ROS-Induced Oxidative Stress Causes Enrichment

and Emergence of Antibacterial Resistance in

Conventional Activated Sludge Process

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ABSTRACT

Conventional activated sludge (CAS) processes represent significant reservoirs of antibacterial

resistance genes (ARGs). However, the mechanisms governing ARG proliferation (enrichment

and emergence) in CAS remain poorly understood. We hypothesize that Reactive Oxygen

Species (ROS), generated by oxygenation in aeration tanks, govern ARG proliferation. Shotgun-

metagenomics assessment of ARG abundance across treatment stages of two plants handling

hospital effluents revealed a substantial reduction in total ARGs in treated effluent. However,

specific ARGs conferring resistance to beta-lactams, tetracyclines, aminoglycosides, rifamycins,

macrolides, and other drug classes showed marked increase in CAS, exhibiting correlations with

ROS stress response genes. Notably, many of these ARGs emerged in the CAS and were absent

in upstream processes. Proof-of-concept experiments demonstrated increased ROS and ARG

levels under high dissolved oxygen (8 mg/L) compared to low oxygen (2 mg/L) conditions. Both

ESKAPE and non-ESKAPE host pathogens in CAS harbored multiple shared ARGs, indicating

horizontal gene transfer facilitated proliferation. ROS-induced stress responses, including

antioxidant defense activation, cell damage, DNA mutation, metabolic adaptation, efflux pump

activation, and enhanced transformation, could contribute to ARG enrichment/emergence. This

study highlights that oxygenation in CAS inadvertently promotes ARG proliferation,

underscoring the need to manage oxygen levels to control ROS-induced ARG proliferation and

environmental dissemination.

KEYWORDS: activated sludge, oxygenation, antibiotic resistance, reactive oxygen species,

horizontal gene transfer, stress response, metagenomics

INTRODUCTION

Antibiotic resistance (AR) poses an escalating threat to global public health, responsible for approximately 1.27 million fatalities worldwide and contributing to almost an estimated 4·95 million (3·62–6·57) deaths associated with bacterial antimicrobial resistance in 2019 [1]. In the United States alone, infections afflict over 2.8 million individuals annually, leading to the deaths of more than 35,000 people [2]. The financial burden of managing infections caused by six prevalent antibiotic-resistant bacteria (ARB)[1] (*Escherichia coli*, *Staphylococcus aureus*, *Klebsiella pneumoniae*, *Streptococcus pneumoniae*, *Acinetobacter baumannii*, and *Pseudomonas aeruginosa*), especially within healthcare settings exceeds $4.6 billion each year [3]. Antibiotic resistance is a critical One Health issue [4], profoundly affecting humans, animals, and the environment, impacting even the most fundamental resource for life: water. The presence of antibiotic-resistant genes (ARGs) and ARBs often can be attributed to effluents from hospitals, farms, and sewage systems, illustrating the interconnectedness of human activities and the broader environmental impact of AR [5–8]. As wastewater treatment plants (WWTPs) operate at the human-environment interface, they play a crucial role in both managing and understanding the global health effects of AR [9–11]. WWTPs receive and process diverse influents, including hospital effluents laden with high concentrations of antibiotics, ARBs, and ARGs [12–14]. This influx creates a potent selection pressure for ARGs and facilitates horizontal gene transfer (HGT) among bacterial communities [14–16]. Notably, hospital wastewaters harbor ARG concentrations several orders of magnitude higher than municipal wastewater [17], highlighting their alarming potential to seed and propagate resistance within WWTPs treating such influents [18–20]. Moreover, recent studies reported that different WWTP treatment processes could exhibit varying impacts on ARG and ARB dynamics [21–23]. Interestingly, the authors identified

significant alteration in the ARGs composition in the sludge compared to the of the influent

resistome and suggested seasonality in antimicrobial use to be key influencing factor behind this

transition [21]. Another study combining multi-omics approach also revealed that both absolute

and relative ARG transcript abundance of the conventional activated sludge (CAS) resistome

were significantly enriched with specific ARGs drug classes (multidrug, beta-lactam,

aminoglycoside, tetracycline, vancomycin etc.) compared to the influent ARG composition,

suggesting bacterial community composition and biomass concentration to be key factors [24].

Many other reports also identified CAS processes as a potential hotspots for ARG enrichment

[23,25–27]. These enriched ARGs in CAS contribute significantly to the effluent resistome which is

disseminated into the receiving environment [21,23,24]. However, the specific mechanisms driving

the enrichment and emergence of ARGs within CAS processes remain poorly understood.

Consequently, understanding the underlying mechanisms behind the emergence of ARGs in CAS

processes has become an urgent need.

