Temporal and spatial variability in availability bias has consequences for marine bird abundance estimates during the non-breeding season

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Abstract

1. To effectively navigate and monitor how marine ecosystems are being reshaped by anthropogenic pressures, we require an understanding of species abundances and distributions. Due to their social, economic, and ecological value, predatory species are often at the forefront of survey efforts. However, survey data are only valuable if they can reliably be converted into estimates of underlying distributions.
2. We consider at-sea surveys of marine predators that often inform ecological impact assessments of offshore windfarms. These surveys are subject to 'availability bias' whereby individuals which are submerged below the surface are consequently 'unavailable' for detection. Although existing correction factors are commonly used in these surveys, they are currently based on limited data that may not be species-, time-, or area-specific. Here, we use time-depth-recorder data to investigate variation in marine bird availability bias.
3. We found that the proportion of diving marine birds submerged below the sea surface during daylight hours, and therefore unavailable to be counted during surveys, varied by species, month, and area, and was different to comparable values previously used to correct for the availability bias of four focal species wintering around northwest Europe: Atlantic puffin, common guillemot, razorbill, and red-throated diver. We therefore present new availability bias correction factors that are species- and month-specific to the areas the study populations of our four species use during their non-breeding seasons: the North Sea, the north and west coasts of the UK, the Baltic Sea, and Icelandic coastal waters.
4. *Synthesis and applications:* Variation in the proportion of daylight hours that marine birds spent submerged lead to differences in availability bias correction factors, thereby impacting estimations of marine bird abundances. We encourage the use of correction factors that use data from the species, marine area, and month during which surveys are conducted to aid the provision of more accurate abundance estimates. Using more relevant correction factors will result in increasingly accurate abundance and distribution estimates of diving marine birds, with huge relevance for a range of applications including planning for offshore windfarm developments, the designation and monitoring of protected areas, and understanding long-term environmental change.

KEYWORDS: *Alca torda;* Diving; Environmental impact assessment; *Fratercula arctica; Gavia stellata;* Offshore windfarm; Seabirds; *Uria aalge*

Introduction

Anthropogenic pressures are rapidly reshaping terrestrial, aquatic, and marine ecosystems (Delong et al., 2018; Henson et al., 2017). As we navigate the challenges of the Anthropocene, strive to monitor and manage changes to these different ecosystems effectively, and reduce further harm, an understanding of species abundances and distributions, and how these are changing, is essential (Boivin et al., 2016; Halpern et al., 2015). Examples of this include when researching and documenting the influence of environmental change on species distributions (Jetz et al., 2019), the consideration and monitoring of protected areas (Edgar et al., 2014), and the execution of environmental impact assessments (EIAs) for oil and gas licensing and offshore windfarms (Dierschke et al., 2016). However, if methods used to establish species abundances and distributions are flawed, then this is likely to lead to poor decision making and perhaps even exacerbate anthropogenic environmental problems (Guisan et al., 2013).

When devising management and conservation plans, apex predators tend to be highly regarded due to their social, economic, and ecological value (Sergio et al., 2006). Despite their significance, estimating the abundances and distributions of marine predators is challenging. This is an issue of high importance, since, in an effort to improve energy security and reduce human reliance on fossil fuels, there has been expansion in the development of marine renewable energy technologies that harness natural resources in the habitats occupied by these species (Soares-Ramos et al., 2020). As part of offshore windfarm EIAs, the at-sea abundances and distributions of marine apex predators are typically assessed using either visual observations, from ships or aircrafts, or digital aerial surveys (Buckland et al., 2012; Thaxter & Burton, 2009). Whilst ideally these surveys would capture and record the presence of all individuals, many taxa, including marine reptiles, mammals, and birds, spend extended periods below the water's surface during which they are undetectable by the methods used to survey them. Historically, this 'availability bias' has been corrected for in these taxa by following methods derived to estimate densities of harbour porpoises *Phocoena phocoena* (Barlow et al., 1988), despite the fact that marine birds can engage in additional relevant behaviours, most notably flight. Thus whilst the quantification of at-sea marine bird abundances and distributions is an important issue, it is also imperative that the approaches used to derive correction factors that account for availability bias are appropriate for this taxa (Certain & Bretagnolle, 2008; Winiarski et al., 2014).

Diving marine birds such as Atlantic puffins *Fratercula arctica* (hereafter 'puffins'), common guillemots *Uria aalge* (hereafter 'guillemots'), razorbills *Alca torda*, and red-throated divers *Gavia stellata* are potentially sensitive to industrial activities, including offshore windfarm developments, across their ranges (Dierschke et al., 2016). Accurately surveying the abundances and distributions of these species is therefore key and historical estimates of at-sea marine birds were likely to have been substantial

underestimations (Harris & Wanless, 2011). Over the last decade, offshore windfarm EIAs have attempted to account for availability bias, by considering the time that individuals from particular locations at specific times of year spend submerged, as a proportion of total time spent on water. At present, correction factors are applied only to birds visible on the water's surface during surveys, with birds in flight being added later (e.g., Harker et al., 2022). Often only a single correction factor is used for each species, this factor frequently having been measured during the breeding season and then applied over the whole annual cycle. For example, when accounting for availability bias within monthly at-sea digital aerial surveys of marine birds in waters off south-east Scotland across an annual cycle, Harker et al. (2022) followed other similar assessments in using values for puffins derived from those breeding at Petit Manan Island, Maine, USA (Spencer, 2012), for guillemots and razorbills from those breeding at the Isle of May, Scotland (Thaxter et al., 2010), and for red-throated divers, they used a value from a different species, great northern divers *Gavia immer,* wintering off the coast of Rhode Island, USA (Winiarski et al., 2014). However, the ever increasing number of biologging studies from across ocean basins (e.g. Lescure et al., 2023), means that there is no longer a need to rely on these commonly used factors from specific areas. Indeed, continued advances in biologging provide opportunities to quantify the time that different species of diving marine birds spend above or below the surface of the sea at different times of year (Duckworth et al., 2021), enabling the derivation of correction factors for availability bias throughout the annual cycle. Together, this increased understanding now enables us to fine tune availability estimates for variation in behaviour in both space and time.

