina latissima* (sugar kelp)

**Short title**

Salinity alters *Saccharina* bacteria

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As climate change progresses, the intensity and variability of freshwater outflow into the ocean is predicted to increase. The resulting increase in low salinity events will be a source of stress for *Saccharina latissima* (sugar kelp) and potentially *Saccharina's* associated bacterial community. Bacteria are important to the health of their host because host-associated bacteria can facilitate or hinder host survival in stressful abiotic conditions. Therefore, understanding how bacteria change under abiotic stress is critical to our understanding of how host physiology will be affected by abiotic stress. In this study, sampled the bacterial community associated with *Saccharina* and the surrounding environment across two years of field sampling at four sites with contrasting salinity profiles around Vancouver, Canada during the 2021 and 2022 spring freshet (562 samples) coupled with salinity manipulation experiments in the lab repeated eight times in 2022, concurrently with field sampling (269 samples). Our illumina data show that salinity is a significant factor in shaping the overall bacterial community of *Saccharina*. We find significant increases in bacterial community diversity and paired with a significant reduction in the relative abundance of core bacteria in low salinity across lab and field data. Our findings suggest that host-filtering is significantly impaired in low salinity, raising questions surrounding how resilient some existing

59    *Saccharina* populations may be to more intense and more variable

60    freshwater influx.

Kelps (*Laminariales*, brown algae) are marine foundation species that form dense underwater forests across the globe, providing food and habitat for other organisms [1]. Through ecosystem services and commercial harvest, kelps contribute an estimated $674 billion CAD per year to the global economy [2]. However, increased frequency and intensity of severe weather events, higher ocean temperatures, lower pH [3], increased coastal water turbidity [4], and lower ocean salinity [4–6] are associated with reductions in kelp recruitment, reproduction, and survival. Thus, kelps and the ecosystem services they provide are vulnerable to climate change [3, 4, 7].

There is a large body of evidence showing that low salinity is harmful to kelp. Field observations from Australia [5] and Norway [6] find up to 100% mortality of kelp sporophytes after freshwater floods. Lab studies on the effects of low salinity on kelps agree with field data. Research of the microscopic gametophyte stage finds lower spore settlement and gemination rates in low salinity [8]. For the macroscopic sporophyte stage, lower photosynthetic efficiency [9, 10], nitrogen uptake rates [11], and growth rates [12] are reported in low salinity treatments. As anthropogenic climate change progresses, lower ocean salinity is predicted by modeling studies to be one of the most intense [3, 4, 7] and least predictable abiotic stressors that kelps will experience [3].

One unknown is how changes in abiotic stressors, including

low salinity, will influence the kelp bacterial community, and

whether these changes exacerbate or ameliorate kelp stress

tolerance. The bacterial community of macroalgae can influence

growth and morphology [13, 14] and tolerance of abiotic conditions

[15].

Both free living and host-associated bacterial communities

change with abiotic conditions. Seasonal cycling can drive

predictable change in environmental [16, 17] and host-associated

[18] bacterial communities. Additionally, local abiotic conditions,

including salinity, can significantly shape the bacterial community

in the water column [19, 20] and on non-kelp algal hosts [21–23].

Temperature and temperature significantly alter the bacterial

community of the kelp *Ecklonia radiata* in the lab [24], but it is not

known how salinity influences the bacterial community of kelps.

Across host-microbe systems, changes in bacterial

community composition in response to abiotic stress are correlated

with negative, positive, or no effects on host condition. In some

marine hosts, including anemones [25], corals [26], and sponges

[27], a change in the bacterial community is associated with

increased host tolerance to abiotic stress. In other cases, including

in green algae (*Ulva sp.*) from localities with different salinity [23]

and transplanted seagrasses [28], changes in the bacterial

community composition are not associated with changes in host-

health. Finally, declines in host condition in response to stress have

been associated with two patterns: directional shifts in community

composition and a destabilization of a typically stable bacterial

community. In kelp, a change in the bacterial community

composition is correlated with decreased host health [24, 29]. For

the brown algae *Ectocarpus sp.* a single bacterial strain is

responsible for the algae's ability to grow in freshwater [15],

showing how destabilization of a small part of the bacterial

community can severely impact the host's ability to tolerate

particular abiotic conditions. Destabilisation has also been observed

systems such as corals and sponges, and is characterized by declines

in the core microbiome, increased alpha diversity, and increased

dispersion in community composition (Anna Karenina principle)

[30]. Presumably, destabilization is a result of the disruption of the

host filtering that typically maintains a stable, low diversity, core

community. However, in most cases, associations between host

condition and bacterial community changes are correlative.

