Title: Recurrent drought increases grassland community seasonal synchrony

Running head: Recurrent drought increases synchrony

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Abstract

Climate change leads to an increase in the frequency and severity of drought events with strong repercussions on grassland ecosystems. While the effects of single drought events on ecosystem structure and functioning are well understood, it is largely unknown whether and how drought frequency modifies ecosystem responses to drought. Here, we assessed how the increase in frequency of severe, annual summer drought affected grassland communities. We examined these effects in a species-rich sub-alpine mountain meadow with a drought frequency of one, three, and 13 years, as well as ambient conditions. We found that high drought frequency increased seasonal plant community synchrony through a reduction in species richness, a shift of plant functional groups, a loss of early-seasonal plant species, and the constrained establishment of new species and individuals throughout the growing season. These changes are associated with a decreased fraction of biomass as drought frequency increased. We conclude that with the ongoing increase in drought frequency due to climate change, grassland plant communities will increase in seasonal synchrony. We suggest that this will leave ecosystems highly vulnerable to future disturbances, as asynchrony is a critical component of stability. Furthermore, we show that negative drought effects were enhanced with an increasing drought frequency. We therefore conclude that single and low-frequency drought studies may not adequately predict longer-term changes in our rapidly shifting climate. Moreover, we found that negative drought

effects on plant communities outweighed the weak adaptive effects of species. We therefore conclude that plant communities are unlikely to be able to adapt to recurrent drought.

Keywords: recurrent drought, drought frequency, seasonal synchrony, synchrony, asynchrony, grassland, species richness, early-seasonal species, functional groups, drought

1. Introduction

Grasslands cover 40 % of the Earth's surface and are crucial for carbon and nutrient cycling, biodiversity conservation, and ecosystem services (Petermann & Buzhdygan 2021; Blair *et al.* 2014; Dixon *et al.* 2014). However, up to 49 % of the world's grasslands are currently degraded, with the climate crisis predicted to exacerbate this number (Bardgett *et al.* 2021; Gibbs & Salmon 2015). Among the detrimental climate scenarios, the increase in frequency and severity of drought events stand out as particularly damaging (IPCC 2023, 2021; Wang *et al.* 2021; Spinoni *et al.* 2018). Drought events alter grassland nutrient cycling, disrupt interactions between species, result in species extinction, and therewith cause severe disruptions in ecosystem structure and function (Smith *et al.* 2024; Oram *et al.* 2023; Bardgett & Caruso 2020; Ingrisch *et al.* 2020; Tello-García *et al.* 2020). In the long-term, the predicted increase in the frequency of drought events therefore threatens to irrevocably change grassland ecosystems, including the fundamental services and biodiversity they conserve (Müller & Bahn 2022; Xu *et al.* 2021; Schirpke *et al.* 2019). At the same time, potential adaptations of species and communities may dampen these detrimental effects to some extent (Müller & Bahn 2022; Canarini *et al.* 2021; Coleman & Wernberg 2020). This occurs for example through drought-induced legacy effects that alter ecosystem structure and function after drought has subsided, and which results in plant communities being able to cope better with a subsequent drought event (Müller & Bahn 2022). Yet, while effects of single drought events are comparatively well studied, our understanding of the effects of a long-term increase in drought frequency on grassland community structure and function remains limited.

Multiple interlinked drought-induced changes, for example in plant species composition, diversity, complementarity, as well as species loss may have detrimental effects by affecting the **synchrony** of grassland communities (He *et al.* 2022; Muraina *et al.* 2021; de Mazancourt *et al.* 2013). Single drought was found to increase the synchrony of plant communities between years (Zhang *et al.* 2019), similarly as recurrent drought of three

(He *et al.* 2022) and four years (Muraina *et al.* 2021). With greater synchrony between years, species communities exhibit increased similarity in their response to environmental changes, the swiftness of their responses, or their functional complementarity throughout various seasons (Loreau & de Mazancourt 2013). These changes in grassland community synchrony likely have detrimental effects on long-term ecological functioning, especially under a future climate with a high drought frequency. Increases in grassland community synchrony between years has been shown to decrease community stability (Wilcox *et al.* 2017; Xu *et al.* 2015; Loreau & de Mazancourt 2013). This is because asynchronous communities stabilise ecosystems through different responses to disturbances (de Mazancourt *et al.* 2013). For drought disturbances, species asynchrony was suggested to be the dominant driver to stabilise community productivity under drought (Luo *et al.* 2023b; Muraina *et al.* 2021; Craven *et al.* 2018). Community synchrony between years is thus influenced by drought and is suggested to shape a communities' response to subsequent droughts by decreasing its stability. However, besides changing community synchrony *between* seasons, drought is also likely to affect community synchrony *within* the growing season. Therewith, it might be a critical mechanism in shaping responses to recurrent drought. However, we currently lack evidence on the effects of recurrent drought on community *seasonal* synchrony. In the presence of more frequent droughts, we hypothesize that the plant community will become more synchronous within the season, which also leads to reduced stability in the face of disturbances.

Drought events alter plant community composition, phenology, and productivity (Smith *et al.* 2023; Müller & Bahn 2022). Drought leads to a decrease in plant community **productivity**, by limiting plant growth through reduced water availability, or by mortality of species under severe drought conditions (Sippel *et al.* 2018; Frank *et al.* 2015). Additionally, after a drought event, soil rewetting results in a temporarily increased availability of nitrogen (Ingrisch *et al.* 2018; Karlowsky *et al.* 2018). This surge in nitrogen supports the recovery and, in some cases, leads to a short-term overshoot in productivity (Griffin-Nolan *et al.* 2018;

Hofer *et al.* 2016), which can compensate for the lower plant community productivity during the drought event (Hahn *et al.* 2021). This response is mainly based on fast-growing grass species, which makes the community sensitive to changes in composition post drought (Mackie *et al.* 2019; Stampfli *et al.* 2018). An important issue is that drought-induced changes in productivity and composition in grasslands may take a long time to come back to pre-disturbance level (Müller & Bahn 2022). If a subsequent drought falls within this recovery timespan of a previous drought, it may lead to constantly impaired ecosystems (Schwalm *et al.* 2017), and consequently to a more profound effect on productivity under a high frequency of drought (Xu *et al.* 2021). Hence, a potential accumulation of effects can emerge when multiple droughts occur annually. However, as recurrent drought events are typically studied in two-year experiments and only rarely extend beyond this timeframe, we lack critical knowledge on the productivity loss that builds up under a high drought frequency.