Marine bird diving behaviour varies spatially, temporally, and interspecifically, being influenced by foraging strategies, energy requirements, life cycle stages, and environmental conditions (Phillips et al., 2017). Puffins, guillemots, razorbills, and red-throated divers breed once annually during the late spring – summer at sites around the coastlines of the Atlantic and Pacific oceans, flying and diving intensively during this period as they forage both to feed themselves and their chicks (Dunn et al., 2019; Eriksson et al., 1990). These species then spend the rest of the annual cycle (the non-breeding season) largely at sea, flying less and often experiencing harsher weather conditions and heightened energetic costs, with consequences for their diving behaviour (Duckworth et al., 2022; Dunn et al., 2020). Furthermore, whilst these marine birds are highly adapted to foraging on benthic and pelagic prey (Kleinschmidt et al., 2019; Linnebjerg et al., 2013), they differ in their ability to forage at night, with the diving behaviour of puffins and red-throated divers, in particular, largely being limited to daylight hours (Duckworth et al., 2020; Shoji et al., 2015). Throughout the autumn and winter, as daylight hours become increasingly constrained, varying latitudinally and therefore between non-breeding areas, the proportion of daylight hours that these birds spend submerged may increase.

Variation in behaviour caused by these various drivers will therefore result in differing magnitudes of availability bias in estimates of marine bird abundances, meaning that estimates calculated based on data from certain locations, during the breeding season are undesirable, being associated with increased uncertainty.

Here, we use published biologging data on the dive behaviour of puffins, guillemots, razorbills, and red-throated divers wintering across northwest Europe to quantify broad scale temporal and spatial variation in the proportion of time that they spend below the water's surface. From these data, we calculate new species-, month- and area- specific correction factors for puffins and razorbills in the North Sea, guillemots in the North Sea or off the west coast of the UK, and red-throated divers in the North Sea and Baltic Sea, within eastern Icelandic coastal waters, and around the coast of northern Scotland in the months following their breeding seasons. With these examples, we show the importance of accounting for temporal and spatial differences in availability bias.

Materials and methods

Biologging deployments

Breeding adult puffins, guillemots, razorbills, and red-throated divers were captured from their breeding sites at sites in Scotland, Finland, and Iceland (Figure 1) during the breeding season (for details, see Table 1). All birds were fitted with time-depth-recorders (G5, CEFAS, Lowestoft UK, 31 × 8 mm, 2.7 g in air) attached to Darvic leg-rings. In all cases the attachment process took <10 min and the weight of the logger was <2% of the total body mass of the bird. Birds were recaptured during subsequent breeding seasons and the devices were removed. For more details on the puffin and razorbill deployments see Dunn et al. (2019), for guillemot deployments see Buckingham et al. (2023), and for red-throated diver deployments see Duckworth et al. (2021). Sampling periods varied (Table 1) due to some TDRs failing during the autumn and winter.

Table 1. Details of the time-depth-recorder loggers deployed on and retrieved from Atlantic puffins, common guillemots, razorbills, and red-throated divers at breeding sites across northwest Europe. For more details on the puffin and razorbill deployments see (Dunn et al., 2019), for guillemot deployments see (Buckingham et al., 2023), and for red-throated diver deployments see (Duckworth et al., 2021).

Species	Winter area	Breeding site	Deployment year	N deployed and retrieved	Sampling regime	Sampling Period
Atlantic puffin	North Sea	Isle of May, Scotland	2008	30, 12	Either 3 s for a 24 h period every 10 d, or 30 s for a 24 h period every day	July – December
Common guillemot	UK West Coast	Colonsay and Treshnish, Scotland	2019	30, 19	Every 12 s for a 24 h period every 5 days	July – March
	North Sea	Whinnyfold and Isle of May, Scotland	2019	31, 28	Every 12 s for a 24 h period every 5 days	July – March
Razorbill	North Sea	Isle of May, Scotland	2008	24, 12	Either 3 s for a 24 h period every 10 d, or 30 s for a 24 h period every day	June – January
Red-throated diver	North Sea and Baltic Sea	Southern Finland	2018	31, 8	6 s resolution over a 24 h period every 5 days	September – December
	Icelandic coastal waters	Northeast Iceland	2018	12, 8	6 s resolution over a 24 h period every 5 days	September – December
	Northern Scotland	Orkney and Shetland, Scotland	2018	31, 7	6 s resolution over a 24 h period every 5 days	September – December

