The limits of predicting maladaptation to future environments with genomic data

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Abstract

Anthropogenically driven changes in land use and climate patterns pose unprecedented challenges to species persistence. To understand the extent of these impacts, genomic offset methods have been used to forecast maladaptation of natural populations to future environmental change. However, while their use has become increasingly common, little is known regarding their predictive performance across a wide array of realistic and challenging scenarios. Here, we evaluate four offset methods (Gradient Forests, the Risk-Of-Non-Adaptedness, redundancy analysis, and LFMM2) using an extensive set of simulated datasets that vary demography, adaptive architecture, and the number and spatial patterns of adaptive environments. For each dataset, we train models using either *all*, *adaptive*, or *neutral* marker sets and evaluate performance using *in silico* common gardens by correlating known fitness with projected offset. Using over 4,850,000 of such evaluations, we find that 1) method performance is largely due to the degree of local adaptation across the metapopulation ($LA_{\Delta\text{SA}}$), 2) *adaptive* marker sets provide minimal performance advantages, 3) within-landscape performance is variable across gardens and declines when offset models are trained using additional non-adaptive environments, and 4) despite (1), performance declines more rapidly in novel climates for metapopulations with higher $LA_{\Delta\text{SA}}$ than lower $LA_{\Delta\text{SA}}$. We discuss the implications of these results for management, assisted gene flow, and assisted migration.

1 | Introduction

The impacts of climate change, habitat loss, and extreme weather events pose urgent challenges to the management of species, communities, habitats, and ecosystem services (Bonan, 2008; Doney et al., 2012; Hoegh-Guldberg & Bruno, 2010). Traditional methods used to infer environmental suitability, such as reciprocal transplants and common gardens, require time and resources that may not be available or feasible for many organisms of management concern, particularly for long-lived organisms where reproductive stages occur after several decades of development. Ecological forecasting models have therefore become increasingly germane to support environmental decision making by managers across both terrestrial and marine systems.

In the context of population viability in the face of environmental change, many of these models rely on theoretical expectations that the limits of species' distributions are primarily determined by the distribution of environmental conditions (e.g., Good 1931), and that occupancy of highly suitable habitat enables increased abundance through greater survival and reproduction (i.e., fitness) of individuals (Brown, 1984). Such methods, termed species distribution models or ecological niche models (see Elith & Leathwick, 2009 for a discussion on terminology) are correlative approaches that are often used to predict (relative) habitat suitability for a single species (Lee-Yaw et al., 2022). This information is used to understand potential impacts on the species from future climate change. However, these methods often ignore aspects of the species' evolutionary history that could be important for predicting long-term population persistence, such as the environmental drivers of local adaptation or spatial patterns of adaptive genetic variation (Waldvogel et al., 2020).

Subsequent methods, termed genomic offsets (reviewed in Capblancq et al., 2020; Rellstab et al., 2021), have attempted to address these shortcomings by modeling relationships between environmental and genetic variation to predict maladaptation of natural populations to either future climates *in situ*, or to predict the relative suitability of these populations for the specific environment of a restoration site. Empirical attempts to confirm predictions from genomic offset models are rare and, compared to attempts *in silico* (Láruson et al., 2022), have found relatively weaker relationships between predicted maladaptation to common garden climates and the measurement of phenotypic proxies for fitness from individuals grown in these same environments (e.g., Capblancq & Forester, 2021; Fitzpatrick et al., 2021; Lind et al., 2024). Even so, these empirical results have consistently shown the expected negative relationship between predicted offset and common garden performance. Further, many of these studies found that genomic offsets often perform better than climate or geographic distance alone (e.g., Capblancq & Forester, 2021; Fitzpatrick et al., 2021; Láruson et al., 2022; Lind et al., 2024).

Across empirical and *in silico* studies, little difference in performance was found between

models trained using only adaptive
markers (i.e., known *in silico*, or
candidates from empirical genotype-
environment [GEA] associations) and
those chosen at random, suggesting that
genome-wide data may be sufficient to
capture signals relevant to
environmental adaptation.

Together, these results suggest that
genomic offset methods may provide
valuable insight for management. Little
is known, however, about how robust
these methods are across a wide array
of realistic empirical scenarios, nor the
extent to which independent methods
will arrive at similar conclusions when
analyzing the same data. Indeed,
concerns regarding the accuracy of
ecological forecasting models present a
primary limitation towards
incorporating inferences from these
models into management (Clark et al.,
2001; Schmolke et al., 2010) and
genomic offset models are no exception.
Major questions still remain about how
performance is affected by aspects of
the evolutionary history of sampled
populations, the type of signals in
putatively ideal datasets that may
mislead offset inference, the importance
of identifying environmental drivers of
local adaptation *a priori*, and the
consistency of predictive performance
across contemporary environmental
space. Finally, because novel climates
with no recent analog are expected to
increase in the future (Lotterhos et al.,
2021; Mahony et al., 2017) there is also
uncertainty regarding the performance
of forecasting models when predictions
are made to environments that

drastically differ from those used to train and
build the models themselves (Fitzpatrick et
al., 2018; Lind et al., 2024).

While much uncertainty remains
regarding the predictive performance of
genomic offsets, the domain of applicability
(i.e., the circumstances under which a method
is acceptably accurate) for these methods can
be more precisely defined using simulated
data (Lotterhos et al., 2022). Simulated data,
where there is no error in the estimation of
allele frequencies, environmental variables,
individual fitness, or the knowledge regarding
the drivers of local adaptation, present ideal
circumstances for understanding the limits of
genomic offsets and the circumstances under
which data from naturally occurring taxa will
provide useful inference. To provide relevant
inference regarding the domain of
applicability, simulations should capture the
complexities of empirical data with biological
realism (e.g., clinal or patchy environments),
present contrasting cases of differing scenarios
while controlling for important features of the
data (e.g., varying population connectivity
but controlling for mean differentiation), and
challenge methods using adversarial scenarios
that capture extreme characteristics of
empirical data (e.g., prediction to novel
environments with no current analog
available for model training; Lotterhos et al.
2022).

Here, we use a wide array of previously
published biologically realistic, contrasting,
and adversarial simulations from Lotterhos
(2023) in an attempt to more precisely define
the limits of predictive performance of five
genomic offset methods (Table 1): Gradient
Forests (GF$_{offset}$; *sensu* Fitzpatrick & Keller,
2015), the Risk Of Non-Adaptedness (RONA,
Rellstab et al., 2016), Latent Factor Mixed

Models (LFMM2$_{\text{offset}}$, *sensu* Gain &
François, 2021, and redundancy analysis
(RDA$_{\text{offset}}$, *sensu* Capblancq & Forester,
2021). The main goal of this study was
to understand how the evolutionary and
experimental parameters used in the
training and evaluation of offset
methods affect the accuracy of the
methods' projections of maladaptation
under ideal empirical scenarios (i.e.,
using data with no inherent error).
Using these scenarios, we ask the
following six questions: 1) Which
aspects of the past evolutionary history
affect performance of offset methods? 2)
How is offset performance affected by
the proportion of loci with clinal alleles
in the data? 3) Is method performance
driven by causal loci or by genome-wide
patterns of isolation-by-environment?
4) What is the variation of model
performance across the landscape? 5)
How does the addition of non-adaptive
nuisance environments in training affect
performance? 6) How well do offset
models extrapolate to novel
environments outside the range of
environmental values used in training?

222 2 | Methods

Throughout this manuscript we will
be citing code used to carry out specific
analyses in-line with the text.
Supplemental Notes S1-S2 outlines and
describes the sets of scripts or, most
often, jupyter notebooks, used to code
analyses. Scripts and notebooks are
both referenced as Supplemental Code
(SC) using a directory numbering
system (e.g., SC 02.05). More

information regarding the numbering system,
archiving, and software versions can be found
in the Data Availability section.

