Age-specificity in territory quality and spatial structure in a wild bird population

Joe P. Woodman[1], Josh A. Firth[1,2], Ella F. Cole[1] & Ben C. Sheldon[1]

[1] Edward Grey Institute of Field Ornithology, Department of Biology, University of Oxford, Oxford, UK

[2] School of Biology, University of Leeds, Leeds, UK

ABSTRACT

Age influences behaviour, survival, and reproduction; hence variation in population age structure can
affect population-level dynamics. The extent of spatial age structure may be important in driving
spatially-variable demography, particularly when space-use is linked to reproduction, yet it is not well
understood. Here, we use long-term data from a wild passerine bird population to quantify covariance
between territory quality and age and assess spatial age structure. We find associations between age
and aspects of territory quality, but little evidence for spatial age structure compared to territory quality
and reproductive spatial structure. We also report little between-year repeatability of spatial age
structure compared to structure in reproductive output. We suggest that high breeding site fidelity and
frequent territory turnover by younger breeders, driven by high mortality and immigration rates, limits
the association between age and territory quality and weakens overall spatial age structure. Greater
spatial structure and repeatability in reproductive output compared to age suggests that in this system
habitat quality may be more important in driving spatially-variable demography than age. We suggest
that the framework developed here can be used in many other populations to assess the causes and
consequences of spatial age structure.

Key words

Age structure; breeding density; *Parus major*; spatially-variable demography; territory quality; territory
choice

(1) Introduction

Age affects many aspects of life, as resources are allocated into processes and traits at different points
throughout lifespan to maximise fitness [1] and individuals gain in experience as they age. Further,
variation in population age structure may affect population dynamics, via effects on reproductive output
and social organisation [2–5]. There is widespread evidence of temporal variation in age structure in
wild populations [4,6,7], but the way in which age structure varies across space seems to be much
less well understood. Some work identifies spatial variation in age structure in fish populations, often
linked to human harvesting of different age-cohorts in given areas [8,9]; and in ungulate and bird
populations, associated with age-specificity in habitat use [10,11]. Given that age is often important
for individuals' fitness, spatial age structure is particularly interesting as it might drive local age-related
population dynamics. For example, if there is covariance between habitat quality and the age of

individuals occupying sites, both could lead to spatially-variable demography, thus potentially biassing

estimates of the effects of age on reproduction or, conversely, estimates of environmental effects on

trait variation.

One social system where covariance between age and habitat might lead to local age structure is

where individuals defend breeding territories. Here, age structure might be particularly important in

driving spatial age-related dynamics due to the tight association between breeding territories and

reproductive output [12–14]. In many animal taxa, individuals undergo reproductive attempts while

defending a territory in which resources are used for breeding [15,16]. This is particularly prevalent in

socially monogamous bird species [17,18]. In such cases, spatial age structure could develop as non-

randomness arises when individuals of either similar or dissimilar age breed in closer proximity to each

other than expected from chance.

Several mechanisms might affect spatial age structure in territorial animals; we outline four

mechanisms here. (1) First, temporal variation in population age structure might influence how age is

arranged in space. Temporal variation in wild population age structure is common, particularly in short-

lived species where the proportion of populations consisting of the youngest age-cohorts may vary

greatly between breeding seasons [7,19]. Such variation might affect how age is arranged in space

through passive mechanisms whereby clusters of territories occupied by same-age individuals will be

more likely to arise when the age distribution is skewed towards this age-cohort. For example, recent

work demonstrates how fluctuations in population age structure determine how age is structured within

breeding pairs in a short-lived bird population, where there is greater age-assortative pairing when the

proportion of yearlings is higher [19]. Thus, between-year age distribution is likely to passively affect

spatial age structure as clusters of territories occupied by same-age individuals are more likely to arise

when much of the population exists in a single age-cohort (Figure 1a).

However, age-specific biases in settlement may drive spatial age structure over-and-above that

expected from temporal variation in age distribution alone. This may occur if there is spatial structure

in territory quality (which may be generated for example through spatial clustering of resources), and

there is covariance between age and territory quality, which may arise through two mechanisms. (2a)

In many species, older individuals are dominant [20,21], thus potentially leading to age-specificity in

territory use where older individuals outcompete younger ones to acquire higher quality territories [22–

25] (Figure 1b). (2b) Alternatively, covariance between age and territory quality might arise if higher

quality territories elevate survival, and individuals therefore persist for longer in such territories over

time [26]. Additionally, individuals might be more likely to remain on higher quality territories, which

will generate a similar association between length of tenure and hence age (Figure 1b). Therefore, if

older individuals are more likely to occupy higher quality territories through either of these
mechanisms, spatial structure in territory quality might drive spatial age structure and repeatability in
this over time.

(3) Age-specific territory occupation might also be generated through mechanisms that do not directly
link to how ageing affects territory acquisition. For example, if there is individual-specificity in territory
acquisition mediated by states or phenotypes other than age, but there is covariance between these
and age, then association may arise between age and the type of territory (despite no causal direct
link between the two). For example, territory-specific occupation depending on prior dispersal is
relatively common across taxa [27]. Specifically, in fragmented landscapes, sites have environmental
'edges', such as forest edges which provide an interface between internal woodland and the external
environment [28]. In edge environments, there is often a higher incidence of individuals that disperse
into the site [29,30]. Dispersing between birth and first breeding, known as natal dispersal, generally
involves greater distances than dispersal between two breeding attempts [31,32]. Thus, individuals
moving into a new environment are often younger [18,33,34]. In such cases, if edge territories are
more commonly occupied by younger immigrant individuals, then this might provide a mechanism
through which spatial age structure develops (Figure 1c). Edge territories provide a particularly
interesting mechanism by which spatial age structure is generated, not only because of immigrant-
specific occupation, but also because they are often of poorer quality [30,35,36], thus they might also
directly generate spatial age structure if there is covariance between age and territory quality
(mechanisms 2a and 2b).

Figure 1 – Overview of mechanisms that might generate spatial age structure across territories
referred to in the introduction. The theoretical populations consist of 200 individuals, each of which are
associated with a territory, illustrated across Wytham Woods, Oxford, as an example. Yellow and
green territories are occupied by 'young' and 'old' individuals respectively. In (a), individuals occupy

territories randomly, but at time *t* the population consists of a 1:1 ratio of old/young individuals; and at

time *t* + 1 a 1:3 ratio of old/young individuals. In (b), the population consists of a 1:1 ratio of old/young

individuals, but older individuals occupy the territories of highest quality. In (c), territories which do not

border the edge are occupied by a 3:1 ratio of old/young individuals, whereas territories with an edge

boundary are occupied by a ratio 1:3 of old/young individuals.

These mechanisms lead to *a priori* expectations that spatial age structure might develop across

breeding territories; however, empirical evidence of this is currently sparse. Additionally, although

these mechanisms might act concurrently on individuals in time, the areas in space that bias age-

specific settlement may differ from each other. For example, if different aspects of territory quality are

independent then the relative strength of these resource distributions, if any, in influencing age-specific

settlement is unknown. Further, how such mechanisms bias age-specific settlement might be

influenced by sex given the different roles that males and females will often play in territory acquisition

[18,37]. In short, given the role that age structure might play in spatially-structured demography, it is

important that we advance our understanding on the evidence for spatial age structure and the

mechanisms that drive it.