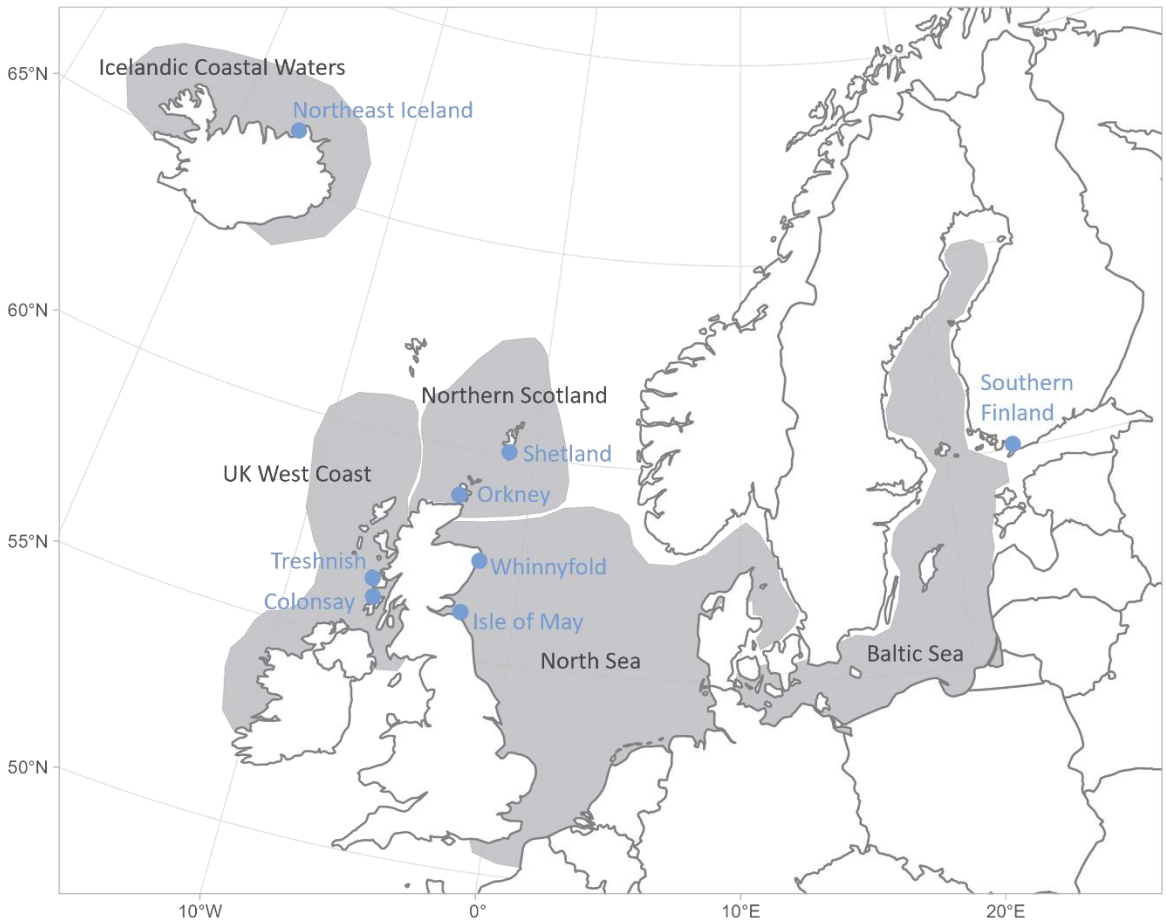

Figure 1. Schematic map illustrating the locations of the breeding sites (shown as blue points and labelled with blue text) where time-depth-recorder loggers were deployed on and retrieved from Atlantic puffins, common guillemots, razorbills, and red-throated divers. Wintering areas are shaded in grey and labelled with black text and are adapted from (Buckingham et al., 2023; Duckworth et al., 2022; Harris et al., 2010).

Data processing

All data processing and analyses were conducted in R version 4.2.2 (R Core Team, 2022). Dive incidences were classified as occasions where time-depth-recorder depth values exceeded 1 m. To categorise dive incidences as occurring during either daylight, twilight, or nighttime, we extracted time- and population-specific area estimates. To do this, we obtained centroids of each population's monthly non-breeding distribution from previous work using light-based geolocator devices (Buckingham et al., 2022, 2023; Duckworth et al., 2022; Harris et al., 2010) (Supporting Information S1). We used the '*oce*' R package to extract sun elevation angles at these areas and categorised each dive incidence as occurring during daylight (sun was >0 degrees above the horizon), twilight (sun was 0–15 degrees below the horizon), or nighttime (sun was >15 degrees below the horizon) accordingly. We used a cut-off of 15 degrees below the horizon to account for uncertainty in our area estimates. We used the same approach to estimate the duration of daylight hours for each day.

We calculated the duration of each incidence of submergence and the time of day at which it took place. We extracted submergence incidences that occurred during daylight hours, summed their durations, and divided them by the duration of daylight to calculate the proportion of daylight hours during which each individual was submerged. We excluded twilight and nighttime hours from our analyses as visual marine bird surveys occur during daylight hours only (NatureScot, 2023b). This metric therefore encompasses all behaviour and should be applied to the total count of birds typically captured during daytime aerial surveys, including both those on the surface and in flight. By doing this, we assumed that the proportions of time that birds spend engaged in their key behaviours (i.e., flying, diving and on water) are equivalent to the proportion of individual birds counted (or not, when they are submerged) engaging in these behaviours during surveys. This assumption underpins similar corrections of survey data which are made using time-budget data (Barlow et al., 1988).

Statistical analyses

We composed a set of generalised linear mixed effects models to quantify the proportion of daylight hours spent submerged by each species based on the explanatory terms available for that species, using the Stan programming language via the '*brms*' R package (Bürkner, 2017). As we were modelling proportional data, we used the Beta distribution that provided flexibility but was limited to between 0 and 1 (Heiss, 2021). For the razorbill and puffin models we included an explanatory term of month only (because data were only available for one area; Tables 1 & 2), whereas for the red-throated diver and guillemot models we include month, area, and their interaction (Table 2). We included month as a categorical representation of time within our models, rather than as a continuous covariate, because marine bird surveys are typically conducted at monthly intervals (NatureScot, 2023a). For all models we included individual bird ID as a random effect.