While antibiotics are relatively recent human inventions, microbes have developed resistance

mechanisms over thousands of years through their continuous evolutionary struggle for survival

[28]. Recent evidence suggests a unifying mechanism behind the action of many bactericidal

antibiotics, where reactive oxygen species (ROS) generation plays a key role behind lethality [29–

31]. ROS like hydrogen peroxide (H_2O_2), hydroxyl radical ($OH^{\cdot}$), singlet oxygen ($^1O_2$), and

superoxide anion (O_2^-) are inherent byproducts of bacterial aerobic respiration [32,33]. Furthermore,

ROS-mediated killing can trigger a secondary stress response beyond the initial antibiotic-

induced damage, involving factors like activation of response regulators (*e.g.*, *ArcA*), metabolic

reprogramming of the TCA cycle, and disruption of the nucleotide pool with increased ATP

demand [34]. Notably, ROS-mediated killing is also a weapon in the arsenal of various innate host

defense systems [35]. To counter this oxidative stress and survive, bacteria have evolved diverse

antioxidant defenses through metabolic remodeling. These include overproduction of ROS-

scavengers, activation of alternative metabolic pathways like the glyoxylate shunt (GS) and

pentose phosphate pathway (PPP), and enhanced substrate-level phosphorylation [34]. However,

bacterial efforts to develop ROS tolerance often come at a cost. The increased oxidative stress

can lead to direct DNA damage, potentially triggering novel mutations, and ultimately

contributing to the emergence of AR [36–39]. Since ROS are common to both antibiotic action and

bacterial stress responses, resistance to ROS could be a key shared characteristic underlying

ARG development.

The significance of ROS in CAS processes can be attributed to the constant aeration inherent to

these systems, which creates a unique environment characterized by high dissolved oxygen (DO)

levels and, consequently, elevated ROS generation. While existing research hints at the

connection between ROS and ARG emergence in the context of WWTP, it primarily focuses on

lab experiments and specific antibiotics or stressors. One study observed a significant increase in

antibiotic resistance in *Escherichia coli* exposed to aeration tanks compared to primary or final

sedimentation tanks [40]. Another study linked triclosan-induced oxidative stress to a rise in ARGs,

particularly tetracycline and multidrug resistance genes [41]. Similarly, ROS generated by phenolic

compounds have been proven to promote resistance through increased membrane permeability

and HGT [42]. Moreover, antidepressant fluoxetine induced resistance to multiple antibiotics (*e.g.*,

fluoroquinolone, aminoglycoside, β-lactams, tetracycline and chloramphenicol) in *Escherichia*

coli via ROS-mediated mutagenesis [39]. Despite these findings, the critical question of whether

aeration-induced ROS in real-world WWTPs shape the multifaceted resistome in CAS processes,

remains unanswered.

We hypothesize that oxidative stress resulting from high DO levels plays a pivotal role in

fostering the development of antibiotic resistance within the CAS processes. We propose to

assess the hypothesis by investigating how aeration influences ROS generation leading to ARGs

induction and impact on the ARGs within CAS compared to hospital effluents and WWTP

influent. In recent years, metagenomic approaches have revolutionized WWTP surveillance,

enabling comprehensive profiling of resistomes [8,24,43,44]. By analyzing shotgun metagenomic data

from two full-scale WWTPs treating hospital effluents, we assess whether aeration in CAS

processes influences alteration in ARG abundance and, crucially, whether ROS stressors

correlate with the specific enriched and emerged ARGs.

MATERIALS AND METHODS

Sample data collection

In this work, we considered paired-end shotgun metagenomics samples obtained using a HiSeq

2500 instrument from two full-scale WWTPs processing university hospital effluents in

Göttingen (WWTP$_{Gö}$), and Greifswald (WWTP$_{Gw}$), Germany [45]. The samples were collected

from five different stages (abbreviated as S) from both the WWTPs, *i.e.*, S1: Hospital effluent (a

storage tank next to the university hospital), S2: WWTP Influent, S3: Primary Sludge, S4:

conventional activated sludge (CAS), and S5: WWTP effluent as described elsewhere [45,46]. The

details of the sample names, and NCBI SRA accession used in this study, can be found in the

Table S1 and Table S2. The raw metagenomes can be accessed in the NCBI Sequence Read

Archive under the BioProject accession number PRJNA524094 [46].

Metagenomics Analysis

We performed preprocessing steps to eliminate low-quality sequences using Trimmomatic [47]

with specific quality thresholds: LEADING:3, TRAILING:3, SLIDINGWINDOW:10:15, and

MINLEN:50 with the following settings: TruSeq3-PE-2.fa:2:30:10:2:keepBothReads [48]. For

metagenomic taxonomic and functional profiling, we employed the SqueezeMeta v1.5.2 pipeline

[49]. The read statistics per sample is provided in Table S1 and Table S2. Co-assembly was

performed using Megahit [50], and short contigs (<200 bps) were filtered out using prinseq [51].

Within the SqueezeMeta pipeline, Barrnap [52] tool was used for predicting RNAs, while Prodigal

[53] was utilized for predicting ORFs. Similarity searches against NCBI GenBank nr database [54]

and Kyoto Encyclopedia of Genes and Genomes [55], were performed for taxonomic and

functional assignments, respectively, using Diamond [56] with a maximum e-value threshold of 1e-

03. Moreover, the Comprehensive Antibiotic Resistance Database [57] release 3.2.4 was used for

annotating the ARGs and identifying the ARBs in metagenomics data. Bowtie2 [58] was used for

read mapping against the contigs. The metagenomics data was analyzed using the SQMtools R

package [59].