Assigning causality in the link between host condition and change

in host-associated bacterial communities in response to stress is an

open challenge [27, 31, 32].

In this study, we asked how low salinity stress affects the

subcanopy kelp *Saccharina* ((L.) C.E. Lane, C. Mayes, Druehl, and

G. W. Saunders) and its bacterial community. We define stress as a

condition that adversely affects growth via damage and/or resource

allocation associated with damage prevention and cellular repair

[33, 34]. For our two-year field survey from April to July in 2021

and 2022, and the concurrent lab study in 2022, we predicted that (i)

salinity would shape the bacterial community of *Saccharina* as

observed in other algal systems [22, 23] and that (ii) alpha and beta-

diversity would increase in lower salinity, accompanied by a

decrease in relative abundance core bacteria, as observed in other

marine host systems experiencing abiotic stress [27, 30, 32].

Our extensive dataset shows that salinity has a small but

significant effect on the overall bacterial community composition

and that lower salinity significantly decreases the relative abundance

of the core bacterial community while increasing mean community

dissimilarity, suggesting that host-filtering by *Saccharina* is

impaired in low salinity.

**Methods**

*Field Site Description*

Five field sites near Vancouver, Canada were visited across two successive years. Both years, sites were sampled every two weeks at low tide, from mid-April until mid-July (2021) or early August (2022) during the time of the annual freshwater influx, due to snow melt in late spring, hereafter called the spring freshet.

In 2021, we collected data at Lighthouse Park (lhp; 49.329, -123.264), Starboat Cove Park (scp; 49.333; -123.261), Third Beach (tb; 49.302, -123.158); and Girl in a Wetsuit (gw; 49.304, -123.126; Fig. 1). In 2022, lhp was replaced by Cates Park (cp; 49.300, -122.958; Fig. 1). The West-Vancouver sites (lhp, scp, and cp) were always sampled on the same day and the Vancouver sites (tb and gw) were sampled the subsequent day. It was not possible to sample at all four sites on the same day due to the tide height required and travel time.

*Site-Conditions*

In 2021 and 2022 we recorded water temperature and salinity at the start of each site visit (Fig. 1). In 2021, all measurements were conducted with a YSI ProQuatro Multiparameter Metre. On two sequential samplings in 2022, salinity only measurements were taken with a refractometer because

the YSI was not available. In all cases, abiotic measurements were

repeated three times per site visit and the measurement instrument

was calibrated per manufacture instructions prior to each sampling.

*Saccharina field collection*

In 2021, six *Saccharina* individuals were collected

haphazardly at each site visit, including individuals that were out of

the water, in the wave splash zone, or fully submerged in the ocean

or tidepools. Given the strong influence of *Saccharina* position seen

in the 2021 microbial data (supplementary results; Fig. S3), in 2022,

only individuals that were submerged in the ocean were sampled

(six per site). At tb in 2022, six additional individuals were collected

from tidepools starting in May 2022 in case only tidepool

*Saccharina* were accessible at tb later during the summer and to

better understand the differences between tidepool and ocean

*Saccharina* bacterial communities (supplementary results; Fig. S4).

*Field bacterial data collection*

In both years two water samples, two rock swabs, and at least

six *Saccharina* meristem swabs (bottom 10 cm of the thallus) were

collected each site visit to capture changes in the bacterial

community throughout the spring freshet. Samples were taken

wearing gloves sprayed with 70% ethanol between samples.

We swabbed the meristem because it is the newest and most

selective tissue [35], presumably hosting the bacterial community

most indicative of *Saccharina.* We gently rinsed *Saccharina* and

rocks with 0.22 μm filter-sterilized seawater; then swabbed the

surface for 10 s with cotton tipped swab (VWR, CA10805-154). The

swab was then broken off into a cryovial (VWR, CA66021-993).

Water column bacterial samples were collected by pre-

filtering seawater through a 150 μm mesh before filtering the water

through a 0.22 μm membrane (MilliporeSigma, Sterivex™ Filter

Unit) until the filter clogged or 500 mL of water had been sampled.

The Sterivex$^{TM}$ was stored in a Whirl-Pak (VWR, 13500-390).

All samples were placed in a cooler with ice packs at -20°C

until returning to the lab (within three hours of sample collection),

where samples were stored at -70°C until extraction.