We adjusted the priors of the coefficients so that they had a normal distribution centred at 0 with a standard deviation of 1 and kept the default brms random effect priors. For the model intercepts, we incorporated prior knowledge in the form of values that are currently used to correct for availability bias in these species (Harker et al., 2022). We therefore logit-transformed the inverse of the proportion of time spent at the surface in puffins (Spencer, 2012), guillemots (Thaxter et al., 2010), razorbills (Thaxter et al., 2010), and red-throated divers (using a value for great northern divers; Winiarski et al., 2014). To eliminate divergent transitions that caused a bias in the posterior draws of the puffin model, we removed the December data (n_{days} = 1) and one outlier value (<4 minutes submerged). Whilst we initially ran all models with 4 MCMC chains and with 2,000 iterations per chain, including warmup, we increased the number of iterations of the puffin model to 3,000. We confirmed model convergence via visual inspection of the chains, posterior predictive check plots, and by

calculating a Gelman-Ruban convergence statistic ($\hat{R}$) of 1 (McElreath, 2020). We used the '*marginaleffects*' R package (Arel-Bundock, 2023) to calculate the marginal effects of each unit change in the explanatory variables. We used these values to make comparisons between our model estimates of the proportions of daylight hours that the species spent submerged, and the proportion of time on water submerged previously used within offshore EIAs (Harker et al., 2022). We note that these values are therefore not identical in their derivation, but that they remain comparable due to the very small proportion of time (typically <2%) that these species spend engaged in flight during the non-breeding season (Duckworth, 2023; Dunn et al., 2020).

Table 2. Details of the generalised linear models run to model the proportion of daylight hours spent submerged by Atlantic puffins, common guillemots, razorbills, and red-throated divers from breeding sites across northwest Europe during the non-breeding season.

Model	Explanatory variables		Categories	Intercept prior
Puffin	Single variable:	Month	Jul, Aug, Sep, Oct, Nov	$N(-1.802, 0.1)$
Guillemot	Two-way interaction:	Month	Jul, Aug, Sep, Oct, Nov, Dec, Jan, Feb, Mar	$N(-1.150, 0.1)$
		Area	UK East Coat, UK West Coast	
Razorbill	Single variable:	Month	Jun, Jul, Aug, Sep, Oct, Nov, Dec, Jan	$N(-1.504, 0.1)$
Red-throated diver	Two-way interaction:	Month	Sep, Oct, Nov, Dec, Jan, Feb, Mar	$N(-0.847, 0.1)$
		Area	Baltic Sea and North Sea, Icelandic coastal waters, Northern Scotland	

We provide values of $Pr(being\ visible)$, the probability of an animal being at the water's surface or in flight and therefore available to be recorded on a survey, for the months of the non-breeding season for puffins in the North Sea, guillemots in the North Sea and the UK west coast, razorbills in the North Sea, and red-throated divers within Icelandic coastal waters, those that move between the Baltic Sea and the North Sea, and those off the north coast of Scotland (Table 3). These values were derived by subtracting the posterior distributions of the models' outputs (i.e., the mean estimate ± lower and upper 95% confidence intervals) from 1. Values of $Pr(being\ visible)$ can be used to correct relative abundance estimates of birds by dividing the density of birds observed sitting on the sea and in flight by $Pr(being\ visible)$ (Barlow et al., 1988). Thus, the values generated by this study can be used in offshore windfarm assessments in a similar way that existing availability bias correction factors are currently used.

Results

There was temporal variation in the proportion of daylight hours that puffins, guillemots, razorbills, and red-throated divers spent submerged over the months that followed the end of the species' breeding seasons. Birds spent an average of 1.5 hours submerged each day, this varying between 1% and 59% of daylight hours spent submerged.

The proportion of daylight hours that puffins in the North Sea spent submerged was lowest in July and August (Fig 2A). In July, the confidence intervals encompassed the proportion of time on water submerged previously used to account for availability bias in this species (Spencer, 2012). Although there were wide confidence intervals surrounding estimates of the proportion of daylight hours that puffins spent submerged, this value increased throughout September, October, and November, again overlapping with the previously used value in October before exceeding it in November (Fig 2A).

The temporal pattern in the proportion of daylight hours submerged was similar in razorbills wintering around the east coast of the UK; they spent little time submerged below the water's surface during the daytime periods that immediately followed the breeding season (June – August), but this increased over the course of the winter, only overlapping with a previously used value in January (Thaxter et al., 2010).