236 2.1 | Explanation of Simulations and
237 Training Data

To train offset methods we used single
nucleotide polymorphism (SNP) and
environmental data from a set of previously
published simulations (225 levels with 10
replicates each) of a Wright-Fisher
metapopulation of 100 demes on a 10 x 10 grid
evolving across a heterogeneous landscape
(Lotterhos, 2023). Each dataset was
simulated under a combination of the
following four evolutionary parameters: i)
three landscapes (10 populations x 10
populations) that varied in vicariance and
environmental gradients (*Estuary - Clines*;
Stepping Stone - Clines; and *Stepping Stone -
Mountain*), ii) five demographies that varied
population size and migration rates across the
landscape, iii) three genic levels that varied in
the effect size and number of mutations
underlying adaptation (mono-, oligo-, and
polygenic), and iv) five pleiotropy levels that
varied the number of quantitative traits under
locally stabilizing selection ($n_{\text{traits}} \in \{1, 2\}$),
presence of pleiotropy (when $n_{\text{traits}} = 2$), and
variability of selection strength across
individual traits (see Fig. 1 in Lotterhos
2023).
The adaptive trait(s) were under selection
by a different environmental variable, where
the optimum trait value was given by the
local environment on the landscape. The
adaptive trait(s) undergoing selection
responded to either a latitudinal temperature
gradient (*temp*; $n_{\text{traits}} = 1$), or to both *temp*
and a longitudinal "*Env2*" gradient ($n_{\text{traits}} =$

2). *Env2* represented distinct biological analogies depending on the context: in the *Stepping Stone - Mountain* landscape *Env2* was analogous to elevation (e.g., as with tree species), whereas in the *Estuary - Clines* landscape the *Env2* environment was analogous to gradients of salinity within coastal inlets connected only by the outer marine (ocean) environment (e.g., as with stickleback or oyster species).

Twenty independent linkage groups were simulated. Of these, mutations that had effects on one or more phenotypes under selection (i.e., quantitative trait nucleotides, QTNs) were allowed to evolve on only ten linkage groups, and neutral mutations were added to all 20 linkage groups with tree sequencing (for details see Lotterhos 2023). Adaptive traits were determined additively by effects of QTNs.

In all simulations, phenotypic clines evolved between each trait and the selective environment (Lotterhos, 2023), where populations became locally adapted to their environment, measured at the metapopulation level as the mean difference of demes in sympatry minus allopatry ($LA_{\Delta SA}$, Blanquart et al., 2013). $LA_{\Delta SA}$ equates to the average levels of local adaptation at the deme level which can be calculated for each deme by both home-away ($LA_{\Delta HA}$) and local-foreign ($LA_{\Delta LF}$) measures.

These simulations represent a wide array of realistic, contrasting, and adversarial scenarios in which we could more precisely define the domain of applicability of offset methods. For instance, in the *Stepping Stone - Mountain* landscape, geographic distance and environmental distance were not strongly correlated, whereas in the *Stepping Stone - Clines* and *Estuary - Clines* they were. Additionally, the proportion of mutations with monotonic frequency gradients (i.e., allelic clines) underlying local adaptation varied across the simulated datasets (Lotterhos, 2023), which may also affect offset performance. These simulations also presented demographic scenarios in which selection was confounded with genetic drift or population genetic structure.

For each simulation, ten individuals were randomly chosen per population for a total of 1000 individuals. Individual genotypes were coded as counts of the derived allele. Alleles with global minor allele frequency (MAF) $<$ 0.01 were removed. Using all 100 populations, population-level derived allele frequencies and current environmental values were used as input to train offset methods.

In addition to the 2250 simulated Wright-Fisher datasets (225 levels * 10 replicates), we also included a non-Wright-Fisher case with range expansion from three refugia and secondary contact (Lotterhos 2023). This simulation evolved variable degrees of admixture across the landscape. Six moderately polygenic environmental traits ($n_{\text{traits}} = 6$) were under selection from the environment. Environments were based on six weakly correlated environmental variables taken from Bioclim environmental measures of western Canada. The simulation evolved local adaptation at all six traits with unconstrained pleiotropy. For more details on simulations, see (Lotterhos, 2023).

2.2 | Evaluation of Offset Methods

We investigated the performance of five implementations of four genomic offset methods (Table 1): GF_{offset}, RDA_{offset}, $LFMM2_{offset}$, and RONA. While GF_{offset}, RDA_{offset}, and $LFMM2_{offset}$ can use multivariate environmental data to train models, RONA can only account for a single environment at one time (Table 1). Additionally, while GF_{offset} and RONA do not apply correction for population genetic structure, $LFMM2_{offset}$ does by default, and structure correction with RDA_{offset} is optional. We thus evaluate RDA_{offset} with (RDA-corrected) and without (RDA-uncorrected) population genetic structure correction (Table 1). For additional specifics related to the implementation of each offset method, see Supplemental Note S1.1-S1.4 and Fig. S1, Fig. S2, Fig. S3.

We varied construction of genomic offset training datasets for each replicate of the 1-, 2-, and 6-trait simulations by varying the marker set used in model training (Fig. 1A, Table 2; see *Q3* below). Each model was trained using genetic and environmental data from all 100 populations. The environmental variables used were only those imposing selection pressure. We predict offset from each model for each population to all 100 within-landscape common gardens from a full factorial *in silico* reciprocal transplant design (Fig. 1B). For each common garden, we quantified offset model performance as the rank correlation (Kendall's **τ**) between the population mean fitness (averaged over sampled individuals, Equation 3 in Lotterhos 2023) and projected population offset (Fig. 1C). Strong negative relationships between fitness and predicted offset indicate higher performance of the method (note y-axes of Kendall's **τ** are inverted within figures to show more intuitive performance relationships, Fig. 1C-11). We refer to the preceding processing of data as the *Adaptive Environment* workflow (Fig. 1, Table 2).

To explore the impact of the choice of environmental variables used (see *Q5* below), we used a workflow similar to the *Adaptive Environment* workflow, except instead of using only adaptive environmental variables, we used additional non-adaptive (i.e., nuisance) environmental variables in training and prediction (second row, Table 2). These nuisance variables had relatively weak correlation structure with adaptive environments and each other (Fig. S4). We refer to each of these nuisance levels by the number of traits under selection and the number of nuisance environments used (e.g., *1-trait 3-nuisance*). We refer to this workflow as the *Nuisance Environment* workflow .

Finally, to contrast with within- landscape evaluations, we explored predictive performance of *Adaptive Environment* offset models in novel environments that are beyond the range of values of those used in training (see *Q6* below). In these novelty cases, we use 11 common gardens, each progressively more distant from the average environment used in training (i.e., climate center) and evaluate performance in each garden. We refer to this workflow as the *Climate Novelty* workflow. See Supplemental Note S3 and Fig. S5 for details regarding the choice of environmental values for novelty scenarios.

2.3 | Study Questions

Q1 - Which aspects of the past evolutionary history affect within-landscape performance of offset methods?