Drought shifts plant community **composition** towards more slower growing and more drought adapted-species (Wilcox *et al.* 2021). For example, drought may lead to a dominance of grasses over forbs (Xu *et al.* 2021; DeBoeck *et al.* 2018; Hoover *et al.* 2014). This may be driven by a higher post-drought resource use of grasses compared with the conservative resource use of forbs (Stampfli *et al.* 2018) or by a more efficient resource acquisition of water of grasses compared with forbs (Tello-García *et al.* 2020). Furthermore, drought-induced changes in plant community composition are driven by a reduction in plant species performance and an increase in mortality of plants both during and after a drought event (Sippel *et al.* 2018; Frank *et al.* 2015). In the long term, plant compositional shifts due to high drought frequency are likely to relate additionally to drought-induced changes in plant reproductive effects, e.g. reduction in reproductive shoots, seeds (Zeiter *et al.* 2016), belowground bud density (Qian *et al.* 2022), as well as altered seed banks (Basto *et al.* 2018), and regeneration from seed (Stampfli & Zeiter 2004). Through these drought-induced changes in plant species composition, recurrent drought events may in the long-term result in permanent loss of plant species with potentially irreversible consequences for the

113 functioning of plant communities. Furthermore, under recurrent drought, plant species also
114 have the capacity to adapt (Müller & Bahn 2022), which dampens drought effects under a
115 higher drought frequency (Backhaus *et al.* 2014; Walter *et al.* 2011). Therefore, a high
116 drought frequency is expected to result in plant communities consisting of drought-tolerant
117 species, a dominance of grasses, plant species profiting from the recovery dynamics of
118 drought events, as well as plant species that adapt to recurrent drought conditions. However,
119 due to the lack of studies investigating high drought frequency, long-term compositional
120 shifts might have been overlooked to date.

121 Drought alters **phenology** by shifting the beginning or end of the growing season
122 (Yuan *et al.* 2023; Ge *et al.* 2022; Rihan *et al.* 2022; Wang *et al.* 2022; Ji *et al.* 2021; Zeng *et*
123 *al.* 2021; Yuan *et al.* 2020; Berwaers *et al.* 2019; Kang *et al.* 2018) and advancing end-of-
124 season leaf senescence (Hoover *et al.* 2021; Berwaers *et al.* 2019; Peng *et al.* 2019; Kang
125 *et al.* 2018). Furthermore, drought alters species flowering phenology by advancing the
126 flowering date and extending the flowering length, whereby effects differ strongly between
127 species (Castillioni *et al.* 2022; Vorkauf *et al.* 2021; Forrest & Miller-Rushing 2010; Jentsch
128 *et al.* 2009). Hence, recurrent drought is expected to move the duration and timing of the
129 growing season and flowering with effects on community composition, though the directions
130 and magnitudes of these changes under recurrent drought are still poorly understood.

131 We lack a profound understanding of the implications under a high frequency of drought
132 events on the temporal shifts in community composition, productivity, and phenology. Here,
133 we assessed how the increase in frequency of severe summer drought affected grassland
134 community structure, function, and seasonal dynamics. We examined this in a species-rich
135 sub-alpine mountain meadow with increasing annual summer drought frequencies of up to
136 13 years. We quantified species abundances in the field throughout the growing season and
137 sampled plant species-specific biomass and necromass at peak drought and drought
138 recovery. We asked the following questions: How did an increase in drought frequency affect
139 (1) plant community productivity, species richness, and flower production during peak

drought and recovery, respectively, as well as (2) plant community seasonal synchrony. We examined whether these patterns were underlain by shifts in (3) plant community composition, (4) plant functional groups, and (5) plant community traits. Taken together, we present the effects of a high frequency of drought events on plant communities and discuss underlying mechanisms.

2. Material and Methods

2.1 Study Site

The studied sub-alpine grassland is located in the Austrian Central Alps in the Stubai valley at 1820 m a.s.l. (47°70 45″N, 11°180 20″E). The site is exposed southeast with an inclination of ~ 20° and has an average annual temperature of 3 °C and an annual precipitation of 1100 mm. The soil type is a dystric cambisol (top soil pH = 5.5). The vegetation community is classified as *Trisetetum flavescentis*, species-rich with on average 49 species per m^2 and consists largely of perennial grasses and forbs. The growing season lasts from the end of April until mid-October. The sub-alpine grassland is manured every 2-3 years, mown once in early August at peak biomass, and sporadically grazed in spring and autumn. The studied area, however, has not been fertilised and is excluded from grazing since 2007. For further details on the site see Ingrisch *et al.* (2018), Hasibeder *et al.* (2015), and Schmitt *et al.* (2010).

2.2 Experimental set-up

Four summer drought treatments were established in the enclosed area at the study site: a single year summer drought, three years of recurrent summer drought, 13 years of recurrent summer drought, and ambient conditions (n = 4, respectively). The treatment of 13 years of recurrent drought was established in 2008, the three-year recurrent drought in 2018, and the single-year drought in 2020. Summer drought was simulated for seven weeks from 17 June until 7 August 2020 using rainout shelters (Fig 1). This is the warmest period in the growing season and therefore the period where a natural drought is most likely to occur. The timing and intensity of the annual drought were kept as similar as possible in previous years since 2008. The ambient treatment was not covered with rainout shelters and thus exposed to ambient precipitation. The rainout shelters of the single and three years of recurrent summer drought had a base area of 3 x 3.5 m^2 and 2.5 m height. For the rainout shelters of the 13 years of recurrent drought, two plots are combined, respectively, and as a result, the shelters had a base area of 3 x 7 m^2. All shelters were covered with a transparent UV-A and UV-B

transmissive plastic foil (Lumisol clear AF, Folitec, Westerburg, Germany, light transmittance ca. 90 %). To allow air circulation, the shelters were open at the bottom (up to 0.5 m) and at the top of the face sides. Within each of the treatments, one permanent 20 cm x 20 cm plot was installed four weeks before summer drought simulation (n = 4; in total 16). Each plot was located at least 50 cm away from the rainout shelter to avoid edge effects. The drought period was ended by removing the shelters and watering of all 20 x 20 cm^2 plots with 31 mm of distilled water. At this time point, the entire site was also mown, and the cuttings were removed.

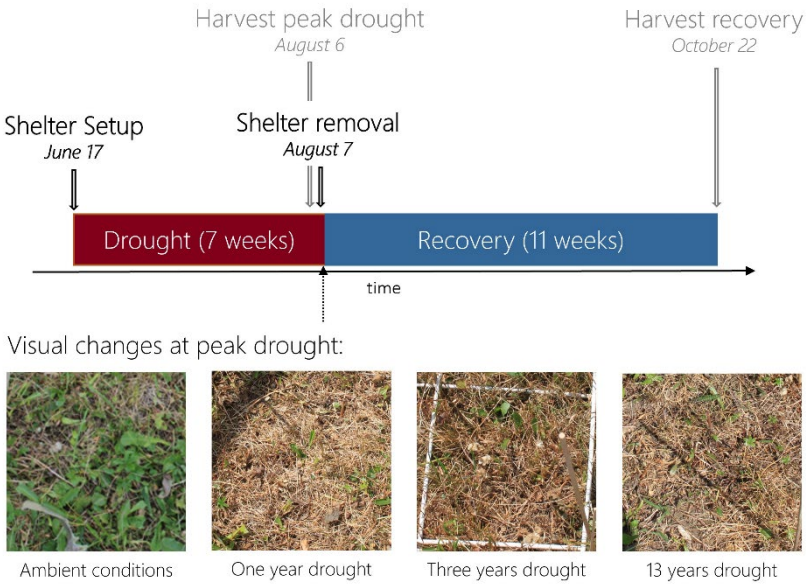


Fig 1 The course of the experiment including setup and removal of the rainout shelters, and peak drought and recovery harvest. The pictures show the drought intensity at peak drought of ambient conditions, one year, three years, and 13 years of recurrent drought. One replicate per treatment is shown.

