Title: Choices to landscapes: Mechanisms of animal movement scale to landscape patterns of space use

Authors and affiliations:
Will Rogers[1], will.rogers@yale.edu https://orcid.org/0000-0001-9899-8064
Scott Yanco[1,2] scott.yanco@yale.edu https://orcid.org/0000-0003-4717-9370
Walter Jetz[1,2], walter.jetz@yale.edu https://orcid.org/0000-0002-1971-7277

[1] Department of Ecology & Evolutionary Biology, Yale University, P. O. Box 208106, New Haven, CT 06520-8106
[2] Center for Biodiversity and Global Change, Yale University, P. O. Box 208106, New Haven, CT 06520-8106

Running title: Selection and movement predict space use

Keywords: Species distribution, Habitat selection, Resource selection, Occurrence, Movement, Step-selection, Ergodicity

Type of article: Letter

Statement of authorship: WR and SY conceptualized the paper. WR wrote the code and carried out the analyses supporting the manuscript. WR wrote the initial draft with contributions from SY. All authors provided feedback on the concept and contributed to the final draft.

Data statement: No new or original data were produced in this study. All data used was simulated or open source with expressed approval by original authors for use in this study.

Abstract word count: 143
Main text word count: 4,871
References: 55
Figures/tables/boxes: 5 figures, 0 tables, 0 boxes

Corresponding author:
Name: Will Rogers
Email: will.rogers@yale.edu, rogerswill47@gmail.com
Mail: Department of Ecology & Evolutionary Biology, Yale University, P. O. Box 208106, New Haven, CT 06520-8106
Address: 165 Prospect Street, New Haven, CT, US 06511
Phone: (203) 432-6138
Fax: (203) 432-5176

42 **Abstract:**

43   Understanding the geographic distributions of animals is central to ecological inquiry and

conservation planning. Movement-based habitat selection models offer a powerful tool for

identifying preferred environmental attributes, yet applying these models to predict animal

geographic distributions faces methodological and computational challenges. Here, we present a

framework that integrates habitat selection and movement behaviors to generate landscape-scale

space use predictions. Through simulations and empirical data, we demonstrate that combining

local selection and movement dynamics yields highly accurate emergent spatial distribution

predictions. Our framework outperforms occurrence-based frameworks across individual,

population, and regional scales. By explicitly addressing the role of movement constraints and

selection patterns in heterogeneous environments, our framework bridges animal movement and

spatial distribution modeling in a scalable manner. This approach offers a new paradigm to link

organism-environment interactions from individual space use to habitat connectivity and

population distributions relevant to policy and conservation.

Introduction:

Understanding how organisms interact with their environments is foundational to ecological inquiry, revealing critical insights into spatiotemporal patterns of biodiversity (Buckley *et al.* 2011; Woodin *et al.* 2013; Evans *et al.* 2015), population dynamics (Bozinovic *et al.* 2011; Madin *et al.* 2012), and genetic connectivity (Cushman & Lewis 2010; Fletcher *et al.* 2022). Beyond theory, accurately linking where organisms occur with the environmental conditions they select is essential for conservation planning and ecological forecasting (Urban 2015). Models based on spatial observation data, which make scalable predictions by correlating species occurrences with environmental conditions, often average over the underlying ecological mechanisms (Elith & Leathwick 2009). While they provide broad applicability, these models risk failure under novel conditions, especially as biotic interactions shift with exposure to changing climates and land use (Pearson & Dawson 2003). Mechanistic models, in contrast, aim to identify concrete environmental relationships driving species occurrence (Kearney & Porter 2009; Buckley *et al.* 2011). However, mechanistic processes are challenging to integrate into spatial predictions, such models may fall short of offering actionable insights and, worse, may even fail compared to pattern-based models. The recent explosion and projected growth of individual-based tracking data (Kays & Wikelski 2023) amplifies this challenge amidst a suite of new opportunities for biodiversity change monitoring (Jetz *et al.* 2022; Davidson *et al.* 2025), underscoring the need for approaches that can translate fine-scale behavioral data into population- or species-level spatial predictions (Nathan *et al.* 2008; Hooten *et al.* 2017). Capturing these interactions requires more than presence data; it demands a nuanced understanding of how animal behaviors like movement and habitat selection shape observed patterns within complex, fragmented landscapes (Schick *et al.* 2008; Matthiopoulos *et al.* 2020).

79   Resolving how to incorporate mechanistic insights of animal movement into scalable, predictive

80   models of space use is critical as we confront the twin global crises of environmental change and

81   biodiversity loss (Powers & Jetz 2019).

82   Inferring relationships between animals and their environment has traditionally relied on

observations of occurrence patterns over some set of environmental contexts which vary

geographically (Boyce & McDonald 1999; Boyce *et al.* 2002), typically by using environmental

dimensions to contrast observations from locations where organisms are putatively absent

(Mackey & Lindenmayer 2001; Johnson *et al.* 2006; Moorcroft & Barnett 2008). For example,

species distribution models (SDMs) compare the patterns of presences as compared to possible

null distributions over large geographic extents (Elith *et al.* 2011; Merow *et al.* 2014). Resource

selection functions (RSFs) - a methodologically similar framework to SDMs (Fieberg *et al.*

2021; Northrup *et al.* 2022) - use logistic regression to model the probability of individual

occurrences within some domain of predefined availability, like individual home ranges (Boyce

& McDonald 1999; Manly *et al.* 2007). Both SDMs and RSFs are occurrence-based frameworks

that capture information about multiple biological mechanisms leading to species or individual

occurrences while accounting for elements of the observation process itself (Keating & Cherry

2004; Phillips *et al.* 2009; Sastre & Lobo 2009; Wisz & Guisan 2009; Newbold *et al.* 2010).

Notably, SDMs and RSFs assume that all environmental conditions within the landscape, an

animal's home range, or other selection domains were instantaneously available for that animal

to select, irrespective of the geographic configuration of these environments or the sequence of

such choices made by the organism.