For each offset method, we used a fixed-effects type II ANOVA model to test for significant differences in the performance from 2-trait *Adaptive Environment* models trained using *all* markers using the following factors: landscape (*Estuary - Clines, Stepping Stone - Clines, Stepping Stone - Mountain*), demography (five levels describing population size and migration patterns across the landscape), genic level of architecture (three levels from oligogenic to polygenic), presence or absence of pleiotropy, proportion of loci with clinal allele frequencies (as defined in Lotterhos, 2023), degree of local adaptation (*ΔSA*), and common garden ID. Specifically,

$$Y_{ij} = L_i + D_i + GL_i + P_i + p_{cQTN,t,i} + p_{cNeut,t,i} + p_{cQTN,Env2,i} + p_{cNeut,Env2,i} + LA_{\Delta SA,i} + G_j$$
(Eq. 1)

where Y_{ij} is the within-landscape performance (Kendall's τ) of a single method for garden j in simulation i, with factors for landscape (L), demography (D), genic level (GL), presence of pleiotropy (P), proportion of QTN or neutral alleles with *temp* clines (respectively $p_{cQTN,t,i}$ and $p_{cNeut,t,i}$), proportion of QTN or neutral alleles with *Env2* clines (respectively $p_{cQTN,Env2,i}$ and $p_{cNeut,Env2,i}$), degree of local adaptation ($LA_{\Delta SA}$), and garden ID (G). The first four factors are illustrated in Fig. 1 of Lotterhos (2023).

Q2 - How is offset performance affected by the proportion of clinal alleles in the data? (Q1B)

Clinal alleles (i.e., alleles with monotonic gradients in frequency across space) that covary with environmental clines could be weighted more heavily in offset models that emphasize loci whose allele frequencies explain significant variation across local environmental values. Using 2-trait models trained using *all* markers from the *Adaptive Environment* workflow, we used an ANOVA model (Eq. 2) to test the hypothesis that clinal alleles differentially impact model performance, independent from the other factors from Eq. 1:

$$Y_{ij} = p_{cQTN,t,i} + p_{cNeut,t,i} + p_{cQTN,Env2,i} + p_{cNeut,Env2,i}$$
(Eq. 2)

The factors representing clinal alleles in Eq. 2 are the same as those in Eq. 1.

Q3 - Is method performance driven by causal loci or by genome-wide patterns of Isolation By Environment? (Q2A)

For each offset method and workflow, we varied the set of input markers for 1-, 2- and 6-trait simulations that were used in training to determine if performance of a method was driven by properties of the evolutionary forces shaping genotype-environment relationships: 1) *adaptive* markers (i.e., QTNs with effects on at least one trait), 2) *neutral* markers (SNPs on linkage groups without QTNs), and 3) *all* markers (union of *adaptive* and *neutral*

markers, as well as non-QTN markers
on the same linkage groups as QTNs).
Only loci that passed MAF filtering
were included in marker sets. If offset
performance is determined solely by
adaptive signals in genetic data, offsets
trained using *adaptive* markers should
have better performance than *all* or
neutral markers, and *all* markers should
have better performance than *neutral*
markers.

If the marker set has little impact on
offset performance, this could indicate
that offset methods are giving weight to
genome-wide signals present in the
data. Previously, some (e.g., Lachmuth,
Capblancq, Keller, et al., 2023; Lind et
al., 2024) have postulated that this
signal may be related to isolation by
environment ((IBE, i.e., when genetic
and environmental distances are
positively correlated, independent of
geographic distance; Wang &
Bradburd, 2014).

If IBE is driving patterns of offset
performance, we expect 1) performance
to be similar between offsets estimated
using *adaptive* markers and those
estimated using *neutral* markers; 2) a
greater proportion of variation in
performance to be explained by p_{cNeut}
than p_{cQTN} (from *Q2*); 3) a strong,
positive relationship between
performance and $LA_{\triangle SA}$; and 4) the
difference in IBE between two marker
sets to be positively correlated with the
difference in performance of two models
trained with those markers. We
measure IBE as the rank correlation
(Spearman's ρ) between population
pairwise F_{ST} (Weir & Cockerham, 1984)

and Euclidean climate distance of adaptive
environmental variables.

Q4 - What is the variation of model
performance across the landscape? (Q3a)

Within a landscape, offset methods may
not have high predictive performance at every
site or every environment. Understanding
variabil- ity in the predictive performance of
offset models across the landscape is
particularly relevant when offsets are used for
restoration or assisted gene flow initiatives
(i.e., ranking sources for a given site). If
predictive performance is variable across the
landscape, this may limit the usefulness of
genomic offsets for such purposes even if
model performance is validated in one
common garden. Using the *Adaptive*
Environment workflow, we visualized
variation of 1- and 2-trait within-landscape
performance with boxplots for each common
garden for each method and landscape. To
understand if variation in predictive
performance was a function of the model
quality, we investigated the relationship
between a model's performance variability
(i.e., standard deviation across 100 common
gardens) and the model's median
performance.

Q5 - How does the addition of non-adaptive
nuisance environments in training affect
performance? (Q2B)

In practice, the environments imposing
selection are rarely known *a priori.*
Additionally, the inclusion of environmental
measures that are not correlated with the
main axes of selection may reduce model
performance compared to models trained
using only causal environments. To

investigate the sensitivity of offset methods to environmental input we compared *Adaptive Environment* workflow models from 1-, 2-, and 6-trait simulations – where only the adaptive environment(s) are used in training (*0-nuisance*) – to models from the *Nuisance Environment* workflow trained with the same data but with the addition of nuisance environments (*N-nuisance*, where $N > 0$; Table 2).

We use nuisance environmental variables from Lotterhos (2023) that were real BioClim variables (*TSsd*, *PSsd*, and *ISO*) taken from British Columbia and Alberta, Canada, which have minimal correlation with causal environments and each other (Fig. S4). These three nuisance environments differ from previous implementations of such variables (Láurson et al. 2022) in that they are spatially autocorrelated whereas nuisance environments in Láruson et al. (2022) were not. For 1-trait scenarios, *Env2* was also used as a nuisance environmental variable.

If offset methods are unaffected by the addition of nuisance environmental variables, performance should not differ between *0-nuisance* and *N-nuisance* implementations. Finally, in empirical settings the set of adaptive environments are not known *a priori*. We also explored whether GF would rank adaptive environments higher than nuisance environments using weighted importance output from GF.

Q6 - How well do offset models extrapolate to novel environments outside the range of environments used in training? (Q2C)

Even if offset methods have high within-landscape performance, this does not directly address situations where future environmental conditions are vastly different from the environmental conditions used for training (i.e., novel environments). If performance decreases with increasing environmental novelty relative to training data, this raises questions about the utility of genomic offsets for predicting 1) relative *in situ* vulnerability of populations to future climate change, and 2) the relative suitability of populations to restoration sites that differ drastically than those used in training.

To understand if offset performance degrades with environmental novelty relative to training data, we predicted offset to 10 novel environmental scenarios for the 1-, 2-, and 6-trait simulations using the *Climate Novelty* workflow (Table 2). The novel environmental scenarios were a set of common garden environments, z_E, extending outward from the training populations and exceeding values observed on the landscape for all adaptive environmental variables (Supplemental Note S3). We represent these scenarios as standard deviations from the center of environmental values used in training: $z_E \in \{1.72, 2.35, 2.74, 3.13, 3.53, 3.92, 4.31, 4.70, 5.09, 5.48, 5.88\}$. Fitness in novel environments was estimated assuming that the phenotypic optimum continues to have a linear relationship with the environmental variable (Equation 3 in Lotterhos 2023).

3 | Results

Q1 - Which aspects of the past evolutionary history affect within-landscape performance of offset methods?

The ANOVA model (Eq. 1) indicated that the degree of local adaptation of the metapopulation ($LA_{\Delta SA}$) was the primary factor influencing offset performance, followed by common garden location, demography, and landscape (Table S1; Fig. S6). Within the simulations, $LA_{\Delta SA}$ was impacted by pleiotropy, the relative strength of selection, and landscape, (Fig. S7; see also Figs. S2A, S2B in Lotterhos, 2023), so there may be some confounding among these factors.