2.3 Measurements

2.3.1 Microclimate

A microclimate station at the study site recorded precipitation, air temperature, and relative air humidity (see details in Hasibeder *et al.* (2015)). During the whole season, soil water

content (SWC) (Decagon EC-5 and 5TM, Decagon Devices, Pullman, WA, USA; SM300, Delta-T Devices, Cambridge, UK; Hobo S-SMD-M005, Onset Computer corporation, Bourne, MA, USA) was measured continuously in the main rooting horizon (30-min interval) at 5, 10, and 20 cm of soil depth. The measurements were replicated twice in the single-, and three-years treatment, and three to four times in the ambient treatment. Due to partial sensor failures, no data is available for the 13-years treatment. For results on the microclimate and SWC, see Fig S1.

2.3.2 Measurements throughout the growing season

In all 16 plots, species composition and vegetation height were measured every two weeks from 22 May to 22 October 2020. Plant species composition was estimated by noting for each species present the number of modules and flowering stems. Modules refer to the number of tillers for grasses and graminoids, the number of rosettes for small dicots, and the number of leaves for larger dicots per rooted species, following (in 't Zandt *et al.* 2021; Herben *et al.* 2013; Herben *et al.* 1997). The data was used to calculate the total number of modules, plant species number, community evenness, and Shannon-diversity index throughout the growing season. Counting was accompanied by measurements of mean vegetation height (in cm). Seedling composition in the plots was counted separately every three weeks between the 23 June and the 22 October 2020. A seedling was defined as a plant individual with less than two leaves, and was not part of the module counts. Seedlings were counted per plant functional group: forbs, grasses, and legumes.

2.3.3 Biomass, necromass, and leaf traits

Aboveground biomass (living plant material) and necromass (dead plant material) was sampled destructively in the plots at peak drought on 6 August 2020 and during recovery on 22 October 2020 (Fig 1). For this, plant material was cut at 2 cm above the soil and frozen at -18 °C until further analysis. Samples were split into plant species and separated into leaves, stems, and flowers. This was done for biomass and necromass, respectively. For each species of each plot, the leaf area was determined from thawed leaves that were saturated

with water and scanned (V700 Photo, Epson, WinRHIZO Pro 2012, Regent Instruments). Scanned leaves were dried in the subsequent three days at 60 °C and were weighed. Specific leaf area (SLA) was calculated by dividing the scanned leaf area by the dry weight of the leaves for each species and treatment, respectively. Leaf area index (LAI) was calculated from the leaf surface of all leaves divided by the total unit of ground area of the plot.

We also assessed drought frequency effects on aboveground phytomass fractions. These were calculated as the proportion of biomass, necromass, and flower mass of the total aboveground phytomass in each plot. We did this, because phytomass fractions were in most cases more informative concerning drought frequency effects than absolute values of phytomass (Fig S2) as well as biomass and necromass (data not shown). This is because, most phytomass at the alpine meadow is produced early in the season between May and mid-July (data not shown), which is before the peak drought. Hence, in our study, drought does not impair plant development and therefore has few effects on phytomass production (Fig S2), as is typically the case in shorter-term drought experiments. Increasing drought rather resulted in phytomass loss and/or plant mortality, which is visible in *fractions* of the phytomass, as it illustrates the redistribution of mass. Notably, while biomass refers to productivity, biomass fraction reflects the maintained productivity.

We assessed plant species seasonality, indicating their growth timing during the season. For this, we determined the time point at which the plant species peaked in its number of modules. The typical seasonal behaviour of each plant species was determined based on the ambient treatment plots only. Early seasonal species were defined as species that peaked in modules before the biomass cut and thus reached peak growth between the end of May to end of July. Late seasonal species were defined as species that peaked in modules after the biomass cut and thus reached peak growth between August and October. Shifts in early and late seasonal patterns of species were assessed using values from all treatments.

Because of the low occurrence of flowers, seasonal shifts in flowering was calculated based on flower biomass. For this, we calculated the ln-response ratio between a species' flower dry mass at peak drought and its flower dry mass after the 11-week recovery period. A positive value signifies plant species flowering early in the growing season, i.e., during the drought period. A negative value signifies plant species flowering late in the growing season, i.e., during the recovery phase.

2.3.4 Plant community metrics and composition

Plant community evenness was calculated as Pilou's index. For this, we divided the plant diversity Shannon index by the ln-transformed species richness using the function diversity and specnumber of the vegan package (R Core Team 2021; Oksanen *et al.* 2020). Plant community asynchrony was calculated following Loreau & de Mazancourt (2008), using the package codyn (Hallett *et al.* 2016). Using the same package, plant species loss between subsequent time points was calculated to indicate species loss *within* the growing season. To determine whether communities were dominated by either early or late plant species, we calculated the proportional number of early plant species present in each community at each time point.

We assessed plant community composition using Non-Metric Multidimensional Scaling (NMDS) using metaMDS from the vegan package (Oksanen *et al.* 2020). NMDS on both module counts and biomass were run on Bray-Curtis dissimilarity matrices and with four dimensions to minimise stress values. Significant separation of communities resulting from 'drought frequency' (factor), date (scaled, continuous), and their interaction was tested using a restricted permutation test: samples were permuted *within* plots to take the repeated sampling structure over time into account. The test was performed using adonis from the vegan package with 999 permutations (Oksanen *et al.* 2020). A passive overlay of 'drought frequency' (ln-transformed + 1), 'species richness', 'proportion early species', 'synchrony', species loss between subsequent time points (for NMDS based on counts only), 'LAI', functional group total abundance, flower abundance, and seedling abundance was created

and tested for significance using the function envfit of the vegan package with 999 permutations (Oksanen *et al.* 2020).

2.3.5 Statistics

All calculations and statistical analyses were done in R 4.1.1 (R Core Team 2021). Effects of drought frequency on percentage data were performed using Generalized Linear Mixed Models (GLMM) with a beta distribution and logit-link using the package glmmTMB (Magnusson *et al.* 2017). Drought frequency effects on species richness were tested using Generalised Linear Mixed-Effects Models (GLMER) with a poisson distribution and log-link function using the lme4 package (Bates *et al.* 2015). Drought frequency effects on continuous variables were tested using Linear Mixed Effects Models (LMER) using the lme4 package (Bates *et al.* 2015). Models including ‘date’ (continuous, scaled) were run with ‘plot’ as a random effect to take the repeated sampling design into account. Models on plant species-specific traits were only run for plant species for which at least two data points occurred in each treatment.