Here, we outline processes and their consequences for generating spatial age structure on long-term

breeding data from a natural population of great tits *Parus major* in Wytham Woods, Oxford. We first

assess whether there is covariance between territory quality and the age of individuals in such

territories. Second, we evaluate whether there is non-random spatial arrangement of age within

breeding seasons. Finally, we assess evidence for age-biases in space that persist across time

compared to spatial structure in territory quality and reproductive output. We discuss the relative roles

of spatial age structure versus structure in territory quality for producing spatially-variable demography,

and suggest that the framework presented here can be used across populations when considering

spatial age structure.

(2) Methods

(i) Study system and data collection

The great tit *Parus major* is a passerine bird found in woodlands across Europe, with breeding ages

ranging 1–9, averaging 1.8 years [38–40]. Although there are some continuous changes with age [38],

the main age effects on individual-level traits are captured by two age-classes: first-years (hereafter

juveniles) and older (hereafter adults [39–43]). The species is socially monogamous, with pairs

defending territories during annual breeding seasons [44]. Data used here are from a long-term study

in Wytham Woods, Oxford (51°46'N, 1°20'W), a 385ha mixed deciduous woodland surrounded by

farmland [45]. The tit population has been monitored since 1947, where breeding adults and their

chicks have been marked with unique BTO (British Trust for Ornithology) rings since the 1960s; and

standard reproductive metrics are collected [46]. Individuals breed almost exclusively in the 1026 nest-

boxes which are in fixed positions with known GPS coordinates [47,48]. All chicks are ringed at 14-

days of age, while adults are trapped at nest-boxes and identified by ring number, or marked with a

new ring if they are immigrants. Age is based on year of hatching for local birds, or plumage

characteristics for immigrants [49]. Although immigration rates are high (46%), most are first caught

as yearlings (78%) and can therefore be aged accurately.

(ii) Data selection

We constructed a dataset that assigns the year of hatching to all individuals between 1950–2022,

across which exact age was calculated for 88.8% of 46062 identified breeding individuals. In this study,

we included birds in analyses that attempted to breed between 1978–2022, for which data were more

complete compared to earlier dates (Figure S1). Individuals that were first caught post-fledging are

assumed to be immigrants, as locally-hatched tits are marked as nestlings in nest-boxes and the

proportion of birds hatched in natural cavities is very low [50]. Immigrants that entered the population

with adult plumage were assigned a minimum age of 2, and subsequent age estimates were based

on this (6.7% and 10.0% of breeding females and males). Age was therefore determined for 68.7% of

breeding individuals where at least one egg was laid (due to a combination of nests failing prior to

adult trapping and unsuccessful trapping attempts, there are cases where the identity of parents is

unknown).

(iii) Statistical analyses

Determining breeding territories

We defined annual breeding territories through a Dirichlet tessellation technique that forms Thiessen

polygons [51,52] around each occupied nest-box. The polygon includes all space within the habitat

that is closer to the focal box than any other (with a boundary also imposed by the woodland edge).

This metric of territory has been shown to be biologically meaningful in terms of territory size and

territorial neighbours in tit species and is strongly related to other methods of calculating territories

[47,53–56]. However, a limitation is that unrealistically large polygons are formed in areas where nest-

boxes are placed at great distances from each other. We therefore capped territories at 2ha, which is

a more realistic maximum spatial scale at which individuals use territories, as supported in previous

analytical and field studies [30,47,48,57].

Age and territory quality

We first assessed covariance between the age of individuals and their territories' quality. We measured
territory quality through four measures: the number of oak trees *Quercus* spp. within 75m of the nest-
box; average territory density; the edge distance index; and the long-term nest-box popularity index.
Each of these is justified below.

Great tits predominantly provision offspring with caterpillars collected close to their nests [40], thus
variation in caterpillar availability is directly linked to reproductive success [58–60]. Caterpillars are
found most abundantly on oak trees [40,58], therefore oak proximity, health and abundance is
important for breeding success [47,55,61,62]. A radius of 75m was chosen as the abundance of oaks
within this distance has been shown to be particularly important for breeding [63,64].

The density of conspecifics breeding in proximity may influence resource availability if foraging ranges
overlap, and therefore territory density may also represent an aspect of territory quality. Additionally,
territory density may affect site quality through social mechanisms, such as increased competition and
emergent need for territory defence leading to reduced foraging [65], or conversely mutual benefits
between familiar neighbours [55,66,67]. We calculated average territory density directly from the
Thiessen polygon area produced from tessellation by taking the reciprocal of the mean polygon area.

Territories at woodland edges are associated with lower reproductive success in great tits [30].
Following Wilkin et al. (2007) [30], we defined the edge distance index (EDI) for each nest-box by
multiplying the distance to forest edge by the proportion of woodland habitat within a 75m radius of
the box. Thus, boxes within 75m of the edge have an EDI value in proportion to the amount of
woodland habitat within this radius, therefore considering not only the distance to edge, but also the
number and geometric arrangement of edges relative to nest-box.

Finally, the frequency a territory is occupied in the long-term may provide a measure of quality as
individuals may choose sites that confer reproductive benefits more often, as evidenced in other
species [68–72]. There is evidence of this in Wytham, where the number of times a nest-box has been
occupied positively correlates with the average number of offspring that fledge per breeding attempt
(Figure S2). We therefore calculated the number of times a nest-box has been occupied since 1965.
However, this is related to the number of nest-boxes within close proximity to a focal box (Figure S3),
because in areas of high nest-box density there are multiple unoccupied boxes which would likely be
associated with the same territorial range if they were occupied (thus, in regions of high box density,
birds may re-occupy the same territory over multiple years, but not necessarily the exact same box).
To correct for this, we ran a linear model between the number of boxes within 30m of a focal nest-box

(supporting information) and the number of times said box has been occupied, and took the residuals

as the long-term nest-box popularity index.

We constructed a generalised linear mixed-effects model assuming a binomial error distribution to

analyse the association between these four measures of territory quality and the age of the breeding

individual. We modelled age (juvenile/adult) as the response variable, with the territory quality

measures as explanatory variables, which were z-transformed to compare their relative effects in

predicting age. Individual ID, nest-box ID, and breeding year were included as random effects (Table

S1). We ran three sets of these models: one with all individuals; one with only females; and one with

only males, allowing us to assess potential sex-specific differences in the association between aspects

of territory quality and age. These models were also run to test for covariance between territory quality

and residency status (supporting information). All models were conducted in R statistical software [73]

using the *lme4* package [74].

Spatial age structure

For each year's breeding population, we constructed an individual-by-individual matrix denoting

breeding neighbours i.e. a network of breeding territories, where nodes represent individuals and

edges represent the spatial connectivity of territories. Specifically, edges connected individuals if their

territories share a boundary from the tessellation technique and were weighted relative to the distance

between nest-boxes of neighbouring territories.

