The limits of predicting maladaptation to future environments with genomic data

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Abstract

Anthropogenically driven changes in land use and climate patterns pose unprecedented challenges to species persistence. To understand the extent of these impacts, genomic offset methods have been used to forecast maladaptation of natural populations to future environmental change. However, while their use has become increasingly common, little is known regarding their predictive performance across a wide array of realistic and challenging scenarios. Here, we evaluate four offset methods (Gradient Forests, the Risk-Of-Non-Adaptedness, redundancy analysis, and LFMM2) using an extensive set of simulated datasets that vary demography, adaptive architecture, and the number and spatial patterns of adaptive environments. For each dataset, we train models using either *all*, *adaptive*, or *neutral* marker sets and evaluate performance using *in silico* common gardens by correlating known fitness with projected offset. Using over 4,850,000 of such evaluations, we find that 1) method performance is largely due to the degree of local adaptation across the metapopulation ($LA_{\Delta SA}$), 2) *adaptive* marker sets provide minimal performance advantages, 3) within-landscape performance is variable across gardens and declines when offset models are trained using additional non-adaptive environments, and 4) despite (1), performance declines more rapidly in novel climates for metapopulations with higher $LA_{\Delta SA}$ than lower $LA_{\Delta SA}$. We discuss the implications of these results for management, assisted gene flow, and assisted migration.

1 | Introduction

The impacts of climate change, habitat loss, and extreme weather events pose urgent challenges to the management of species, communities, habitats, and ecosystem services (Bonan, 2008; Doney et al., 2012; Hoegh-Guldberg & Bruno, 2010). Traditional methods used to infer environmental suitability, such as reciprocal transplants and common gardens, require time and resources that may not be available or feasible for many organisms of management concern, particularly for long-lived organisms where reproductive stages occur after several decades of development. Ecological forecasting models have therefore become increasingly germane to support environmental decision making by managers across both terrestrial and marine systems.

In the context of population viability in the face of environmental change, many of these models rely on theoretical expectations that the limits of species' distributions are primarily determined by the distribution of environmental conditions (e.g., Good 1931), and that occupancy of highly suitable habitat enables increased abundance through greater survival and reproduction (i.e., fitness) of individuals (Brown, 1984). Such methods, termed species distribution models or ecological niche models (see Elith & Leathwick, 2009 for a discussion on terminology) are correlative approaches that are often used to predict (relative) habitat suitability for a single species (Lee-Yaw et al., 2022). This information is used to understand potential impacts on the species from future climate change. However, these methods often ignore aspects of the species' evolutionary history that could be important for predicting long-term population persistence, such as the environmental drivers of local adaptation or spatial patterns of adaptive genetic variation (Waldvogel et al., 2020).

Subsequent methods, termed genomic offsets (reviewed in Capblancq et al., 2020; Rellstab et al., 2021), have attempted to address these shortcomings by modeling relationships between environmental and genetic variation to predict maladaptation of natural populations to either future climates *in situ*, or to predict the relative suitability of these populations for the specific environment of a restoration site. Empirical attempts to confirm predictions from genomic offset models are rare and, compared to attempts *in silico* (Láruson et al., 2022), have found relatively weaker relationships between predicted maladaptation to common garden climates and the measurement of phenotypic proxies for fitness from individuals grown in these same environments (e.g., Capblancq & Forester, 2021; Fitzpatrick et al., 2021; Lind et al., 2024). Even so, these empirical results have consistently shown the expected negative relationship between predicted offset and common garden performance. Further, many of these studies found that genomic offsets often perform better than climate or geographic distance alone (e.g., Capblancq & Forester, 2021; Fitzpatrick et al., 2021; Láruson et al., 2022; Lind et al., 2024).

Across empirical and *in silico* studies, little difference in performance was found between

models trained using only adaptive markers (i.e., known *in silico*, or candidates from empirical genotype-environment [GEA] associations) and those chosen at random, suggesting that genome-wide data may be sufficient to capture signals relevant to environmental adaptation.

Together, these results suggest that genomic offset methods may provide valuable insight for management. Little is known, however, about how robust these methods are across a wide array of realistic empirical scenarios, nor the extent to which independent methods will arrive at similar conclusions when analyzing the same data. Indeed, concerns regarding the accuracy of ecological forecasting models present a primary limitation towards incorporating inferences from these models into management (Clark et al., 2001; Schmolke et al., 2010) and genomic offset models are no exception. Major questions still remain about how performance is affected by aspects of the evolutionary history of sampled populations, the type of signals in putatively ideal datasets that may mislead offset inference, the importance of identifying environmental drivers of local adaptation *a priori*, and the consistency of predictive performance across contemporary environmental space. Finally, because novel climates with no recent analog are expected to increase in the future (Lotterhos et al., 2021; Mahony et al., 2017) there is also uncertainty regarding the performance of forecasting models when predictions are made to environments that drastically differ from those used to train and build the models themselves (Fitzpatrick et al., 2018; Lind et al., 2024).

While much uncertainty remains regarding the predictive performance of genomic offsets, the domain of applicability (i.e., the circumstances under which a method is acceptably accurate) for these methods can be more precisely defined using simulated data (Lotterhos et al., 2022). Simulated data, where there is no error in the estimation of allele frequencies, environmental variables, individual fitness, or the knowledge regarding the drivers of local adaptation, present ideal circumstances for understanding the limits of genomic offsets and the circumstances under which data from naturally occurring taxa will provide useful inference. To provide relevant inference regarding the domain of applicability, simulations should capture the complexities of empirical data with biological realism (e.g., clinal or patchy environments), present contrasting cases of differing scenarios while controlling for important features of the data (e.g., varying population connectivity but controlling for mean differentiation), and challenge methods using adversarial scenarios that capture extreme characteristics of empirical data (e.g., prediction to novel environments with no current analog available for model training; Lotterhos et al. 2022).