Heterogeneity of the mountain meadow was taken into account as a random effect in statistical models. For models on plant community measures, ‘block’ was added as a random effect. Models on plant species-specific traits included the mean vegetation height of each plot before the setup of the rainout shelters as a random effect. This variable was better at capturing variation *within* drought treatments than ‘block’ and, importantly, was not significantly affected by previous drought events (data not shown).

For GLMM and GLMER models, we tested over- and under-dispersion of the model residuals using the package DHARMa (Hartig 2022). For LMER models, we tested assumptions of homogeneity of variances, and normal residual distributions following Zuur *et al.* (2010). Data was ln- or sqrt-transformed when model residuals did not adhere to model assumptions.

3. Results

3.1 Drought frequency decreased biomass and species richness and increased necromass

In a species-rich sub-alpine mountain meadow, we assessed how the annual frequency of severe summer drought affected plant community phytomass, composition, and seasonal dynamics. At the peak of drought, 7 weeks after the setup of the rainout shelters, we found that drought frequency significantly decreased the fraction of community biomass and increased the fraction of community necromass (Fig 2AB). The increase in necromass was primarily related to an increase in the necromass fraction of grass and legume species (Fig S3). Eleven weeks after soil rewetting (recovery), these effects were still present, albeit to a lesser extent (Fig S4AB). In the recovery phase, the fraction of forb biomass decreased, while the fraction of grass biomass and necromass increased with drought frequency (Fig S3).

Drought frequency did not significantly affect the fraction of flower biomass at peak drought (Fig 2C). During recovery, however, the fraction of flower biomass was significantly decreased with increasing drought frequency (Fig S4C).

Furthermore, drought frequency significantly decreased plant species richness at both peak drought and recovery (Fig 2D; S4D). On average, 13 years of recurrent summer drought resulted in a 35 % and 46 % loss of plant species compared with ambient conditions at peak drought and recovery, respectively. Despite this loss in plant species, drought frequency did not significantly affect plant community evenness at both peak drought and recovery (Fig 2E; S4E).

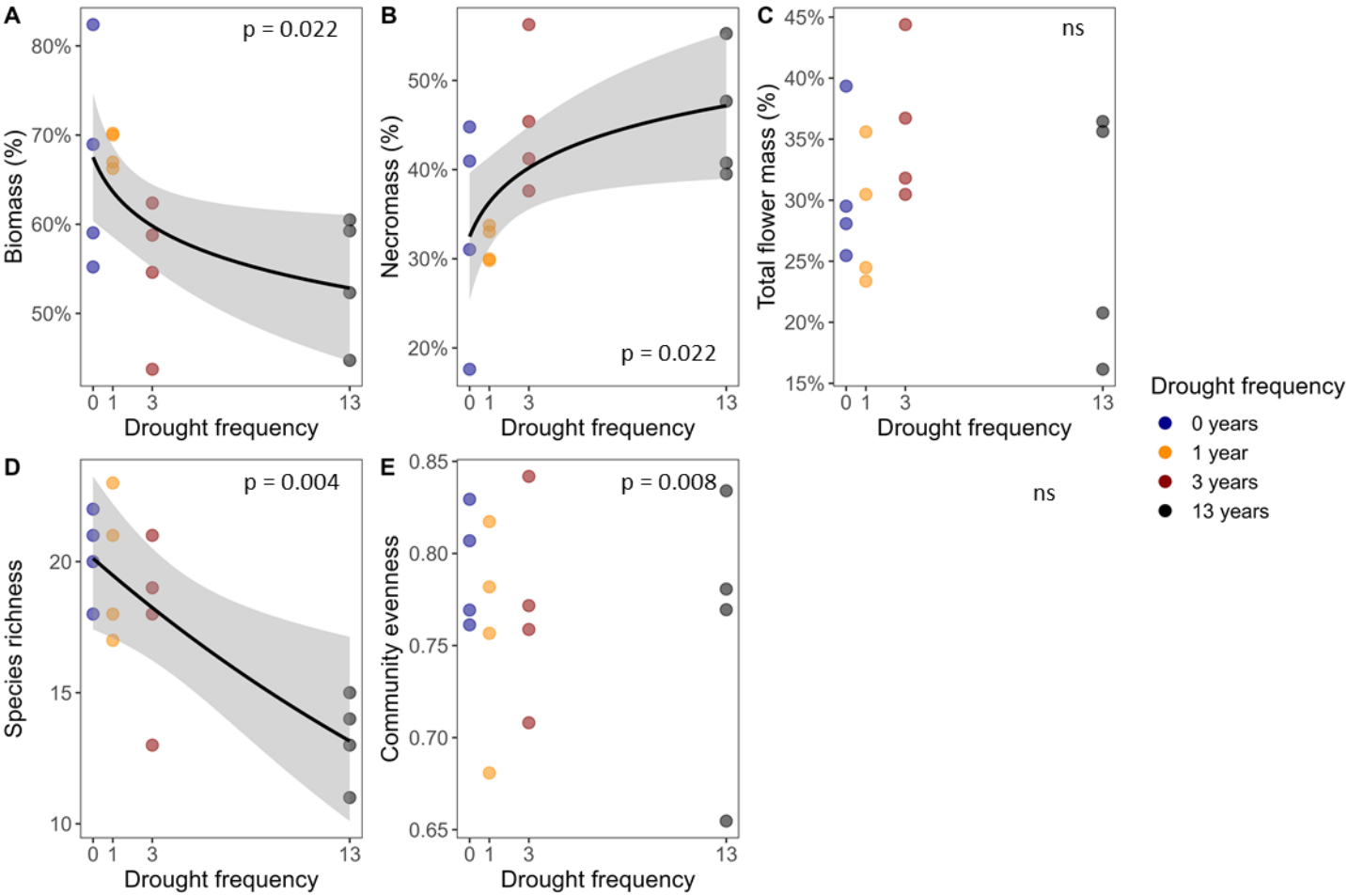


Fig 2 Drought frequency effects at peak drought on (A) the fraction of plant community biomass (living plant material), (B) the fraction of plant community necromass (dead plant material), (C) the fraction of total flower mass, (D) plant community species richness, and (E) plant community evenness. A-C are expressed as percentages of the phytomass (bio- and necromass) in each plot. For significant relations, the black line indicates the mean relation with the shading showing the 95% confidence interval. P-values of significant relations are shown (ns: no significant effect, $p > 0.05$, $n = 4$ per drought frequency level). For effects of drought frequency on overall plant community parameters in the recovery phase, see Fig S4.