In line with the ANOVA model, the performance of specific offset methods generally increased with increasing $LA_{\Delta SA}$ (Fig. 2), but there were some interesting differences among methods. For instance, GF$_{offset}$, LFFM2$_{offset}$, RDA-uncorrected, and RONA$_{temp}$ all improved as $LA_{\Delta SA}$ increased, while RDA-corrected and RONA$_{Env2}$ showed relatively weaker relationships.

Across landscapes, offset methods generally had higher performance in *Stepping Stone - Clines* landscapes than *Stepping Stone - Mountain* landscapes (Fig. 2B) despite similar levels of $LA_{\Delta SA}$ (Fig. 2A). Offset methods also generally performed better in the two *Stepping Stone* landscapes than the *Estuary - Clines* landscape (Fig. 2B). However, there were some interactions between method and landscape (Fig. 2C). For instance, RDA-corrected performed better in the *Estuary - Clines* compared to the two *Stepping Stones* landscapes, while the RDA-uncorrected showed the opposite pattern: performance was higher in the two *Stepping Stones* landscapes compared to *Estuary - Clines*.

The performance of methods was similar across genic levels but increased slightly as the number of QTNs underlying adaptation became more polygenic (Fig. S8). Additionally, while demography primarily influenced population differentiation across the landscape with little impact on $LA_{\Delta SA}$ within simulations (Table S2 in Lotterhos 2023), migration breaks between populations and latitudinal clines in population size generally decreased offset performance for LFMM2$_{offset}$, GF$_{offset}$, and RDA- uncorrected (Fig. S9).

Q2 - How is offset performance affected by the proportion of clinal alleles in the data? (Q1B)

The sum of squares from Eq. 1 indicated that the proportion of clinal alleles did not account for meaningful variation in offset performance (Table S1). Even so, results from an ANOVA model with just the proportion of clinal loci as explanatory variables (Eq. 2) indicated that p_{cNeut} accounted for 4.14-9.65 times the variation than did p_{cQTN} for GF$_{offset}$, LFMM2$_{offset}$, and RDA-corrected. For GF$_{offset}$ and RDA-uncorrected, $p_{cNeut,Env2}$ accounted for >16% of the sum of squares (Table S2, Fig. S10).

Overall, relationships between performance and p_{cNeut} (second column, Fig. S11) had stronger relationships than between performance and p_{cQTN} (first column, Fig.

S11). However, sometimes performance
increased with p_{cNeut} and sometimes it
decreased, depending on the method
(Fig. S11), indicating that each method
is differentially sensitive to clinal alleles
in the data. Ultimately, strong
population genetic structure along
environmental clines in 2-trait
simulations (Fig. S12) drove
relationships with p_{cNeut} (Fig. S13)
which in turn drove relationships with
performance (Fig. S14, Fig. S11).

Q3 - Is method performance driven by
causal loci or by genome-wide patterns
of Isolation-By-Environment? (Q2A)

Overall, 1- and 2-trait *Adaptive*
Environment models had relatively
similar performance among marker sets.
For instance, models trained using *all* or
neutral markers had similar
performance while models trained using
adaptive markers performed slightly
higher than the other sets. The median
increase in performance from *adaptive*
compared to *all* or *neutral* models was
less than 3%. In total, using *adaptive*
markers outperformed 68% of models
using *neutral* markers and 67% of
models using *all* markers, while 74% of
models using *all* markers outperformed
neutral models (Fig. 3A-C). For RDA-
corrected the *neutral* markers
performed slightly better than either
adaptive or *all* markers in 2-trait
evaluations (Fig. 3E). A*daptive*
markers from 6-trait evaluations
provided varied performance
advantages across methods (Fig. 4).

The *adaptive* marker sets had relatively
elevated levels of *IBE* compared to sets of
neutral or *all* markers in 1- and 2-trait
simulations, but levels of *IBE* were
nonetheless quite similar between marker sets
(Fig. S15). Consequently, performance of
models trained with *adaptive* markers
generally had stronger relationships with *IBE*
than $LA_{\Delta SA}$ but this was not the case for
models trained with either *all* or *neutral*
markers (Fig. S16).
Intriguingly, levels of *IBE* found within a
landscape (Fig. S17A) did not correspond to
the degree of $LA_{\Delta SA}$ that developed (Fig.
S17B). Even so, while *IBE* was generally
unrelated to $LA_{\Delta SA}$ across all simulations,
there were generally positive relationships
between *IBE* and $LA_{\Delta SA}$ when controlling for
the number of traits and differences in
strengths of selection (Fig. S18). As such, *IBE*
from *all* markers explained very little
variation in performance when added as a
factor to the ANOVA model from Eq. 1 (SC
02.02.01), but accounted for some variation in
ANOVA models with only $LA_{\Delta SA}$ and *IBE* as
explanatory variables (0-34% for *IBE* vs 0-
74% for $LA_{\Delta SA}$; Table S3). Except for RONA,
the differences in performance between two
models trained with different marker sets was
generally unrelated to the differences in *IBE*
between the two marker sets used to train the
models (Fig. S19).
Together these results indicate that while
higher degrees of local adaptation may lead to
increased levels of *IBE* in the genome, the
signal of *IBE* of input markers generally has
minimal and varied impact on performance
differences for the scenarios evaluated here.
Alternatively, the levels of *IBE* present in the
simulated genomes may exceed a minimum
threshold of *IBE*, beyond which differences in

performance between marker sets are minimized.

Q4 - What is the variation of model performance across the landscape? (Q3a)

All 1- and 2-trait models exhibited variation in the predictive performance across gardens within a landscape, from essentially no predictive performance to very high predictive performance (Fig. S20, Fig. S21, Fig. S22, Fig. S23). Variation in performance was also observed for 6-trait models (Fig. 4).

While there was variability in predictive performance of 1- and 2-trait models within each landscape, in many cases the best performing models had the lowest levels of performance variation (Figs. S24, S25, S26). Ultimately we found no strong indicator for predicting when a model will be highly variable. Indeed, while performance generally increased with $LA_{\Delta SA}$ (Fig. 2), variability in performance was not strongly related to the variability in deme-level LA on the landscape (Figs. S27, S28, S29). Despite $LA_{\Delta SA}$ driving performance more generally (from Q1), this indicates that variation in model performance across the landscape is not strongly driven by metapopulation levels of, nor deme-level variation in, LA.

Q5 - How does the addition of non-adaptive nuisance environments in training affect performance? (Q2B)

Training offset models with the addition of non-adaptive nuisance environmental variables generally reduced offset method performance (Fig. 5). This decline was most dramatic for offset trained on 1-trait simulations (Fig. 5A) compared to the decline observed for 2-trait (Fig. 5B) and 6-trait (Fig. 5C) simulations. The only instances for which median performance did not decrease monotonically with nuisance level were for 2-trait simulations evaluated with GF_{offset} (Fig. S30).

Overall, landscape had the most influence over performance differences due to non-adaptive nuisance environments (Fig. S30), whereas there was little difference across other simulation parameters (not shown except in SC 02.02.06). Even so, *adaptive* markers seemed to provide some advantages in the presence of nuisance environments, particularly for 1-trait datasets where the advantages were more substantial compared to 2-trait datasets (Fig. S31, Fig. S32).

In some cases, the rankings of weighted environmental importance output from GF ranked nuisance variables higher than at least one adaptive environment (Table S4). Across *1- and 2-trait N-nuisance* models trained with *all* markers, GF incorrectly ranked environmental drivers in 26.9% (133/495) of the cases. Rankings improved somewhat for models trained with *adaptive* markers, incorrectly ranking environmental variables in 20.6% (102/495) of the cases (Table S4).