Bioinformatics Analysis and Visualization

We used Cumulative Sum Scaling [60] (CSS) to normalize both taxa and gene raw read sequence

counts. To assess the diversity of ARG composition across the different treatment stages, we

calculated a Bray-Curtis dissimilarity matrix with PERMANOVA to determine the significance

of clustering, using MicrobiomeAnalyst [61]. A heatmap was plotted to visualize changes in the

abundance of each ARG, utilizing a z-score with TBtools [62]. Furthermore, differentially

enriched/emerged ARGs in CAS (S4) were identified using EdgeR [63] with an adjusted p-value

cutoff of 0.05. The correlation between ROS stressor genes and differentially abundant ARGs

was performed using a Mantel test (p-value < 0.05), while the co-association network between

the ARGs and ARB pathogens was constructed using the Sparse Correlations for Compositional

data (SparCC) [64] correlation (r > 0.6 and p-value < 0.05) method. Both correlation analyses were

conducted using OmicStudio tools (https://www.omicstudio.cn/). The co-association network

was visualized with Gephi [65] version 1.10.1. Graphics were generated using the ggplot2 [66]

package in R [67] version 4.2.1.

Batch reactor experiments with different DO levels

In order to validate our hypothesis of higher ROS generation under more oxidative stress leading

to increased ARGs, we performed batch reactor experiments with CAS samples under low (2

mg/L) and high (8 mg/L) DO levels. The CAS samples were obtained from the Māngere WWTP

in Auckland, New Zealand. The experiments were conducted in reactors with a capacity of 1 L,

at a temperature of 20 °C. Each liter of the microbial reaction solution contained 2.75 g/L of

mixed liquor suspended solids (MLSS) activated sludge, 3.84 g/L of sodium bicarbonate as the

inorganic carbon source, and 1 mL of trace element solution (Table S3). Over the course of the

48-hour experiment, a total of 50 mL of concentrated artificial wastewater (Table S4) was

uniformly and continuously added into the system using a syringe pump and mixed using a

magnetic stirrer. Detailed methodology for ROS measurement and DNA extraction for

metagenomics analysis are provided in the Supporting Information. The raw metagenome

samples from these experiments can be accessed in the European Nucleotide Archive under the

BioProject accession number PRJEB74089 as provided in Table S5.

RESULTS AND DISCUSSION

Figure 1. Alteration of ARG abundance across different WWTP stages. (a) Schematic representation of the WWTP stages studied, including S1: Hospital effluent, S2: WWTP influent, S3: Primary sludge, S4: Activated sludge, and S5: WWTP effluent. The variation in ROS stressor levels across these stages is depicted below. (b) and (c) show the total relative ARG abundance at each stage for WWTP$_{Gö}$ and WWTP$_{Gw}$, respectively. Heatmaps display the trend of change in each ARG relative abundance across the stages in (d) WWTP$_{Gö}$ and (e) WWTP$_{Gw}$ with a prominent set of enriched (or emerged) ARGs (red) in both S4 and S5. RAS: Recycled activated sludge.

Comparison of ARG abundances across different treatment stages

In this study, we systematically analyzed and compared metagenomic samples from various stages of wastewater treatment processes to assess changes in the resistome profile. These stages included hospital effluents (S1), WWTP influent (S2), primary treatment (S3), CAS (S4), and WWTP effluents (S5) (Figure 1a). Our findings showed a significant decrease in total ARG abundances at different stages of WWTP$_{Gö}$ compared to the hospital effluents: 24.38% in S2, 37.95% in S3, 46.04% in S4, and 63.89% in S5 (Figure 1b). Moreover, the number of ARGs also exhibited reductions: 28.27% in S2, 29.39% in S3, 51.92% in S4, and 63.74% in S5 (Figure S1a). A similar analysis conducted in the WWTP$_{Gw}$ also revealed a reduction in ARG abundances of 34.48% in S2, 35.86% in S3, 24.45% in S4, and 53.50% in S5, compared to S1 (Figure 1b). The numbers of ARGs detected at these stages also decreased by 25.24% in S2, 30.95% in S3, 67.78% in S4, and 45.40% in S5 (Figure S1c). The highest abundance of ARGs was observed in the hospital effluent (S1) for both WWTPs, which is expected due to the prevalent use of antibiotics in hospital settings, leading to a higher presence of ARGs. Across the treatment stages, a consistent trend of ARG reduction was noted, indicating the effectiveness of the WWTP processes in decreasing the overall ARG abundance in the effluent (S5). This pattern of ARG reduction in CAS and effluents are in agreement with previous metagenomics-based studies [21,26,44].