*Lab experiment protocol*

We performed eight experimental rounds manipulating

salinity through the 2022 field season, corresponding to each field

sampling event at gw. After collecting the field bacterial samples,

18 *Saccharina individuals* were collected (the holdfast, stipe, and

bottom 15 cm of the blade), numbered with flagging tape wrapped

around the stipe, and transported in a cooler filled with seawater

from gw (approximately 1 h transit).

*Saccharina* were cultured in seawater at 10°C on a 12h

light:12h dark photoperiod with bubblers to induce water motion.

Each culturing round included six 8 L aquaria (two per treatment),

with three *Saccharina* individuals per aquaria (Fig. 1), in one of

three salinity treatments: 10, 20, and unaltered sea water (full

strength hereafter). The full strength seawater is pumped from 30 m

depth in Burrard Inlet and brought to UBC by truck. The salinity

fluctuated between 31 and 32 depending on the culturing round

(Table S1).

Most experimental rounds lasted four days with samples

taken on day zero (field collection day), day one (24 h after

collection), and day four. Our first culturing round lasted six days;

on day six the *Saccharinas* in 10 salinity were green and fell apart

to the touch, so we shortened the culturing time to three days for

culturing round two. We did not observe significant *Saccharina*

damage on day three, so we extended the culturing duration to four

days for the remaining experimental rounds. For culturing round

five, there are no dry weight measurements on day four (Fig. 2)

because we extended the culturing time to fourteen days, to see if

the *Saccharinas* in the 10 salinity would fall apart in as they did in

round one. They did not.

*Lab bacterial collection protocol*

All lab bacterial sampling (Fig. 1) followed the protocol in

the field section with the following modifications. The substrate

samples were taken by swabbing the airline tubing of the bubbler

rather than rocks and the aquaria water was not pre-filtered before

taking bacterial samples. Note, the day zero *Saccharina* swabs are

the gw field samples. Day zero water and substrate samples were

taken prior to adding the *Saccharinas* to the aquaria. Lab *Saccharina*

samples (day one and day four) are swabs of all three *Saccharinas*

in the same aquaria, because we assumed the bacterial community

was homogenized. In total, we collected one of each sample type per

aquaria, per lab sampling day (Fig. 1).

*Lab Saccharina condition*

After taking bacterial samples, *Saccharina* condition was

assessed    on    an    individual    basis    (n=6    per    salinity

treatment/timepoint/experimental round) by quantifying blistering

occurrence, wet weight, estimated quantum yield, and ratio of dry

weight to wet weight on the last day of each experimental round.

Blistering occurrence was recorded as in other kelp studies

[5, 24]. Estimated quantum yield, measurements were taken with a

Junior-PAM Chlorophyll Fluorometer as described in the user

manual. Briefly, we calibrated the PAM and dark acclimated the

*Saccharinas* for five minutes on a damp cloth. Then, we placed the

optical fiber over the base of *Saccharina* (just above the stipe) and

took the measurement. Wet weight measurements were taken after

blotting dry the *Saccharina* before promptly returning the individual

to its respective aquaria.

Dry weights were obtained as follows. After taking wet

weight measurements on the last experimental day, stipes were cut

off from the *Saccharinas* [9]. Then, we weighed the *Saccharinas*

again, dried them in at 60°C for at least 24 h, and re-weighed them

[9, 36].

*DNA extraction and sequencing*

In 2021, all DNA was extracted with QUIAGEN DNeasy

PowerSoil Pro Kit 96-well plate; 47017 or single tube; 47014

following the manufacturer's protocol. In 2022, 36% of field swabs

(rock and *Saccharina*) samples were extracted with QIAGEN 96-

well kits and 64% with ZymoBIOMICS 96 MagBead DNA/RNA

Kit (Zymo, R2136). All water samples from the 2022 field season

and all lab samples in 2022 were extracted with the Zymo kit (see

metadata). The combination was a result of global supply chain

issues. For the extractions with the Zymo kit, the QIAGEN

PowerSoil Max beads (QIAGEN, 12988-10) were used for the

initial bead-beating step, and then the Zymo kit was used following

the manufacturers protocol. Each extraction plate included a blank

swab or Sterivex[TM] membrane (extraction blank, n=12).