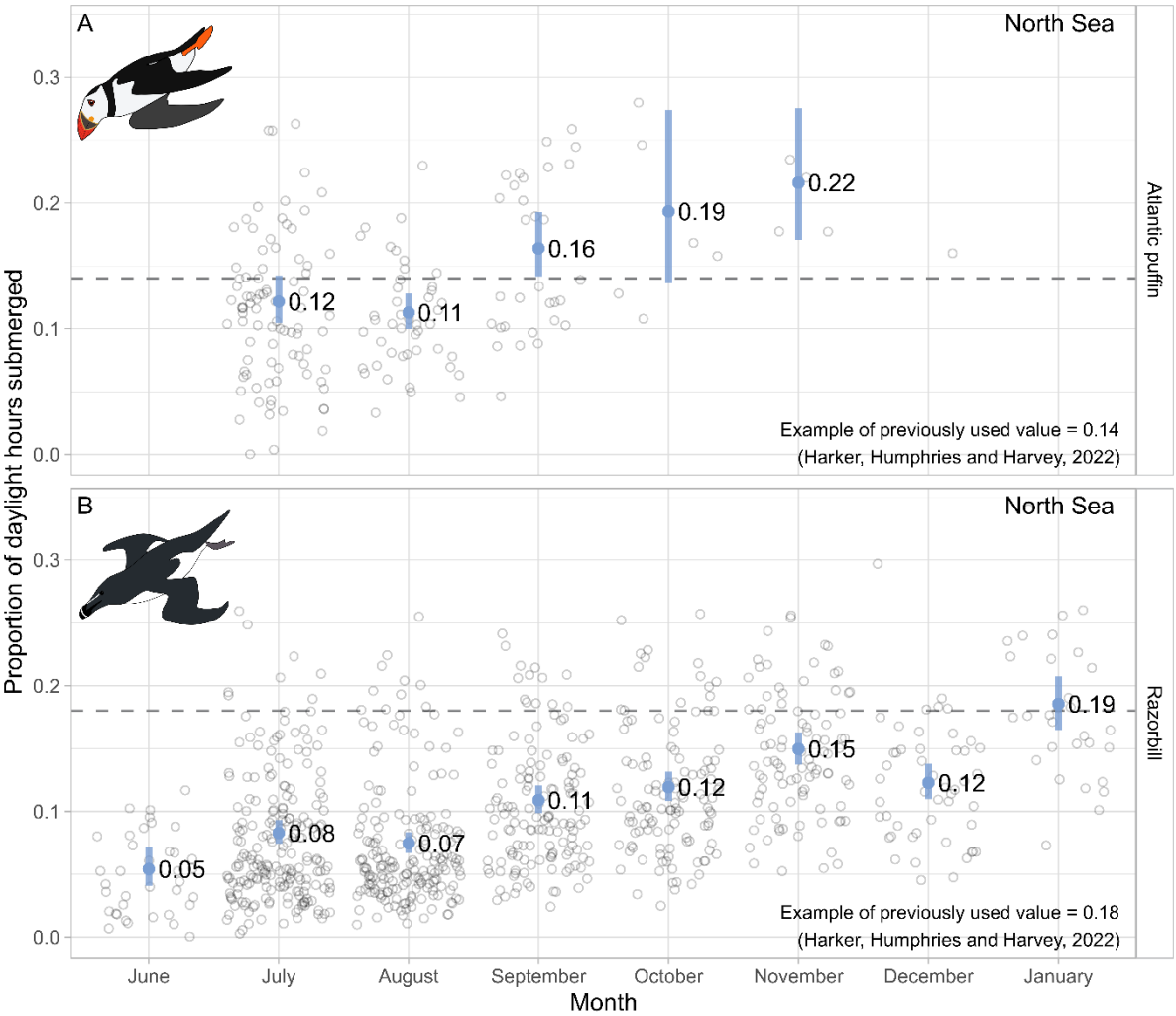

Figure 2. Marginal model predictions of the proportion of daylight hours spent submerged by Atlantic puffins (A) and razorbills (B) wintering within the North Sea, averaged by month, between July and November and June and January for puffins and razorbills, respectively. Blue points illustrate mean marginal predictions and shaded lines illustrate 95% confidence intervals, and the raw data are shown as faint open circles behind. The horizontal dashed lines indicate values of the proportion of time on water submerged previously used during environmental impact assessments for these species (see text for details).

The proportion of daylight hours that guillemots spent submerged during their non-breeding season was similar between birds wintering in both the North Sea and off the UK west coast (Fig 3). Initially, following their breeding season, birds spent a relatively low proportion of daylight hours submerged beneath the water's surface. Indeed, values between July – October were lower than values of the proportion of time on water submerged previously used to correct for availability bias, only exceeding this value in November for birds wintering in the North Sea, and December for birds wintering off the UK west coast (Fig 3). The proportion of daylight hours spent submerged then increased for both groups throughout the sampling period until March (Fig 3).