Furthermore, the diversity of resistome profiles across various stages of the WWTP$_{Gö}$ and WWTP$_{Gw}$ was evaluated using principal coordinate analysis (PCoA) with a Bray-Curtis dissimilarity matrix (Figure S1b, d). The PCoA demonstrated clear clustering of resistomes corresponding to the five treatment stages. In WWTP$_{Gö}$, the analysis accounted for 79.33% of the variation (R^2: 0.79329), with significant separation observed (PERMANOVA F-value: 14.392,

p-value: 0.001), indicating distinct ARG profiles at each treatment stage. Similarly, WWTP$_{Gw}$'s

resistome profiles were markedly segregated, explaining a higher variance of 83.56% (R^2:

0.83563) and demonstrating substantial differentiation (PERMANOVA F-value: 19.064, p-

value: 0.001). The PCoA results showed that the resistomes from hospital effluent (S1), WWTP

influent (S2), and primary treatment (S3) stages clustered closely, suggesting similar ARG

compositions, possibly due to the direct contribution of hospital effluent to the influent. In

contrast, samples from the CAS (S4) and treated effluent (S5) stages were distinctly separate

from the preceding stages. This delineation reflects a significant transition in the ARG profile,

attributable to the biological processes and selection pressure within the activated sludge system

and secondary clarifiers. The distinct clusters for the S4 and S5 stages also suggest that the both

the CAS, and effluent harbors unique ARG composition, with potential environmental

implications upon its release. Similar variation in resistome in influent, activated sludge and

effluent have been reported in previous studies [21,23].

In addition to the overall trend analysis, a detailed heatmap of all ARGs based on normalized

abundance z-scores was generated to further elucidate the specific changes in ARG profiles at

each treatment stage, with a focus on the CAS (S4) process (Figure 1d, e). The heatmap

illustrated that while some ARGs were consistently abundant across all stages, others were more

stage specific. This abundance pattern reinforces the notion of stage-specific selective pressures

and microbial community dynamics shaping the ARG profile. Interestingly, despite the general

trend of decreasing total ARG abundance (Figure 1b, c), certain ARGs exhibited notable

persistence and remarkable enrichment/emergence in the CAS stage (S4) in both the WWTPs

(Figure 1d, e). This relative enrichment of ARGs in the CAS was previously reported using

combined metagenomics and metatranscriptomics study [24].

Identifying the enriched and emerged ARGs in CAS process

To enhance our understanding of the enriched or emerged ARGs within the CAS process, we conducted a detailed analysis of ARGs that were differentially higher abundant in CAS than the upstream stages. Specifically, in the WWTP$_{Gw}$ facility, we observed that 111, 115, and 116 ARGs exhibited significantly (FDR < 0.05) higher abundance in CAS when compared to hospital effluent (S1), influent (S2), and primary treatment stages (S3), respectively, as illustrated in Figure 2a. Similarly, for the WWTP$_{Gö}$ facility, 152, 138, and 114 ARGs were found to be more abundant in CAS relative to the same comparator stages (hospital effluent, influent, and primary treatment) (Figure S2a). Additionally, certain ARGs were identified as being significantly more abundant in S4 compared to all other stages (S1, S2, and S3), with some ARGs being uniquely abundant to specific treatment stages (Figure S3 and Supplemental Data 1). To facilitate downstream analyses, we compiled a list of ARGs that were significantly abundant in the CAS process, selecting those genes that displayed a higher average abundance in CAS compared to all preceding treatment stages. This analysis yielded a total of 139 and 106 highly abundant ARGs for WWTP$_{Gö}$ and WWTP$_{Gw}$, respectively, as detailed in Supplemental Data 2a. Notably, 17.27% (WWTP$_{Gö}$) and 19.81% (WWTP$_{Gw}$) of the ARGs were not detected at any of the upstream stages, suggesting they emerged under the selection pressure of the CAS process (Supplemental Data 2b). These emerged ARGs exhibited resistance to aminoglycoside (*AAC(3)-IIIc*, *AAC(6')-Iih*), beta-lactam (*OXA-34*, *OXA-660*, *OXA-4*, *GOB-3*, *GOB-4*, *IDC-2*, *IMP-1*, *IMP-13*, *IMP-28*), rifamycin (*iri*, *arr-1*), tetracycline (*tap*, *tet(V)*), glycopeptide (*vanD*, *vanI*), and other antibiotics. Additionally, among the enriched ARGs in both WWTPs were those belonging to various beta-lactamases, such as *OXA* (including *OXA-2*, *OXA-45*, *OXA-101*, *OXA-129*, *OXA-198*) beta-lactamases. These ARGs confer resistance to antibiotics like carbapenems,

cephalosporins, and penams, highlighting a significant selection pressure for beta-lactam

resistance. This is especially concerning given the extensive use of beta-lactam antibiotics and

the potential for these resistance mechanisms to spread into the environment. Furthermore, we

identified multiple ARGs conferring resistance to other antibiotic classes, including tetracycline

(*tet(35)*, *tet(49)*, *tet(G)*, *tet(O/32/O)*, *tet(T)* etc.), glycopeptide (*vanC*, *vanF*, *vanS_in_vanC_cl*,

etc.), and aminoglycoside (*AAC(3)-Id*, *AAC(6')-32*, *AAC(6')-Ic*, *AAC(6')-IIa*, *AAC(6')-IIb*,

AAC(6')-IIc, *aadA*, *amrB*, etc.), among others. Resistance to rifamycin (*rpoB2*, *Bado_rpoB_RIF*,

Nfar_rox, *rphA*, *rphB,*), macrolide (*EreA2*, *EreB*, *mphA*), fluoroquinolone (*QepA4*, *QnrS6*,

QnrS8), and sulfonamide (*sul2*, *sul4*) antibiotics were also observed, alongside ARGs resistant to

multiple antibiotics, such as *AxyY*, *efrB*, *MexF*, *mexP*, *mexQ*, *mexY*, *opmE*, *smeE*, and more, as

provided in Supplemental Data 2a. Consistent with our observations, a significant enrichment of

various resistance genes, including those encoding for beta-lactamases, macrolides, and

tetracyclines, was identified in recirculated activated sludge in comparison to primary sludge [26].