Q6 - How well do offset models extrapolate to novel environments outside the range of environments used in training? (Q2C)

The datasets that had the greatest within-landscape performance (i.e., those with higher levels of $LA_{\Delta SA}$) were also those that experienced the steepest decline in

performance with increasing climate
novelty (red shade, Fig. 6).
Importantly, declines in performance
for datasets with greater $LA_{\Delta SA}$ were
not due to instances where all
populations had zero fitness (and thus
performance was undefined and
manually set to 0; Supplemental Note
S4, Fig. S33). Despite little change in
the median performance for datasets
with low levels of LA, most performance
scores from these datasets were below
Kendall's **τ**=0.5, and therefore had
little predictive value in novelty
scenarios.
Advantages of *adaptive* marker sets
were much less prevalent across
methods for *Climate Novelty* scenario
performance than either *Adaptive*
Environment or *Nuisance Environment*
scenarios (Fig. S34).

4 | Discussion

Solutions are needed to mitigate the
negative impacts of global change on
biodiversity. In the last decade,
genomic offset methods have been
identified as a complement to other
ecological forecasting models because
they incorporate intraspecific variation
(Keller & Fitzpatrick, 2015; Capblancq
et al., 2020; Rellstab et al., 2021). Our
evaluations show that offset methods
are differentially impacted by both the
evolutionary history of sampled
populations as well as the decisions
made during model training. Our
analyses emphasize the importance of
sampling locally adapted populations,

identifying the drivers underlying
environmental selection pressures *a priori*,
and restricting offset projections to climates
similar to those used in training. Below, we
discuss the implications of these findings
towards restoration, conservation, and the
management of biodiversity.

4.1 | *The importance of local adaptation*

A basic assumption of genomic offset
methods is that the sampled populations are
adapted to their local environment (Rellstab
et al., 2016, 2021), but this assumption has
not been formally tested. Our analyses show
that indeed the degree of local adaptation
($LA_{\Delta SA}$) is one of the primary factors that
determine model performance for most
methods. A value of $LA_{\Delta SA} \sim 0.5$ indicates
that fitness in demes is on average 50% higher
in sympatry than allopatry. Values of $LA_{\Delta SA}$
represent the average deme-level magnitudes
of $LA_{\Delta HA}$ and $LA_{\Delta LF}$ across the
metapopulation (Blanquart et al., 2013).
Previous metaanalyses of studies measuring
local adaptation of natural populations have
used different measures of LA from the ones
we calculate here, but do show that some
species evolve large fitness differences among
populations (Hereford, 2009; Leimu &
Fischer, 2008). Given the prevalence of LA
found previously (Hereford, 2009; Leimu &
Fisher, 2010), we may therefore expect some
genomic offset methods to do reasonably well
when local adaptation in the metapopulation
is high (i.e., when $LA_{\Delta SA} > 0.5$).

4.2 | *The importance of the signals within genomic marker sets*

Because of the assumption that locally
adapted populations will be necessary for

satisfactory model performance, initial
implementations of genomic offset
models focussed on putatively adaptive
markers where this signal may be
strongest (Keller & Fitzpatrick, 2015;
Rellstab et al., 2016). More recently,
investigators have varied the set of
markers used to train models but have
found little influence on performance
(Fitzpatrick et al., 2021; Lachmuth,
Capblancq, Keller, et al., 2023; Láruson
et al., 2022; Lind et al., 2024). Our
results are similar to previous
investigations, finding that the *adaptive*
marker sets provide minimal advantage
over *all* or *neutral* marker sets, but not
universally or by great margins.

One hypothesis put forth as to why
adaptive marker sets perform similarly
to all markers is that genome-wide data
captures sufficient signatures of IBE
(Lachmuth, Capblancq, Keller, et al.,
2023; Lind et al., 2024). Our analysis
found weak positive relationships
between performance and levels of *IBE*
within marker sets. Even so, and except
for RONA, there were no universal
relationships within methods between
the difference in *IBE* of marker sets and
the difference in performance of the
models trained with these markers.

While we found little impact of
levels of *IBE* on overall performance,
the way in which we measured IBE may
have masked causative relationships.
For instance, in our measure of IBE we
correlated environmental distance with
pairwise F_{ST}. In doing so, our measure
of IBE distills genetic distance down to
a single value from a large number of
loci, and gives less weight to loci with

rare alleles. In future studies, creating a
marker set by ranking loci by single-locus
measures of IBE offers another opportunity to
understand the impact of IBE on
performance. Such marker sets could be used
to compare to performance from putatively
adaptive marker sets or marker sets composed
of all or random loci. Empirical datasets will
also be able to specifically address
geographical distances while quantifying IBE
(e.g., Bradburd et al., 2013).

While measures of IBE are one signal
remaining to be explored in future analyses,
the proportion of clinal neutral loci within
marker sets was shown to impact
performance, sometimes being positively
related to performance and sometimes
negatively depending on the context. These
and other signals within data that could
improve or mislead offset models also warrant
further investigation.

4.3 | The importance of adaptive environmental variables

In empirical settings, the environ- mental
drivers of local adaptation are rarely known *a*
priori. Even so, our results emphasize the
importance of identifying these variables
before training offset models, as there were
often declines in performance between models
trained using only adaptive environmental
variables (*0-nuisance*) and those trained using
additional non-adaptive nuisance
environmental variables (*N-nuisance*).

The importance of identifying these
selective environments may be particularly
germane to two general empirical scenarios. In
the first empirical scenario,, sparsely sampling
an environmentally heterogeneous range may
enrich genetic signals (e.g., coincident

population structure) most correlated
to environmental variables that
maintain a gradient across this extent,
and miss signals relevant to more local
scales. In the second empirical scenario,
identifying the environmental variables
underlying selection is particularly
important when a specific genomic
offset method is ill-suited to
differentiate importance among input
variables. For instance, RDA (and
therefore RDA_{offset}) assumes that the
environmental variables used to build
models are not collinear; (as
implemented here; Capblancq &
Forester, 2021; Legendre & Legendre,
2012). Because of this, empirical
datasets must be limited to a subset of
available environmental measures. The
process of excluding environmental
variables in this way may omit signals
of adaptive drivers (particularly when
true drivers are not well measured), or
perhaps incorporate environmental
variables that do not coincide with
drivers of selection. In these cases,
performance is likely to decline. As
such, this may indicate that methods
such as RDA_{offset} are likely to perform
worse in, or less uniformly across,
realistic empirical settings than what
our current findings suggest.

On the other hand, users of GF may
be tempted to include a large number
of environmental variables in training,
hoping that GF can accurately
attribute the correct environmental
variation to adaptive genetic structure.
Our results show that it is not
necessarily the case that GF will give
the highest importance values to the

true adaptive environmental variables.
Indeed, weighted feature importance scores
from GF models still incorrectly ranked the
adaptive environments below neutral
environments in 20%-27% of the datasets,
depending on which marker set was used.
These importance values ultimately affect the
model predictions. Including all available
environmental variables may therefore
negatively impact GF_{offset} performance, and
could have weakened overall performance in
previous empirical evaluations that used a
large number of environmental measures in
training (e.g., Lind et al., 2024).

There are some differences between the
nuisance environmental variables
implemented here and those that have been
implemented previously. For instance,
Láruson et al. (2022) created nuisance
variables by randomly sampling a
multivariate normal distribution. In contrast
to findings here, Láruson et al. (2022) found
that model performance was relatively
unaffected with the addition of nuisance
variables. The minimal influence of nuisance
variables on performance found by Láruson et
al. (2022) may differ from the performance
declines reported here because the nuisance
variables we used were spatially
autocorrelated, while those from Láruson et
al. (2022) were not. Inclusion of nuisance
variables that are spatially autocorrelated
may mislead offset models more generally
than variables with little spatial
autocorrelation because of the spurious
relationship between environmental structure
and genetic structure.