3.2 Drought frequency decreased forb and legume seedlings

To assess the impact of drought frequency on plant community seasonal development, we monitored the number of plant species modules and seedlings at biweekly intervals

throughout the season. Drought frequency affected the number of forb and legume seedlings, but not of grass seedlings. During drought and recovery, the number of forb and legume seedlings was consistently lowest with 13 years of recurrent summer drought (Fig 3). Additionally, while in the ambient treatment and one- year drought plots, many forb and legume seedlings appeared after the mowing event, these peaks were lower in plots with three years and especially 13 years of recurrent summer drought (Fig 3). At the same time, plant community LAI decreased with drought frequency at both peak drought and recovery (Fig S5AB). The latter indicates that the soil surface was less covered by leaves of the established plant community with an increase in drought frequency.

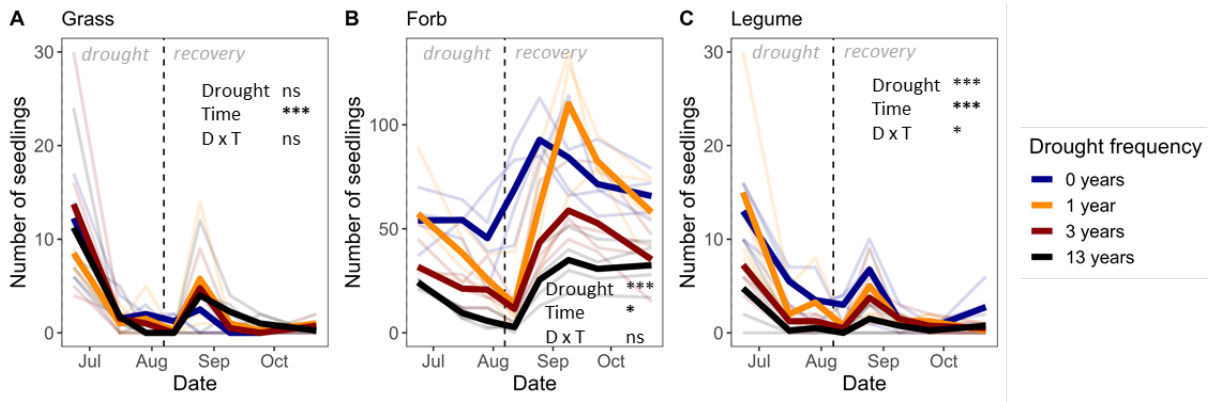
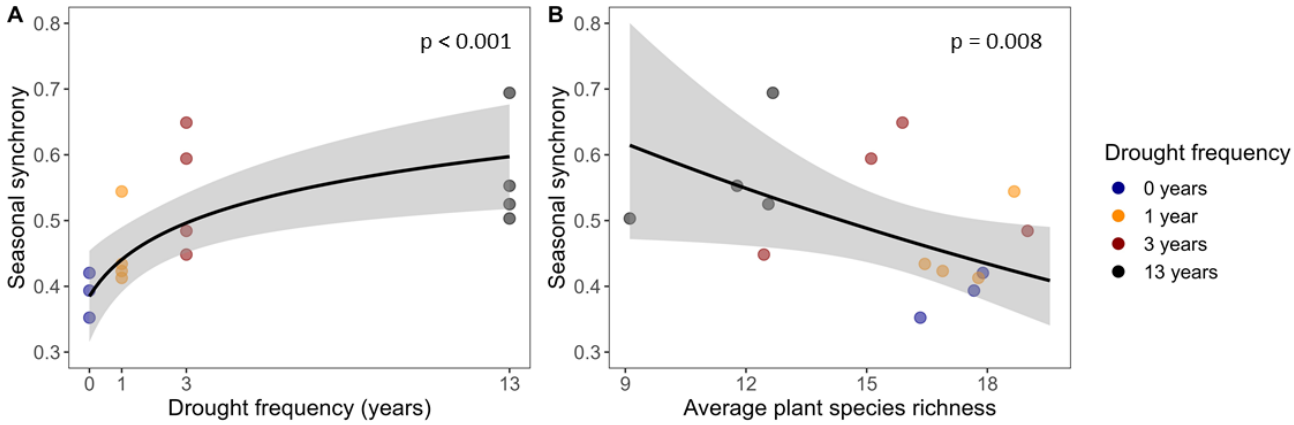


Fig 3 Drought frequency effects on the number of seedlings of (A) grass, (B) forb, and (C) legume species throughout the growing season. The dotted line indicates the timepoint at which drought ended, and the plots were mown and rewetted. Solid lines indicate mean seedling numbers, and faded lines indicate seedling numbers of each plot (n = 4 per drought frequency treatment). Results from analyses of variances are shown with '***' indicating $p < 0.001$, '*' indicating $0.01 > p < 0.05$, and 'ns' indicating no significant effect ($p > 0.05$).

3.3 Drought frequency increased plant community seasonal synchrony

The total number of plant modules in all treatments fluctuated throughout the growing season (Fig S6A). However, it was not significantly affected by drought frequency throughout the season (Fig S6A). Furthermore, plant species richness before, during, and after drought was consistently lowest in plots subjected to 13 years of drought, followed by three years,

352 and one year of drought (Fig S6B). Plant species richness across all treatments dropped
353 sharply after peak drought and the mowing event but increased to levels similar to their
354 respective pre-drought conditions within a month of recovery (Fig S6B).
355 Using the module counts of each plant species over the season, we calculated plant
356 community seasonal synchrony. We found that drought significantly increased seasonal
357 synchrony, that is, drought synchronised the timing of plant species growth during the
358 season. Increasing drought frequency enhanced this synchronising effect (Fig 4A).
359 Furthermore, seasonal synchrony was negatively related to the average plant species
360 richness throughout the season, that is, synchrony was high when average plant species
361 richness was low (Fig 4B).



362

363 **Fig 4** (A) Effects of increasing drought frequency on seasonal synchrony (sensu Loreau & de
364 Mazancourt (2008)) based on plant module counts across the growing season from mid-May to mid-
365 October, and (B) seasonal synchrony in relation to the average plant species richness (n = 4 per
366 drought frequency treatment). The black line indicates the mean relation with the shading showing the
367 95% confidence interval. P-values of significant relations are shown.