100   Movement constraints and habitat arrangement limit an organism's access to certain

environmental conditions within its geographic range (Fig. 1A), as isolated patches of preferred

habitat may be unreachable and, therefore, unused (Soberon & Peterson 2005; Matthiopoulos *et*

al. 2020; Van Moorter *et al.* 2023). These limitations shape spatial occurrence patterns,

especially in heterogeneous or movement-restricted landscapes (Fig. 1B) (Barnett & Moorcroft

2008; Signer *et al.* 2017). This is an important problem overlooked by SDMs and RSFs, which

unwittingly average over movement and dispersal as a driver of selection. Step-selection

functions (SSFs) address this deficiency by using GPS data to model selection based on quasi-

realistic, stepwise availability conditioned by movement constraints and even estimating those

constraints concomitantly (Avgar *et al.* 2016). However, SSFs lack a natural method to directly

predict geographic distributions because they infer preference conditioned on each step,

functionally limiting spatial predictions to comparisons against an (often arbitrarily) chosen

baseline. This relativized projection of SSFs also ignores the movement constraints inherent in

model parametrization and risks misrepresenting actual use across a landscape (Signer *et al.*

2017). Solutions such as SSF-based agent-based simulations (Signer *et al.* 2017; Potts & Borger

2023) or Markov chain Monte Carlo adaptations of movement frameworks (Michelot *et al.* 2019)

provide compelling frameworks for projecting local movement mechanisms, but these methods

are computationally intensive and lack guaranteed convergence. Multi-stage frameworks, which

separately model resource selection (location quality) and step selection (connectivity), offer

scalable alternatives but require complex discretization of movement and selection mechanisms

(Van Moorter *et al.* 2023). Bridging movement constraints and environmental selection in

geographic projections remains critical to linking fine-scale organism-environment interactions

to broader spatial distributions for ecological and conservation applications (Schick *et al.* 2008;

Matthiopoulos 2022).

To address this problem, we introduce an approach that leverages SSFs to model

concurrent movement and selection processes, enabling geographically explicit predictions of

habitat use that (i) mechanistically capture organism-environment interactions and (ii) improve

accuracy over occurrence-based frameworks by incorporating movement constraints.

Specifically, we fit SSFs to estimate local-scale movement and retrieve stable-state distributions

(SSDs) through eigendecomposition within an ergodic framework. We then compare our

combined framework to occurrence-based (RSF) and selection-only (SSF) models to evaluate

predictive accuracy across varying animal selection patterns, landscape heterogeneity, and three

empirical cases of increasingly out-of-sample prediction.

Methods:

We introduce a way to isolate expected population-level space use from individually fit

SSFs, capitalizing on locally estimated selection and movement patterns. In this framework,

local movement probabilities are defined through both fine-scale habitat selection and distance-

weighted movement. These local movement patterns are used to generate population-level space

use through eigendecomposition of movement probabilities, generating a steady-state

distribution of probability of use for the landscape.

Landscapes as transition matrices

We first considered a landscape, G, made up of n cells along with environmental data

(Fig. 2A). We treated the occupancy of these cells as discrete states, between which individuals

might transition according to rates described by a stochastic transition matrix, A, of n rows and

columns. We defined each value in this matrix, $A_{i,j}$, as the probability of moving from one cell

(G_i) to another (G_j) in the landscape. For an animal in G_i, the sum of all transition probabilities to

all other cells in G (including G_i) must sum to one:

$$P(G_i|G_i) + \sum_{k=1, k \neq i}^{n} P(G_{k=j}|G_i) = 1. \qquad \text{Eq. 1}$$

In this way, Matrix A contains the directional probabilities of movement between all cells

in G based on beginning in any given cell, G_i, parametrized by later movement modeling and

creating a row-stochastic matrix. The inverse of this matrix, A^{-1}, defines a column-stochastic

matrix with a first principal eigenvector (associated with an eigenvalue of one) defining the

probabilities of the stable-state distribution across G (Fletcher *et al.* 2019).

SSFs can parametrize transition matrices

To gather these transition probabilities, we used predicted relative risk informed by fitted

SSFs, a quantity derived by comparing baseline conditions (at G_i) to a set of available conditions

(G) and often described as a relative strength of selection (Avgar et al. 2017). Relative risks are

not equivalent to probabilities but can be kernelized when choices are finite. To do this, we

divide Eq. 1 by the probability of moving from G_i to G_i:

$$\frac{P(G_i|G_i)}{P(G_i|G_i)} + \sum_{k=1, k \neq i}^{n} \frac{P(G_{k=j}|G_i)}{P(G_i|G_i)} = \frac{1}{P(G_i|G_i)}. \qquad \text{Eq. 2A}$$

161   Therefore, Eq. 2A shows that the sum of relative risks for moving to all cells in G is

162   equal to the inverse probability of remaining in the current cell. Or, simplified, the inverse of the

163   sum of all relative selection strengths is equal to the probability of remaining in G_i (Eq. 2B):

$$\frac{1}{1 + \sum_{k=1, k \neq i}^{n} \frac{P(G_{k=j}|G_i)}{P(G_i|G_i)}} = P(G_i|G_i). \qquad \text{Eq. 2B}$$

164   More importantly, Eq. 2B now describes how to link a limited set of relative selection

165 strengths to the probabilities of remaining in place. Therefore, we could solve for the

166 probabilities of moving to all comparison cells by multiplying the relative selection strength

167 returned by the SSF predictions by the probability of staying in place.

169 *Extension to Integrated Step Selection Functions*

170   SSFs can also explicitly incorporate behavioral elements of movement within habitat

171 selection models. Integrated SSFs (iSSFs) include terms to capture movement kernels, resulting

172 in a distributional description of the likely movement distances around a focal position (Avgar *et*

173 *al.* 2016). Functionally, this movement kernel can force the probability of selection for very

174 long-distance movements to zero even when habitat conditions are otherwise suitable. In this

175 way, iSSF models can define local neighborhoods of functionally available habitat for animal

176 movement. By incorporating distance into SSF models, we can confidently compare G_i to cells

177 in G within a biologically informed limit on neighborhood distances (herein the 95th percentile of

178 observed step distances, Fig. 2G). This substantially reduces the information contained in the

179 transition matrix A, treating it as a sparse matrix containing only transitions to cells within a

180 reasonable biological limit of movement (Fig. 2H) and enabling stable-state distribution

181 calculations (Fig. 2K). We then compared SSD predictions to pattern-only (RSF, Figs. 2B, 2D,

182 2I) and selection-only predictions (SSF, Figs. 2F, 2J).

184 *Simulation*

185   We stimulated animal movements based on an SSF framework proposed by Potts and

186 Borger (2023). We generated random environmental variables by simulating four environmental

variables (100 cells x 100 cells), each with 10 different scales of spatially autocorrelated

Gaussian random fields (Hanson *et al.* 2024) (S1 Figs. 1-2). We also specified theoretical

animals with 10 different magnitudes of resource selection across the four environmental

variables (β_{1-4}, S1 Table 1). For each set of spatial autocorrelation scales (landscape

heterogeneity) and animal selection strength, we multiplied selection coefficients (β_{1-4}) by

respective environmental surfaces (1-4) to generate a selection surface (S1 Fig. 3). We repeated

this process, creating five random landscapes per level of heterogeneity. We used these selection

surfaces, a step length constraint, and a tendency to maintain step angles across steps to simulate

animal movements (See S1 for more details). Two sets of movements were generated per

simulation: a focal set of 10,000 steps used in modeling (S1 Fig. 4) and a non-focal set of 500

random animals each for 1,000 steps (500,000 total) for validation. We then compared kernelized

selection and simulated space use across levels of landscape heterogeneity and animal selection

strength (Fig. 1C-E).