This enrichment was suggested due to the selective pressure within the wastewater treatment

facility. Similarly, previous research documented an escalation in the relative abundance of a

wide array of ARGs, spanning multiple drug classes such as beta-lactamases, aminocoumarins,

glycopeptides, rifamycins, and sulfonamides, from influent to effluent in a conventional WWTP

[23]. Furthermore, another study reported an increase in both the transcript and gene abundance of

various ARG classes, including Multidrug, beta-lactamases, aminoglycosides, tetracyclines,

trimethoprim, and vancomycin, in secondary effluent compared to the influent of post-primary

clarifiers [24]. Additionally, it was found that aerobic biomass exhibited a higher overall abundance

of ARGs compared to anaerobic and anoxic biomass, with a notable increase in effluent

abundances of fluoroquinolone and tetracycline ARGs [27]. These observations underscore the

diverse spectrum of resistant ARGs emerging within the CAS process, indicating that while the

CAS process effectively reduces the overall ARG load, it also inadvertently facilitates the

proliferation (enrichment and emergence) of multiple concerning ARGs.

Figure 2. Differentially enriched (or emerged) ARGs and their correlation network with ROS stressors in the CAS
process. a) Volcano plots illustrating the differentially abundant ARGs in the CAS stage compared to S1: Hospital
effluent, S2: WWTP influent, S3: Primary sludge in the WWTP_Gw. The grey and red colors indicate the

271 nonsignificant and significant ARGs, respectively. b) A network analysis depicting the correlations between selected
272 enriched/emerged ARGs and ROS stress response genes in CAS process in WWTP_{Gw}.

Assessing the correlation between the ROS stressors and enriched or emerged ARGs

We hypothesized that ROS generated from aeration in the CAS contributes to ARG enrichment/emergence in the CAS process. To explore this, we examined the relationship between key genes involved in ROS stress response and the enriched/emerged ARGs found in CAS of both the WWTPs. We focused on five crucial ROS-related genes [68]: two transcription factors, *oxyR* and *soxR*, which play key roles in the cellular response to oxidative stress, and three genes encoding enzymes that help detoxify ROS, namely *sodAB* (Superoxide dismutase), *katG* (Catalase-peroxidase), and *ahpCF* (alkyl hydroperoxide reductase). Notably, previous reports revealed *oxyR* [69,70] and *soxRS* [71] contributed to virulence in *E. coli*, *Pseudomonas aeruginosa*, and *Salmonella enterica*. Our analysis revealed a significant correlation between these ROS stress response genes and numerous ARGs in both WWTPs (Figure 2b, Figure S2b, Supplemental Data 3). Specifically, in the WWTP_{Gö}, the number of ARGs positively correlated with each ROS stressor were as follows: *oxyR* (24), *soxR* (42), *sodAB* (65), *katG* (65), and *ahpCF* (35). Meanwhile, in WWTP_{Gw}, the numbers of ARGs linked to the same stressors were *oxyR* (81), *soxR* (71), *sodAB* (63), *katG* (92), and *ahpCF* (25). These findings indicate that the mechanisms bacteria use to cope with oxidative stress, including the activation of specific genes and detoxification enzymes, may play a role in the proliferation of ARGs in CAS environment. In the case of WWTP_{Gö}, a total of 14 ARGs were identified as having correlations with all ROS stressors (Figure S4a). These ARGs encompass a diverse array of resistance mechanisms, including multidrug resistance (*e.g.*, *mexY*), beta-lactamases (*OXA-4*, *PAU-1*, *AQU-3*) which provide resistance to beta-lactam antibiotics, tetracycline resistance (*tet(T)*), macrolide resistance (*mtrA*), and various others (*lmrC*, *rpoB2*, *Bado_rpoB_RIF*, *rphB*, *dfrB6*, *sul4*, *adeF*, *arr-1*,

vanR_in_vanE_cl). Furthermore, a significant cohort of 23 ARGs, including *OXA* variants (*OXA-*

9, *OXA-21*, *OXA-129*, *OXA-37*), metallo-beta-lactamases (*IMP-13*, *IMP-28*), and others (*MYO-1*,

GOB-4, *vanS* in various clusters, *vanD*, *aadA8*, *AxyY*, *efrB*, *Cstr_tetA*, *rsmA*, *AAC(6')-IIa*, *dfrB2*,

AAC(6')-Iih, *PEDO-1*, *BRP(MBL)*, *opcM*), showed correlation with the antioxidant enzymes

sodAB and *katG*, indicating a broad spectrum of antibiotic resistance mechanisms being

influenced by oxidative stress responses. Besides, within the WWTP_Gw, 17 ARGs were

correlated with all ROS stressors, including genes responsible for multidrug resistance (*MuxB*,