Primers for PCR were the 515F (5'-

GTGYCAGCMGCCGCGGTAA-3') and 806R (5'-

GGACTACNVGGGTWTCTAAT-3'), which target the V4 region

of the 16S rRNA gene. Both primers have an Illumina adaptor and

golay barcode on the 5' [37, 38]. PCR reactions were as follows: 1

µL extracted DNA (for extracted samples) or water (for PCR blanks,

1 per PCR plate, n=12), 15 µL Phusion™ High-Fidelity DNA

Polymerase (Thermo Fisher, F530S), 2.4 µL of forward primer, 2.4

µL reverse primer (primers are from a 2.5 µM working stock

solution), to a final reaction volume of 30 µL with molecular water

(Thermo Fisher, SH30538FS). PCR reactions consisted of 30 s

initial denaturation at 98°C, 25 cycles of amplification (30 s at 98°C,

30 s at 55°C, and 20 s at 72°C), and a final elongation step at 72°C

for 10 min, which are the conditions recommended for the

Phusion[TM] polymerase. Samples where a band was not visible on a

DNA gel after two attempts with 25 cycles were amplified with 35

cycles (see metadata).

PCR reactions were cleaned with the QIAGEN QIAquick

PCR Purification Kit (28104 or 28181) and quantification was

performed with the Quan-IT PicoGreen assay kit (Thermo Fisher,

P7589), following the manufacturers protocol. Samples were pooled

to equal concentrations and sent for Bioanalyzer at the UBC

Sequencing Facility, Vancouver, British Columbia. In 2021, the

library was sent for Illumina MiSeq at the Hakai Institute, Campbell

River. In 2022, both libraries were sent for illumina MiSeq at the

UBC Sequencing Facility at the same time. In all cases, libraries

were constructed with the Illumina, MS-1023003 kit (MiSeq v3,

2x300).

*Sequence data processing*

Raw, demultiplexed reads were downloaded from the

Illumina hub and imported into RStudio (R version 4.2.2, RStudio

v2022.10.31; [39, 40]). Reads were processed following the dada2

pipeline (v1.24.0) [41]. Briefly, primers were removed, reads were

truncated to maintain high quality for downstream analysis with the

filterAndTrim function (275 bp for forward and 200 bp for reverse

in 2021 and 230 bp for forward and 175 bp for reverse in 2022),

error rates were calculated, paired forwards and reverse reads were

merged, a sequence table was constructed, and chimeric reads were

removed. The 2021 and 2022 sequence tables were merged

(mergeSequenceTables) and amplicon sequence variants (ASVs)

that differed by only the end base pair were collapsed

(collapseNoMismatch), prior to assigning taxonomy with SILVA

v138 [42] formatted for the dada2 pipeline [43].

The merged sequence table, the taxonomy table, and

metadata were grouped into a phyloseq object for filtering (v1.40.0)

[44]. Unassigned taxa and taxa assigned to chloroplasts,

mitochondria, or eukaryotes were removed in addition to sequences

assigned as *Pseudomonas*. *Pseudomonas* were present at high

relative abundance in multiple extraction blanks and was abundant

in nearly all samples extracted with the Zymo kit, but present in only

1 sample extracted with the Qiagen kits across both years; hence,

*Pseudomonas* is very likely a kit contaminant. Next, samples with

less than 1000 reads were removed; then, ASVs representing less

than 0.001% of total reads in the dataset were removed; counts in

the ASV table that were five or less were converted to zero (per

sample filtering, to minimize the effect of barcode switching), and

ASVs found in less than two samples were removed. Finally,

samples with more than 200,000 reads were removed (four samples)

and all samples from one 10 salinity aquaria in experimental round

five were removed (four samples total) because we noticed that there

was *Desmerestia viridis* lodged in the holdfast. *Desmerestia sp.*

produce sulfuric acid [45] and pH alters the bacterial community of

kelps [24]. In total, filtering retained 16,643,912 of 24,862,382 total

paired reads, 3,016 of an initial 29,105 unique ASVs, and 838

samples from the initial 1,095 samples (Table S2 and S3), with a

mean of 19,861 reads per sample.

We converted the data to an iNEXT (v3.0.0) [46] compatible

format with the metagMisc package (v0.0.4) [47] to perform

coverage-based rarefaction. The sample coverage was set to 0.8 and

iterated 1000 times.

Following filtering and coverage-based rarefaction, we

noticed a large difference in richness between samples extracted

with the Zymo kit compared to the Qiagen kit. We investigated to

see if this was due to the kits detecting a different suite of taxa. This

is not the case (see supplementary methods and results).

*Statistical analysis*

To compare blistering incidence between salinity treatments,

we used a Fisher's exact test with a Benjamini-Hochberg correction

with the package rstatix (v0.7.1) [48]

To compare means between treatment groups, we ran an

analysis of variance (ANOVA) followed by a Tukey post-hoc test.