For each of the 500 simulations, we converted the focal 10,000 steps into movement

tracks, jittered locations slightly to create a positive distribution of step lengths, and sampled 100

random steps per observed step using the "amt" package (Signer *et al.* 2019). Based on observed

step lengths, we sampled random steps from a uniform distribution with a minimum of 0 and a

maximum 25% larger than the maximum observed step length. This increased random sampling

of step distances both close and far from the animal mover. We fit SSFs by treating the step-

terminus values of the four input environmental layers as predictors and including step length

and log-step length as predictors using the "amt" package (S1 Table 2) (Signer *et al.* 2019). We

used the fitted SSF models to parametrize *A* and generate stable-state distributions for each

simulation (S1 Fig. 8). We then generated opposing selection-only predictions by using SSFs to

predict the selection probability for all cells in the landscape using a random cell in the landscape

as a baseline (S1 Fig. 9). Finally, we fit RSFs using the "amt" package, sampling 10 random

points per occurrence within the minimum convex polygon of use locations and treating the four

input rasters as covariates (S1 Table 2) (Signer *et al.* 2019). We then used these RSFs to generate

opposing spatial predictions to SSD and SSF for each of the 500 simulations (S1 Fig.10).

215   To isolate the effect of landscape heterogeneity and animal selection patterns on

inferences about animal selection (S1 Figs. 11-12) and spatial predictions (S1 Figs. 12-25), we

primarily used Shannon entropy, relative entropy (or Kullback–Leibler divergence), and patterns

of cumulative use probability. Shannon entropy, defined as $\sum_{i=1}^{n} p(x_i)\ln(p(x_i))$ where $p(x_i)$ is

the probability of use of a cell in the landscape, is a measure of heterogeneity in probability

distributions. As Shannon entropy decreases, probability distribution outcomes become more

certain; for landscapes, this means that animal space use is heterogeneously concentrated in

relatively few locations. Relative entropy, or Kullback–Leibler divergence, was measured as

$\sum_{i=1}^{n} p(x_{observed})\ln\left(\frac{p(x_{observed})+1\times10^{-10}}{p(x_{predicted})+1\times10^{-10}}\right)$ where $p(x_{observed})$ is the probability of use of a cell in

the landscape in the validation 500,000 steps and $p(x_{predicted})$ is the probability of use of a cell in

the landscape in a spatial prediction method. The constant 1×10^{-10} ensures that sites which are

unvisited in validation movements are defined. As relative entropy increases, the worse that a

predicted probability distribution approximates the observed (or true) probability distribution.

Cumulative use probability was calculated by ordering probability distributions from highest to

lowest relative probability of use and then calculating the cumulative sum. For comparisons, we

specifically quantified the total area of cells required to generate a cumulative probability of use

of 0.5, or the minimum area in which an animal is expected to use 50% of the time. Further

description of the simulation and more extensive methods are provided in the Supplement (S1).

Case Study #1 - Intra-individual prediction

To demonstrate the predictive utility of SSD for individual space use, we tested the

predictive accuracy of SSD vs. SSF and RSF using a red deer (*Cervus elaphus*) dataset for a

single individual included in the "amt" package (Signer *et al.* 2019), using distance to forest and

forest cover (calculated from a binary forest raster provided in the "amt" package, aggregated to

a resolution of 125m by 125m, S2 Table 1). We separated the first (chronologically) 75% of

observed data as an in-sample training dataset and retained the latter 25% of observed data as an

out-of-sample testing dataset (S2 Fig. 1). As above, we sampled random steps from a wider

uniform distribution with 100 random points per observed step. We included an interaction

between distance to forest and forest cover in models along with step length and log-step length

in SSF models (S2 Table 2). We generated the SSD as above using a neighborhood defined by

the 95th percentile of observed step lengths (S2 Fig. 2). We then generated the RSF and SSF

probabilities as in simulations using the same predictor structures (S2 Table 2, S2 Figs. 3-4).

We then compared the heterogeneity and accuracy of each prediction. To quantify

heterogeneity, we calculated the Shannon entropy of each predicted probability distribution. In

this context, lower Shannon entropy describes predictions that the intensity of animal use is

concentrated in fewer locations. To quantify accuracy, we used two main comparison heuristics.

First, following Boyce *et al.* (2002), binned prediction surfaces into 10 bins of equal area and

then calculated the Spearman rank correlation between observed use intensity and bin order.

Second, as all prediction kernels summed to one, we calculated the geometric mean of predicted

use across use locations to discriminate more accurate predictions. Spearman rank correlations

are common in resource selection validation (Morris *et al.* 2016), but they are sensitive to

discrete prediction surfaces and our simulations showed that Spearman rank correlations may be

biased to select homogeneous prediction surfaces (S1 Fig. 23). The geometric mean of predicted

use, in contrast, was better able to accurately recover the difference in performance between

predictions and was generally more conservative (S1 Fig. 24). To ensure that comparisons were

not sensitive to the underlying out-of-sample datasets, we bootstrapped the observed set of GPS

points (sampling with replacement) 100 times, generating 100 estimates of Spearman rank

correlations and mean log-probability of use per predicted surface. We calculated the difference

between SSD and RSF or SSF predictions to isolate comparisons. We report geometric mean

results in the main text and Spearman rank comparisons in S2.

Case Study #2 - Inter-individual predictions

To demonstrate the predictive utility of SSD across individuals, we used a roe deer

(*Capreolus capreolus*) dataset for five individuals in Northern Italy (Cagnacci *et al.* 2011) to

assemble the same SSD, SSF, and RSF predictions. We resampled tracks to 4 hours (the median

interval in GPS transmissions), and as above, we isolated the first 75% of individual movements

as in-sample, and the next 25% as out-of-sample testing datasets (S2 Fig. 5). Because the dataset

contained migratory movements, we decided to model only summer movements of deer (defined

as Julian day 120-290, S2 Fig. 5). We included square-root distance to forest (calculated from

2006 CORINE landcover classes of conifer, mixed, and broadleaf forests) and slope (calculated

from STRM elevation models, aggregated at a scale of 129m by 185.2m) as covariates in our

models and predictions (S2 Fig. 6) with step length and log-step length, while the RSF included

the square-root of distance to forest and slope (S2 Table 2). We generated SSD (S2 Fig. 7), SSF

(S2 Fig. 8), and RSF predictions as above (S2 Fig. 9). We performed out-of-sample evaluations

as above with 100 bootstrapped geometric mean predicted use and Spearman rank correlations

between surface deciles and observed use intensity in 25% out-of-sample for individuals.