Here, we use a wide array of previously published biologically realistic, contrasting, and adversarial simulations from Lotterhos (2023) in an attempt to more precisely define the limits of predictive performance of five genomic offset methods (Table 1): Gradient Forests (GF$_{offset}$; *sensu* Fitzpatrick & Keller, 2015), the Risk Of Non-Adaptedness (RONA, Rellstab et al., 2016), Latent Factor Mixed

194 Models (LFMM2$_{offset}$, *sensu* Gain & 195 François, 2021, and redundancy analysis 196 (RDA$_{offset}$, *sensu* Capblancq & Forester, 197 2021). The main goal of this study was 198 to understand how the evolutionary and 199 experimental parameters used in the 200 training and evaluation of offset 201 methods affect the accuracy of the 202 methods' projections of maladaptation 203 under ideal empirical scenarios (i.e., 204 using data with no inherent error). 205 Using these scenarios, we ask the 206 following six questions: 1) Which 207 aspects of the past evolutionary history 208 affect performance of offset methods? 2) 209 How is offset performance affected by 210 the proportion of loci with clinal alleles 211 in the data? 3) Is method performance 212 driven by causal loci or by genome-wide 213 patterns of isolation-by-environment? 214 4) What is the variation of model 215 performance across the landscape? 5) 216 How does the addition of non-adaptive 217 nuisance environments in training affect 218 performance? 6) How well do offset 219 models extrapolate to novel 220 environments outside the range of 221 environmental values used in training?

222 2 | Methods

223   Throughout this manuscript we will 224 be citing code used to carry out specific 225 analyses in-line with the text. 226 Supplemental Notes S1-S2 outlines and 227 describes the sets of scripts or, most 228 often, jupyter notebooks, used to code 229 analyses. Scripts and notebooks are 230 both referenced as Supplemental Code 231 (SC) using a directory numbering 232 system (e.g., SC 02.05). More

233 information regarding the numbering system, 234 archiving, and software versions can be found 235 in the Data Availability section.

236 *2.1 | Explanation of Simulations and* 237 *Training Data*

238   To train offset methods we used single 239 nucleotide polymorphism (SNP) and 240 environmental data from a set of previously 241 published simulations (225 levels with 10 242 replicates each) of a Wright-Fisher 243 metapopulation of 100 demes on a 10 x 10 grid 244 evolving across a heterogeneous landscape 245 (Lotterhos, 2023). Each dataset was 246 simulated under a combination of the 247 following four evolutionary parameters: i) 248 three landscapes (10 populations x 10 249 populations) that varied in vicariance and 250 environmental gradients (*Estuary - Clines*; 251 *Stepping Stone - Clines*; and *Stepping Stone -* 252 *Mountain*), ii) five demographies that varied 253 population size and migration rates across the 254 landscape, iii) three genic levels that varied in 255 the effect size and number of mutations 256 underlying adaptation (mono-, oligo-, and 257 polygenic), and iv) five pleiotropy levels that 258 varied the number of quantitative traits under 259 locally stabilizing selection ($n_{traits} \in \{1, 2\}$), 260 presence of pleiotropy (when $n_{traits} = 2$), and 261 variability of selection strength across 262 individual traits (see Fig. 1 in Lotterhos 263 2023).
264   The adaptive trait(s) were under selection 265 by a different environmental variable, where 266 the optimum trait value was given by the 267 local environment on the landscape. The 268 adaptive trait(s) undergoing selection 269 responded to either a latitudinal temperature 270 gradient (*temp*; $n_{traits} = 1$), or to both *temp* 271 and a longitudinal "*Env2*" gradient ($n_{traits} =$

2). *Env2* represented distinct biological analogies depending on the context: in the *Stepping Stone - Mountain* landscape *Env2* was analogous to elevation (e.g., as with tree species), whereas in the *Estuary - Clines* landscape the *Env2* environment was analogous to gradients of salinity within coastal inlets connected only by the outer marine (ocean) environment (e.g., as with stickleback or oyster species).

Twenty independent linkage groups were simulated. Of these, mutations that had effects on one or more phenotypes under selection (i.e., quantitative trait nucleotides, QTNs) were allowed to evolve on only ten linkage groups, and neutral mutations were added to all 20 linkage groups with tree sequencing (for details see Lotterhos 2023). Adaptive traits were determined additively by effects of QTNs.

In all simulations, phenotypic clines evolved between each trait and the selective environment (Lotterhos, 2023), where populations became locally adapted to their environment, measured at the metapopulation level as the mean difference of demes in sympatry minus allopatry ($LA_{\Delta SA}$, Blanquart et al., 2013). $LA_{\Delta SA}$ equates to the average levels of local adaptation at the deme level which can be calculated for each deme by both home-away ($LA_{\Delta HA}$) and local-foreign ($LA_{\Delta LF}$) measures.

These simulations represent a wide array of realistic, contrasting, and adversarial scenarios in which we could more precisely define the domain of applicability of offset methods. For instance, in the *Stepping Stone - Mountain* landscape, geographic distance and environmental distance were not strongly correlated, whereas in the *Stepping Stone - Clines* and *Estuary - Clines* they were. Additionally, the proportion of mutations with monotonic frequency gradients (i.e., allelic clines) underlying local adaptation varied across the simulated datasets (Lotterhos, 2023), which may also affect offset performance. These simulations also presented demographic scenarios in which selection was confounded with genetic drift or population genetic structure.

For each simulation, ten individuals were randomly chosen per population for a total of 1000 individuals. Individual genotypes were coded as counts of the derived allele. Alleles with global minor allele frequency (MAF) < 0.01 were removed. Using all 100 populations, population-level derived allele frequencies and current environmental values were used as input to train offset methods.

In addition to the 2250 simulated Wright-Fisher datasets (225 levels * 10 replicates), we also included a non-Wright-Fisher case with range expansion from three refugia and secondary contact (Lotterhos 2023). This simulation evolved variable degrees of admixture across the landscape. Six moderately polygenic environmental traits ($n_{\text{traits}} = 6$) were under selection from the environment. Environments were based on six weakly correlated environmental variables taken from Bioclim environmental measures of western Canada. The simulation evolved local adaptation at all six traits with unconstrained pleiotropy. For more details on simulations, see (Lotterhos, 2023).

2.2 | Evaluation of Offset Methods

We investigated the performance of five implementations of four genomic offset methods (Table 1): GF_{offset}, RDA_{offset}, $LFMM2_{offset}$, and RONA. While GF_{offset}, RDA_{offset}, and $LFMM2_{offset}$ can use multivariate environmental data to train models, RONA can only account for a single environment at one time (Table 1). Additionally, while GF_{offset} and RONA do not apply correction for population genetic structure, $LFMM2_{offset}$ does by default, and structure correction with RDA_{offset} is optional. We thus evaluate RDA_{offset} with (RDA-corrected) and without (RDA-uncorrected) population genetic structure correction (Table 1). For additional specifics related to the implementation of each offset method, see Supplemental Note S1.1-S1.4 and Fig. S1, Fig. S2, Fig. S3.