We created three networks per year: one with all individuals (but removing edges within breeding pairs

that occupy the same territory); one with only females; and one with only males. Edges connecting

individuals of unknown identity were removed. Across these networks, we calculated the assortativity

coefficient of age (juvenile/adult), which measures the correlation between individuals' age and that of

their territorial neighbours accounting for edge weight (proximity of neighbouring nest-boxes) and the

relative proportion of the two age classes across the network. Through this technique, the age-

composition of neighbourhoods of birds that are likely to interact during territorial and foraging

behaviour contribute to the emergent quantitative signal of spatial age structure to a greater degree

than they would if spatial autocorrelation were calculated through pairwise distance of all individuals

across the system. Thus, this method allows us to assess evidence for spatial age structure at a

biologically relevant scale (analyses were also run treating age as a discrete and continuous trait, as

well as calculating assortment of residency status; supporting information). We also ran analyses to

compare spatial age structure with spatial structuring of territory quality and reproductive output. To

calculate territory quality structure, we assigned the value from the previously described four measures

of quality to each node (associated with an occupied nest-box) and calculated the territory quality

assortativity coefficient across the network. For spatial reproductive output structure, we ran parallel

analyses, but calculated the assortativity in clutch size, chick number, fledgling number, and binary

success (where 0 is no fledglings and 1 is at least one fledgling) associated with each nest-box. These

analyses were run using the *assortnet* package [75].

Temporal repeatability of spatial age structure

Finally, we tested whether spatial regions show temporal repeatability in age-composition, territory

quality, and reproductive output. To assess this, within each year, we defined a radius around each

occupied nest-box that corresponded to an area of 25, 50, 100, 150 and 200ha, representing

neighbourhoods of breeding individuals of variable population sizes. Within each radius we calculated:

the proportion of individuals that were adults; the mean number of oaks within 75m of the focal boxes;

mean territory density; mean edge distance index; mean nest-box popularity index; mean clutch size;

mean chick number; mean fledgling number; and proportion of boxes with binary success. We then

calculated the same metrics for all nest-boxes outside of the focal radius. From this, we calculated a

ratio index of within versus outside radius for each calculated measure, where a value of one

represents the same annual average measure within and outside the radius. We then tested the

repeatability of the ratio index associated with each spatial scale across the 45 years of data by

calculating the intraclass correlation coefficient (ICC) derived from a linear mixed-effects model, where

the response variable is the ratio index, and year and radius area are fitted as categorical grouping

variables.

(3) Results

Age and territory quality

Across 20121 breeding individuals, representing 11167 breeding attempts, we found that high density

territories weakly predicted occupation by adults compared to juveniles (odds ratio = 1.108, 95%

confidence intervals = 1.065–1.153, $p < 0.001$), in addition to territories with a greater long-term

popularity index (OR = 1.084, 95% CIs = 1.045–1.124, $p < 0.001$), with little evidence that the number

of oaks within 75m or EDI were linked to age (Figure 2; Table S2). When assessing the influence of

territory quality on the age of individuals separately for both sexes, similar patterns were found, except

territories that have a greater long-term popularity index more strongly predicted occupation by adult

males (OR = 1.143, 95% CIs = 1.092–1.195, $p < 0.001$), but were weakly if at all associated with

female age (OR = 1.041, 95% CIs = 0.997–1.086, $p = 0.066$).

Figure 2 – Point estimates of odds ratios obtained from generalised linear mixed-effects models. The

analysis examines the association between territory quality on the odds of the breeding individual

being an adult. Each point corresponds to a specific level of the fixed effects (on the y-axis), and error

bars denote 95% confidence intervals. Green points are from analysis assessing all individuals, purple

are only females, and blue are only males.

Spatial age structure

Generally, we found weak positive age assortment across breeding territories, but there was temporal

variation with many years having little evidence for spatial age structure and some having significant

disassortment (i.e. birds of similar age less likely to be associated), whether considering the whole

population (median r age assortativity coefficient = 0.014, 51% of annual standard errors overlap zero;

Figure 3), females (median r = 0.002; 58% SEs overlap zero), or males (median r = -0.004; 38% SEs

overlap zero). In contrast, there was much greater spatial structure in territory quality (oak abundance

median r = 0.782; territory density median r = 0.645; EDI median r = 0.817; and nest-box popularity

index median r = 0.258, no annual SEs overlap zero). There is greater evidence for spatial

reproductive output structure compared to age, but the signal of this structure is still relatively weak

(clutch size median r = 0.059, 22% SEs overlap zero; chick number median r = 0.047, 29% SEs

overlap zero; fledgling number median r = 0.075, 22% SEs overlap zero; binary success median

r = 0.057, 24% SEs overlap zero; Figure 3; Table S3 for all results).

Figure 3 – Annual age and fledgling number assortativity *r* coefficients across territories. Large purple

points represent age assortativity values where the standard error does not overlap zero, and large

green points are fledgling number assortativity values where standard error does not overlap zero.

The dotted purple line denotes the median age assortativity value, and the green line denotes the

median fledgling number assortativity value.

Temporal repeatability of spatial age structure

There was low temporal repeatability in the age-composition of spatial regions at all assessed scales

25–200ha (ICC range: 0.036–0.057; Figure 4). In contrast, as would be expected, there is very high

temporal repeatability in all assessed territory quality measures (Figure 4). Compared to age-

composition, there is higher temporal repeatability in average reproductive measures of

neighbourhoods, with mean clutch size (ICC range: 0.234–0.447) displaying the highest repeatability

(Figure 4; Table S4 for all results).

contributor to spatially variable demography and why there is not greater spatial age structure in this

system.

Age and territory quality

Across all breeding individuals, we show that older birds are more likely to acquire territories at higher

densities that have a greater long-term popularity, while there is a lack of evidence for a relationship

between age and the abundance of oak trees or distance to edge. Considering females and males

separately, we find similar associations between age and territory quality, except that the popularity
index of a nest-box more strongly predicts occupation by adult males than females.

It is interesting that territories at higher average densities are more likely to be occupied by adults than
juveniles considering that this may result in reduced territory size, increased chance of encounter with
conspecifics during chick-rearing behaviour, and has previously been linked to lower reproductive
success [47]. This may relate to evidence that adults have a weaker response to density dependence
in great tit populations, where juvenile cohorts have a greater effect in reducing recruitment and
survival, and more negatively respond to density dependence [7]. Together with our results, this may
reflect improved breeding competence of adults compared to juveniles leading to higher territorial
densities. Specifically, older individuals may have increased foraging efficiency [76], thus negating the
need for large territories (indeed, individuals with smaller foraging ranges can be more efficient with
higher reproductive success [77]), therefore generating higher breeding densities among adults and a
weaker response to density dependence. Additionally, increased familiarity between adult territorial
neighbours may reduce competition associated with foraging [55,66], or even promote joint benefits
such as nest defence [67], thus favouring closer breeding between adults. Recent advances in tracking
technology of passerine foraging behaviour [78] might advance insight into whether competition is
reduced among older familiar neighbours by assessing evidence for increased coordination during
foraging compared to unfamiliar neighbours.