4.4 | The effect of environmental novelty

While within-landscape performance generally increased with $LA_{\Delta SA}$, the datasets with the greatest levels of $LA_{\Delta SA}$ were also the datasets where performance declined most readily with climate novelty. This occurred because locally adapted metapopulations were under strong selection to be fine-tuned to their environment, and as a result most individuals suffered severe fitness declines with environmental change. In contrast, less locally adapted metapopulations were under weaker selection, and suffered less steep fitness declines with environmental change. This result highlights an interesting paradox: offset methods that have the highest performance in common garden transplants under current climates (because of strong local adaptation) may have the lowest performance in predicting "genomic vulnerability" as the range of climate variables become more novel compared to the ranges used in training the model.

Thus, genomic offset models are likely ill-suited for estimating fitness ranks of populations in environments that differ drastically from those used to train the models themselves. This is particularly relevant for applications of offset methods that attempt to estimate *in situ* risk of climate change to years or climate scenarios where the environment is expected to be increasingly novel. While climate novelty is often measured with respect to historical variability (e.g., Lotterhos et al., 2021; Mahony et al., 2017; Williams et al., 2007), indices of local climate change indicate that local environments in terrestrial systems could experience change in excess of three standard deviations relative to historic values (Williams et al., 2007). Similar indices in marine systems indicate potential for even greater novelty (Lotterhos et al., 2021). We observed performance declines below the analogous z_E=3.13 *Climate Novelty* scenario, indicating offset predictions will likely be inaccurate in many real-world climate change predictions. These issues are also germane to measures derived from offset values (Gougherty et al., 2021; Lachmuth, Capblancq, Keller, et al., 2023; Lachmuth, Capblancq, Prakash, et al., 2023), which currently do not consider the degree of climate novelty in the prediction (but see DeSaix et al., 2022).

Our results present a best-case scenario for predicting performance in novel environments, as in many cases there will be biological reasons as to why climate-fitness relationships will differ in future environments from relationships measured within the contemporary climate space (see Fig. 5 in Capblancq et al., 2020). For the simulations here, the relationship between contemporary and novel environments with fitness was the same.

4.5 | Genomic offsets in practice

Our evaluations show that genomic offset methods hold promise for predicting maladaptation to environmental change within a historical baseline, in metapopulations that evolve strong local adaptation. However, our analyses also emphasize the limits of these methods in some

systems or scenarios. In practice,
species that are locally adapted to
measurable environmental variables
will be best suited for offset methods
when predicting the relative
performance of populations in a
contemporary common garden, but
paradoxically these species may be least
suited to using these methods to predict
their vulnerability to novel climates.

Together, these results indicate that
some genomic offset methods may be
suited to guide initiatives such as near-
term assisted gene flow, where targeted
restoration sites within a species range
have climates that are similar to those
used to train offset models. Even so, our
results also show that the performance
of these methods are often variable
across a landscape, indicating that high
performance at one site does not mean
the offset model will perform well at
another. While genomic offset methods

may be suitable for assisted gene flow
initiatives, they may be less suited for assisted
migration programs where populations are
moved outside of their native range and
environments differ from training data.

Before genomic offsets can be incorporated
into management plans, considerable thought
must be put into the sensitivity of model
outcomes from input data (Lind et al., 2024),
the uncertainty inherent in environmental or
climate forecasts (Lachmuth, Capblancq,
Keller, et al., 2023), as well as the degree of
novelty of future climates (DeSaix et al., 2022,
this study). While accurate predictions are
limited for novel climates of the future, these
offset methods could still be used to guide
management in the intervening time in a
stepwise manner where experiments can be
used to validate model performance in
practice. Using simulations tailored to the life
history of target species also presents a
promising avenue to understand limitations of
these methods for specific management cases.

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Author Contributions

1259

1260 KEL received funding. KEL and BML conceptualized the project and methodology. With
1261 input, editing, and feedback from KEL, BML wrote code to train and evaluate offset models,
1262 created figures, curated coding and records for archiving, and wrote the manuscript.

Conflict of Interests

1263

1264 The authors declare no conflicts of interest.

Data Availability

1265

We reference the analysis code in the text of our documents by designating Supplemental
Code (SC) using a directory numbering system from our servers (as opposed to the order
listed in the manuscript). Supplemental Code includes both executable scripts (*.R, *.py) as
well as jupyter notebooks (*.ipynb). For example, for Script 3 in Directory 1, we refer to SC
01.03; for Notebook 5 in Subfolder 3 of Directory 2, we will refer to SC 02.03.05. Each
Directory will be archived on Zenodo.org and include a citation below, which will also link
to the GitHub repository. Notebooks are best viewed within a local jupyter or jupyter lab
session (to enable cell output scrolling / collapsing), but can also be viewed at
nbviewer.jupyter.org using the web link in the archive's README on GitHub. Analyses
were carried out primarily using python v3.8.5 and R v3.5.1 and v4.0.3. Exact package and
code versions are available at the top of each notebook. More information on coding
workflows and coding environments can be found in Supplemental Note S1 and Supplemental
Note S2.

1279

1280 All directories, notebooks, and scripts can be found on GitHub, and will be archived on
1281 Zenodo.
1282 https://github.com/ModelValidationProgram/MVP-offsets

References

Blanquart, F., Kaltz, O., Nuismer, S. L., & Gandon, S. (2013). A practical guide to
measuring local adaptation. *Ecology Letters*, *16*(9), 1195–1205.
https://doi.org/10.1111/ele.12150

Bonan, G. B. (2008). Forests and Climate Change: Forcings, Feedbacks, and the
Climate Benefits of Forests. *Science*, *320*(5882), 1444–1449.
https://doi.org/10.1126/science.1155121

Bradburd, G. S., Ralph, P. L., & Coop, G. M. (2013). Disentangling the Effects of
Geographic and Ecological Isolation on Genetic Differentiation. *Evolution*, *67*(11),
3258–3273. https://doi.org/10.1111/evo.12193

Brown, J. H. (1984). On the Relationship between Abundance and Distribution of
Species. *The American Naturalist*, *124*(2), 255–279.
https://doi.org/10.1086/284267

Capblancq, T., Fitzpatrick, M. C., Bay, R. A., Exposito-Alonso, M., & Keller, S. R.
(2020). Genomic Prediction of (Mal)Adaptation Across Current and Future
Climatic Landscapes. *Annual Review of Ecology, Evolution, and Systematics*,
51(1), 245–269. https://doi.org/10.1146/annurev-ecolsys-020720-042553

Capblancq, T., & Forester, B. R. (2021). Redundancy analysis: A Swiss Army Knife for
landscape genomics. *Methods in Ecology and Evolution*.
https://doi.org/10.1111/2041-210x.13722

Clark, J. S., Carpenter, S. R., Barber, M., Collins, S., Dobson, A., Foley, J. A., Lodge,
D. M., Pascual, M., Jr., R. P., Pizer, W., Pringle, C., Reid, W. V., Rose, K. A.,
Sala, O., Schlesinger, W. H., Wall, D. H., & Wear, D. (2001). Ecological
Forecasts: An Emerging Imperative. Science, *293*(5530), 657–660.
https://doi.org/10.1126/science.293.5530.657

DeSaix, M. G., George, T. L., Seglund, A. E., Spellman, G. M., Zavaleta, E. S., &
Ruegg, K. C. (2022). Forecasting climate change response in an alpine specialist
songbird reveals the importance of considering novel climate. *Diversity and*
Distributions, *28*(10), 2239–2254. https://doi.org/10.1111/ddi.13628

Doney, S. C., Ruckelshaus, M., Duffy, J. E., Barry, J. P., Chan, F., English, C. A.,

Galindo, H. M., Grebmeier, J. M., Hollowed, A. B., Knowlton, N., Polovina, J.,

Rabalais, N. N., Sydeman, W. J., & Talley, L. D. (2012). Climate Change

Impacts on Marine Ecosystems. *Annual Review of Marine Science*, *4*(1), 11–37.

https://doi.org/10.1146/annurev-marine-041911-111611

Elith, J., & Leathwick, J. R. (2009). Species Distribution Models: Ecological

Explanation and Prediction Across Space and Time. Annual Review of Ecology,

Evolution, and Systematics, *40*(1), 677–697.

https://doi.org/10.1146/annurev.ecolsys.110308.120159

Fitzpatrick, M. C., Blois, J. L., Williams, J. W., Nieto-Lugilde, D., Maguire, K. C., &

Lorenz, D. J. (2018). How will climate novelty influence ecological forecasts?