We varied construction of genomic offset training datasets for each replicate of the 1-, 2-, and 6-trait simulations by varying the marker set used in model training (Fig. 1A, Table 2; see *Q3* below). Each model was trained using genetic and environmental data from all 100 populations. The environmental variables used were only those imposing selection pressure. We predict offset from each model for each population to all 100 within-landscape common gardens from a full factorial *in silico* reciprocal transplant design (Fig. 1B). For each common garden, we quantified offset model performance as the rank correlation (Kendall's **τ**) between the population mean fitness (averaged over sampled individuals, Equation 3 in Lotterhos 2023) and projected population offset (Fig. 1C). Strong negative relationships between fitness and predicted offset indicate higher performance of the method (note y-axes of Kendall's **τ** are inverted within figures to show more intuitive performance relationships, Fig. 1C-11). We refer to the preceding processing of data as the *Adaptive Environment* workflow (Fig. 1, Table 2).

To explore the impact of the choice of environmental variables used (see *Q5* below), we used a workflow similar to the *Adaptive Environment* workflow, except instead of using only adaptive environmental variables, we used additional non-adaptive (i.e., nuisance) environmental variables in training and prediction (second row, Table 2). These nuisance variables had relatively weak correlation structure with adaptive environments and each other (Fig. S4). We refer to each of these nuisance levels by the number of traits under selection and the number of nuisance environments used (e.g., *1-trait 3-nuisance*). We refer to this workflow as the *Nuisance Environment* workflow .

Finally, to contrast with within- landscape evaluations, we explored predictive performance of *Adaptive Environment* offset models in novel environments that are beyond the range of values of those used in training (see *Q6* below). In these novelty cases, we use 11 common gardens, each progressively more distant from the average environment used in training (i.e., climate center) and evaluate performance in each garden. We refer to this workflow as the *Climate Novelty* workflow. See Supplemental Note S3 and Fig. S5 for details regarding the choice of environmental values for novelty scenarios.

2.3 | Study Questions

Q1 - Which aspects of the past evolutionary history affect within-landscape performance of offset methods?

For each offset method, we used a fixed-effects type II ANOVA model to test for significant differences in the performance from 2-trait *Adaptive Environment* models trained using *all* markers using the following factors: landscape (*Estuary - Clines, Stepping Stone - Clines, Stepping Stone - Mountain*), demography (five levels describing population size and migration patterns across the landscape), genic level of architecture (three levels from oligogenic to polygenic), presence or absence of pleiotropy, proportion of loci with clinal allele frequencies (as defined in Lotterhos, 2023), degree of local adaptation (ΔSA), and common garden ID. Specifically,

$$Y_{ij} = L_i + D_i + GL_i + P_i +$$
$$p_{cQTN,t,i} + p_{cNeut,t,i} + p_{cQTN,Env2,i} +$$
$$p_{cNeut,Env2,i} + LA_{\Delta SA,i} + G_j$$
$$\text{(Eq. 1)}$$

where Y_{ij} is the within-landscape performance (Kendall's τ) of a single method for garden j in simulation i, with factors for landscape (L), demography (D), genic level (GL), presence of pleiotropy (P), proportion of QTN or neutral alleles with *temp* clines (respectively $p_{cQTN,t,i}$ and $p_{cNeut,t,i}$), proportion of QTN or neutral alleles with *Env2* clines (respectively $p_{cQTN,Env2,i}$ and $p_{cNeut,Env2,i}$), degree of local adaptation ($LA_{\Delta SA}$), and garden ID (G). The first four factors are illustrated in Fig. 1 of Lotterhos (2023).

Q2 - How is offset performance affected by the proportion of clinal alleles in the data?

Clinal alleles (i.e., alleles with monotonic gradients in frequency across space) that covary with environmental clines could be weighted more heavily in offset models that emphasize loci whose allele frequencies explain significant variation across local environmental values. Using 2-trait models trained using *all* markers from the *Adaptive Environment* workflow, we used an ANOVA model (Eq. 2) to test the hypothesis that clinal alleles differentially impact model performance, independent from the other factors from Eq. 1:

$$Y_{ij} = p_{cQTN,t,i} + p_{cNeut,t,i} + p_{cQTN,Env2,i} +$$
$$p_{cNeut,Env2,i}$$
$$\text{(Eq. 2)}$$

The factors representing clinal alleles in Eq. 2 are the same as those in Eq. 1.

Q3 - Is method performance driven by causal loci or by genome-wide patterns of Isolation By Environment?

For each offset method and workflow, we varied the set of input markers for 1-, 2- and 6-trait simulations that were used in training to determine if performance of a method was driven by properties of the evolutionary forces shaping genotype-environment relationships: 1) *adaptive* markers (i.e., QTNs with effects on at least one trait), 2) *neutral* markers (SNPs on linkage groups without QTNs), and 3) *all* markers (union of *adaptive* and *neutral*

markers, as well as non-QTN markers
on the same linkage groups as QTNs).
Only loci that passed MAF filtering
were included in marker sets. If offset
performance is determined solely by
adaptive signals in genetic data, offsets
trained using *adaptive* markers should
have better performance than *all* or
neutral markers, and *all* markers should
have better performance than *neutral*
markers.

If the marker set has little impact on
offset performance, this could indicate
that offset methods are giving weight to
genome-wide signals present in the
data. Previously, some (e.g., Lachmuth,
Capblancq, Keller, et al., 2023; Lind et
al., 2024) have postulated that this
signal may be related to isolation by
environment ((IBE, i.e., when genetic
and environmental distances are
positively correlated, independent of
geographic distance; Wang &
Bradburd, 2014).

If IBE is driving patterns of offset
performance, we expect 1) performance
to be similar between offsets estimated
using *adaptive* markers and those
estimated using *neutral* markers; 2) a
greater proportion of variation in
performance to be explained by p_{cNeut}
than p_{cQTN} (from *Q2*); 3) a strong,
positive relationship between
performance and $LA_{\Delta SA}$; and 4) the
difference in IBE between two marker
sets to be positively correlated with the
difference in performance of two models
trained with those markers. We
measure IBE as the rank correlation
(Spearman's ρ) between population
pairwise F_{ST} (Weir & Cockerham, 1984)

and Euclidean climate distance of adaptive
environmental variables.

Q4 - What is the variation of model
performance across the landscape?