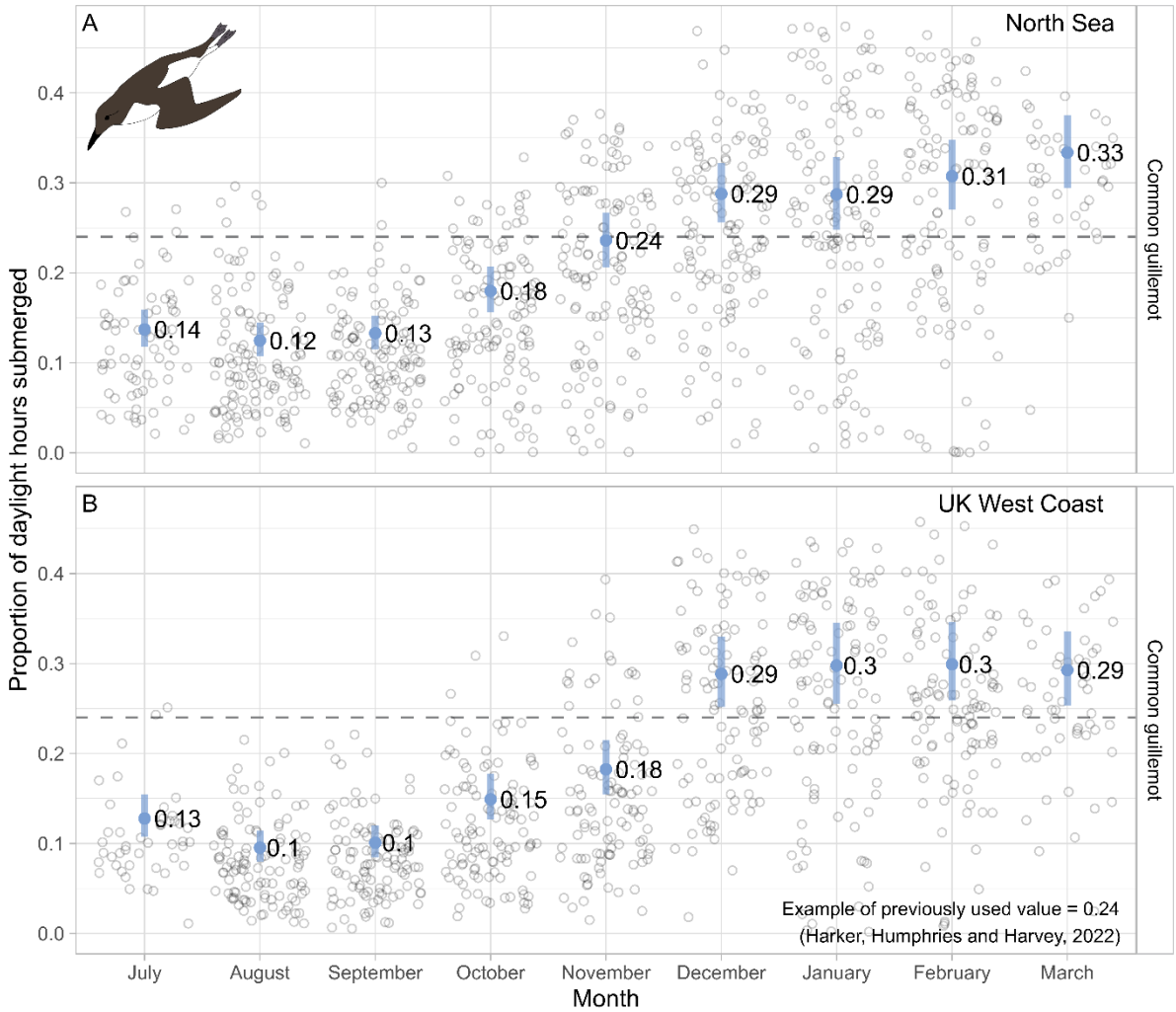

Figure 3. Marginal model predictions of the proportion of daylight hours spent submerged by common guillemots wintering (A) in the North Sea and (B) off the UK west coast, averaged by month, between July and March. Blue points illustrate mean marginal predictions and shaded lines illustrate 95% confidence intervals, and the raw data are shown as faint open circles behind. The horizontal dashed lines indicate a value of the proportion of time on water submerged previously used during environmental impact assessments for this species (see text for details).

Red-throated divers wintering in different areas spent differing proportions of daylight hours submerged. Birds that initially wintered within the Baltic Sea, before moving to the southern North Sea during October, spent an increasing amount of time below the water's surface following their breeding seasons (Fig. 4C). Although a similar pattern was observed in red-throated divers that wintered around north Scotland, with the proportion of daylight hours submerged increasing between September and December, there was a slight decrease in time spent submerged in January (Fig. 4B). A similar dip in the proportion of daylight hours submerged in December was observed in red-throated divers that wintered in Icelandic coastal waters, with high values being observed in January and February, before another decrease in March (Fig. 4A). The majority of our model predictions for the

proportion of daylight hours spent submerged by red-throated divers are lower than the value of time on water submerged previously used to correct for availability bias (Fig 4).