In all cases, we tested the assumption of equal variance between

groups with Levene's test in the package car (v3.1-1) [49] and

validated the assumption of normality with qq-plots. When the

assumptions of equal variance was violated, we performed a Welsh

ANOVA with the function oneway.test with "var.equal = FALSE"

followed by a Games-Howell post hoc test in the package biostat

(v0.0.20) [50]. When the assumption of normality was violated, we

ran a Kruskal-Wallis test followed by a post-hoc Wilcoxon test with

a Benjamini-Hochberg correction [40].

We ran permutational analyses of variance (PERMANOVA)

on rarefied data with the adonis2 function in the package vegan

(v2.6-4) [51] with the distance metric set to Bray-Curtis. Marginal

PERMANOVAs were ran the same way, adding the by = "margin"

argument. In all cases, we tested for equal dispersion with the

betadispers function (equal unless stated otherwise) and ran a post-

hoc Pairwise Adonis test in the pairwise.adonis package(v0.4.1)

[52]

To calculate the Bray-Curtis dissimilarity index between

samples, we used the divergence function in the package

microbiome (v1.22.0) [53] as in [54]. Pearson's correlation tests

were performed using the cor.test() function and linear regression

models were performed with the lm() function and assumptions

were assessed graphically.

To identify the core *Saccharina* ASVs, we selected all field

samples collected when the ocean salinity was 20 or greater and ran

an indicator species analysis (IndVal) on non-rarefied data with the

function multipatt from the package indicspecies (v1.7.12) [55] with

999 permutations, comparing substrate types (*Saccharina*, water,

rock). The core ASVs are those assigned to *Saccharina* with an

IndVal statistic $\geq 0.7$ and a p-value $\leq 0.05$.

*Reanalysis of published data*

Raw data from King et al. (2023) were downloaded,

processed with the dada2 pipeline, and filtered as described above

and in their original publication. Core *Saccharina* ASVs were

identified with IndVal analysis and used a prevalence ≥ 0.8,

replicating methods in King et al. (2022)

## Results

*Lab salinity treatments differentially induced stress in Saccharina.*

We verified that lab salinity treatments differentially induced stress in *Saccharina* by measuring published stress-associated kelp phenotypes. We assessed blistering incidence on day four by Fisher's Exact Test followed by a pairwise Fisher post-hoc test, which shows that our lowest salinity treatment had significantly greater blistering incidence on day four than the other salinity treatments (Fig. 2A). Experimental round one lasted six days and experimental round two lasted three days. We did not observe any blistering on the last day of these rounds; however, on day six of round one, the *Saccharina* in the10 salinity were green and disintegrated when touched, indicating death.

Next, we compared the dry weight to wet weight ratio on day four by ANOVA (Fig. 2B) followed by a Tukey Post-Hoc test. The 10 salinity has a significantly lower dry weight to wet weight ratio compared to the other salinity treatments. Then, we compared the ratio of wet weight on day four compared to day zero (field collection day) by ANOVA. There are no significant differences between salinity treatments (Fig. 2C). These results paired with our dry weight data suggest that *Saccharina* is accumulating moisture but not additional biomass in low salinity. Finally, we analyzed photosynthetic efficiency found no significant difference between

salinities on day four by ANOVA (Fig. 2D). Hence, *Saccharina*

remains alive in our lab cultures for at least four days.

*Salinity treatments in the lab shapes the bacterial community.*

We assessed changes in bacterial community diversity and

composition in response to salinity treatments. We ran a

PERMANOVA to test if salinity alters the bacterial community for

each lab substrate type on day four (Table S4) and made NMDS

plots (Fig. 3, Fig. S5). The bacterial community composition is

significantly different between the full-strength and 10 salinity

treatments on day four for *Saccharina* and water (Fig. 3, Table S4),

but not for substrates (Fig. S5, Table S4) samples.

To compare variance in the bacterial community

dissimilarity between salinity treatments, we calculated the Bray-

Curtis dissimilarity index between samples from the same salinity

treatment and sample type on day 4 (Fig. 4, Fig. S6). Then, we

compared the mean dissimilarity between treatments and found

significant differences for *Saccharina* (20 more dissimilar that full-

strength) and water (full strength more dissimilar than 10 and 20)

samples (Fig. 4); substrate samples did not differ (Fig. S6). We also

compared the mean Shannon-Weiner diversity index between

salinity treatments found that by ANOVA, salinity does not

significantly alter alpha diversity in the lab for any sample type (Fig.

4, Fig. S6). These data show that *Saccharina* bacteria are more

dissimilar between individuals, but not more diverse, in lower

salinity treatments.