Within a landscape, offset methods may
not have high predictive performance at every
site or every environment. Understanding
variabil- ity in the predictive performance of
offset models across the landscape is
particularly relevant when offsets are used for
restoration or assisted gene flow initiatives
(i.e., ranking sources for a given site). If
predictive performance is variable across the
landscape, this may limit the usefulness of
genomic offsets for such purposes even if
model performance is validated in one
common garden. Using the *Adaptive*
Environment workflow, we visualized
variation of 1- and 2-trait within-landscape
performance with boxplots for each common
garden for each method and landscape. To
understand if variation in predictive
performance was a function of the model
quality, we investigated the relationship
between a model's performance variability
(i.e., standard deviation across 100 common
gardens) and the model's median
performance.

Q5 - How does the addition of non-adaptive
nuisance environments in training affect
performance?

In practice, the environments imposing
selection are rarely known *a priori*.
Additionally, the inclusion of environmental
measures that are not correlated with the
main axes of selection may reduce model
performance compared to models trained
using only causal environments. To

investigate the sensitivity of offset methods to environmental input we compared *Adaptive Environment* workflow models from 1-, 2-, and 6-trait simulations – where only the adaptive environment(s) are used in training (*0-nuisance*) – to models from the *Nuisance Environment* workflow trained with the same data but with the addition of nuisance environments (*N-nuisance*, where $N > 0$; Table 2).

We use nuisance environmental variables from Lotterhos (2023) that were real BioClim variables (*TSsd*, *PSsd*, and *ISO*) taken from British Columbia and Alberta, Canada, which have minimal correlation with causal environments and each other (Fig. S4). These three nuisance environments differ from previous implementations of such variables (Láurson et al. 2022) in that they are spatially autocorrelated whereas nuisance environments in Láruson et al. (2022) were not. For 1-trait scenarios, *Env2* was also used as a nuisance environmental variable.

If offset methods are unaffected by the addition of nuisance environmental variables, performance should not differ between *0-nuisance* and *N-nuisance* implementations. Finally, in empirical settings the set of adaptive environments are not known *a priori*. We also explored whether GF would rank adaptive environments higher than nuisance environments using weighted importance output from GF.

Q6 - How well do offset models extrapolate to novel environments outside the range of environments used in training?

Even if offset methods have high within-landscape performance, this does not directly address situations where future environmental conditions are vastly different from the environmental conditions used for training (i.e., novel environments). If performance decreases with increasing environmental novelty relative to training data, this raises questions about the utility of genomic offsets for predicting 1) relative *in situ* vulnerability of populations to future climate change, and 2) the relative suitability of populations to restoration sites that differ drastically than those used in training.

To understand if offset performance degrades with environmental novelty relative to training data, we predicted offset to 10 novel environmental scenarios for the 1-, 2-, and 6-trait simulations using the *Climate Novelty* workflow (Table 2). The novel environmental scenarios were a set of common garden environments, z_E, extending outward from the training populations and exceeding values observed on the landscape for all adaptive environmental variables (Supplemental Note S3). We represent these scenarios as standard deviations from the center of environmental values used in training: $z_E \in \{1.72, 2.35, 2.74, 3.13, 3.53, 3.92, 4.31, 4.70, 5.09, 5.48, 5.88\}$. Fitness in novel environments was estimated assuming that the phenotypic optimum continues to have a linear relationship with the environmental variable (Equation 3 in Lotterhos 2023).

3 | Results

Q1 - Which aspects of the past evolutionary history affect within-landscape performance of offset methods?

The ANOVA model (Eq. 1) indicated that the degree of local adaptation of the metapopulation ($LA_{\Delta SA}$) was the primary factor influencing offset performance, followed by common garden location, demography, and landscape (Table S1; Fig. S6). Within the simulations, $LA_{\Delta SA}$ was impacted by pleiotropy, the relative strength of selection, and landscape, (Fig. S7; see also Figs. S2A, S2B in Lotterhos, 2023), so there may be some confounding among these factors.

In line with the ANOVA model, the performance of specific offset methods generally increased with increasing $LA_{\Delta SA}$ (Fig. 2), but there were some interesting differences among methods. For instance, GF_{offset}, $LFFM2_{offset}$, RDA-uncorrected, and $RONA_{temp}$ all improved as $LA_{\Delta SA}$ increased, while RDA-corrected and $RONA_{Env2}$ showed relatively weaker relationships.

Across landscapes, offset methods generally had higher performance in *Stepping Stone - Clines* landscapes than *Stepping Stone - Mountain* landscapes (Fig. 2B) despite similar levels of $LA_{\Delta SA}$ (Fig. 2A). Offset methods also generally performed better in the two *Stepping Stone* landscapes than the *Estuary - Clines* landscape (Fig. 2B). However, there were some interactions between method and landscape (Fig. 2C). For instance, RDA-corrected performed better in the *Estuary - Clines* compared to the two *Stepping Stones* landscapes, while the RDA-uncorrected showed the opposite pattern: performance was higher in the two *Stepping Stones* landscapes compared to *Estuary - Clines*.

The performance of methods was similar across genic levels but increased slightly as the number of QTNs underlying adaptation became more polygenic (Fig. S8). Additionally, while demography primarily influenced population differentiation across the landscape with little impact on $LA_{\Delta SA}$ within simulations (Table S2 in Lotterhos 2023), migration breaks between populations and latitudinal clines in population size generally decreased offset performance for $LFMM2_{offset}$, GF_{offset}, and RDA- uncorrected (Fig. S9).

Q2 - How is offset performance affected by the proportion of clinal alleles in the data? (Q1B)

The sum of squares from Eq. 1 indicated that the proportion of clinal alleles did not account for meaningful variation in offset performance (Table S1). Even so, results from an ANOVA model with just the proportion of clinal loci as explanatory variables (Eq. 2) indicated that p_{cNeut} accounted for 4.14-9.65 times the variation than did p_{cQTN} for GF_{offset}, $LFMM2_{offset}$, and RDA-corrected. For GF_{offset} and RDA-uncorrected, $p_{cNeut,Env2}$ accounted for >16% of the sum of squares (Table S2, Fig. S10).

Overall, relationships between performance and p_{cNeut} (second column, Fig. S11) had stronger relationships than between performance and p_{cQTN} (first column, Fig.