Territories that were more popular in the long-term were more likely to be occupied by adults than
juveniles. The use of frequency of occupancy as a measure of territory quality is well-established in
avian systems [68–72], and may capture information on cues of quality which individuals respond to
in the long-term which cannot be directly measured. It is particularly interesting that the association
between more popular territories and age is weakened when only assessing females, but strengthened
in males. This may be explained by sex differences where male great tits are more important in initial
acquisition and defence of territories [39,48,79,80], which is widespread in other avian [37,81–83] and
taxonomic groups [84–86]. Thus, this could suggest that older males with greater competitive ability
[43] outcompete younger individuals to occupy higher quality sites. This may be driven either by novel
acquisition of high-quality territories, or by persistence and defence of a territory between years
[42,87]. In the latter case, males may occupy a high-quality territory in their first breeding attempt,
which supports successful breeding, and then retain this territory in future breeding. Thus, if survival
is high between years, particularly among individuals in high-quality territories, then covariance
between male age and territory quality may arise as younger first-year breeders are forced into low-
quality territories.

There is little evidence that the abundance of oaks close to the nest predicts age. Oak abundance at

this scale has been linked to great tit breeding success [47,55,61–64], thus increased experience and

competitive ability of adults [43] might lead to the assumption that older individuals will acquire

territories with more oaks. However, although higher oak abundance might provide better quality

territories when considering entire populations, little is known about how this interacts with

phenological synchrony at the individual-level [63] and how the degree of phenological matching

depends on age. Specifically, older great tits typically breed earlier [88,89], possibly due to improved

breeding competence leading to greater synchrony between chick-rearing and peak caterpillar

abundance. Further, the rate at which caterpillars are provisioned to chicks strongly correlates with

the caterpillar biomass available [90]. Thus, if a greater proportion of juveniles mistime their breeding

compared to adults, then they may be unable to monopolise caterpillar resources even if in oak-rich

territories due to temporal mismatch. In addition to greater phenological synchrony, older birds may

improve their foraging ability in a broader sense through greater site familiarity from previous

occupation [27,34,76], utilisation of a wider variety of food resources [40], and improved foraging

coordination with pair-bonded partners [91,92]. In this sense, the overall abundance of oaks may not

be a crucial cue when acquiring territories, rather there is stronger selection on phenological matching

to the local environment [61,63] and general foraging ability. Additionally, there is little evidence that

the edge distance index is associated with age, which has been linked to reduced reproductive output

[30], and thus there is no apparent age-specificity in how edge effects mediate territory acquisition.

The lack of an effect of oak abundance and edge distance on age-specific territory occupation, as well

as the general weak association with other territory attributes, may suggest that territory acquisition in

adults is predominantly driven by retaining previous territories in which they successfully bred,

regardless of quality. This may be driven by a relative lack of available territories compared to overall

population size, thus providing strong selection for the retention of a previously occupied territory

[93,94]. High breeding site fidelity is evident in great tits [87]. In this case, the cue which individuals

respond to when acquiring a territory may be experience of previous breeding success, as opposed

to current environmental quality cues. Consequently, any covariance between older age and high

territory quality might break down through two mechanisms: either previous breeding occurred in a

low-quality territory, but breeding was successful and thus the territory is retained as an adult; or

previous breeding occurred in a high-quality territory, but the attempt was unsuccessful, and thus

individuals move to new poorer quality territories in future years. Breeding success in poor quality

territories can regularly occur, for example if intraspecific competition is low, or when individual or

partner quality is high. In such cases, older individuals might utilise their higher competitive ability to

defend territories in which they have had success regardless of quality [87,95,96]. Conversely,

unsuccessful breeding may happen in areas of high-quality, through stochastic environmental effects

(such as extreme weather or predation), or low individual or partner quality. In these cases, despite
being in high-quality territories, pairs may divorce in an attempt to improve future breeding [97] and
disperse to new territories [87]. Through this process, older individuals will not necessarily occupy the
best quality territories, thus leaving unoccupied territories of higher quality for juveniles to acquire.
409
Given our findings, future research should aim to disentangle the relative effects of territory quality and
previous reproductive success on site fidelity, as well as assessing behavioural differences in the
degree of territory defence in individuals that re-occupy sites of varying quality. Further, examining
age-specific behaviour following divorce and acquisition of a new territory might shed light on how age
relates to territory quality when removing the effect of site fidelity. For instance, there may be variation
in territory quality across divorced individuals that newly settle depending on their own age and that of
their new partner. Finally, more work should assess age-specificity in territory acquisition while
considering non-breeding individuals that fail to acquire a territory. In great tit populations, there are
often more individuals than available territories [39], leaving non-breeders in the population during the
spring. In avian species, non-breeders tend to be younger [82,98] and the ability to defend a territory
against intrusions from non-breeders increases with age [99]. Thus, experiments to identify non-
breeders during territorial activity [100] and accounting for this cohort might reveal further age-
specificity in territory quality.
423

Spatial age structure
Our study reveals more frequent positive annual spatial age structure, but this structure is weak and
often absent, with no clear sex differences. In contrast, there is strong evidence for spatial territory
quality structure, with particularly strong assortment in oak abundance, territory density and edge
distance (the latter of which is spatially necessitated), and moderate assortment in the long-term
popularity index of a nest-box. There is greater evidence for spatial structure in reproductive output
compared to age, with fledgling number displaying the greatest structure.
431

It is interesting that there is not greater spatial age structure not only due to our evidence of covariance
between some attributes of territory quality and age, and spatial structure in such territory quality, but
also due to social mechanisms that might generate spatial age structure. Winter social associations in
great tits are positively age-assorted, albeit fairly weakly [41], and social connections carryover into
breeding as individuals establish territories closer to previous social associates [56]. This is likely
because having familiar neighbours can reduce the need for territory defence, increase cooperation,
and enhance reproductive success in great tits [55,66,67]. Thus, because older individuals are more
likely to be familiar with their neighbours than younger ones [55], this might be expected to generate
spatial age structure through clusters of older individuals breeding in proximity. However, although this

441 mechanism provides direct benefits at the dyadic-level for closer breeding between adults, there is
442 little overall spatial age structure. This may be partly explained by the species' life-history, where there
443 is high annual mortality (~52% [38]). Together with annual influxes of immigrant birds, this means that
444 the breeding population is often composed of a large proportion of individuals that are first-time
445 breeders (46–83% [19]), thus only making it possible for a small proportion of the population to have
446 familiar neighbours. In short, despite benefits for older breeders to retain previous neighbours, high
447 mortality rates lead to a high chance that at least some previous territorial associates will have died,
448 and high competition for breeding sites mean these will be acquired by new and likely juvenile breeders
449 [39,79,80], thus reducing overall spatial age structure. This hypothesis should be tested through
450 comparative analyses assessing whether there is greater spatial age structure in territorial species
451 with lower mortality rates and thus a greater opportunity to retain previous neighbours, particularly
452 considering familiarity between neighbours is adaptive in many taxonomic groups [101–103].