*Factors shaping the bacterial community of Saccharina in the field.*

We examined how salinity shapes bacterial community

diversity and composition the field in comparison to temperature

and Julian Day. Note that salinity is significantly negatively

correlated with Julian Day and water temperature, while water

temperature and Julian Day are positively correlated (Fig. S1). We

ran marginal PERMANOVAs including water salinity, temperature,

and Julian day for each field substrate type to assess the unique

explanatory power provided by each variable (Table 1). Julian day

explains the most variation in bacterial community composition,

followed by temperature, then salinity for all substrate types.

Consistent with observations in the lab, salinity does not

significantly shape the bacterial community of rocks (Fig. S5, Table

S5). Salinity is significant for *Saccharina* and water samples (Fig.

3, Table 1, S5), although the explanatory power of salinity is

relatively low.

*Community dissimilarity increases in low ocean salinity, but not*

*diversity.*

We asked whether community dissimilarity and diversity

change as salinity changes. We calculated the Bray-Curtis

dissimilarity index for each field sample, comparing the same

sample type within the same site and sampling visit (salinity, temp,

and Julian day were the same). We also calculated the Shannon's

diversity index for each field sample. In both cases, linear regression

models tests show that for *Saccharina* samples, Bray-Curtis

dissimilarity increases as ocean salinity decreases. We find no

significant relationship between salinity and Shannon's diversity

index (Fig. 4, Table S6). We found no significant correlation

between diversity metrics and salinity for water (Fig. 4, Table S6)

or rock samples (Fig. S6, Table S6).

*The relative abundance of the Saccharina core bacteria negatively*

*correlated with salinity.*

We then identified the core bacterial community of

*Saccharina* by conducting an IndVal analysis comparing

*Saccharina* samples to water and rock samples from all sampling

events where ocean salinity was greater or equal to 20. We used an

IndVal statistic threshold of 0.7, which required core ASVs to be a

>50% prevalence and enriched on *Saccharina* compared to water

and rocks. We identified six core ASVs: *Persicirhabdus sp.* (ASV1),

*Litorimonas sp.* (ASV3), *Maribacter sp.* (ASV4), *Cocleimonas sp.*

(ASV5), *Robignitomaculum sp.* (ASV7), and and unidentified

G*ammaproteobacteria* (ASV15; Fig. 5, Table S7). We ran

Pearson's correlation tests on the relative abundance of all core

ASVs combined and each core ASV alone. The total core relative

abundance, *Cocleimonas sp.*, and *Maribacter sp.* are significantly

correlated with salinity in the field (Fig. 5).

We then asked if these changes are observed in the more

controlled environment of the lab. We again find that in lower

salinity, the relative abundance of the core overall is significantly

lower and find that the relative abundance of different ASVs

(*Litorimonas sp.* and *Robiginitomaculum sp.*) is significantly lower

in low salinity (Fig. 5). Hence, the relative abundance of the core

overall consistently decreases as salinity decreases, but core ASVs

driving this pattern are inconsistent between lab and field.

The decline in core relative abundance could be a result of a

different core bacterial community colonizing or blooming in low

salinity. We tested for a low salinity core as above using samples

ocean salinity was less than 20. There were no new core ASVs

specific to low salinity. In the lab, we observed an increase in the

relative abundance of Alteromonadales at low salinity compared to

higher salinity in the lab (Fig. S10), but which ASVs were

enriched differed across sampling rounds was not consistent.

*Core bacteria are conserved across very large spatial and temporal*

*scales.*

To compare our *Saccharina* core to data in published in the

literature, we downloaded reads from a study performed in the

United Kingdom from September and August 2015 [56]. We then

performed another IndVal using a prevalence of ≥0.8 on *Saccharina*

as a threshold to identify the *Saccharina* indicator ASV from the

King *et al.* (2022) study. Then, we compared these ASV sequences

to our own and found that our *Maribacter* (ASV4) and unidentified

*Gammaproteobacteria* (ASV15) are perfect sequence matches to a

ASVs identified as *Saccharina* indicator in their data. Our other core

ASVs are not perfect matches to theirs, but all genera that are

represented in our core are present in the King *et al. Saccharina*

indicator taxa. These findings suggest that depending on the filtering

threshold, *Saccharina* may have a global core bacterial ASV

community.

**Discussion**

The predicted [3, 4] and ongoing die back of kelp forests [3, 5, 57] due to low salinity represents a global concern because of the environmental, economic, and cultural value that these foundation species provide [2, 3]. Our research shows that salinity significantly alters the bacterial community of *Saccharina,* both in the lab and in the field.