MuxC, *AxyY*), beta-lactamase activity (*OXA-4*, *OXA-5*, *OXA-145*, *IMP-5*, *PAU-1*), tetracycline

resistance (*tet(T)*, *tetB(P)*), and rifamycin resistance (*rphB*, *rpoB2*), among others (*rsmA*, *adeF*,

BEL-1, *aadA8*, *novA*) (Figure S4b). Additionally, a noteworthy observation was the correlation

of a considerable number of ARGs with multiple ROS stressors, specifically 27 ARGs

correlating with *soxR*, *oxyR*, *sodAB*, and *katG*, and 20 ARGs correlating with *soxR*, *oxyR*, and

katG. This pattern underscores the intricate relationship between oxidative stress response

mechanisms and the proliferation of antibiotic resistance within these CAS processes. Previous

research reported that *Escherichia coli K12* resistant to multiple antibiotics (chloramphenicol,

amoxicillin and tetracycline) produced significantly higher levels of ROS compared to the wild-

type [39]. Additionally, the resistant strains also exhibited significant overexpression of ROS

scavengers including *sodA* and *ahpF*. A recent study showed that exposure to bactericidal

antibiotics (amoxicillin, enrofloxacin, and kanamycin) increased ROS production in the *oxyR*

knocked out *E. coli* compared to the wild-type [36]. Moreover, heme induced electron transport

chain enhanced the ROS production leading to *de novo* acquired resistance in *Lactobacillus*

lactis [36]. These reports support our findings on the pattern of correlations and underscore the

potential link between ROS stressors and ARG emergence. A notable aspect of our findings is

the identification of common and unique ARGs across different ROS stressors. Certain ARGs

appear frequently across multiple ROS stressor correlations, suggesting that these ARGs might

be more universally responsive to ROS-induced stress. In contrast, unique ARGs correlating with

specific ROS stressors indicate targeted responses to particular oxidative stress conditions. This

analysis hinges on our hypothesis that the ROS generated during the aeration process in CAS are

instrumental in inducing the enrichment of ARGs. Through a meticulous examination of various

ROS stressor genes and their correlations with the differentially enriched ARGs, we have

uncovered patterns that reinforce this hypothesis and provide insights into the mechanisms

driving the ARG enrichment/emergence.

Experimental support for increased ARG abundance and ROS under high DO

To further elucidate the potential influence of ROS on the enrichment/emergence of ARGs

within the CAS process, we conducted a series of batch reactor experiments employing CAS

samples derived from the Māngere WWTP in Auckland, New Zealand. This approach was

tailored to mimic the oxidative conditions prevalent in the CAS process, focusing particularly on

the variation of DO concentrations.

Figure 3. Impact of DO levels on ROS production and ARG abundance. a) Bar graph representing the concentration
of H_2O_2 under DO2 and DO8, indicating a statistically significant increase in ROS at higher DO levels (* $p < 0.05$).
b) Comparative total relative ARG abundance showing a moderate increase at elevated DO levels (DO8) in

comparison to lower DO levels (DO2). c) The bar graph illustrates the average relative abundance of oxidative stress
response genes in DO2 and DO8 conditions, reflecting an upregulation in response to increased DO concentration.

Specifically, our experimental design encompassed two distinct DO concentrations, namely,

DO2 (2 mg/L) and DO8 (8 mg/L), to evaluate the impact of elevated DO levels on ROS

generation. Quantitative analysis of hydrogen peroxide (H_2O_2) concentrations yielded average

concentrations of 32.05 ± 2.08 μM and 38.38 ± 0.86 μM for DO2 and DO8 conditions,

respectively (Figure 3a). These results suggest a slight elevation in ROS levels in response to an

increase in DO concentration. Subsequent analysis focused on the total abundance of ARGs

across varying DO levels further underscored similar trends. Specifically, the relative abundance

of total ARGs exhibited a moderate increment under the DO8 condition (11312.57±1065.84)

when compared to the lower DO2 level (9799.61±1442.85) (Figure 3b). Interestingly, a similar

resistome profile observed in CAS of both WWTP$_{Gö}$ and WWTP$_{Gw}$ was detected in the DO

experiments. Among these, prevalent ARG classes included aminocoumarin antibiotics,

aminoglycosides, beta-lactams (encompassing carbapenems, cephalosporins, penams),

fluoroquinolones, macrolides, tetracyclines, peptide antibiotics, rifamycins, and sulfonamides

(Figure S5a). These ARG resistant drug classes were also observed in both the CAS processes in

the WWTP$_{Gö}$, and WWTP$_{Gw}$ (Figure S5 b, c). Notably, a significant proportion of these ARGs

demonstrated enrichment under the DO8 condition in comparison to DO2, with multiple

instances such as beta-lactamases (*OXA-34*, *OXA-145*, *OXA-198*), aminoglycoside antibiotics