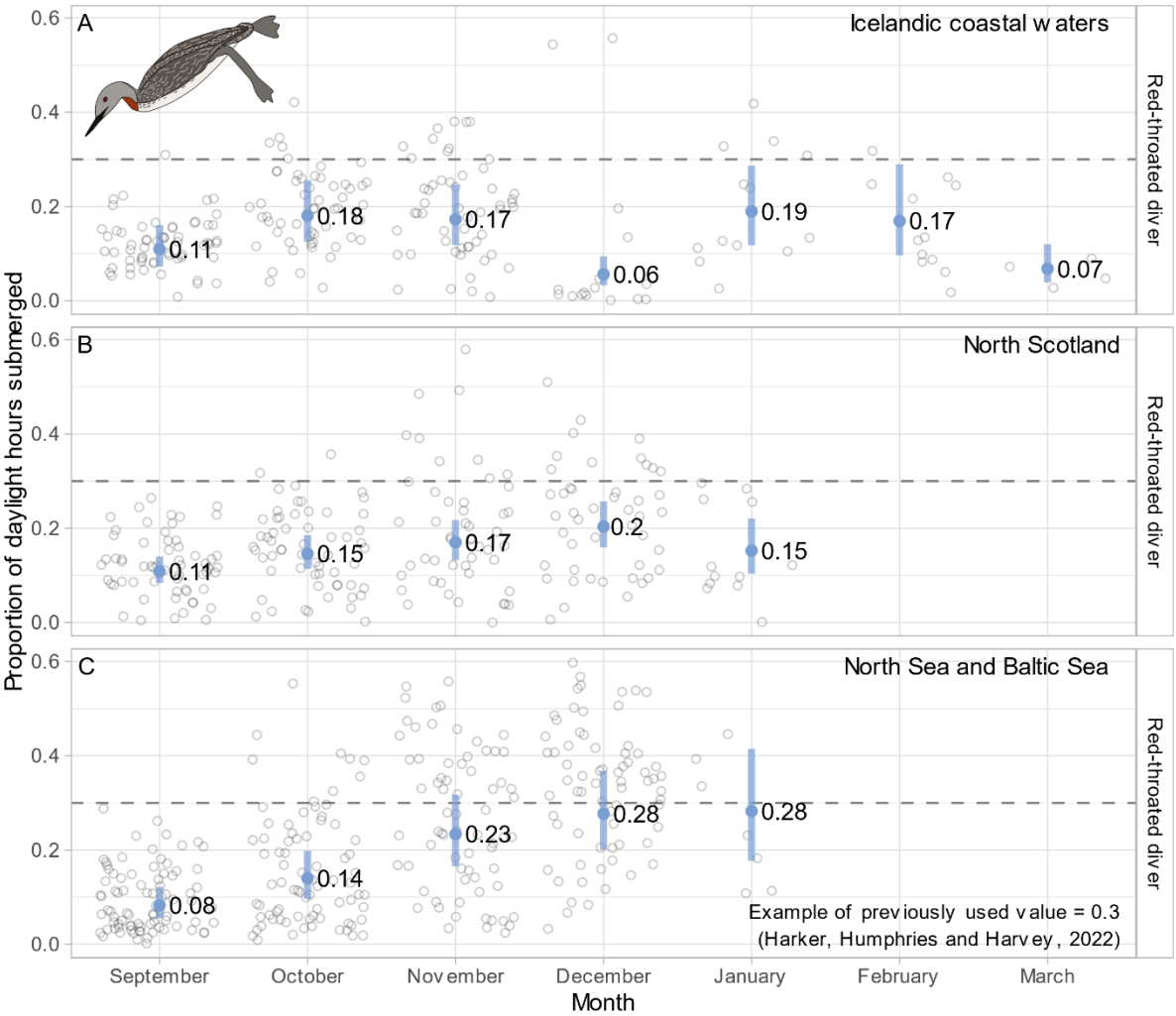

Figure 4. Marginal model predictions of the proportion of daylight hours spent submerged by red-throated divers wintering (A) in Icelandic coastal waters, (B) off north Scotland, averaged by month, between September and March, and (C) in the Baltic Sea during September and October and the North Sea during November – January. Blue points illustrate mean marginal predictions and shaded lines illustrate 95% confidence intervals, and the raw data are shown as faint open circles behind. The horizontal dashed lines indicate a value of the proportion of time on water submerged previously used during environmental impact assessments for this species (see text for details).

Temporal, spatial, and interspecific variation in the proportion of daylight hours spent submerged by the puffins, guillemots, razorbills, and red-throated divers led to variation in values of $Pr(being\ visible)$, the probability of each species being available to be recorded at the water's surface or in flight during surveys (Table 3). These values vary between 0.67 (95% confidence interval = 0.62 – 0.71) for guillemots wintering in the North Sea in March, and 0.95 for razorbills in a similar area in June (95% confidence interval = 0.93 – 0.96).

Table 3. Estimates (and lower and upper 95% confidence intervals in italics) of the probability of being available at the surface or in flight, $Pr(being\ visible)$, for Atlantic puffins, common guillemots, razorbills, and red-throated divers from breeding sites across northwest Europe during the months of the non-breeding season. These correction factors can be applied to survey estimates by dividing estimates of the density or abundance of birds in flight and on water by $Pr(being\ visible)$ to give corrected estimates.

Species	Marine area	$Pr(being\ visible)$									
		Jun	Jul	Aug	Sep	Oct	Nov	Dec	Jan	Feb	Mar
Atlantic puffin	North Sea		0.88 *0.86 – 0.9*	0.89 *0.87 – 0.9*	0.84 *0.81 – 0.86*	0.81 *0.73 – 0.86*	0.78 *0.72 – 0.83*				
Razorbill	North Sea	0.95 *0.93 – 0.96*	0.92 *0.91 – 0.93*	0.93 *0.92 – 0.93*	0.89 *0.88 – 0.9*	0.88 *0.87 – 0.89*	0.85 *0.84 – 0.86*	0.88 *0.86 – 0.89*	0.81 *0.79 – 0.84*		
Common guillemot	North Sea		0.86 *0.84 – 0.88*	0.88 *0.86 – 0.89*	0.87 *0.85 – 0.89*	0.82 *0.79 – 0.84*	0.76 *0.73 – 0.79*	0.71 *0.68 – 0.74*	0.71 *0.67 – 0.75*	0.69 *0.65 – 0.73*	0.67 *0.62 – 0.71*
	UK west coast		0.87 *0.85 – 0.89*	0.9 *0.89 – 0.92*	0.9 *0.88 – 0.92*	0.85 *0.82 – 0.87*	0.82 *0.79 – 0.85*	0.71 *0.67 – 0.75*	0.7 *0.65 – 0.74*	0.7 *0.65 – 0.74*	0.71 *0.66 – 0.75*
Red-throated diver	Icelandic coastal waters				0.89 *0.84 – 0.93*	0.82 *0.74 – 0.87*	0.83 *0.75 – 0.88*	0.94 *0.91 – 0.97*	0.81 *0.71 – 0.88*	0.83 *0.71 – 0.9*	0.93 *0.88 – 0.96*
	Baltic Sea				0.92 *0.88 – 0.94*	0.86 *0.8 – 0.9*					
	North Sea						0.77 *0.68 – 0.83*	0.72 *0.63 – 0.8*	0.72 *0.59 – 0.82*		
	North Scotland				0.89 *0.86 – 0.92*	0.85 *0.82 – 0.89*	0.83 *0.78 – 0.87*	0.8 *0.74 – 0.84*	0.85 *0.78 – 0.9*		

Discussion

There was substantial variation in the proportion of daylight hours spent submerged by puffins, guillemots, razorbills, and red-throated divers in different marine areas throughout their non-breeding seasons. This variation led to differences in the probability of them being available for detection by commonly used survey methods ('availability bias', defined as '$Pr(being\ visible)$'), ultimately influencing estimates of their absolute densities. Our results suggest that previously used correction factors for availability bias (often derived from individuals of different species, or in different areas, or during different seasons) likely provided incorrect estimates of marine bird abundances, the degree of which would have varied by species, time, and area. We therefore encourage the use of correction factors that are estimated from the species, marine area, and month during which the survey is conducted. Our study highlights the importance of using appropriate input data (i.e., data that is as closely related to the surveyed population as possible) when quantifying species abundances and distributions, evidence that is relevant for various applied science studies across ocean basins. We present new values through which to account for variation in availability bias for our four focal species and their corresponding marine areas (Table 3).