S11). However, sometimes performance
increased with p_{cNeut} and sometimes it
decreased, depending on the method
(Fig. S11), indicating that each method
is differentially sensitive to clinal alleles
in the data. Ultimately, strong
population genetic structure along
environmental clines in 2-trait
simulations (Fig. S12) drove
relationships with p_{cNeut} (Fig. S13)
which in turn drove relationships with
performance (Fig. S14, Fig. S11).

Q3 - Is method performance driven by
causal loci or by genome-wide patterns
of Isolation-By-Environment? (Q2A)

Overall, 1- and 2-trait *Adaptive*
Environment models had relatively
similar performance among marker sets.
For instance, models trained using *all* or
neutral markers had similar
performance while models trained using
adaptive markers performed slightly
higher than the other sets. The median
increase in performance from *adaptive*
compared to *all* or *neutral* models was
less than 3%. In total, using *adaptive*
markers outperformed 68% of models
using *neutral* markers and 67% of
models using *all* markers, while 74% of
models using *all* markers outperformed
neutral models (Fig. 3A-C). For RDA-
corrected the *neutral* markers
performed slightly better than either
adaptive or *all* markers in 2-trait
evaluations (Fig. 3E). A*daptive*
markers from 6-trait evaluations
provided varied performance
advantages across methods (Fig. 4).

The *adaptive* marker sets had relatively
elevated levels of *IBE* compared to sets of
neutral or *all* markers in 1- and 2-trait
simulations, but levels of *IBE* were
nonetheless quite similar between marker sets
(Fig. S15). Consequently, performance of
models trained with *adaptive* markers
generally had stronger relationships with *IBE*
than $LA_{\Delta SA}$ but this was not the case for
models trained with either *all* or *neutral*
markers (Fig. S16).
Intriguingly, levels of *IBE* found within a
landscape (Fig. S17A) did not correspond to
the degree of $LA_{\Delta SA}$ that developed (Fig.
S17B). Even so, while *IBE* was generally
unrelated to $LA_{\Delta SA}$ across all simulations,
there were generally positive relationships
between *IBE* and $LA_{\Delta SA}$ when controlling for
the number of traits and differences in
strengths of selection (Fig. S18). As such, *IBE*
from *all* markers explained very little
variation in performance when added as a
factor to the ANOVA model from Eq. 1 (SC
02.02.01), but accounted for some variation in
ANOVA models with only $LA_{\Delta SA}$ and *IBE* as
explanatory variables (0-34% for *IBE* vs 0-
74% for $LA_{\Delta SA}$; Table S3). Except for RONA,
the differences in performance between two
models trained with different marker sets was
generally unrelated to the differences in *IBE*
between the two marker sets used to train the
models (Fig. S19).
Together these results indicate that while
higher degrees of local adaptation may lead to
increased levels of *IBE* in the genome, the
signal of *IBE* of input markers generally has
minimal and varied impact on performance
differences for the scenarios evaluated here.
Alternatively, the levels of *IBE* present in the
simulated genomes may exceed a minimum
threshold of *IBE*, beyond which differences in

performance between marker sets are minimized.

Q4 - What is the variation of model performance across the landscape? (Q3a)

All 1- and 2-trait models exhibited variation in the predictive performance across gardens within a landscape, from essentially no predictive performance to very high predictive performance (Fig. S20, Fig. S21, Fig. S22, Fig. S23). Variation in performance was also observed for 6-trait models (Fig. 4).

While there was variability in predictive performance of 1- and 2-trait models within each landscape, in many cases the best performing models had the lowest levels of performance variation (Figs. S24, S25, S26). Ultimately we found no strong indicator for predicting when a model will be highly variable. Indeed, while performance generally increased with $LA_{\Delta SA}$ (Fig. 2), variability in performance was not strongly related to the variability in deme-level LA on the landscape (Figs. S27, S28, S29). Despite $LA_{\Delta SA}$ driving performance more generally (from Q1), this indicates that variation in model performance across the landscape is not strongly driven by metapopulation levels of, nor deme-level variation in, LA.

Q5 - How does the addition of non-adaptive nuisance environments in training affect performance? (Q2B)

Training offset models with the addition of non-adaptive nuisance environmental variables generally reduced offset method performance (Fig. 5). This decline was most dramatic for offset trained on 1-trait simulations (Fig. 5A) compared to the decline observed for 2-trait (Fig. 5B) and 6-trait (Fig. 5C) simulations. The only instances for which median performance did not decrease monotonically with nuisance level were for 2-trait simulations evaluated with GF_{offset} (Fig. S30).

Overall, landscape had the most influence over performance differences due to non-adaptive nuisance environments (Fig. S30), whereas there was little difference across other simulation parameters (not shown except in SC 02.02.06). Even so, *adaptive* markers seemed to provide some advantages in the presence of nuisance environments, particularly for 1-trait datasets where the advantages were more substantial compared to 2-trait datasets (Fig. S31, Fig. S32).

In some cases, the rankings of weighted environmental importance output from GF ranked nuisance variables higher than at least one adaptive environment (Table S4). Across *1- and 2-trait N-nuisance* models trained with *all* markers, GF incorrectly ranked environmental drivers in 26.9% (133/495) of the cases. Rankings improved somewhat for models trained with *adaptive* markers, incorrectly ranking environmental variables in 20.6% (102/495) of the cases (Table S4).

Q6 - How well do offset models extrapolate to novel environments outside the range of environments used in training? (Q2C)

The datasets that had the greatest within-landscape performance (i.e., those with higher levels of $LA_{\Delta SA}$) were also those that experienced the steepest decline in

performance with increasing climate novelty (red shade, Fig. 6). Importantly, declines in performance for datasets with greater $LA_{\Delta SA}$ were not due to instances where all populations had zero fitness (and thus performance was undefined and manually set to 0; Supplemental Note S4, Fig. S33). Despite little change in the median performance for datasets with low levels of LA, most performance scores from these datasets were below Kendall's **τ**=0.5, and therefore had little predictive value in novelty scenarios.

Advantages of *adaptive* marker sets were much less prevalent across methods for *Climate Novelty* scenario performance than either *Adaptive Environment* or *Nuisance Environment* scenarios (Fig. S34).

4 | Discussion

Solutions are needed to mitigate the negative impacts of global change on biodiversity. In the last decade, genomic offset methods have been identified as a complement to other ecological forecasting models because they incorporate intraspecific variation (Keller & Fitzpatrick, 2015; Capblancq et al., 2020; Rellstab et al., 2021). Our evaluations show that offset methods are differentially impacted by both the evolutionary history of sampled populations as well as the decisions made during model training. Our analyses emphasize the importance of sampling locally adapted populations,

identifying the drivers underlying environmental selection pressures *a priori*, and restricting offset projections to climates similar to those used in training. Below, we discuss the implications of these findings towards restoration, conservation, and the management of biodiversity.