Additionally, we used individual-level models (fit using 75% of individual data) to predict the

distributions of all other individuals' spatial use intensity in the dataset and assessed accuracy

using the same bootstrapping method (referred to as "between individual" predictions).

Case Study #3 - Prediction to novel contexts

Lastly, to explore the predictive capacity of the SSD method to novel landscapes, we

used a dataset of eight fisher (*Pekania pennanti*) from Northern New York, USA (S2 Fig. 10)

(LaPoint *et al.* 2013) to generate SSD, RSF, and SSF probabilities considering a landscape with

square-root distance to forests as well as step length and log-step length (S2 Tables 1-2). Seven

fisher inhabited a similar environmental extent with a mix of forested and human landscapes

(approximately 10km by 15km, S2 Fig. 11), whereas one fisher occupied a completely different

landscape approximately 30km away from the others and with far less human infrastructure and

population density (S2 Fig. 15). We regarded the seven local fisher as focal individuals for

modeling, and we retained the non-local fisher as a unique opportunity to evaluate individual-

level predictive performance across environmental contexts. We calculated distance to forest

using the National Landcover Database from 2006 to identify land classes which were wetland

and deciduous, mixed, and evergreen forest. We dissolved the resulting raster to generate a

distance to forest using QGIS (v. 3.22) and GDAL proximity analysis for computational

efficiency before aggregating spatially to 91.2m by 123.2m resolution. Because individuals in

the fisher dataset had different GPS fix-rates, we resampled tracks to the median fix rate for each

animal (range: 2-15 minutes). We performed the same out-of-sample evaluations within

individuals (75/25 split) and between individual comparisons (one individual predicting the

distribution of all others) for all seven focal fishers using SSD (S2 Fig. 12), SSF (S2 Fig. 13),

and RSF surfaces (S2 Fig. 14). We then compared each of the seven focal fisher movement

models to predict the spatial distribution of the non-local fisher in the novel environmental

context, again using 100 bootstraps of the test dataset to assess the accuracy of SSD (S2 Fig. 16),

SSF (S2 Fig. 17), and RSF surfaces (S2 Fig. 18).

Supplements and code

Description of the method and simulations (S1) and empirical examples (S2) are provided

as supplements. The code used to generate these results is available online

(https://github.com/will-rogers/SSD_SSF.git) and is provided in a compiled document as a

supplement to the main text (S3). All code will be archived on Zenodo upon acceptance.

Results:

We assessed the new framework of stable-state distributions generated from local

selection and movement patterns (SSD) in its ability to predict the distribution of organisms

across landscapes. To demonstrate its performance to existing approaches we compared

predictions to those from occurrence-based (RSF) and selection-alone (SSF) frameworks. To do

so, we first isolated the effect of landscape heterogeneity and animal selection strengths in

simulations. In three subsequent empirical examples, we showed how SSD, RSF, and SSF

predictions vary in ability to predict in- and out-of-sample animal space use.

Simulation

We simulated animal movement in 100 unique combinations of landscape heterogeneity

and animal selection strength in five replicates. We estimated selection coefficients under RSF

and SSF frameworks, and compared the predictive performance of RSF, SSF, and SSD

predictions to simulated long-term space use. We found that animal selection strength created

stark disparities between known patterns of selection and observed patterns of use, particularly at

high levels of landscape heterogeneity (Figs. 1C). In more spatially autocorrelated and

heterogeneous landscapes, the Shannon entropy of selection was largely unaffected (as

autocorrelation affects the spatial structure, not the distribution, of selection values, Fig. 1D).

However, increasing animal selection strength strongly influenced the entropy of selection

surfaces (Fig. 1D). Importantly, as selection strength and landscape heterogeneity increased,

observed use patterns had much lower entropy (more spatially concentrated) than expected by

selection alone (Figs. 1D-E). Comparisons of inferential methods (RSF vs. SSF) revealed that

RSF-estimated selection coefficients were significantly biased both as individual selection

strength increased but particularly as landscape heterogeneity increased, whereas SSF inferences

appeared to generally decrease in bias as landscape heterogeneity increased (Fig. 1B, S1 Figs.

11-12). When comparing prediction surfaces, we found that SSD and, to a lesser degree, RSF

predictions better matched heterogeneity in patterns of observed use (S1 Figs. 13-15).

Uniformly, SSD was the best predictor of observed space use, distantly followed by RSF and

SSF predictions based on relative entropy, and this pattern was amplified at moderate-to-high

landscape heterogeneity and increasing animal selection strengths (Fig. 3C, S1 Figs. 18). Finally,

we found that all prediction surfaces overestimated the core space use (minimum 50% use area)

of animals as compared to observed use, particularly at low levels of habitat heterogeneity;

however, SSD delivered the most spatially concentrated predictions and was consistently the

closest estimate of core space use as compared to observed use (Fig. 3D-E, S1 Figs. 16-17).

Conversely, for heterogeneous landscapes with highly selective animals, selection-only

predictions (either SSF or the true selection kernel) overestimated core 50% use area by more

than 125-150% (Fig. 3D-E, S1 Figs. 16-17).

Case Studies

As in simulations, SSD tended to produce more spatially concentrated predictions than

SSF and RSF for the red deer (*n=1*), roe deer (*n=5*), and fisher (*n=7*) datasets (S2 Figs. 1-18).

Based on the geometric mean of predicted use, SSD was a better predictor of the modeled first

75% of data per individual than SSF for 10 of 13 animals or RSF for nine of 13 animals ("in-

sample", Fig. 4). This persisted to predictions for the next 25% of out-of-sample animal

movements, with SSD favored over SSF for nine of 13 animals and favored over RSF for eight

of 13 animals ("within individual", Fig. 4). For interindividual out-of-sample predictions, SSD

was favored over SSF and RSF for 10 of 12 animals ("between individuals", Fig. 4). Similarly,

out-of-sample predictions in a different environmental context favored SSD over SSF and RSF

for six of seven individuals ("between contexts", Fig. 4). Similar patterns were observed in

comparisons using Spearman rank correlations (S2 Fig. 19). SSD predictions were more spatially

concentrated than SSF and RSF predictions for 100% and 77% of animals, respectively (Fig. 5A-

C). Importantly, this difference between SSD and predictions lacking movement constraints

translated to substantive reductions in predicted core suitable habitat (based on minimum 50%

predicted cumulative use area): SSD predictions estimated 20-40% less core habitat than SSF

and 15-25% less core habitat than RSF.