454 Additionally, although there are benefits of retaining familiar neighbours, this is not necessarily
455 underpinned by an active process where older individuals select territories based on closer breeding
456 to previous associates. Instead, it may be underpinned by a passive process where birds which survive
457 between years occupy the same locality [87], leading to the incidental retention of familiar neighbours
458 that also survive and have high site fidelity. Such a passive mechanism is consistent with recent
459 findings that age-assortative mating within pairs of great tits, despite being adaptive [89], is
460 underpinned by passive processes of pair retention and fluctuations in overall population age structure
461 as opposed to active age-related selection [19]. Thus, even if there are direct advantages of retaining
462 familiar neighbours, if it is not underpinned by active processes, then this might weaken its prevalence
463 and therefore overall spatial age structure.

465 Finally, although there is temporal overlap in when territories are used, there is variation between pairs
466 which will affect how neighbours interact depending on their breeding time synchrony, thus affecting
467 spatial age structure at any single point in time. Firstly, there is spatial variation in the timing of food
468 availability at a spatial scale relevant to individual territories, underpinned by variation in individual tree
469 phenology [104], and this local variation predicts breeding timing in individual pairs [63]. Thus, pairs
470 with neighbouring territories may not interact much during chick-rearing behaviour if neighbouring
471 phenology is highly asynchronous (although there is generally spatial structure in breeding phenology
472 across Wytham [64]). Second, there might be great temporal mismatch in occupation of neighbouring
473 territories when considering initial acquisition and defence behaviour, which begins in January or even
474 earlier, months before chick-rearing [80,105,106]. Thus, interactions between individuals and
475 competition associated with territory defence will be variable depending on the occupation status of
476 neighbouring territories at the time of focal territory acquisition. For example, there may be extended

periods of territorial activity pre-breeding where neighbouring pairs interact, but then a new pair may
acquire a territory which bisects such breeding sites, thus breaking down the territorial boundary
between the two original neighbouring pairs. Previous research in this system has assessed age-
specificity in nest-site visitation patterns prior to breeding in terms of the number and spatial clustering
of nest-boxes visited [100]. Thus, future work could incorporate such techniques with advances in the
tracking of movement [78,107] and other defence behaviour [108] to advance our understanding on
the temporal scale at which individuals interact during territorial behaviour before breeding, how this
is affected as territories are gradually occupied prior to the spring, and age-specific variation in this.
485

Temporal repeatability of spatial age structure
We find low repeatability in the age-composition of spatial regions between years compared to
repeatability in spatial territory quality and reproductive output structure. High repeatability in spatial
territory quality structure is expected, as annual variation in average quality measures across spatial
regions is changed only depending on which nest-boxes are occupied (i.e. if the same boxes were
occupied every year, there would be 100% repeatability). However, there is variation in which boxes
are used each year. Thus, in conjunction with our previous finding of high within-year spatial territory
quality structure, it is interesting then that we report a higher repeatability in the average reproductive
output compared to the age-composition of spatial regions. This suggests that territory quality in this
system may be more important in driving spatially-variable demography than local age structure.
Additionally, the finding of low repeatability in spatial age-composition is consistent with our suggestion
that adult occupation is likely largely driven by persistence in previous territories of varying quality as
opposed to active acquisition of high-quality territories (the latter would lead to higher repeatability as
older birds would frequently use higher quality sites independent of year).
500

(5) Conclusions
Our study highlights association between age and aspects of territory quality in a territorial passerine
species, specifically showing that older birds breed at higher densities and in sites which are more
popular in the long-term. However, we show limited evidence of spatial age structure. We suggest that
age-specific occupation of territories is likely driven by persistence in territories across lifespan (largely
irrespective of quality), leading to passive retention of some older neighbours which also survive
between years. We emphasise the importance of exploring the operation of these mechanisms that
might lead to spatial age structure in species with different life-histories. Specifically, it might be
expected that in species with lower mortality rates, greater spatial age structure would be generated
through passive mechanisms of territorial neighbour retention as more individuals survive between
years. The observed patterns of higher spatial territory quality and reproductive output structure
(compared to spatial age structure), in addition to greater temporal repeatability in these, contributes

valuable insights into the drivers of spatially-variable demography. Namely, given these results, we

suggest that spatial clustering of territory quality is likely to be more important in determining

repeatable spatial structure in reproductive output compared to spatial age structure.

Acknowledgements

Thank you to all Wytham fieldworkers who have worked on the tit study for the past 76 years.

Funding

This work was supported by the Edward Grey Institute of Field Ornithology. J.A.F. was supported by

funding from BBSRC (BB/S009752/1) and NERC (NE/S010335/1 and NE/V013483/1).

References

1. Stearns SC. 1992 *The evolution of life histories.* Oxford: Oxford University Press.
2. Gamelon M, Vriend SJG, Engen S, Adriaensen F, Dhondt AA, Evans SR, Matthysen E,
Sheldon BC, Sæther BE. 2019 Accounting for interspecific competition and age structure in
demographic analyses of density dependence improves predictions of fluctuations in
population size. *Ecol Lett* **22**, 797–806. (doi:10.1111/ele.13237)
3. Siracusa ER, Pereira AS, Brask JB, Negron-Del Valle JE, Phillips D, Platt ML, Higham JP,
Snyder-Mackler N, Brent LJN. 2023 Ageing in a collective: The impact of ageing individuals
on social network structure. *Philosophical Transactions of the Royal Society B: Biological
Sciences* **378**. (doi:10.1098/rstb.2022.0061)
4. Coulson T, Gaillard JM, Festa-Bianchet M. 2005 Decomposing the variation in population
growth into contributions from multiple demographic rates. *Journal of Animal Ecology* **74**,
789–801. (doi:10.1111/j.1365-2656.2005.00975.x)
5. Woodman JP, Gokcekus S, Beck KB, Green JP, Nussey DH, Firth JA. 2024 The ecology of
ageing in wild societies: linking age structure and social behaviour in natural populations.
EcoEvoRxiv (doi:https://doi.org/10.32942/X21893)
6. Hoy SR, MacNulty DR, Smith DW, Stahler DR, Lambin X, Peterson RO, Ruprecht JS,
Vucetich JA. 2020 Fluctuations in age structure and their variable influence on population
growth. *Funct Ecol* **34**, 203–216. (doi:10.1111/1365-2435.13431)
7. Gamelon M, Grøtan V, Engen S, Bjørkvoll E, Visser ME, Sæther BE. 2016 Density
dependence in an age-structured population of great tits: Identifying the critical age classes.
Ecology **97**, 2479–2490. (doi:10.1002/ecy.1442)
8. Fowler AJ, McLeay L, Short DA. 2000 Spatial variation in size and age structures and
reproductive characteristics of the King George whiting (Percoidei : Sillaginidae) in South
Australian waters. *Mar Freshw Res* **51**, 11–22.
9. McIlwain JL, Claereboudt MR, Al-Oufi HS, Zaki S, Goddard JS. 2005 Spatial variation in age
and growth of the kingfish (Scomberomorus commerson) in the coastal waters of the
Sultanate of Oman. *Fish Res* **73**, 283–298. (doi:10.1016/j.fishres.2004.10.020)
10. Ferrer M, Bisson I. 2003 Age and territory-quality effects on fecundity in the Spanish Imperial
Eagle (Aquila adalberti). *Auk* **120**, 180–186. (doi:10.2307/4090153)
11. Albery GF, Clutton-Brock TH, Morris A, Morris S, Pemberton JM, Nussey DH, Firth JA. 2022
Ageing red deer alter their spatial behaviour and become less social. *Nat Ecol Evol* **6**, 1231–
1238. (doi:10.1038/s41559-022-01817-9)
12. Jones JA, Harris MR, Siefferman L. 2014 Physical habitat quality and interspecific competition
interact to influence territory settlement and reproductive success in a cavity nesting bird.
Front Ecol Evol **2**, 1–8. (doi:10.3389/fevo.2014.00071)
13. Groenendijk J, Hajek F, Schenck C, Staib E, Johnson PJ, Macdonald DW. 2015 Effects of
territory size on the reproductive success and social system of the giant otter, south-eastern
Peru. *J Zool* **296**, 153–160. (doi:10.1111/jzo.12231)
14. Kerbiriou C, Gourmelon F, Jiguet F, Le Viol I, Bioret F, Julliard R. 2006 Linking territory quality
and reproductive success in the Red-billed Chough Pyrrhocorax pyrrhocorax: implications for
conservation management of an endangered population. *Ibis* **148**, 352–364.
(doi:10.1111/j.1474-919X.2006.00543.x)
15. Reynolds JD. 1996 Animal breeding systems. *Trends Ecol Evol* **11**, 68–72.
(doi:10.1016/0169-5347(96)81045-7)
16. Davies NB, Krebs JR, West SA. 2012 *An introduction to behavioural ecology*. John Wiley &
Sons.
17. Hinde RA. 1956 Biological significance of the territories of birds. *Ibis* **98**. (doi:10.1016/b978-0-
12-571126-5.50008-x)
18. Greenwood PJ. 1980 Mating systems, philopatry and dispersal in birds and mammals. *Anim
Behav* **28**, 1140–1162. (doi:10.1016/S0003-3472(80)80103-5)