(*aadA3*, *aadA5*, *aadA7*), tetracyclines (*tap*, *tet(36)*, *tet(43)*), sulfonamides (*sul2*, *sul4*),

fluoroquinolones (*adeF*, *emrA*, *smeE*), macrolides (*macB*), rifamycins (*rpoB2*, *rphA*), and

glycopeptides (*vanS*) being particularly pronounced. Additionally, an examination of the relative

abundance of genes associated with ROS stress response (*soxR*, *oxyR*, *sodAB*, *katG*, and *ahpCF*)

across both DO2 and DO8 conditions revealed a consistent upregulation under the higher DO8

condition (Figure 3c). This pattern suggests an intensified oxidative stress response, with the

trend uniformly observed across all analyzed genes, indicating a systemic adaptation to the

elevated ROS levels correlated with higher DO concentrations. These findings helped

particularly in understanding the mechanisms driving ARG dynamics in CAS processes. The

elevated ROS levels under high DO conditions, coupled with the upregulation of oxidative stress

response genes, provide compelling evidence supporting our hypothesis that increased ROS, a

byproduct of intensive aeration in CAS systems, could be a contributing factor to the emergence

and persistence of ARGs.

Identification of host ARBs harboring the enriched/emerged ARGs

In addition to the enriched or emerged ARGs, identifying the host pathogens harboring the

ARGs in the CAS processes is crucial, as these ARBs determine the potential health risks.

Widely adopted correlation-based approaches investigate the relationships between ARGs and

their potential hosts in WWTPs [72], based on the rationale that host ARBs should exhibit similar

abundance trends as the ARGs they carry. Our comprehensive analysis using the SparCC

correlation approach unraveled the intricate ARG-ARB co-associations in the CAS stage of both

WWTP$_{Gö}$, and WWTP$_{Gw}$. The co-association networks comprised 128 nodes (41 ARBs and 87

ARGs) in WWTP$_{Gö}$ and 87 nodes (34 ARBs and 53 ARGs) in WWTP$_{Gw}$ (Figure 4, Figure S6,

Supplemental Data 4). Notable key hub ARBs in WWTP$_{Gö}$ included *Klebsiella quasipneumoniae*

(associated with 22 ARGs), *Pseudomonas aeruginosa* (17 ARGs), *Burkholderia pseudomallei*

(13 ARGs), *Phocaeicola vulgatus* (8 ARGs), *Streptococcus suis* (8 ARGs), and *Enterococcus*

faecalis (7 ARGs). In WWTP$_{Gw}$, the key hub ARBs were *Klebsiella pneumoniae* (13 ARGs),

Achromobacter ruhlandii (9 ARGs), *Delftia acidovorans* (8 ARGs), *Aeromonas salmonicida* (7

ARGs), *Pseudomonas aeruginosa* (7 ARGs), and *Escherichia coli* (6 ARGs). Notably, multiple

host ARBs harbored many of the same ARGs in both CAS processes. For instance, the *rpoB2*

gene (conferring rifamycin resistance) was associated with 9 ARBs. Genes resistant to various

drug classes, including efflux pump resistant genes (*MexD*, *mexY*, *mexQ*, *novA*, *arlS*, *mtrA*,

rsmA, *muxC*), OXA beta-lactamase genes (*OXA-3*, *OXA-45*, *OXA-101*, *OXA-427*), and

tetracycline genes (*tet(60)*, *tet(35)*), were shared among ARBs (Supplemental Data 4).

Figure 4. Network analysis depicting the co-associations between ARGs and ARBs in the CAS process in
WWTP~Gö~. Nodes represent individual ARGs and ARBs, with edges indicating SparCC correlations.

This sharing of ARGs suggests a specific co-selection mechanism, where host ARBs acquire and

maintain ARGs, reflecting their adaptive strategies in the CAS environment. These shared ARGs

indicate possible HGT process in the CAS, significantly contributing to the ARG enrichment.

Transfer of ARGs between bacteria through HGT is a major challenge in combating AR as it

accelerates the rise of resistant species. Alarmingly, several ARBs belonged to the ESKAPE

family of clinically concerning pathogens in both WWTP$_{Gö}$ (*Staphylococcus aureus*,

Pseudomonas aeruginosa, *Enterobacter roggenkampii*, *Enterobacter cloacae*) and WWTP$_{Gw}$

(*Staphylococcus aureus*, *Klebsiella pneumoniae*, *Acinetobacter baumannii*, *Pseudomonas*

aeruginosa, *Enterobacter cloacae*). These ESKAPE pathogens are known to acquire antibiotic

resistance through HGT and have evolved a significant ability to thrive under high ROS stress

conditions [73], potentially contributing to ARG enrichment/emergence in the CAS processes.

These findings have profound public health implications, as CAS processes may inadvertently

promote the proliferation and dissemination of the host ARBs to the secondary clarify and finally

into the environment. While previous reports proposed inactivating ARGs in WWTP disinfection

systems [74,75], controlling the enrichment or emergence of ARBs within the CAS system should

be considered an alternative strategy to limit the enrichment/emergence and dissemination of

ARGs.