Discussion:

Predicting animal distributions is crucial for ecological insights, conservation, and global policy. Expanding individual-level movement data offers new opportunities for biodiversity monitoring (Jetz *et al.* 2022; Kays & Wikelski 2023; Davidson *et al.* 2025) but also challenges in scaling movement patterns to landscape-level space use. For instance, mechanistic movement-based models like SSFs provide better estimates of habitat selection than pattern-based models like RSFs (Fig. 3B). However, SSFs fail to scale local selection patterns to predict landscape-level space use, resulting in paradoxically poor spatial predictions compared to RSFs (Fig. 3C). We showed that incorporating habitat selection and movement constraints in an ergodic framework (SSD) provides better predictions of animal space use than SSF or RSF in simulations (Fig. 3D) and in most empirical examples - from future movements of modeled animals to those of other animals in analogous or novel environments (Fig. 4A-C). Additionally, by including movement constraints, we demonstrated that SSF and RSF predictions can overestimate usable habitat by 15-40% (Fig. 5D). This work highlights the importance of incorporating both habitat selection and movement constraints to derive landscape-scale insights from individual-level data (Matthiopoulos *et al.* 2020).

Our approach builds on previous frameworks in connectivity modeling (Fletcher *et al.* 2019) as well as excellent theoretical (Barnett & Moorcroft 2008; Matthiopoulos *et al.* 2020; Potts & Borger 2023) and empirical addresses of movement modeling predictions (Signer *et al.* 2017; Michelot *et al.* 2019; Van Moorter *et al.* 2023; Forrest *et al.* 2024). For example, transition matrices underlie *circuitscape* (McRae *et al.* 2008) and more recent movement frameworks (Fletcher *et al.* 2022). Similarly, previous authors have shown that SSF models can be used to iteratively generate spatial distributions via agent-based frameworks (Michelot *et al.* 2019; Potts

& Borger 2023; Signer *et al.* 2024). Our innovation lies in using SSFs to parametrize transition

matrices with animal-estimated movement probabilities and in making eigendecomposition

computationally feasible. Our ergodic framework provides a convenient way to effectively

generate infinite simulations of SSF outputs, resulting in a predicted use definition which is

aligned with traditional occurrence frameworks.

Our SSD approach offers immediate benefits for conservation, movement diagnostics,

and multi-scale ecological inference. For conservation planning, SSD-based predictions

outperform RSF and SSF estimates of space use. Accurate distribution predictions are crucial for

conservation (Wiens *et al.* 2009; Evans *et al.* 2015; Araújo *et al.* 2019). By empirically capturing

movement constraints, SSDs address shortcomings in existing approaches, providing more

accurate depictions of space use. SSDs are also more conservative, indicating SSF and RSF

models likely overestimate usable habitat. As a result, our framework provides a mechanistic

alternative to occurrence-based approaches like SDMs and RSFs.

In connectivity modeling, SSDs provide an empirical method to generate transition

matrices using SSFs, where other methods rely on transformations of predicted use surfaces. We

expect this approach to result in more biologically realistic connectivity predictions, with

implications for identifying conservation priorities, movement corridors, and disease spread in

wildlife epidemiology. Further, SSD's ability to account for animal traits in underlying

movement modeling can enhance spatial prioritization efforts tailored to specific ages, sexes, or

behavioral phenotypes (e.g., migratory).

As a diagnostic of animal movement modeling, SSD-based approaches offer the ability to

spatially visualize SSF model predictions, otherwise only available through simulation from

fitted SSFs (Potts & Borger 2023; Signer *et al.* 2024). Diagnosing model parametrization issues

remains a large challenge in animal movement modeling (Fieberg *et al.* 2018; Signer *et al.*

2024). SSDs offer computationally efficient ways to map use probabilities, facilitating checks for

missing variables or overfitting. These maps provide insights into the biological effects of habitat

selection or movement constraints, currently available only via relative risk comparison (Avgar

et al. 2017). SSD may not be feasible in all scenarios, but they add a valuable tool to refine

movement models.

Finally, we envision SSDs will provide richer ecological insights into animal movement

data. For example, SSD approaches can effectively create null models of animal space use based

only on habitat selection and movement. By comparing observed space use to these simplistic

realizations of movement models, we may reveal how factors like social interactions, learning,

predators, parasites, or competitors shape individual ecology. Similarly, hierarchically organized

movement data (e.g., individual, population, species) and inferential methods (e.g., Muff *et al.*

2019) can harness SSD frameworks to address cross-population or cross-species predictions at

large spatial scales. This may provide insights into how individual heterogeneity, local

adaptation, or species differences affect space use.

Future work based on this framework should consider expanded spatial scales of

prediction and extensions to variable movement and selection dynamics. We considered

landscapes with 4,000-22,000 cells, which ran in <5 minutes on a personal computer. Larger-

scale predictions may require optimizing cell neighborhood size to balance computational speed

and accuracy. Similarly, our framework currently omits turn angles due to computational

constraints. Transition matrices model movements between states (cells) but not directional

angles, complicating Markovian processes in SSFs. Though step distance often correlates with

turn angle, this simplification warrants further exploration. Additionally, we estimated

movement kernels using step length and log-step length, sampling unused steps uniformly.

Alternative sampling distributions can improve habitat selection estimates but require correcting

predictions for the underlying sampling distributions (Fieberg *et al.* 2021; Signer *et al.* 2024).

Finally, our framework assumes static behavioral states and environmental conditions.

Incorporating behavioral or environmental variability may require simulation frameworks, such

as Forrest *et al.* (2024). Nonetheless, power iteration and similar matrix approaches could

address temporally dynamic behaviors and landscapes.

In a changing world, animal-environment interactions are also likely to shift, challenging

pattern-based models whose projections may falter when decoupled from mechanisms.

Mechanistic models, though adaptable, lose accuracy if generating mechanisms are omitted. Our

findings suggest combining movement and selection data in spatial projections can bridge these

gaps, enhancing the utility of individual movement data for ecological and conservation

applications.