*Saccharina* hosts a core bacterial community, which is stable across large spatial and temporal scales (King *et al.*, 2023 and this study). We show that in low salinity, the relative abundance of core bacteria significantly decreases in conjunction with a significant increase in bacterial community dissimilarity. Together, these findings suggest that more opportunistic colonizers associate with *Saccharina* in lower salinity conditions in the field, pointing to decreased host-filtering by *Saccharina* in low salinity [58]. Overall, our results shows that although *Saccharina* is persisting in the field and remains photosynthetically active throughout the lab experimental period, *Saccharina* is experiencing stress in low salinity.

We show that salinity is a relatively weak but significant factor in shaping the bacterial community of *Saccharina*, which agrees with other studies investigating the effect of salinity on the bacterial community of other algae [21, 23]. Salinity explains more

variation in the bacterial community of *Saccharina* in the lab (9.8%

total variation) compared to the field (0.66% unique variation, 1.7%

total variation). In the field, Julian day (day of the year; 3.48% of

unique variation, 7.16% of total explained) is much better than both

salinity and temperature at explaining the variation in the bacterial

community of *Saccharina*. This is likely due to combined effects of

many unmeasured variables corelated with Julian day including:

dissolved organic carbon, pH [16], irradiance, and nitrogen

availability [59]. The large discrepancy in the explanatory power of

salinity between lab and field highlights the importance of pairing

field observations, which provide a more holistic view of the system,

with lab experiments, which permit the manipulation of a single

variable (salinity) without confounding effects of other co-variates

present in the field.

The variation in the bacterial community observed herein is

driven by a significant decrease in relative abundance of core ASVs,

paired with a significant increase in community dissimilarity in low

salinity, but no significant difference in diversity, although the

trends we observe suggest minor increases in diversity as salinity

decreases. This suggests that there are not more bacterial ASVs on

*Saccharina* in low salinity (alpha-diversity), but that the variation

between *Saccharina* individuals increases in low salinity (beta-

diversity). Overall, our data fit the Anna Karenina Principle, which

posits increased alpha and beta-diversity in a "disrupted microbiota"

[30], which has been observed in corals [32] and sponges [27].

However, stating that the beta-diversity change we observed

indicates a disrupted microbiota would be a narrow view [60]. The

word disruption itself has a negative connotation [61], which is not

always biologically supported. A disrupted microbiota is often

associated with unhealthy hosts [62], and is proposed to make the

algal host more susceptible to disease [31]. By disruption, what is

often meant, is a different microbiota under different conditions. In

fact, a variable bacterial community can be beneficial, including

increasing host-tolerance to various abiotic conditions [25–27, 63].

Our study is not designed to test if the changes we observed in the

bacterial community of *Saccharina* are beneficial or not to

*Saccharina*'s ability to tolerate low salinity.

Our two-year field study and manipulative lab experiment,

along with existing research showing that increased water

temperatures and decreased pH significantly alter the bacterial

community of kelp in the lab [24], and climate models predicting

that the intensity and frequency of abiotic stressors will increase

with climate change [3], suggests that the bacterial community of

kelps will be affected by climate change. We propose that in

stressful abiotic conditions, kelps will exhibit decreased host-

filtering of their bacterial community. If kelps do not filter their

bacteria as effectively in the future, it is unclear how kelp health and

resilience to abiotic stress will be affected in natural kelp forests and

in mariculture operations. In particular, the significant decrease in

core bacteria, which often perform functions that are beneficial for

the host [27], highlights the importance of considering the

microbiota [64] when predicting host performance and overall

biodiversity under climate change.

**652 Data availability**

Raw reads for this study are available on the European Nucleotide

Database (ENA) under project PRJEB60884. Code and metadata

are available on Borealis

https://borealisdata.ca/privateurl.xhtml?token=96717e7c-3f4d-

4c3b-8301-d91bf887606d.

The data King et al. (2023) were downloaded from ENA

(PRJEB50679) and

https://doi.org/10.6084/m9.fgshare.19453889.v1.

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## Contributions

Siobhan Schenk: conceptualization (equal), data collection (equal), writing (equal), editing (equal), data analysis (equal). Connor G. Wardrop: data collection (equal), editing (equal). Laura W. Parfrey: conceptualization (equal), data collection (equal), writing (equal), editing (equal), data analysis (equal).

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 **Main text figure and table legends**

 **Figure 1**. A) Map of the sampling sites with black arrows

 representing the path of main freshwater sources in the area, B)

 conductivity, and C) water temperature profiles across sampling

 years. The error bars are +/- one standard deviation around the mean.