Our increased understanding of marine behaviour outside the breeding season is highlighting a mounting number of differences when compared to behaviour during the breeding season (Drummond et al., 2021). In this study, the proportion of daylight hours that all four species spent submerged below the water's surface varied across the non-breeding season, generally demonstrating an increase over this time (Fig 2-4). This temporal variation is likely to be representative of fluctuations in climatic conditions, and the birds' subsequent energy requirements and activity budgets (Buckingham et al., 2023; Clairbaux et al., 2021; Fort et al., 2009). Differences in dive behaviour can also be compounded by variation in daylight availability impacting the ability of these visual foragers to gain the energy needed to meet their energetic demands, therefore driving changes in their time activity budgets (Duckworth et al., 2021; Dunn et al., 2020). Indeed, the time that marine birds spend submerged varies over the course of the diurnal cycle (Jardine, *in press*), and should be considered within future work. Other life history events such as feather moult (Cherel et al., 2016) and migratory travel can also influence marine bird activity budgets, with migratory populations of red-throated divers (Fig 4C) spending a higher proportion of time submerged than more residential populations (Fig 4A-B). This variation in activity is likely due to a requirement to adjust their behaviour as they moved from one wintering area to another in addition to needing to meet the high energetic costs often associated with travel in marine birds (Dunn et al., 2023).

Here we show that accounting for temporal and spatial variation in behaviour is critical to enhancing our understanding of the consequences of anthropogenic change. In particular, seasonal change in

availability bias would strongly affect estimates of seasonal abundance. For example, for common guillemots off the west of Scotland, the correction factor was three times higher in February than in September (Figure 3). It is important that EIA and other assessments do not incorrectly estimate marine bird abundances, something that can occur if inappropriate availability bias values continue to be used. Our results suggest that a hypothetical set of sequential surveys might show a decreasing trend in abundance if a fixed correction factor was used, but in reality, the trend could be stable if the appropriate varying proportion of time submerged (and hence correction factor) was used. This reinforces the recommendation to use correction factors that acknowledge these temporal trends to garner more accurate estimates of the at-sea abundances of these species outside their breeding seasons.

EIAs for offshore renewable projects typically derive corrected estimates of marine bird abundances and distributions using corrections for availability bias, but here we show that current practices may not be dependable. Currently, raw counts of birds on the water's surface are divided by $Pr(being\ visible)$, therefore increasing the estimate above the original count, before adding the number of flying birds. For example, if 3,000 common guillemots were recorded on a digital aerial survey in the North Sea during September, previously used correction factors would have estimated that there were 3,947 birds, with no associated availability bias error. Instead, our species-, time-, and area-specific correction factor (outlined in Table 3) would have produced an estimate of the number of birds in the area being 3,448 individuals (95% confidence interval = 3,370 – 3,529). Furthermore, if in January the survey recorded 3,000 common guillemots, 10% of which were in flight, previously used correction factors would have estimated 3,938 birds, with no associated availability bias error. Contrastingly, our correction factor (Table 3) would instead suggest that the number of birds likely in the area was 4,225 individuals (95% confidence interval = 4,000 – 4,478). These examples demonstrate that previous estimates of the abundances and distributions of these marine birds during winter may well have been incorrect, with implications for assessing the potential consequences of anthropogenic activities such as offshore windfarm developments.

Our study shows how incorporating our increasing understanding of temporal and spatial variation in the behaviour of non-breeding marine birds has consequences for the assessment of anthropogenic activities within the marine environment. The data presented here are most relevant to EIA within the context of offshore wind development but would also have relevance for protected area designation and the investigation of long-term change across marine systems. We encourage those making assessments elsewhere to use and/or collect the most appropriate information to their surveyed areas or studied populations, as was originally conceived in attempts to correct for availability bias (Barlow et al., 1988). In doing so, we recommend the use of larger sample sizes, longer term logger

deployments, and improved non-breeding area estimates to enhance future insights, something that is becoming increasingly tangible with the continued modernisation and miniaturisation of biologging devices. As these developments continue, it will be important to also consider longer-term interannual temporal variation, and the movements of juvenile and immature individuals (Fayet et al., 2024; Merkel & Strøm, 2023). All of this is particularly timely during a period when marine birds are suffering catastrophic declines (Dias et al., 2019) in tandem with our oceans becoming increasingly intensively used for renewable energy developments, oil and gas drilling, transport, and sand extraction (Halpern et al., 2019). Ultimately, when seeking to monitor species abundances and distributions we must harness and appropriately interpret relevant datasets, embracing the nuances of these data and the underlying ecological processes, rather than seeking simplistic solutions that may detriment conservation and management strategies (Chadwick et al., 2023).

Authors' contributions

Sue O'Brien, Jonathan Green, and James Duckworth conceived the ideas; Ruth Dunn, James Duckworth and Lila Buckingham managed the data; Ruth Dunn analysed the data; Ruth Dunn led the data visualisation with marine bird illustrations provided by Lila Buckingham; Ruth Dunn led the writing of the manuscript with support from Jonathan Green. All authors contributed critically to the drafts.

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All applicable institutional and/or national guidelines for the care and use of animals were followed.

Conflict of Interest

The authors declare no conflict of interest.

Data availability statement

No new data were collected for the study.

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Data sources

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Red-throated diver dive data are detailed in Duckworth et al. (2021). Data will be held at the public JNCC repository at https://hub.jncc.gov.uk/ email enquiries can be sent to data@jncc.gov.uk.