Acknowledgments:

We thank Johannes Signer, Francesca Cagnacci, and Scott LaPoint for their generosity in

providing the red deer, roe deer, and fisher datasets (respectively) as open-access resources for

the scientific community, and for their permission to use those datasets in our study. We also

thank John Fieberg for initial discussion and members of the Jetz and Ezenwa labs for their

generous feedback on various stages of the project. WR acknowledges funding from the Gruber

Foundation. SY acknowledges funding from the Max Planck – Yale Center for Biodiversity

Movement and Global Change.

References:

1.
Araújo, M.B., Anderson, R.P., Márcia Barbosa, A., Beale, C.M., Dormann, C.F., Early, R. *et al.* (2019). Standards for distribution models in biodiversity assessments. *Science Advances*, 5, eaat4858.

2.
Avgar, T., Lele, S.R., Keim, J.L. & Boyce, M.S. (2017). Relative Selection Strength: Quantifying effect size in habitat- and step-selection inference. *Ecol Evol*, 7, 5322-5330.

3.
Avgar, T., Potts, J.R., Lewis, M.A. & Boyce, M.S. (2016). Integrated step selection analysis: Bridging the gap between resource selection and animal movement. *Methods in Ecology and Evolution*, 7, 619-630.

4.
Barnett, A.H. & Moorcroft, P.R. (2008). Analytic steady-state space use patterns and rapid computations in mechanistic home range analysis. *J Math Biol*, 57, 139-159.

5.
Boyce, M.S. & McDonald, L.L. (1999). Relating populations to habitats using resource selection functions. *Trends in Ecology & Evolution*, 14, 268-272.

6.
Boyce, M.S., Vernier, P.R., Nielsen, S.E. & Schmiegelow, F.K.A. (2002). Evaluating resource selection functions. *Ecological Modelling*, 157, 281-300.

7.
Bozinovic, F., Bastías, D.A., Boher, F., Clavijo-Baquet, S., Estay, S.A. & Jr., M.J.A. (2011). The Mean and Variance of Environmental Temperature Interact to Determine Physiological Tolerance and Fitness. *Physiological and Biochemical Zoology*, 84, 543-552.

8.
Buckley, L.B., Waaser, S.A., MacLean, H.J. & Fox, R. (2011). Does including physiology improve species distribution model predictions of responses to recent climate change? *Ecology*, 92, 2214-2221.

9.
Cagnacci, F., Focardi, S., Heurich, M., Stache, A., Hewison, A.J.M., Morellet, N. *et al.* (2011). Partial migration in roe deer: migratory and resident tactics are end points of a behavioural gradient determined by ecological factors. *Oikos*, 120, 1790-1802.

10.
Cushman, S.A. & Lewis, J.S. (2010). Movement behavior explains genetic differentiation in American black bears. *Landscape Ecology*, 25, 1613-1625.

11.
Davidson, S.C., Cagnacci, F., Newman, P., Dettki, H., Urbano, F., Desmet, P. *et al.* (2025). Establishing bio-logging data collections as dynamic archives of animal life on Earth. *Nature Ecology & Evolution*, 9, 204-213.

12.
Elith, J. & Leathwick, J.R. (2009). Species Distribution Models: Ecological Explanation and Prediction Across Space and Time. *Annual Review of Ecology, Evolution, and Systematics*, 40, 677-697.

13.

Elith, J., Phillips, S.J., Hastie, T., Dudík, M., Chee, Y.E. & Yates, C.J. (2011). A statistical
explanation of MaxEnt for ecologists. *Diversity and Distributions*, 17, 43-57.
14.
Evans, T.G., Diamond, S.E. & Kelly, M.W. (2015). Mechanistic species distribution modelling
as a link between physiology and conservation. *Conservation Physiology*, 3.
15.
Fieberg, J., Signer, J., Smith, B. & Avgar, T. (2021). A 'How to' guide for interpreting
parameters in habitat-selection analyses. *J Anim Ecol*, 90, 1027-1043.
16.
Fieberg, J.R., Forester, J.D., Street, G.M., Johnson, D.H., ArchMiller, A.A. & Matthiopoulos, J.
(2018). Used□habitat calibration plots: A new procedure for validating species
distribution, resource selection, and step□selection models. *Ecography*, 41, 737-752.
17.
Fletcher, R.J., Jr., Sefair, J.A., Wang, C., Poli, C.L., Smith, T.A.H., Bruna, E.M. *et al.* (2019).
Towards a unified framework for connectivity that disentangles movement and mortality
in space and time. *Ecol Lett*, 22, 1680-1689.
18.
Fletcher, R.J., Sefair, J.A., Kortessis, N., Jaffe, R., Holt, R.D., Robertson, E.P. *et al.* (2022).
Extending isolation by resistance to predict genetic connectivity. *Methods in Ecology and
Evolution*, 13, 2463-2477.
19.
Forrest, S.W., Pagendam, D., Bode, M., Drovandi, C., Potts, J.R., Perry, J. *et al.* (2024).
Predicting fine-scale distributions and emergent spatiotemporal patterns from temporally
dynamic step selection simulations. *Ecography*, n/a, e07421.
20.
Hanson, J.O., Schuster, R., Strimas-Mackey, M., Morrell, N., Edwards, B.P.M., Arcese, P. *et al.*
(2024). Systematic conservation prioritization with the prioritizr R package.
Conservation Biology, e14376.
21.
Hooten, M.B., Johnson, D.S., McClintock, B.T. & Morales, J.M. (2017). *Animal movement:
Statistical models for telemetry data*.
22.
Jetz, W., Tertitski, G., Kays, R., Mueller, U., Wikelski, M., Åkesson, S. *et al.* (2022). Biological
Earth observation with animal sensors. *Trends in Ecology & Evolution*, 37, 293-298.
23.
Johnson, C.J., Nielsen, S.E., Merrill, E.H., McDonald, T.L. & Boyce, M.S. (2006). Resource
Selection Functions Based on Use-Availability Data: Theoretical Motivation and
Evaluation Methods. *The Journal of Wildlife Management*, 70, 347-357.
24.
Kays, R. & Wikelski, M. (2023). The Internet of Animals: what it is, what it could be. *Trends in
Ecology & Evolution*.
25.
Kearney, M. & Porter, W. (2009). Mechanistic niche modelling: combining physiological and
spatial data to predict species' ranges. *Ecology Letters*, 12, 334-350.
26.