 D) Representative schematic of 2022 lab experiment. Note, in C,

 missing temperature data from 2022 are from days where a

 refractometer, rather than the YSI were used to measure water

 parameters. Also note, in D, the n = 1 for *Saccharina* because all

 individuals in the same aquaria are swabbed with the same cotton

 swab. Plots showing the correlation between salinity, temperature,

 and Julian Day are in figure S1. Full temperature profiles (Fig. S2)

 show the same patterns across sites and time.

 **Figure 2.** *Saccharina* stress phenotypes by lab salinity treatment. A)

 The presence (orange) or absence (grey) of blistering by lab culture

 day, experimental round, and salinity treatment by individual.

 Fisher's Exact Test (p < 0.001) followed by a Pairwise Fisher Test

 of the day four data, show that the 10 salinity has significantly more

 blistering than the 20 and full-strength salinity on day four (p ≤

 0.0357). B) The ratio of dry weight to wet weight on lab day four.

 C) The ratio of wet-weight on lab day four to wet-weight on day

zero. D) The estimated effective quantum yield on lab day four by

salinity. For B, C, and D, ANOVA output comparing the different

salinity treatments are indicated in the corresponding panel. Note,

round one and two have no day four observations for all panels. In

addition, in panel B only, there are no day four observations for

culture round five. Day one comparisons for panels A, C, and D are

in the supplementary results.

**Figure 3.** NMDS plot of A) *Saccharina* and B) water samples on

day four colored by salinity treatment along with C) *Saccharina* and

D) water samples from the field. Note, corresponding

PERMANOVA outputs are in Table 1, S4, and S5.

**Figure 4.** The per sample Bray-Curtis dissimilarity index (beta-

diversity) of lab (A,B) and field (E,F), *Saccharina* and water

samples. Beta-diversity is calculated within the same substrate type

either by comparing samples within a salinity treatment (lab) or the

same sampling site visit (field). Note, the full output of the linear

regression models containing main effects and all possible

interactions for salinity, water temperature, and Julian Day (E-F) are

in Table S6. The t-values listed here are for the main effect of

salinity in the linear regression model. The Shannon-Weiner

diversity index of lab (C,D) or field (G,H) *Saccharina* and water

samples. For lab samples (A to D), p-values are indicated in if the

salinity treatments are significantly different by ANOVA (Tukey

post-hoc test) or Welsh ANOVA (Games-Howell post-hoc test).

**Figure 5.** Relative abundance of core ASVs in the field (A, B) and

in the lab on day four (C, D) on *Saccharina*. The output of the

Pearson correlations test assessing the relative abundance of core

ASVs as a percent of total reads by ocean salinity are indicated in

the corresponding panels (A, B). For lab day four samples, the

output comparing the mean relative abundance of core ASVs as a

percent of total reads by salinity treatment is also indicated in the

facets (C, D). We conducted either an ANOVA (F) or Kruskal-

Wallis rank sum test (X) if the assumptions of the ANOVA were

violated. When required, we conducted a Tukey post-hoc test for

ANOVAs or post-hoc Wilcoxon rank sum test with a Benjamini-

Hochberg correction for Kruskal-Wallis rank sum tests. Note, the

different y-axes between the field and the lab. Also note, the mean

relative abundance core ASVs on day zero (field collection day) is

$0.5114 \pm 0.2493$ (sd), which is not significantly different from any

salinity treatment on day four ($X_3^2 = 410.48$, p = 0.5672).

**Table 1.** Output of marginal PERMANOVA marginal showing the

unique explanatory power marginal explanatory of salinity, water

temperature, and Julian day for field samples. Output of

PERMANOVA showing the total explanatory power of each

variable are in Table S5.

| Substrate | Variables in model | F | $R^2$ | p |
|---|---|---|---|---|
| *Saccharina* (n=269) | Salinity | 2.213 | 0.007 | 0.012 |
| | Temperature | 2.437 | 0.008 | 0.005 |
| | Julian Day | 10.289 | 0.035 | 0.001 |
| **Water** (n=91) | Salinity | 3.384 | 0.031 | 0.001 |
| | Temperature | 2.952 | 0.027 | 0.001 |
| | Julian Day | 4.094 | 0.037 | 0.001 |
| **Rock** (n=101) | Salinity | 0.833 | 0.008 | 0.787 |
| | Temperature | 1.505 | 0.014 | 0.020 |
| | Julian Day | 1.499 | 0.014 | 0.001 |