Using the Quaternary to assess future reliability. *Global Change Biology*, *24*(8),

3575–3586. https://doi.org/10.1111/gcb.14138

Fitzpatrick, M. C., Chhatre, V. E., Soolanayakanahally, R. Y., & Keller, S. R. (2021).

Experimental support for genomic prediction of climate maladaptation using the

machine learning approach Gradient Forests. *Molecular Ecology Resources*.

https://doi.org/10.1111/1755-0998.13374

Fitzpatrick, M. C., & Keller, S. R. (2015). Ecological genomics meets community-level

modelling of biodiversity: mapping the genomic landscape of current and future

environmental adaptation. Ecology Letters, 18(1), 1–16.

https://doi.org/10.1111/ele.12376

Gain, C., & François, O. (2021). LEA 3: Factor models in population genetics and

ecological genomics with R. Molecular Ecology Resources, 21(8), 2738–2748.

https://doi.org/10.1111/1755-0998.13366

Good, R. D. (1931). A THEORY OF PLANT GEOGRAPHY. *New Phytologist*, *30*(3),

149–149. https://doi.org/10.1111/j.1469-8137.1931.tb07414.x

Gougherty, A. V., Keller, S. R., & Fitzpatrick, M. C. (2021). Maladaptation, migration

and extirpation fuel climate change risk in a forest tree species. *Nature Climate*

Change, 1–15. https://doi.org/10.1038/s41558-020-00968-6

Hereford, J. (2009). A Quantitative Survey of Local Adaptation and Fitness Trade-Offs.

The American Naturalist, *173*(5), 579–588. https://doi.org/10.1086/597611

Hoegh-Guldberg, O., & Bruno, J. F. (2010). The Impact of Climate Change on the
World's Marine Ecosystems. Science, *328*(5985), 1523–1528.
https://doi.org/10.1126/science.1189930

Lachmuth, S., Capblancq, T., Keller, S. R., & Fitzpatrick, M. C. (2023). Assessing
uncertainty in genomic offset forecasts from landscape genomic models (and
implications for restoration and assisted migration). *Frontiers in Ecology and*
Evolution, *11*, 1155783. https://doi.org/10.3389/fevo.2023.1155783

Lachmuth, S., Capblancq, T., Prakash, A., Keller, S. R., & Fitzpatrick, M. C. (2023).
Novel genomic offset metrics account for local adaptation in climate suitability
forecasts and inform assisted migration. *BioRxiv*, 2023.06.05.541958.
https://doi.org/10.1101/2023.06.05.541958

Láruson, Á. J., Fitzpatrick, M. C., Keller, S. R., Haller, B. C., & Lotterhos, K. E.
(2022). Seeing the forest for the trees: Assessing genetic offset predictions from
gradient forest. *Evolutionary Applications*, *15*(3), 403–416.
https://doi.org/10.1111/eva.13354

Lee-Yaw, J. A., McCune, J. L., Pironon, S., & Sheth, S. N. (2022). Species distribution
models rarely predict the biology of real populations. *Ecography*, *2022*(6).
https://doi.org/10.1111/ecog.05877

Legendre, P., & Legendre, L. (2012). *Numerical Ecology* (Vol. 24). Elsevier.
Leimu, R., & Fischer, M. (2008). A meta-analysis of local adaptation in plants.
PLoS ONE, *3*(12), e4010. https://doi.org/10.1371/journal.pone.0004010.s001

Lind, B. M., Candido-Ribeiro, R., Singh, P., Lu, M., Vidakovic, D. O., Booker, T. R.,
Whitlock, M., Isabel, N., Yeaman, S., & Aitken, S. N. (2024). How useful is
genomic data for predicting maladaptation to future climate? *In Press at Global*
Change Biology. https://doi.org/10.1101/2023.02.10.528022

Lotterhos, K. E. (2023). The paradox of adaptive trait clines with nonclinal patterns in
the underlying genes. *Proceedings of the National Academy of Sciences*, *120*(12).
https://doi.org/10.1073/pnas.2220313120

Lotterhos, K. E., Fitzpatrick, M. C., & Blackmon, H. (2022). Simulation Tests of
Methods in Evolution, Ecology, and Systematics: Pitfalls, Progress, and
Principles. Annual Review of Ecology, Evolution, and Systematics, 53(1), 113–

136. https://doi.org/10.1146/annurev-ecolsys-102320-093722

Lotterhos, K. E., Láruson, Á. J., & Jiang, L.-Q. (2021). Novel and disappearing climates
in the global surface ocean from 1800 to 2100. *Scientific Reports*, *11*(1), 15535.
https://doi.org/10.1038/s41598-021-94872-4

Mahony, C. R., Cannon, A. J., Wang, T., & Aitken, S. N. (2017). A closer look at novel
climates: new methods and insights at continental to landscape scales. Global
Change Biology, 23(9), 3934–3955. https://doi.org/10.1111/gcb.13645

Rellstab, C., Dauphin, B., & Exposito-Alonso, M. (2021). Prospects and limitations of
genomic offset in conservation management. Evolutionary Applications, 14(5),
1202–1212. https://doi.org/10.1111/eva.13205

Rellstab, C., Zoller, S., Walthert, L., Lesur, I., Pluess, A. R., Graf, R., Bodénès, C.,
Sperisen, C., Kremer, A., & Gugerli, F. (2016). Signatures of local adaptation in
candidate genes of oaks (Quercus spp.) with respect to present and future
climatic conditions. Molecular Ecology, *25*(23), 5907–5924.
https://doi.org/10.1111/mec.13889

Schmolke, A., Thorbek, P., DeAngelis, D. L., & Grimm, V. (2010). Ecological models
supporting environmental decision making: a strategy for the future. *Trends in*
Ecology & Evolution, *25*(8), 479–486. https://doi.org/10.1016/j.tree.2010.05.001

Waldvogel, A.-M., Feldmeyer, B., Rolshausen, G., Exposito-Alonso, M., Rellstab, C.,
Kofler, R., Mock, T., Schmid, K., Schmitt, I., Bataillon, T., Savolainen, O.,
Bergland, A., Flatt, T., Guillaume, F., & Pfenninger, M. (2020). Evolutionary
genomics can improve prediction of species' responses to climate change.
Evolution Letters, *4*(1). https://doi.org/10.1002/evl3.154
Wang, I. J., & Bradburd, G. S. (2014). Isolation by environment. Molecular
Ecology, 23(23), 5649–5662. https://doi.org/10.1111/mec.12938

Weir, B. S., & Cockerham, C. C. (1984). Estimating F-statistics for the analysis of
population structure. Evolution, 38(6), 1358–1370.