19.   Woodman JP, Cole EF, Firth JA, Perrins CM, Sheldon BC. 2022 Disentangling the causes of
575   age-assortative mating in bird populations with contrasting life-history strategies. *Journal of*
576   *Animal Ecology* (doi:10.1111/1365-2656.13851)
577   20.   Piper WH. 1997 Social Dominance in Birds. In *Current Ornithology* (eds V Nolan, ED
578   Ketterson, CF Thompson), pp. 125–187. Boston, MA: Springer US. (doi:10.1007/978-1-4757-
579   9915-6_4)
580   21.   Wilson EO. 2000 *Sociobiology: the new synthesis*. 25th anniv. Cambridge, Mass. ; London:
581   Belknap Press of Harvard University Press.
582   22.   Brewer SK, DiStefano RJ, Rabeni CF. 2009 The influence of age-specific habitat selection by
583   a stream crayfish community (Orconectes spp.) on secondary production. *Hydrobiologia* **619**,
584   1–10. (doi:10.1007/s10750-008-9577-8)
585   23.   Ferrer M, Bisson I. 2003 Age and territory-quality effects on fecundity in the Spanish Imperial
586   Eagle (Aquila adalberti). *Auk* **120**, 180–186. (doi:10.2307/4090153)
587   24.   Garneau DE, Boudreau T, Keech M, Post E. 2008 Habitat use by black bears in relation to
588   conspecifics and competitors. *Mammalian Biology* **73**, 48–57.
589   (doi:10.1016/j.mambio.2006.11.004)
590   25.   Stepanuk J, Heywood E, Lopez J, DiGiovanni Jr R, Thorne L. 2021 Age-specific behavior and
591   habitat use in humpback whales: implications for vessel strike. *Mar Ecol Prog Ser* **663**, 209–
592   222. (doi:10.3354/meps13638)
593   26.   Kokko H, López-Sepulcre A, Morrell LJ. 2006 From hawks and doves to self-consistent
594   games of territorial behavior. *American Naturalist* **167**, 901–912. (doi:10.1086/504604)
595   27.   Piper WH. 2011 Making habitat selection more 'familiar': A review. *Behav Ecol Sociobiol* **65**,
596   1329–1351. (doi:10.1007/s00265-011-1195-1)
597   28.   Saunders DA, Hobbs RJ, Margules CR. 1991 Biological Consequences of Ecosystem
598   Fragmentation: A Review. *Conservation Biology* **5**, 18–32.
599   29.   Krauss J, Steffan-Dewenter I, Tscharntke T. 2003 Local species immigration, extinction, and
600   turnover of butterflies in relation to habitat area and habitat isolation. *Oecologia* **137**, 591–
601   602. (doi:10.1007/s00442-003-1353-x)
602   30.   Wilkin TA, Garant D, Gosler AG, Sheldon BC. 2007 Edge effects in the great tit: Analyses of
603   long-term data with GIS techniques. *Conservation Biology* **21**, 1207–1217.
604   (doi:10.1111/j.1523-1739.2007.00767.x)
605   31.   Greenwood PJ, Harvey PH. 1982 The natal and breeding dispersal of birds. *Annu Rev Ecol*
606   *Syst* **13**, 1–21. (doi:10.1146/annurev.es.13.110182.000245)
607   32.   Johst K, Brandl R. 1999 Natal versus breeding dispersal: Evolution in a model system. *Evol*
608   *Ecol Res* **1**, 911–921.
609   33.   Verhulst S, Perrins CM, Riddington R. 1997 Natal dispersal of Great Tits in a patchy
610   environment. *Ecology* **78**, 864–872. (doi:10.1890/0012-
611   9658(1997)078[0864:NDOGTI]2.0.CO;2)
612   34.   Slagsvold T, Wiebe KL. 2018 Immigrants and locally recruited birds differ in prey delivered to
613   their offspring in blue tits and great tits. *Anim Behav* **139**, 127–135.
614   (doi:10.1016/j.anbehav.2018.01.007)
615   35.   Loveridge AJ, Valeix M, Elliot NB, Macdonald DW. 2017 The landscape of anthropogenic
616   mortality: how African lions respond to spatial variation in risk. *Journal of Applied Ecology* **54**,
617   815–825. (doi:10.1111/1365-2664.12794)
618   36.   Purcell KL, Thompson CM, Drynan DA. 2012 Response of Brown Creepers to elevation and
619   forest edges in the southern Sierra Nevada, California. *Condor* **114**, 185–196.
620   (doi:10.1525/cond.2012.100181)
621   37.   Brown JL. 1964 The Evolution of Diversity in Avian Territorial Systems. *Wilson Bull* **76**, 160–
622   169.
623   38.   Bouwhuis S, Sheldon BC, Verhulst S, Charmantier A. 2009 Great tits growing old: Selective
624   disappearance and the partitioning of senescence to stages within the breeding cycle.
625   *Proceedings of the Royal Society B: Biological Sciences* **276**, 2769–2777.
626   (doi:10.1098/rspb.2009.0457)

39. Perrins CM. 1979 *British tits*. 1st edn. London: Collins.

40. Gosler Andrew. 1993 *The great tit* . London: Hamlyn.

41. Farine DR *et al.* 2015 The role of social and ecological processes in structuring animal
populations: A case study from automated tracking of wild birds. *R Soc Open Sci* **2**.
(doi:10.1098/rsos.150057)

42. Greenwood PJ, Harvey PH, Perrins CM. 1979 The Role of Dispersal in the Great Tit (Parus
major): The Causes , Consequences and Heritability of Natal Dispersal. *Journal of Animal*
Ecology **48**, 123–142.