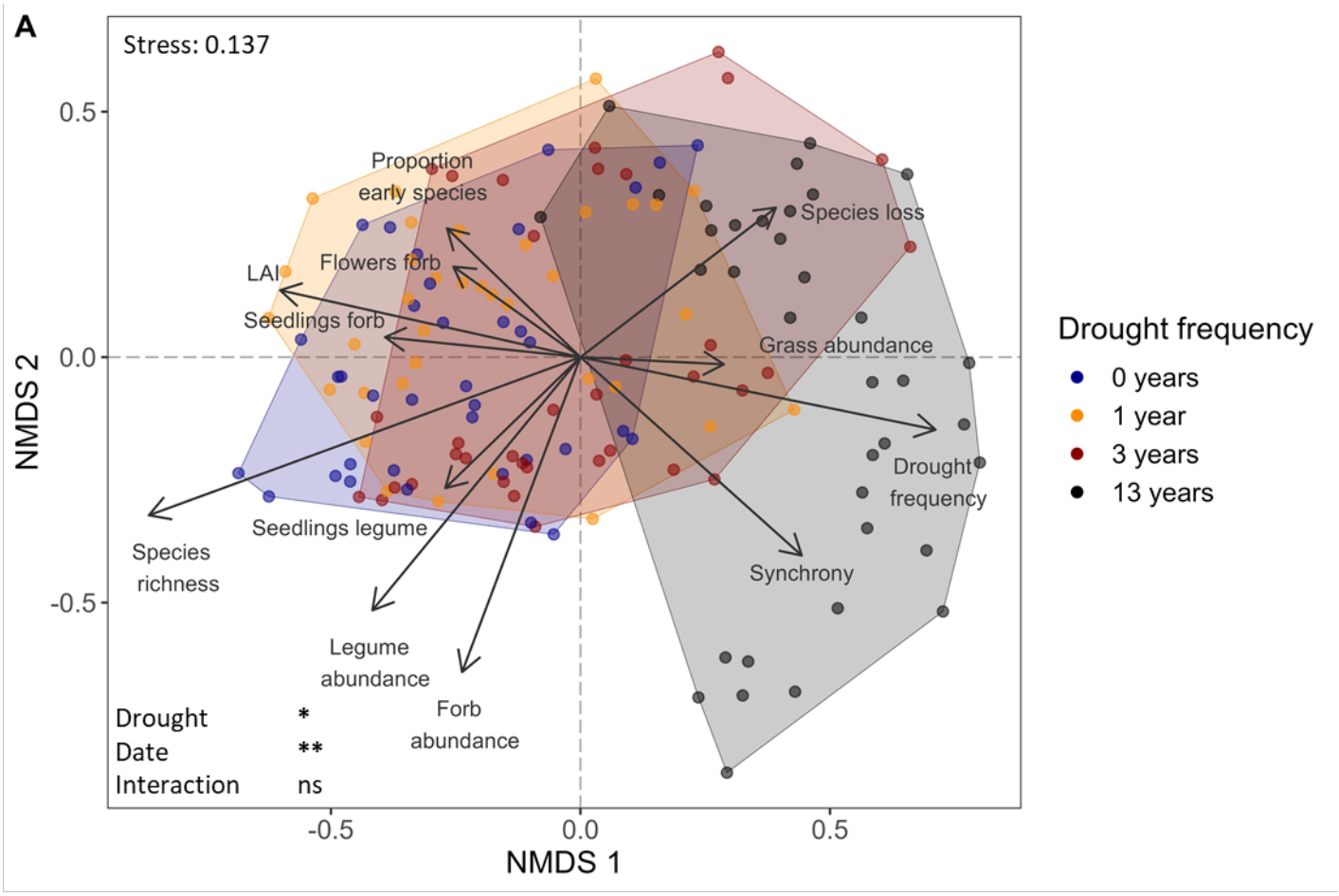
368

369 **3.4 Drought frequency shifts plant community composition towards less early**
370 **seasonal and more grass species**

371 We examined the effect of drought frequency on plant community composition throughout
372 the season using non-metric multidimensional scaling (NMDS). We found significant

differentiation of plant communities based on drought frequency, with communities experiencing no drought and 13 years of drought exhibiting the greatest dissimilarity (Fig 5A, S7). This compositional separation, driven by drought frequency, was primarily captured along the first axes of the NMDS plot. This first axis coincided with the decrease in plant species richness and the increase in community synchrony as drought frequency increased (Fig 5A). Moreover, we observed that the compositional shifts associated with increasing drought frequency were characterised by a high grass abundance and a decrease in early-seasonal plant species. In addition, LAI, the number of forb flowers, and the number of forb seedlings decreased with drought frequency (Fig 5A). These patterns were similar when basing the analyses on plant biomass collected at peak drought and recovery (Fig S8).

During the recovery phase, plant communities in the plots exposed to one and three years of drought shifted along the second NMDS axis. After rewetting, these one- and three-year droughted plots temporarily shifted into the top-right quadrant of the NMDS (Fig 5B). This direction was related to a high loss of species, particularly early-seasonal species (for specification see Table S1), and a decrease in forb and legume abundance (Fig 5B). However, later in the recovery phase, the one- and three-year droughted communities gradually returned to a similar position on the second axis of the NMDS plot as before the recovery phase (Fig 5B). This recovery pattern indicates slower recovery of early-seasonal forb, and legume species within the one- and three-year droughted plots (Fig 5B).


A
Stress: 0.137
Proportion early species
Species loss
LAI
Flowers forb
Seedlings forb
Grass abundance
Drought frequency
Seedlings legume
Species richness
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Forb abundance
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Date **
Interaction ns
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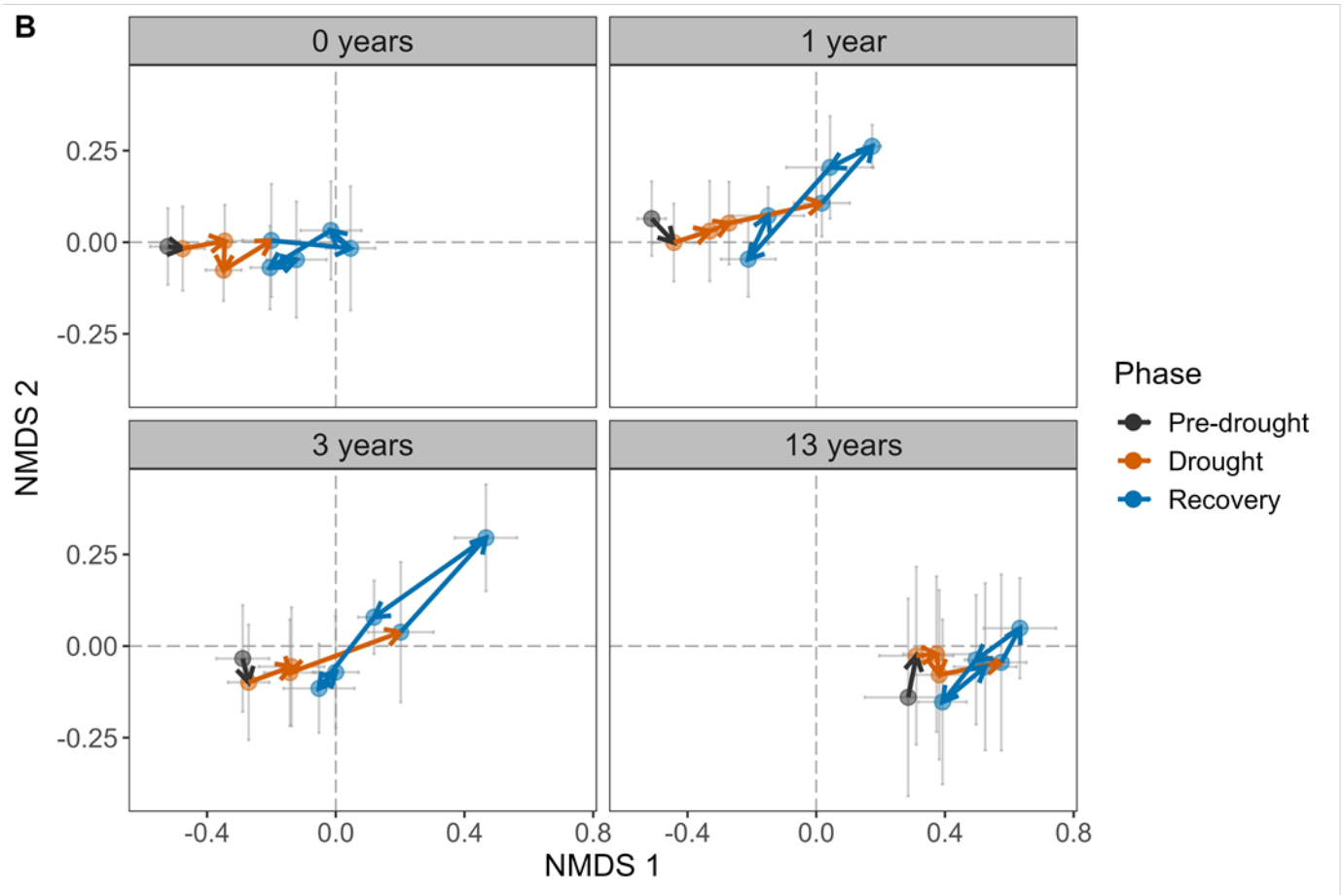
0 years
1 year
3 years
13 years