Mechanism of ROS-induced ARG enrichment/emergence in CAS

Sublethal concentrations of endogenous ROS can trigger a series of direct or indirect cellular

responses that may contribute to the enrichment/emergence of ARGs in the CAS process. These

responses include the activation of transcriptional regulators, expression of oxidative defense

enzymes, metabolic adaptations, DNA damage, activation of efflux pumps, cell damage, and

increased permeability and transformation potential (Figure 5a). Firstly, enhanced expression of

transcription factors like *oxyR* and *soxRS* facilitate the production of ROS-scavenging enzymes

such as superoxide dismutase, catalase, thioredoxins and others [76]. In both the WWTP$_{Gö}$ and

WWTP$_{Gw}$ CAS processes, we observed a high abundance of these transcription factors and ROS

scavengers (Figure 2b and Figure S2b). Secondly, ROS stress can disrupt the TCA cycle, leading

to hyperactivation of respiration, as seen in certain antibiotic treatments [77]. Previous reports also

suggested the disruption of the nucleotide pool, leading to the increase of ATP demand elevating

TCA cycle activity and cellular respiration under quinolones treatments [78]. This disruption may

trigger metabolic adaptations, such as activating the glyoxylate shunt (GS) [79] to reduce oxidative

burden or redirecting flux to the PPP for increased NADPH production, supporting antioxidant

defense systems [80]. Our comparative analysis revealed an enrichment of genes involved in these

metabolic pathways, ATP and NADPH production, and consumption in the CAS system

compared to other treatment processes (Figure 5b, Figure S7). Moreover, sublethal ROS levels

can cause DNA damage and potential mutations, facilitating the emergence of *de novo* antibiotic

resistance [36]. Notably, DNA repair-related genes were abundant in the CAS processes at both

WWTPs. We also identified several ARGs in the CAS process that were not detected in the

upstream processes of both WWTPs (Supplemental Data 2b). This indicates that ROS-induced

mutations could be responsible for their emergence. As a secondary defense, bacteria activate

multidrug efflux pumps to expel antibiotics from the cell, promoting antibiotic tolerance. The

abundance of multidrug efflux pump genes was high in the CAS processes (Figure 5b, Figure

S7). Additionally, oxidative damage can weaken the bacterial cell envelope, potentially leading

to the release of ARGs into the environment. This process, coupled with enhanced

transformation capabilities under oxidative stress, increases the uptake of foreign DNA,

including ARGs, through HGT. Increased cell permeability can promote the uptake of foreign

ARGs and eDNA through transformation [81]. Previous reports suggested that *Streptococcus* sp.

released eDNA under H_2O_2 stress [82]. A recent report revealed that phenolic compounds can

promote ROS-induced increased cell permeability and HGT [42]. These simultaneous events

occurring in the CAS process under oxidative stress conditions may play a significant role in

enriching ARGs or helping ARG emergence. Overall, the proposed mechanisms by which ROS

can contribute to ARG enrichment/emergence in the CAS process include activation of oxidative

stress responses and metabolic adaptations, DNA damage and *de novo* resistance development,

increased efflux pump activity, and enhanced cell permeability promoting horizontal gene

transfer facilitated by oxidative damage. This multifactorial process underscores the complex

interplay between oxidative stress and antibiotic resistance in the CAS process.

Figure 5. Mechanisms of ROS-induced ARG enrichment/emergence in the CAS process. a) A schematic illustration
of the proposed mechanisms by which ROS may contribute to ARG proliferation in the CAS process. The diagram
depicts the cascade of cellular responses triggered by ROS, including (A) activation of transcription factors that (B)
induced expression of ROS-scavenging enzymes, (C) metabolic adaptations to mitigate ROS damage, (D) DNA
damage leading to mutations, (E) cell envelope damage and ARG release, (F) activation of multidrug resistance
efflux pumps, (G) uptake of free ARGs and eDNA. These responses collectively facilitate an environment

conducive to the proliferation and spread of ARGs in the CAS system. b) Heatmap illustrating the change in gene

abundances associated with the cellular processes across various stages of the WWTP$_{Gö}$.

Implications

Our findings shed light on the role of ROS in promoting ARG enrichment and emergence within

the CAS microbiome. Confirming our hypothesis, the study emphasizes the need for oxygen

control in WWTPs to prevent ROS-induced ARG proliferation. Additionally, we observed

significant differences in the ARG composition between hospital effluents or influents and the

WWTP effluents. Notably, the ARG profiles of WWTP effluents are more influenced by the

CAS resistome than the incoming wastewater composition. The majority of highly abundant

ARGs detected in WWTP effluents originate (enrich or emerge) from the CAS process,

indicating that the treatment processes, particularly within CAS systems, play a crucial role in

determining the resistome of discharged effluents. Thus, there is a critical need to closely

examine the ARG proliferation within CAS systems to reduce the ARG load in effluent

discharges. Also, we suggest that ARB-specific targeted approach should be adopted to control

the abundance of the key host pathogens harboring the ARGs in the CAS processes. Our findings

suggest that aeration strategies are required for lowering ARG proliferation without

compromising carbon and nutrient treatment efficiency.

ASSOCIATED CONTENT

Supporting Information

All the supplementary methods, tables and figures are provided in Supporting Information.pdf.

Additionally, detailed information on differentially abundant ARGs, ARG family and drug

classes, ARG correlation with ROS genes, and ARG-ARB co-association networks are provided

in Supporting Data.xlsx.

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