4.1 | *The importance of local adaptation*

A basic assumption of genomic offset methods is that the sampled populations are adapted to their local environment (Rellstab et al., 2016, 2021), but this assumption has not been formally tested. Our analyses show that indeed the degree of local adaptation ($LA_{\Delta SA}$) is one of the primary factors that determine model performance for most methods. A value of $LA_{\Delta SA} \sim 0.5$ indicates that fitness in demes is on average 50% higher in sympatry than allopatry. Values of $LA_{\Delta SA}$ represent the average deme-level magnitudes of $LA_{\Delta HA}$ and $LA_{\Delta LF}$ across the metapopulation (Blanquart et al., 2013). Previous metaanalyses of studies measuring local adaptation of natural populations have used different measures of LA from the ones we calculate here, but do show that some species evolve large fitness differences among populations (Hereford, 2009; Leimu & Fischer, 2008). Given the prevalence of LA found previously (Hereford, 2009; Leimu & Fisher, 2010), we may therefore expect some genomic offset methods to do reasonably well when local adaptation in the metapopulation is high (i.e., when $LA_{\Delta SA} > 0.5$).

4.2 | *The importance of the signals within genomic marker sets*

Because of the assumption that locally adapted populations will be necessary for

satisfactory model performance, initial
implementations of genomic offset
models focussed on putatively adaptive
markers where this signal may be
strongest (Keller & Fitzpatrick, 2015;
Rellstab et al., 2016). More recently,
investigators have varied the set of
markers used to train models but have
found little influence on performance
(Fitzpatrick et al., 2021; Lachmuth,
Capblancq, Keller, et al., 2023; Láruson
et al., 2022; Lind et al., 2024). Our
results are similar to previous
investigations, finding that the *adaptive*
marker sets provide minimal advantage
over *all* or *neutral* marker sets, but not
universally or by great margins.

One hypothesis put forth as to why
adaptive marker sets perform similarly
to all markers is that genome-wide data
captures sufficient signatures of IBE
(Lachmuth, Capblancq, Keller, et al.,
2023; Lind et al., 2024). Our analysis
found weak positive relationships
between performance and levels of *IBE*
within marker sets. Even so, and except
for RONA, there were no universal
relationships within methods between
the difference in *IBE* of marker sets and
the difference in performance of the
models trained with these markers.

While we found little impact of
levels of *IBE* on overall performance,
the way in which we measured IBE may
have masked causative relationships.
For instance, in our measure of IBE we
correlated environmental distance with
pairwise F_{ST}. In doing so, our measure
of IBE distills genetic distance down to
a single value from a large number of
loci, and gives less weight to loci with

rare alleles. In future studies, creating a
marker set by ranking loci by single-locus
measures of IBE offers another opportunity to
understand the impact of IBE on
performance. Such marker sets could be used
to compare to performance from putatively
adaptive marker sets or marker sets composed
of all or random loci. Empirical datasets will
also be able to specifically address
geographical distances while quantifying IBE
(e.g., Bradburd et al., 2013).

While measures of IBE are one signal
remaining to be explored in future analyses,
the proportion of clinal neutral loci within
marker sets was shown to impact
performance, sometimes being positively
related to performance and sometimes
negatively depending on the context. These
and other signals within data that could
improve or mislead offset models also warrant
further investigation.

4.3 | *The importance of adaptive environmental variables*

In empirical settings, the environ- mental
drivers of local adaptation are rarely known *a*
priori. Even so, our results emphasize the
importance of identifying these variables
before training offset models, as there were
often declines in performance between models
trained using only adaptive environmental
variables (*0-nuisance*) and those trained using
additional non-adaptive nuisance
environmental variables (*N-nuisance*).

The importance of identifying these
selective environments may be particularly
germane to two general empirical scenarios. In
the first empirical scenario,, sparsely sampling
an environmentally heterogeneous range may
enrich genetic signals (e.g., coincident

population structure) most correlated to environmental variables that maintain a gradient across this extent, and miss signals relevant to more local scales. In the second empirical scenario, identifying the environmental variables underlying selection is particularly important when a specific genomic offset method is ill-suited to differentiate importance among input variables. For instance, RDA (and therefore RDA$_{offset}$) assumes that the environmental variables used to build models are not collinear; (as implemented here; Capblancq & Forester, 2021; Legendre & Legendre, 2012). Because of this, empirical datasets must be limited to a subset of available environmental measures. The process of excluding environmental variables in this way may omit signals of adaptive drivers (particularly when true drivers are not well measured), or perhaps incorporate environmental variables that do not coincide with drivers of selection. In these cases, performance is likely to decline. As such, this may indicate that methods such as RDA$_{offset}$ are likely to perform worse in, or less uniformly across, realistic empirical settings than what our current findings suggest.

On the other hand, users of GF may be tempted to include a large number of environmental variables in training, hoping that GF can accurately attribute the correct environmental variation to adaptive genetic structure. Our results show that it is not necessarily the case that GF will give the highest importance values to the true adaptive environmental variables. Indeed, weighted feature importance scores from GF models still incorrectly ranked the adaptive environments below neutral environments in 20%-27% of the datasets, depending on which marker set was used. These importance values ultimately affect the model predictions. Including all available environmental variables may therefore negatively impact GF$_{offset}$ performance, and could have weakened overall performance in previous empirical evaluations that used a large number of environmental measures in training (e.g., Lind et al., 2024).

There are some differences between the nuisance environmental variables implemented here and those that have been implemented previously. For instance, Láruson et al. (2022) created nuisance variables by randomly sampling a multivariate normal distribution. In contrast to findings here, Láruson et al. (2022) found that model performance was relatively unaffected with the addition of nuisance variables. The minimal influence of nuisance variables on performance found by Láruson et al. (2022) may differ from the performance declines reported here because the nuisance variables we used were spatially autocorrelated, while those from Láruson et al. (2022) were not. Inclusion of nuisance variables that are spatially autocorrelated may mislead offset models more generally than variables with little spatial autocorrelation because of the spurious relationship between environmental structure and genetic structure.