B
0 years
1 year
3 years
13 years
NMDS 2
NMDS 1
Phase
Pre-drought
Drought
Recovery

392

Fig 5 Non-metric multidimensional scaling (NMDS) of plant species module counts throughout the growing season in permanent plots receiving no drought (0 years), one year, three years, and 13 years of recurrent summer drought. (A) Drought frequency effects on plant community composition with a passive overlay of community variables significantly relating to the NMDS patterns. Each point represents a plot at a point in time. Hulls group all plots and time points per drought frequency treatment together. Length of arrows indicate the strength of the relation. 'Proportion early species' refers to the proportion of early species versus the total number of species present in a plot. 'Grass', 'forb', and 'legume abundance' refer to the total number of modules summed per functional group. 'Synchrony' refers to community seasonal synchrony sensu Loreau & de Mazancourt (2008). 'Species loss' refers to the percentage of species lost *within* the growing season between two subsequent time points. 'LAI' refers to leaf area index. Results of a restricted permutation test is shown with '**' indicating $0.001 > p < 0.01$, '*' indicating $0.01 > p < 0.05$, and 'ns' indicating no significant effect ($p > 0.05$) ($n = 4$ plots per drought frequency treatment). (B) Community compositional development over time. Points are mean values ± standard errors. Note that axes scales differ between A and B to improve readability. For NMDS axes 3 and 4, see Fig S7. For position of plant species along the axes, see Fig S9.

3.5 Weak drought frequency effects on individual plant species responses

For 10 plant species that were present in sufficient abundances in all three drought frequency treatments and the ambient treatment, we tested whether drought frequency affected SLA at peak drought and recovery, as well as the timing of their growth in the season. At peak drought, we found that drought frequency significantly decreased SLA of *Alchemilla vulgaris, Leontodon hispidus, Plantago lanceolata* (marginally)*, Trifolium repens,* and *Veronica chamaedrys* (Fig S10)*.* During recovery, these patterns had disappeared (Fig S11)*.*

Drought frequency only significantly affected the timing of *Trifolium pratense,* which peaked in number of modules earlier in the season with a high drought frequency (Fig S12). Drought frequency only significantly affected the timing of flowering of *Leontodon hispidus,* which shifted its flower production to earlier in the season with a high drought frequency (Fig S13).

4. Discussion

We investigated the effects of recurrent summer drought of up to 13 years on plant communities in a sub-alpine mountain meadow, as well as their underlying mechanisms. We found that drought frequency increased plant community seasonal synchrony. This increase in synchrony with greater drought frequency was underpinned by a decrease in species richness, a loss of early-seasonal plant species, a shift in plant functional groups, and the constrained establishment of new species and individuals throughout the growing season. Moreover, these changes were associated with a decreased biomass fraction as drought frequency increased.

Higher seasonal synchrony with increasing drought frequency due to species loss, compositional shifts, and lower species establishment

We found that with increasing drought frequency, species **synchrony** within the plant community was enhanced. This indicates that the species in the plant community under increasing drought frequency were more similar in their timing of plant species growth throughout the season. Results on increasing synchrony *between* years for grasslands were reported by other studies, finding an increase in synchrony with drought frequency of three years in an alpine meadow (He *et al.* 2022), and four years across grasslands (Muraina *et al.* 2021). Our study suggests that with increasing drought frequency (to up to 13 years of recurrent drought) seasonal synchrony becomes increasingly pronounced. In consequence, multiple droughts may hamper the stability (de Mazancourt *et al.* 2013; Loreau & de Mazancourt 2013), e.g. under future disturbances. This is because synchronous communities are typically less stable to disturbances than asynchronous communities (Luo *et al.* 2023b; Muraina *et al.* 2021; Craven *et al.* 2018; de Mazancourt *et al.* 2013). Hence, an increase in seasonal synchrony with drought frequency suggests that multiple droughts may hamper the stability of communities to future disturbances. We suggest that with the ongoing and projected increases in the frequency of drought events due to climate change (IPCC

2021), grassland ecosystems become increasingly seasonally synchronous and vulnerable to future disturbances.

We identified several underlying mechanisms that contributed to the increase in synchrony with increasing drought frequency. First, the enhancement in synchrony was associated with a **decrease in plant species richness**. This drought-induced species loss aligns with findings from other studies that observed enhanced mortality during and following single (Sippel *et al.* 2018; Stampfli *et al.* 2018; Frank *et al.* 2015), and recurrent drought events (Luo *et al.* 2023a; Wagg *et al.* 2017). Species richness mostly tends to be negatively associated with synchrony (Zhang *et al.* 2018; Hautier *et al.* 2014), but was also found to be positively associated with synchrony (Valencia *et al.* 2020). Furthermore, the species that remained in the community, were more similar in terms of their seasonal development. Under increasing drought frequency, communities were dominated by species growing later in the season, as **early-growing species were lost**. This may have been caused by a reduced recovery potential of early-season species after the drought event: We found that early-seasonal species were taking a longer time to recover under single- and three-year drought. While it was suggested, that early-season species may be more susceptible to drought (Mamolos *et al.* 2001), we found first evidence that early-growing species actually get lost at high drought frequencies. As a result, under recurrent drought, more temporally similar growing species lead to more seasonally synchronized communities.