43. Sandell M, Smith HG. 1991 Dominance, prior occupancy, and winter residency in the great tit
(Parus major). *Behav Ecol Sociobiol* **29**, 147–152. (doi:10.1007/BF00166490)

44. Hinde RA. 1952 The Behaviour of the Great Tit (Parus major) and Some Other Related
Species. *Behaviour* **2**, 1–201. (doi:10.2307/4510389)

45. Savill P, Perrins C, Kirby K, Fisher N. 2010 *Wytham Woods: Oxford's ecological laboratory*.
OUP Oxford.

46. Perrins CM. 1965 Population Fluctuations and Clutch-Size in the Great Tit, Parus major L.
Journal of Animal Ecology **34**, 601–647.

47. Wilkin TA, Garant D, Gosler AG, Sheldon BC. 2006 Density effects on life-history traits in a
wild population of the great tit Parus major: Analyses of long-term data with GIS techniques.
Journal of Animal Ecology **75**, 604–615. (doi:10.1111/j.1365-2656.2006.01078.x)

48. Krebs JR. 1971 Territory and Breeding Density in the Great Tit, Parus Major L. *Ecology* **52**,
2–22.

49. Svensson L. 1992 *Identification guide to European passerines*. 4th, revd. edn. Stockholm :
Thetford: Lars Svensson ; British Trust for Ornithology.

50. Kidd LR, Sheldon BC, Simmonds EG, Cole EF. 2015 Who escapes detection? Quantifying
the causes and consequences of sampling biases in a long-term field study. *Journal of Animal*
Ecology **84**, 1520–1529. (doi:10.1111/1365-2656.12411)

51. Rhynsburger D. 1973 Analytic Delineation of Thiessen Polygons. *Geogr Anal* **5**, 133–144.
(doi:10.1111/j.1538-4632.1973.tb01003.x)

52. Tanemura M, Hasegawa M. 1980 Geometrical models of territory I. Models for synchronous
and asynchronous settlement of territories. *J Theor Biol* **82**, 477–496. (doi:10.1016/0022-
5193(80)90251-9)

53. Schlicht L, Valcu M, Kempenaers B. 2014 Thiessen polygons as a model for animal territory
estimation. *Ibis* **156**, 215–219. (doi:10.1111/ibi.12105)

54. Adams ES. 2001 Approaches to the study of territory size and shape. *Annu Rev Ecol Syst* **32**,
277–303.

55. Gokcekus S, Firth JA, Regan C, Cole EF, Sheldon BC, Albery GF. 2023 Social Familiarity
and Spatially Variable Environments Independently Determine Reproductive Fitness in a Wild
Bird. *American Naturalist* **201**, 813–824. (doi:10.1086/724382)

56. Firth JA, Sheldon BC. 2016 Social carry-over effects underpin trans-seasonally linked
structure in a wild bird population. *Ecol Lett* **19**, 1324–1332. (doi:10.1111/ele.12669)

57. Both C, Visser ME. 2000 Breeding territory size affects fitness: An experimental study on
competition at the individual level. *Journal of Animal Ecology* **69**, 1021–1030.
(doi:10.1046/j.1365-2656.2000.00458.x)

58. Perrins CM. 1991 Tits and their caterpillar food supply. *Ibis* **133**, 49–54. (doi:10.1111/j.1474-
919X.1991.tb07668.x)

59. Keller LF, van Noordwijk AJ. 1994 Effects of local environmental conditions on nestling
growth in the great tit Parus major L. *Ardea* **82**, 349–362.

60. Rytkönen S, Krams I. 2003 Does foraging behaviour explain the poor breeding success of
great tits Parus major in northern Europe? *J Avian Biol* **34**, 288–297. (doi:10.1034/j.1600-
048X.2003.03041.x)

61. Cole EF, Regan CE, Sheldon BC. 2021 Spatial variation in avian phenological response to
climate change linked to tree health. *Nat Clim Chang* **11**, 872–878. (doi:10.1038/s41558-021-
01140-4)

62. Wilkin TA, King LE, Sheldon BC. 2009 Habitat quality, nestling diet, and provisioning
behaviour in great tits Parus major. *J Avian Biol* **40**, 135–145. (doi:10.1111/j.1600-
048X.2009.04362.x)
63. Hinks AE, Cole EF, Daniels KJ, Wilkin TA, Nakagawa S, Sheldon BC. 2015 Scale-dependent
phenological synchrony between songbirds and their caterpillar food source. *American*
Naturalist **186**, 84–97. (doi:10.1086/681572)
64. Wilkin TA, Perrins CM, Sheldon BC. 2007 The use of GIS in estimating spatial variation in
habitat quality: A case study of lay-date in the Great Tit Parus major. *Ibis* **149**, 110–118.
(doi:10.1111/j.1474-919X.2007.00757.x)
65. Ydenberg RC, Krebs JR. 1987 The Tradeoff Between Territorial Defense and Foraging in the
Great Tit (Parus major) 1. *Am Zool* **27**, 337–346.
66. Grabowska-Zhang AM, Wilkin TA, Sheldon BC. 2012 Effects of neighbor familiarity on
reproductive success in the great tit (Parus major). *Behavioral Ecology* **23**, 322–333.
(doi:10.1093/beheco/arr189)
67. Grabowska-Zhang AM, Sheldon BC, Hinde CA. 2012 Long-term familiarity promotes joining in
neighbour nest defence. *Biol Lett* **8**, 544–546. (doi:10.1098/rsbl.2012.0183)
68. Sergio F, Newton I. 2003 Occupancy as a Measure of Territory Quality. *Journal of Animal*
Ecology **72**, 857–865.
69. Johnson MD. 2007 Measuring habitat quality: A review. *Condor* **109**, 489–504.
(doi:10.1650/8347.1)
70. Potti J, Camacho C, Canal D, Martinez-Padilla J. 2018 Long-term occupancy of nest boxes as
a measure of territory quality for Pied Flycatchers. *J Field Ornithol* **89**, 337–347.
(doi:10.1111/jofo.12266)
71. Smith RJ, Moore FR. 2005 Arrival timing and seasonal reproductive performance in a long-
distance migratory landbird. *Behav Ecol Sociobiol* **57**, 231–239. (doi:10.1007/s00265-004-
0855-9)
72. Steenhof K, Newton I. 2007 Assessing nesting success and productivity. *Raptor research and*
management techniques
73. R Core Team. 2021 R: a language and environment for statistical computing. Vienna: R
Foundation for Statistical Computing.
74. Bates D, Mächler M, Bolker BM, Walker SC. 2015 Fitting linear mixed-effects models using
lme4. *J Stat Softw* **67**. (doi:10.18637/jss.v067.i01)
75. Farine DR. 2014 Measuring phenotypic assortment in animal social networks: Weighted
associations are more robust than binary edges. *Anim Behav* **89**, 141–153.
(doi:10.1016/j.anbehav.2014.01.001)
76. Perrins CM, McCleery RH. 1985 The effect of age and pair bond on the breeding success of
Great Tits Parus major. *Ibis* **127**, 306–315. (doi:10.1111/j.1474-919X.1985.tb05072.x)
77. Cole EF, Morand-Ferron J, Hinks AE, Quinn JL. 2012 Cognitive ability influences reproductive
life history variation in the wild. *Current Biology* **22**, 1808–1812.
(doi:10.1016/j.cub.2012.07.051)
78. Baldan D, van Loon EE. 2022 Songbird parents coordinate offspring provisioning at fine
spatio-temporal scales. *Journal of Animal Ecology* , 1316–1326. (doi:10.1111/1365-
2656.13702)
79. Björklund M, Westman B, Allander K. 1989 Song in Swedish Great Tits: Intra-or Intersexual
Communication? *Behaviour* **111**, 257–269.
80. Krebs JR. 1977 Song and territory in the great tit Parus major. In *Evolutionary Ecology* (eds B
Stonehouse, CM Perrins), London: MacMillan.
81. Eikenaar C, Richardson DS, Brouwer L, Bristol R, Komdeur J. 2009 Experimental evaluation
of sex differences in territory acquisition in a cooperatively breeding bird. *Behavioral Ecology*
20, 207–214. (doi:10.1093/beheco/arn136)
82. Stutchbury BJ. 1991 Floater behaviour and territory acquisition in male purple martins. *Anim*
Behav **42**, 435–443.
83. Beletsky LD. 1992 Social Stability and Territory Acquisition in Birds. *Behaviour* **123**, 290–313.