4.4 | The effect of environmental novelty

While within-landscape performance generally increased with $LA_{\Delta SA}$, the datasets with the greatest levels of $LA_{\Delta SA}$ were also the datasets where performance declined most readily with climate novelty. This occurred because locally adapted metapopulations were under strong selection to be fine-tuned to their environment, and as a result most individuals suffered severe fitness declines with environmental change. In contrast, less locally adapted metapopulations were under weaker selection, and suffered less steep fitness declines with environmental change. This result highlights an interesting paradox: offset methods that have the highest performance in common garden transplants under current climates (because of strong local adaptation) may have the lowest performance in predicting "genomic vulnerability" as the range of climate variables become more novel compared to the ranges used in training the model.

Thus, genomic offset models are likely ill-suited for estimating fitness ranks of populations in environments that differ drastically from those used to train the models themselves. This is particularly relevant for applications of offset methods that attempt to estimate *in situ* risk of climate change to years or climate scenarios where the environment is expected to be increasingly novel. While climate novelty is often measured with respect to historical variability (e.g., Lotterhos et al., 2021; Mahony et al., 2017; Williams et al., 2007), indices of local climate change indicate that local environments in terrestrial systems could experience change in excess of three standard deviations relative to historic values (Williams et al., 2007). Similar indices in marine systems indicate potential for even greater novelty (Lotterhos et al., 2021). We observed performance declines below the analogous z_E=3.13 *Climate Novelty* scenario, indicating offset predictions will likely be inaccurate in many real-world climate change predictions. These issues are also germane to measures derived from offset values (Gougherty et al., 2021; Lachmuth, Capblancq, Keller, et al., 2023; Lachmuth, Capblancq, Prakash, et al., 2023), which currently do not consider the degree of climate novelty in the prediction (but see DeSaix et al., 2022).

Our results present a best-case scenario for predicting performance in novel environments, as in many cases there will be biological reasons as to why climate-fitness relationships will differ in future environments from relationships measured within the contemporary climate space (see Fig. 5 in Capblancq et al., 2020). For the simulations here, the relationship between contemporary and novel environments with fitness was the same.

4.5 | Genomic offsets in practice

Our evaluations show that genomic offset methods hold promise for predicting maladaptation to environmental change within a historical baseline, in metapopulations that evolve strong local adaptation. However, our analyses also emphasize the limits of these methods in some

1209 systems or scenarios. In practice, 1210 species that are locally adapted to 1211 measurable environmental variables 1212 will be best suited for offset methods 1213 when predicting the relative 1214 performance of populations in a 1215 contemporary common garden, but 1216 paradoxically these species may be least 1217 suited to using these methods to predict 1218 their vulnerability to novel climates.

1219 Together, these results indicate that 1220 some genomic offset methods may be 1221 suited to guide initiatives such as near- 1222 term assisted gene flow, where targeted 1223 restoration sites within a species range 1224 have climates that are similar to those 1225 used to train offset models. Even so, our 1226 results also show that the performance 1227 of these methods are often variable 1228 across a landscape, indicating that high 1229 performance at one site does not mean 1230 the offset model will perform well at 1231 another. While genomic offset methods

1232 may be suitable for assisted gene flow 1233 initiatives, they may be less suited for assisted 1234 migration programs where populations are 1235 moved outside of their native range and 1236 environments differ from training data.

1237 Before genomic offsets can be incorporated 1238 into management plans, considerable thought 1239 must be put into the sensitivity of model 1240 outcomes from input data (Lind et al., 2024), 1241 the uncertainty inherent in environmental or 1242 climate forecasts (Lachmuth, Capblancq, 1243 Keller, et al., 2023), as well as the degree of 1244 novelty of future climates (DeSaix et al., 2022, 1245 this study). While accurate predictions are 1246 limited for novel climates of the future, these 1247 offset methods could still be used to guide 1248 management in the intervening time in a 1249 stepwise manner where experiments can be 1250 used to validate model performance in 1251 practice. Using simulations tailored to the life 1252 history of target species also presents a 1253 promising avenue to understand limitations of 1254 these methods for specific management cases.

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Author Contributions

KEL received funding. KEL and BML conceptualized the project and methodology. With input, editing, and feedback from KEL, BML wrote code to train and evaluate offset models, created figures, curated coding and records for archiving, and wrote the manuscript.

Conflict of Interests

The authors declare no conflicts of interest.

Data Availability

We reference the analysis code in the text of our documents by designating Supplemental Code (SC) using a directory numbering system from our servers (as opposed to the order listed in the manuscript). Supplemental Code includes both executable scripts (*.R, *.py) as well as jupyter notebooks (*.ipynb). For example, for Script 3 in Directory 1, we refer to SC 01.03; for Notebook 5 in Subfolder 3 of Directory 2, we will refer to SC 02.03.05. Each Directory will be archived on Zenodo.org and include a citation below, which will also link to the GitHub repository. Notebooks are best viewed within a local jupyter or jupyter lab session (to enable cell output scrolling / collapsing), but can also be viewed at nbviewer.jupyter.org using the web link in the archive's README on GitHub. Analyses were carried out primarily using python v3.8.5 and R v3.5.1 and v4.0.3. Exact package and code versions are available at the top of each notebook. More information on coding workflows and coding environments can be found in Supplemental Note S1 and Supplemental Note S2.

All directories, notebooks, and scripts can be found on GitHub, and will be archived on Zenodo.
https://github.com/ModelValidationProgram/MVP-offsets

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1405

Figures for:

The limits of predicting maladaptation to future environments with genomic data

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Running Title: Limits of genomic offsets

Keywords: genomic offset, environmental change, climate change, assisted gene flow, genomic forecasting, restoration

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Method	abbr.	Multivariate?	Structure correction?
Gradient Forests[1]	GF_{offset}	Yes	No
Redundancy Analysis[2] with population structure correction	RDA-corrected	Yes	Yes, with axes loadings from PCA*
Redundancy Analysis[2] without population structure correction	RDA-uncorrected	Yes	No
Latent factor mixed model from Landscape and Ecological Association Studies R package[3]	$LFMM2_{offset}$	Yes	Yes, with latent factors
Risk Of Non-Adaptedness[4]	RONA	No	No

* principal component analysis

Table 1 Genomic offset methods used for evaluation. Genomic offset methods differ in their capability to use multivariate environmental data in training as well as whether a correction for population genetic structure is applied. Superscripts apply to the following reference citations: 1 - Fitzpatrick & Keller, 2015; 2 - Capblancq & Forester, 2021; 3 - Gain & François, 2021; 4 - Rellstab et al., 2016.