Furthermore, the seasonal synchronization of species may have been driven by the observed shift within plant **functional groups** under higher drought frequency. We found that the community composition showed a reduction in forb and legume species, and more grass species with increasing drought frequency. The loss of species within these functional groups with increasing drought frequency may be explained by the observed lower forb and legume seedlings throughout the season and the slower recovery of forbs and legumes post drought. Hence, our findings support the notion that forb and legume species are less robust to recurrent drought compared with grass species. This is in line with studies on single

drought (DeBoeck *et al.* 2018; Hoover *et al.* 2014) and recurrent drought events (Luo *et al.* 2023b; Luo *et al.* 2023a; Sun *et al.* 2023; Xu *et al.* 2021), which also suggest grass species to be less sensitive to drought compared to forbs. Overall, the observed shifts under recurrent drought towards less diversity in functional groups can enhance synchrony by increasing the similarity of seasonal growth within the community. This similarity in functional groups can furthermore decrease the variety in responses to drought throughout the growing season (Zhang *et al.* 2020) and thereby contribute to an enhanced vulnerability under future disturbances.

The increase in synchrony with higher drought frequency may also be driven by the **constrained establishment of new species and individuals** throughout the growing season: As drought frequency increased, we observed a **decline in seedling numbers**. This occurred during the entire vegetation period and was most pronounced with high drought frequency of 13 years. At the same time, less soil surface area was covered by leaves under high drought frequency. Hence, the seedling decline took place despite a diminished light limitation, and thus likely a higher chance of seedling survival in the 13-year recurrent drought plots (Eskelinen *et al.* 2022). Moreover, we observed that plant species **reproductive output** was inhibited with increasing drought frequency: In the recovery phase, the fraction of flower biomass was decreased with increasing drought frequency. A negative drought effect on reproductive output was also found by Zeiter *et al.* (2016). Reasons might be, that species experiencing drought were unable to produce flowers due to preceding and ongoing water limitation. Consequently, in the communities with these high drought frequencies, fewer seedlings appeared in the recovery phase.
Overall, in our study, flower development and consequently seedlings, were severely inhibited with higher drought frequencies. This reduces the establishment of new species and individuals and leads to species loss, with effects on community composition and synchrony. During long-term recurrent drought, effects of drought on reduced plant reproductive output and resultant seedlings likely exacerbated plant species loss and the

increase in seasonal synchrony. However, single- and low-frequency experiments do not cover drought effects that go beyond the lifetime of the plant species. We therefore propose that, to comprehensively comprehend changes in plant communities arising from high drought frequency, the implementation of long-term experiments is essential.

Enhanced loss of biomass fraction with increasing drought frequency

Drought did not reduce productivity in our study (Fig S2), which is consistent with findings on productivity under five years of drought by Jentsch *et al.* (2011) and shoot biomass under four years of drought by Fuchslueger *et al.* (2016). Nonetheless, other low drought frequency studies reported a continuous decline in aboveground net primary productivity under recurrent drought of up to four years (Luo *et al.* 2023a; Xu *et al.* 2021). The divergent results may be attributed to the different frequencies of drought events or plant communities. However, we found that increasing drought frequency gradually decreased the fraction of plant community biomass and enhanced the fraction of necromass at peak drought and recovery, respectively. Hence, during the growing season, including drought and recovery, the maintained productivity is lower, and senescence is higher with increasing drought frequency. The higher senescence at peak drought and recovery supports our observation of increasing species loss throughout the growing season. Reductions in maintained productivity might be caused by the lasting repercussions of drought on plant community composition and synchrony.

Under recurrent drought, plants have the capacity to develop **adaptations** that can dampen drought effects (Müller & Bahn 2022). This can occur e.g., by adjusting photosynthesis, osmosis (Li *et al.* 2022), root biomass (Legay *et al.* 2018), epigenetics, transcription factors and rate (Jacques *et al.* 2021; Ding *et al.* 2012; Bruce *et al.* 2007), as well as transgenerational changes (Chen *et al.* 2022). Within our study, we found only weak species adaptations under recurrent drought. For example, increasing drought frequency did not result in adaptations of the duration of the growing season and only shifted the timing of growth for one legume species to earlier in the season. Also, the timing of plant species

529 flowering was only shifted earlier with increasing drought frequency for one forb species.
530 Moreover, the lower SLA at peak drought of some of the most abundant forb species and
531 one legume species with increasing drought frequency can hint towards higher water-use
532 efficiency, resource conservation, and reduced plant growth, which can enhance short-term
533 stress tolerance (Wellstein *et al.* 2017). This short-term adaptation could lead to less
534 vulnerability under drought. Overall, adaptations did not play a major role in compensating
535 drought effects in our study.

536 Chen *et al.* (2022) suggested that adaptations *along generations* under recurrent
537 drought lead to improved ecosystem responses. However, we suggest that the lower
538 reproduction might overrule any potential adaptive effects of single plant species along
539 generations through species loss. Plant communities are therefore likely not able to adapt to
540 the predicted rapid increase in drought frequency with the ongoing change in climate. To
541 gain further understanding of the repercussions on plant community due to increasing
542 drought frequencies, long-term drought experiments, including transgenerational effects,
543 could yield important insights.

5. Conclusion

We show that long-term annual drought frequency in grasslands increases plant community seasonal synchrony. Synchrony increased through a reduction in species richness, a loss of early-seasonal plant species, a shift in plant functional groups, and the constrained establishment of new species and individuals throughout the growing season. These changes were associated with a decreased fraction of biomass with increasing drought frequency.

Overall, we conclude that with the increase in frequency and severity of drought events due to climate change, grassland plant communities will increase in seasonal synchrony. As asynchrony is a critical component of community stability, we suggest that this will lead to lasting effects far beyond the drought event, by increasing the vulnerability of ecosystems to future disturbances. Moreover, we show that negative drought effects were enhanced with an increasing in drought frequency. We therefore conclude that single and low-frequency drought studies may not adequately predict longer-term changes in our rapidly shifting climate. Furthermore, on the long-term, negative effects, mostly associated with species loss, will likely overrule the comparatively weak adaptation of species. Hence, with the current speed of climate change, communities will be severely and probably irrevocably affected the moment species are lost from the ecosystem. We therefore conclude that plant communities are unlikely to be able to adapt to recurrent drought. Importantly, long-term experiments will be essential in understanding the severity of these climate change impacts.

6. Sample CRediT author statement

Lena M. Müller: Conceptualization, Investigation, Project administration, Writing - Original Draft (lead), Visualization, Funding acquisition; **Michael Bahn**: Conceptualization, Resources, Funding acquisition, Writing - Review & Editing; **Maximillian Weidle**: Investigation, Writing - Review & Editing; **Georg Leitinger**: Writing - Review & Editing, Funding acquisition; **Dina in 't Zandt**: Conceptualization, Investigation, Formal Analysis, Data Curation, Visualization, Writing - Review & Editing, Writing - Original Draft, Supervision

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8. Data Availability Statement

Upon acceptance, all data and code (R scripts) will be made publicly available via the Zenodo data repository, will be referenced using a stable object identifier (DOI) in the Materials and Methods section, and cited in the reference list.

9. Conflict of interest

The authors declare no conflict of interest.

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