Keating, K.A. & Cherry, S. (2004). USE AND INTERPRETATION OF LOGISTIC
REGRESSION IN HABITAT-SELECTION STUDIES. *The Journal of Wildlife*
Management, 68, 774-789.
27.
LaPoint, S., Gallery, P., Wikelski, M. & Kays, R. (2013). Animal behavior, cost-based corridor
models, and real corridors. *Landscape Ecology*, 28, 1615-1630.
28.
Mackey, B.G. & Lindenmayer, D.B. (2001). Towards a hierarchical framework for modelling
the spatial distribution of animals. *Journal of Biogeography*, 28, 1147-1166.
29.
Madin, J.S., Hoogenboom, M.O. & Connolly, S.R. (2012). Integrating physiological and
biomechanical drivers of population growth over environmental gradients on coral reefs.
Journal of Experimental Biology, 215, 968-976.
30.
Manly, B., McDonald, L., Thomas, D.L., McDonald, T.L. & Erickson, W.P. (2007). *Resource*
selection by animals: statistical design and analysis for field studies. Springer Science &
Business Media.
31.
Matthiopoulos, J. (2022). Defining, estimating, and understanding the fundamental niches of
complex animals in heterogeneous environments. *Ecological Monographs*, 92, e1545.
32.
Matthiopoulos, J., Fieberg, J., Aarts, G., Barraquand, F. & Kendall, B.E. (2020). Within Reach?
Habitat Availability as a Function of Individual Mobility and Spatial Structuring. *Am Nat*,
195, 1009-1026.
33.
McRae, B.H., Dickson, B.G., Keitt, T.H. & Shah, V.B. (2008). Using circuit theory to model
connectivity in ecology, evolution, and conservation. *Ecology*, 89, 2712-2724.
34.
Merow, C., Smith, M.J., Edwards Jr, T.C., Guisan, A., McMahon, S.M., Normand, S. *et al.*
(2014). What do we gain from simplicity versus complexity in species distribution
models? *Ecography*, 37, 1267-1281.
35.
Michelot, T., Blackwell, P.G. & Matthiopoulos, J. (2019). Linking resource selection and step
selection models for habitat preferences in animals. *Ecology*, 100, e02452.
36.
Moorcroft, P.R. & Barnett, A. (2008). Mechanistic home range models and resource selection
analysis: A reconciliation and unification. *Ecology*, 89, 1112-1119.
37.
Morris, L.R., Proffitt, K.M. & Blackburn, J.K. (2016). Mapping Resource Selection Functions in
Wildlife Studies: Concerns and Recommendations. *Appl Geogr*, 76, 173-183.
38.
Muff, S., Signer, J. & Fieberg, J. (2019). Accounting for Individual☐specific Variation in
Habitat☐selection Studies: Efficient Estimation of Mixed☐effects Models Using
Bayesian or Frequentist Computation. *Journal of Animal Ecology*.
39.

Nathan, R., Getz, W.M., Revilla, E., Holyoak, M., Kadmon, R., Saltz, D. *et al.* (2008). A
movement ecology paradigm for unifying organismal movement research. *Proceedings of*
the National Academy of Sciences of the United States of America, 105, 19052-19059.
40.
Newbold, T., Reader, T., El-Gabbas, A., Berg, W., Shohdi, W.M., Zalat, S. *et al.* (2010). Testing
the accuracy of species distribution models using species records from a new field
survey. *Oikos*, 119, 1326-1334.
41.
Northrup, J.M., Vander Wal, E., Bonar, M., Fieberg, J., Laforge, M.P., Leclerc, M. *et al.* (2022).
Conceptual and methodological advances in habitat-selection modeling: guidelines for
ecology and evolution. *Ecological Applications*, 32.
42.
Pearson, R.G. & Dawson, T.P. (2003). Predicting the impacts of climate change on the
distribution of species: are bioclimate envelope models useful? *Global Ecology and*
Biogeography, 12, 361-371.
43.
Phillips, S.J., Dudík, M., Elith, J., Graham, C.H., Lehmann, A., Leathwick, J. *et al.* (2009).
Sample selection bias and presence-only distribution models: implications for
background and pseudo-absence data. *Ecological Applications*, 19, 181-197.
44.
Potts, J.R. & Borger, L. (2023). How to scale up from animal movement decisions to
spatiotemporal patterns: An approach via step selection. *J Anim Ecol*, 92, 16-29.
45.
Sastre, P. & Lobo, J.M. (2009). Taxonomist survey biases and the unveiling of biodiversity
patterns. *Biological Conservation*, 142, 462-467.
46.
Schick, R.S., Loarie, S.R., Colchero, F., Best, B.D., Boustany, A., Conde, D.A. *et al.* (2008).
Understanding movement data and movement processes: Current and emerging
directions. *Ecology Letters*, 11, 1338-1350.
47.
Signer, J., Fieberg, J. & Avgar, T. (2017). Estimating utilization distributions from fitted step-
selection functions. *Ecosphere*, 8, e01771.
48.
Signer, J., Fieberg, J. & Avgar, T. (2019). Animal movement tools (amt): R package for
managing tracking data and conducting habitat selection analyses. *Ecology and*
Evolution, 9, 880-890.
49.
Signer, J., Fieberg, J., Reineking, B., Schlägel, U., Smith, B., Balkenhol, N. *et al.* (2024).
Simulating animal space use from fitted integrated Step-Selection Functions (iSSF).
Methods in Ecology and Evolution, 15, 43-50.
50.
Soberon, J. & Peterson, A.T. (2005). Interpretation of Models of Fundamental Ecological Niches
and Species' Distributional Areas. *Biodiversity Informatics*, 2.
51.
Urban, M.C. (2015). Accelerating extinction risk from climate change. *Science*, 348, 571-573.
52.

Van Moorter, B., Kivimaki, I., Panzacchi, M., Saura, S., Brandao Niebuhr, B., Strand, O. *et al.* (2023). Habitat functionality: Integrating environmental and geographic space in niche modeling for conservation planning. *Ecology*, e4105.

53.

Wiens, J.A., Stralberg, D., Jongsomjit, D., Howell, C.A. & Snyder, M.A. (2009). Niches, models, and climate change: Assessing the assumptions and uncertainties. *Proceedings of the National Academy of Sciences*, 106, 19729-19736.

54.

Wisz, M.S. & Guisan, A. (2009). Do pseudo-absence selection strategies influence species distribution models and their predictions? An information-theoretic approach based on simulated data. *BMC Ecology*, 9, 8.

55.

Woodin, S.A., Hilbish, T.J., Helmuth, B., Jones, S.J. & Wethey, D.S. (2013). Climate change, species distribution models, and physiological performance metrics: predicting when biogeographic models are likely to fail. *Ecology and Evolution*, 3, 3334-3346.