84. Milner RNC, Booksmythe I, Jennions MD, Backwell PRY. 2010 The battle of the sexes? Territory acquisition and defence in male and female fiddler crabs. *Anim Behav* **79**, 735–738. (doi:10.1016/j.anbehav.2009.12.030)

85. López P, Martín J. 2002 Effects of Female Presence on Intrasexual Aggression in Male Lizards, Podarcis hispanicus. *Aggressive Behavior: Official Journal of the International Society for Research on Aggression* **28**, 491–498. (doi:10.1002/ab10026)

86. Clutton-Brock TH. 1989 Mammalian Mating Systems. *Proceedings of the Royal Society B: Biological Sciences* **236**, 339–372.

87. Harvey PH, Greenwood PJ, Perrins CM. 1979 Breeding Area Fidelity of Great Tits (Parus major). *Journal of Animal Ecology* **48**, 305–313.

88. McCleery RH, Perrins CM. 1998 ...temperature and egg-laying trends. *Nature* **391**, 30–31. (doi:10.1111/j.1748-7692.1991.tb00109.x)

89. Harvey PH, Greenwood PJ, Perrins CM, Martin AR. 1979 Breeding success of great tits Parus major in relation to age of male and female parent. *Ibis* **121**, 216–219.

90. Naef-Daenzer L, Naef-Daenzer B, Nager RG. 2000 Prey selection and foraging performance of breeding Great Tits Parus major in relation to food availability. *J Avian Biol* **31**, 206–214. (doi:10.1034/j.1600-048X.2000.310212.x)

91. Mariette MM, Griffith SC. 2015 The adaptive significance of provisioning and foraging coordination between breeding partners. *Am Nat* **185**. (doi:10.5061/dryad.vs7v6)

92. Griffith SC. 2019 Cooperation and Coordination in Socially Monogamous Birds: Moving Away From a Focus on Sexual Conflict. *Front Ecol Evol* **7**. (doi:10.3389/fevo.2019.00455)

93. Harts AMF, Jaatinen K, Kokko H. 2016 Evolution of natal and breeding dispersal: When is a territory an asset worth protecting? *Behavioral Ecology* **27**, 287–294. (doi:10.1093/beheco/arv148)

94. Clark CW. 1994 Antipredator behavior and the asset-protection principle. *Behavioral Ecology* **5**, 159–170.

95. de Vos GJ. 1983 Social behaviour of black grouse. *Ardea* **71**, 1–103.

96. Yasukawa K. 1979 Territory Establishment in Red-Winged Blackbirds: Importance of Aggressive Behavior and Experience. *Condor* **81**, 258–264.

97. Culina A, Radersma R, Sheldon BC. 2015 Trading up: The fitness consequences of divorce in monogamous birds. *Biological Reviews* **90**, 1015–1034. (doi:10.1111/brv.12143)

98. Sergio F, Blas J, Hiraldo F. 2009 Predictors of floater status in a long-lived bird: A cross-sectional and longitudinal test of hypotheses. *Journal of Animal Ecology* **78**, 109–118. (doi:10.1111/j.1365-2656.2008.01484.x)

99. Arcese P. 1987 Age, intrusion pressure and defence against floaters by territorial male song sparrows. *Anim Behav* **35**, 773–784.

100. Firth JA, Verhelst BL, Crates RA, Garroway CJ, Sheldon BC. 2018 Spatial, temporal and individual-based differences in nest-site visits and subsequent reproductive success in wild great tits. *J Avian Biol* **49**, 1–11. (doi:10.1111/jav.01740)

101. Frostman P, Sherman PT. 2004 Behavioral response to familiar and unfamiliar neighbors in a territorial cichlid, Neolamprologus pulcher. *Ichthyol Res* **51**, 283–285. (doi:10.1007/s10228-004-0223-9)

102. Siracusa ER, Boutin S, Dantzer B, Lane JE, Coltman DW, McAdam AG. 2021 Familiar Neighbors, but Not Relatives, Enhance Fitness in a Territorial Mammal. *Current Biology* **31**, 438-445.e3. (doi:10.1016/j.cub.2020.10.072)

103. Temeles EJ. 1994 The role of neighbours in territorial systems: when are they 'dear enemies'? *Anim Behav* **47**, 339–350.

104. Cole EF, Sheldon BC. 2017 The shifting phenological landscape: Within- and between-species variation in leaf emergence in a mixed-deciduous woodland. *Ecol Evol* **7**, 1135–1147. (doi:10.1002/ece3.2718)

105. Krebs JR. 1978 Song repertoires and territory defence in the great tit. *Nature* **271**, 539–542.

106. Krebs JR. 1982 Territorial Defence in the Great Tit (Parus major): Do Residents Always Win? *Behavioral Ecology and Sociobiology* **11**, 185–194.

786   107.   Levin II, Zonana DM, Burt JM, Safran RJ. 2015 Performance of encounternet tags: Field tests
787   of miniaturized proximity loggers for use on small birds. *PLoS One* **10**.
788   (doi:10.1371/journal.pone.0137242)
789   108.   Merino Recalde N. 2023 pykanto: A python library to accelerate research on wild bird song.
790   *Methods Ecol Evol* **14**, 1994–2002. (doi:10.1111/2041-210X.14155)
791