Workflow	n_{traits}	(1) Simulations Levels (replicates per level)	(3, 7) Environmental Data	Training and Prediction?	(6) Within-landscape Evaluation?	(9) Total Performance Evaluations
Adaptive Environment (AE)	1-trait	45 (10)	temp	Yes	Yes	675,000
	2-trait	180 (10)	temp + Env2			3,240,000
	6-trait	1 (1)	MAT + MTwetQ + MTDQ + PDM + PwarmQ + PWM			3,000
Nuisance Environment (NE)**	1-trait	45 (1)	[AE$_{1\text{-trait}}$ environments + Env2] $_{+/-}$ [ISO + PSsd] $_{+/-}$ [TSsd]	Yes	Yes	175,500
	2-trait	180 (1)	[AE$_{2\text{-trait}}$ environments + ISO + PSsd] $_{+/-}$ [TSsd]			432,000*
	6-trait	1 (1)	[AE$_{6\text{-trait}}$ environments + ISO + PSsd + TSsd]			1,200*
Climate Novelty (CN)†	1-trait	45 (10)	AE$_{1\text{-trait}}$ environments	Prediction only (using AE training models)	No	64,800*
	2-trait	180 (10)	AE$_{2\text{-trait}}$ environments			259,200*
	6-trait	1 (1)	AE$_{6\text{-trait}}$ environments			144*

* excludes RONA
** The set of population values for each unique nuisance environment were the same across traits and landscapes
† includes evaluation of climate center and eleven *Climate Novelty* scenarios

Table 2 Workflows used to process simulation data for the evaluation of genomic offset methods. Numbers given in column names refer to locations in schematic of Fig. 1. The *Adaptive Environment* workflow processes all population data from 1- and 2-trait (example shown in Fig. 1) as well as 6-trait simulations using only adaptive environmental variables in training, and evaluates performance in each garden on the metapopulation landscape. The *Nuisance Environment* workflow processes 1-, 2-, and 6-trait simulations similarly to the *Adaptive Environment* workflow, except in addition to adaptive environmental variables used in training, non-adaptive (i.e., nuisance) environmental variables are also used - each bracketed set of environmental variables indicate a distinct nuisance level (e.g., "*1-trait 1-nuisance*" = [AE$_{1\text{-trait}}$ environments + Env2] and "*1-trait 4-nuisance*" = [AE$_{1\text{-trait}}$ environments + Env2 + ISO + PSsd + TSsd]). The *Climate Novelty* workflow uses trained models from the *Adaptive Environment* workflow (Fig. 1A-5) and evaluates offset in 11 novel environments relative to the range of environments used in training. See Supplemental Note S3 for details regarding the choice of *Climate Novelty* environmental values and visualizations of climate data in principal component space. See Supplemental Notes S1-S2 for descriptions of coding workflows. Counts of evaluations were tabulated in SC 02.10.01.

Figure 1 Analysis of 1-, 2-, and 6-trait simulations included three main phases: A) model training, B) model prediction, and C) evaluation of models. The *Adaptive Environment* workflow is shown as an example of the processing of 1- and 2-trait simulation data for genomic offset evaluation. In total, three general workflows are used to evaluate genomic offset methods (Table 2). Subpanels of this schematic are numbered for referencing in Table 2 and the main text.

(Fig. 2)

(Fig. 2 continued)

Figure 2   Predictive performance of genomic offset models (y-axes) is driven by the degree of local adaptation (A) and the spatial patterns of adaptive environments across the landscape (B, C). For each model, a median value from performance scores from 100 common gardens is shown for A and B; C shows scores across all common gardens for each model (note that y-axes are inverted, as more negative values have higher performance). Data included in these figures was processed through the *Adaptive Environment* workflow but only includes models trained using 2-trait simulations and *all* loci. Code to create (A) and (B) can be found in SC 02.02.02; code to create (C) can be found in SC 02.02.01.

(Fig. 3)

Figure 3 Comparison of marker choice across genomic offset methods for 1- and 2-trait simulations. A-C are scatterplots of pairwise comparisons of performance between marker sets (histograms in each margin) from both 1- and 2-trait models where density of points is indicated by color in legend (note color scale is different for each figure to accentuate patterns in data). D-E are boxplots from the same data in A-C separated by individual traits. Data included in these figures is from all 1- and 2-trait models from the *Adaptive Environment* workflow. Code to create these figures can be found in SC 02.02.03.

Figure 4   Comparison of marker choice across genomic offset methods for the 6-trait simulation. A-C are scatterplots of pairwise comparisons of performance between marker sets (RONA is not shown, except in SN 02.05.10). D are boxplots from the same data in A-C (RONA$_{6\text{-traits}}$ is the combined performance across all six environmental models). Data included in this figure is from the 6-trait models processed through the *Adaptive Environment* workflow. Note there is only one 6-trait replicate, and variation within figures represents the performance across 100 common gardens for each method. Code to create these figures can be found in SN 02.05.10.

(Fig. 5)

Figure 5   Effect of non-adaptive nuisance environmental variables on offset performance. Shown are evaluations of offsets from 1- and 2-trait models trained using only adaptive environments (*0-nuisance*) or with adaptive environments and the addition of *N*>0 non-adaptive environmental variables (*N-nuisance*). RONA is not shown because it is univariate with respect to environmental variables. Nuisance variables are listed in Table 2. Code to create figures can be found in SC 02.02.06 and SC 02.02.08.

Figure 6   Performance decays with climate novelty relative to training data. Shown is model performance (y-axes) across methods at climate center (A) and across common gardens each representing increasing degrees of climate novelty relative to training data (x-axis of B) where all 100 populations have been transplanted. The standard deviation values (x-axis, B) are applicable to all environments for all landscapes except for *Env2* in the *Stepping Stone - Mountain* landscape; the corresponding standard deviations are 1.55, 2.12, 2.47, 2.82, 3.18, 3.53, 3.88, 4.24, 4.60, 4.95, 5.3. When fitness for all transplanted individuals was zero, a model's performance was undefined and manually set to 0; no method predicted a single offset value for all populations in these situations. Setting undefined performance to 0 did not substantially impact patterns between performance and climate novelty, and is explored in Supplemental Text S3. Data included in this figure are from models trained using 1- and 2-trait simulations from the *Climate Novelty* workflow, which excludes both $RONA_{temp}$ and $RONA_{Env2}$. Code used to create this figure can be found in SC 02.04.05.