Williams, J. W., Jackson, S. T., & Kutzbach, J. E. (2007). Projected distributions of
novel and disappearing climates by 2100 AD. Proceedings of the National
Academy of Sciences, *104*(14), 5738–5742.
https://doi.org/10.1073/pnas.0606292104

1405

Figures for:

The limits of predicting maladaptation to future environments with genomic data

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Running Title: Limits of genomic offsets

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Method	abbr.	Multivariate?	Structure correction?
Gradient Forests[1]	GF_{offset}	Yes	No
Redundancy Analysis[2] with population structure correction	RDA-corrected	Yes	Yes, with axes loadings from PCA*
Redundancy Analysis[2] without population structure correction	RDA-uncorrected	Yes	No
Latent factor mixed model from Landscape and Ecological Association Studies R package[3]	$LFMM2_{offset}$	Yes	Yes, with latent factors
Risk Of Non-Adaptedness[4]	RONA	No	No

* principal component analysis

Table 1   Genomic offset methods used for evaluation. Genomic offset methods differ in their capability to use multivariate environmental data in training as well as whether a correction for population genetic structure is applied. Superscripts apply to the following reference citations: 1 - Fitzpatrick & Keller, 2015; 2 - Capblancq & Forester, 2021; 3 - Gain & François, 2021; 4 - Rellstab et al., 2016.

Workflow	n_{traits}	(1) Simulations Levels (replicates per level)	(3, 7) Environmental Data	Training and Prediction?	(6) Within-landscape Evaluation?	(9) Total Performance Evaluations
Adaptive Environment (AE)	1-trait	45 (10)	temp	Yes	Yes	675,000
	2-trait	180 (10)	temp + Env2			3,240,000
	6-trait	1 (1)	MAT + MTwetQ + MTDQ + PDM + PwarmQ + PWM			3,000
Nuisance Environment (NE)**	1-trait	45 (1)	[AE$_{1\text{-trait}}$ environments + Env2] $_{+/-}$ [ISO + PSsd] $_{+/-}$ [TSsd]	Yes	Yes	175,500
	2-trait	180 (1)	[AE$_{2\text{-trait}}$ environments + ISO + PSsd] $_{+/-}$ [TSsd]			432,000*
	6-trait	1 (1)	[AE$_{6\text{-trait}}$ environments + ISO + PSsd + TSsd]			1,200*
Climate Novelty (CN)†	1-trait	45 (10)	AE$_{1\text{-trait}}$ environments	Prediction only (using AE training models)	No	64,800*
	2-trait	180 (10)	AE$_{2\text{-trait}}$ environments			259,200*
	6-trait	1 (1)	AE$_{6\text{-trait}}$ environments			144*

* excludes RONA
** The set of population values for each unique nuisance environment were the same across traits and landscapes
† includes evaluation of climate center and eleven *Climate Novelty* scenarios

Table 2   Workflows used to process simulation data for the evaluation of genomic offset methods. Numbers given in column names refer to locations in schematic of Fig. 1. The *Adaptive Environment* workflow processes all population data from 1- and 2-trait (example shown in Fig. 1) as well as 6-trait simulations using only adaptive environmental variables in training, and evaluates performance in each garden on the metapopulation landscape. The *Nuisance Environment* workflow processes 1-, 2-, and 6-trait simulations similarly to the *Adaptive Environment* workflow, except in addition to adaptive environmental variables used in training, non-adaptive (i.e., nuisance) environmental variables are also used - each bracketed set of environmental variables indicate a distinct nuisance level (e.g., "*1-trait 1-nuisance*" = [AE$_{1\text{-trait}}$ environments + Env2] and "*1-trait 4-nuisance*" = [AE$_{1\text{-trait}}$ environments + Env2 + ISO + PSsd + TSsd]). The *Climate Novelty* workflow uses trained models from the *Adaptive Environment* workflow (Fig. 1A-5) and evaluates offset in 11 novel environments relative to the range of environments used in training. See Supplemental Note S3 for details regarding the choice of *Climate Novelty* environmental values and visualizations of climate data in principal component space. See Supplemental Notes S1-S2 for descriptions of coding workflows. Counts of evaluations were tabulated in SC 02.10.01.

Figure 1 Analysis of 1-, 2-, and 6-trait simulations included three main phases: A) model training, B) model prediction, and C) evaluation of models. The *Adaptive Environment* workflow is shown as an example of the processing of 1- and 2-trait simulation data for genomic offset evaluation. In total, three general workflows are used to evaluate genomic offset methods (Table 2). Subpanels of this schematic are numbered for referencing in Table 2 and the main text.

(Fig. 2)

(Fig. 2 continued)

Figure 2   Predictive performance of genomic offset models (y-axes) is driven by the degree of local adaptation (A) and the spatial patterns of adaptive environments across the landscape (B, C). For each model, a median value from performance scores from 100 common gardens is shown for A and B; C shows scores across all common gardens for each model (note that y-axes are inverted, as more negative values have higher performance). Data included in these figures was processed through the *Adaptive Environment* workflow but only includes models trained using 2-trait simulations and *all* loci. Code to create (A) and (B) can be found in SC 02.02.02; code to create (C) can be found in SC 02.02.01.

(Fig. 3)

Figure 3 Comparison of marker choice across genomic offset methods for 1- and 2-trait simulations. A-C are scatterplots of pairwise comparisons of performance between marker sets (histograms in each margin) from both 1- and 2-trait models where density of points is indicated by color in legend (note color scale is different for each figure to accentuate patterns in data). D-E are boxplots from the same data in A-C separated by individual traits. Data included in these figures is from all 1- and 2-trait models from the *Adaptive Environment* workflow. Code to create these figures can be found in SC 02.02.03.

Figure 4   Comparison of marker choice across genomic offset methods for the 6-trait simulation. A-C are scatterplots of pairwise comparisons of performance between marker sets (RONA is not shown, except in SN 02.05.10). D are boxplots from the same data in A-C (RONA$_{6\text{-traits}}$ is the combined performance across all six environmental models). Data included in this figure is from the 6-trait models processed through the *Adaptive Environment* workflow. Note there is only one 6-trait replicate, and variation within figures represents the performance across 100 common gardens for each method. Code to create these figures can be found in SN 02.05.10.

(Fig. 5)

Figure 5 Effect of non-adaptive nuisance environmental variables on offset performance. Shown are evaluations of offsets from 1- and 2-trait models trained using only adaptive environments (*0-nuisance*) or with adaptive environments and the addition of *N*>0 non-adaptive environmental variables (*N-nuisance*). RONA is not shown because it is univariate with respect to environmental variables. Nuisance variables are listed in Table 2. Code to create figures can be found in SC 02.02.06 and SC 02.02.08.

Figure 6 Performance decays with climate novelty relative to training data. Shown is model performance (y-axes) across methods at climate center (A) and across common gardens each representing increasing degrees of climate novelty relative to training data (x-axis of B) where all 100 populations have been transplanted. The standard deviation values (x-axis, B) are applicable to all environments for all landscapes except for *Env2* in the *Stepping Stone - Mountain* landscape; the corresponding standard deviations are 1.55, 2.12, 2.47, 2.82, 3.18, 3.53, 3.88, 4.24, 4.60, 4.95, 5.3. When fitness for all transplanted individuals was zero, a model's performance was undefined and manually set to 0; no method predicted a single offset value for all populations in these situations. Setting undefined performance to 0 did not substantially impact patterns between performance and climate novelty, and is explored in Supplemental Text S3. Data included in this figure are from models trained using 1- and 2-trait simulations from the *Climate Novelty* workflow, which excludes both RONA$_{temp}$ and RONA$_{Env2}$. Code used to create this figure can be found in SC 02.04.05.