660   **Figures**

662   **Figure 1: Habitat preferences do not equal realized use in simulations**

665   **Figure 2: Movement data-based landscape predictions**

Figure 3: Predictions incorporating movement mechanisms recover simulated patterns of

use

Figure 4: Predictions incorporating movement mechanisms better predict out-of-sample

animal movements

Figure 5: Predictions incorporating movement mechanisms are more spatially

concentrated than those based on selection alone

Figure Legends:

Figure 1: Habitat preferences do not equal realized use in simulations

(**A**) The theoretical frequency of use of a single-dimensional environment based on the

preference (potential use, red) vs. the realized frequency of use given movement constraints in a

heterogeneous habitat (blue). (**B**) Corresponding depictions of the frequency of use of locations

in the landscape without (red) and with movement constraints (blue). (**C**) The effect of landscape

heterogeneity (Gaussian random fields scale) and habitat selection strength multiplier (2 – weak;

8 - strong) on the relative difference between habitat selection and true probability of use in a

single set of simulations, calculated as the relative selection kernel minus true use probability

kernel divided by the relative selection kernel. Differences in surfaces are red (where realized

selection overpredicts potential use probability) and blue (where realized selection underpredicts

potential use probability) and contours highlight where transitions occur. (**D**) The mean Shannon

entropy across 5 simulation iterations per level of selection strength multiplier (y-axis) and

landscape heterogeneity (x-axis) for the relative selection kernel and the true probability of use.

(**E**) For two sets of simulations (with specialist selection [multiplier = 8] and high heterogeneity

[scale = 0.6] or generalist selection [multiplier = 2] and low heterogeneity [scale = 0.01];

highlighted by black squares in **C** and **D**), density plots of relative probability of use based on

selection kernels (red) and actual probability of use in simulations (blue).

Figure 2: Movement data-based landscape predictions

Converting animal movements into spatial predictions of use can occur in three main ways. First,

(**A**) animal movements are compared to environmental contexts in landscapes made of discrete

cells, *G*. (**B**) Animal locations can then be placed into use vs. available frameworks conditional

on the landscape, suitable for resource selection analyses (RSFs). (**C**) Alternatively, tracking data

can be converted into use vs. available frameworks conditional on steps, suitable for step

selection analyses (SSFs). (**D**) RSFs predict the probability of use over a landscape, which can be

kernelized to return (**I**) probabilities of use over G. (**E**) SSFs predict the relative probability

between comparisons within a context, a distinct metric from the probability of use. (**F**) It is

common to make an arbitrary comparison between some either observed or unobserved

environmental contexts in G to all possible contexts in G. (**J**) These spatial relative risks are then

kernelized, representing something akin to RSF output, though mathematically inequivalent. (**G**)

Harnessing the ability of SSFs to parametrize local movement patterns, we can algebraically

solve the probability of moving within a set of local environmental contexts (G_j) by defining that

the relative selection strength for a focal cell (G_i) compared to itself must equal 1. (**H**)

Aggregating patterns of in- and out-selection over all neighborhoods of all G_i resolves a sparse

stochastic matrix for all reasonable movements. (**K**) Solving for the principal eigenvector of the

stochastic matrix returns the stable state distribution - the probability of ending in some G_i given

all possible starts over the landscape, G.

Figure 3: Predictions incorporating movement mechanisms recover simulated patterns of

use

Within simulations, we generated (**A**) a selection kernel (the product of selection strengths for

four randomly generated environmental variables), simulated use probability from 500,000 steps

of animals with constraints on step length and turn angle, and predictions generated from a focal

set of 10,000 steps for resource selection function (RSF), kernelized step-selection function

(SSF) relative selection strength, and the stable state distribution (SSD). (**B**) Five sets of

simulations with different levels of Gaussian random fields scale (x-axis) and strengths of animal

selection (multipliers, color scale) revealed differences between the estimated selection

coefficient for each environmental variable and the true value of the selection coefficient based

on RSF or SSF frameworks. We converted errors and 95% confidence intervals (CIs) such that

directions of comparison are aligned (e.g., positive differences imply more extreme inferences,

irrespective of positive or negative selection). We summarized the performance of predictions

relative to simulated use via (**C**) relative entropy and (**D**) top 50% use area of known selection

kernels and RSF, SSF, and SSD predictions vs. simulated use across varying levels of

heterogeneity (x-axis) and animal selection strengths (horizontal facet). (**E**) SSD, SSF, and RSF

resulted in distinct predictions about animal space use, highlighted by classification (fill color) of

habitat into top 50% use area vs. simulated use for RSF, SSF, and SSD predictions (vertical

facet) and different levels of selection strength (horizontal facet).

Figure 4: Predictions incorporating movement mechanisms better predict out-of-sample

animal movements

(**A-C**) Median and 95% quantiles in 100 bootstraps of differences between stable state

distribution (SSD) and resource selection function (RSF) or step-selection function (SSF)

predictions in geometric mean of predicted probability of use at observed use sites. Bootstraps

were performed by resampling modeled and out-of-sample data with replacement for (**A**) red

deer (*Cervus elaphus*), (**B**) roe deer (*Capreolus capreolus*), and (**C**) fisher (*Pekania pennanti*)

datasets. Validations were performed in four sets of comparisons, we used the first

(chronologically) 75% of observed locations for individuals to predict on the modeled data ("In-

Sample"), the next (chronologically) 25% of observed locations per individual ("Within

Individual"), all observed locations for the other individuals in a dataset within the same

geographic context ("Between Individuals"), and all observed locations for an individual in a

completely different environmental context than the individuals modeled ("Between Contexts").

For visualization, the median difference for each set of comparisons is highlighted by a

horizontal grey bar.

Figure 5: Predictions incorporating movement mechanisms are more spatially

concentrated than those based on selection alone

(**A**) Spatial predictions of "use" (as a kernel over the landscape) for a red deer, (**B**) summer use

of a single roe deer ("M10"), (**C**) and a single fisher ("M2") based on output from a resource

selection function (RSF), a step-selection function (SSF), and the stable state distribution (SSD)

from local movement patterns estimated from the SSF. (**D**) For each case study, the relative

difference in minimum area containing 50% of predicted use for SSD vs. RSF (RSF-SSD/RSF)

and SSF predictions (SSF-SSD/SSF) are plotted for individuals. For visualization, the mean and

range of these relative differences are overlaid on individual estimates. These results highlight

that irrespective of accuracy, incorporating movement constraints in SSD always generates more

spatially concentrated predictions than SSF, and SSD tends to also be more spatially

concentrated than most RSF estimates, meaning that less area may be truly usable for animals

after incorporating movement constraints.

773