Non-sulfate-based hydrogen sulfide production in engineered systems

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Conflicts of interest

The authors declare no conflicts of interest.

Abstract

Unwanted microbiological production of hydrogen sulfide (H_2S) is a major challenge in engineered systems, such as sewage treatment plants, landfills, and aquaculture systems. While sulfur-rich amino acids and other substrates for non-sulfate-based H_2S production are commonly present, the ability and potential of different microorganisms to perform sulfate-free H_2S production has not yet been resolved. Here, we reveal the identity, activity, and genomic characteristics of bacteria that degrade cysteine to produce H_2S in anaerobic enrichment bioreactors seeded with material from aquaculture systems. We compared them with canonical sulfate reducing bacteria and found that both sulfur sources led to microbial H_2S production, yet with the process being more rapid with cysteine than with sulfate. 16S rRNA amplicon sequencing and metagenomic analysis showed four bacterial families, *Dethiosulfatibacteraceae*, *Fusobacteriaceae*, *Vibrionaceae*, and *Desulfovibrionaceae* to be centrally involved in non-sulfate H_2S production. Metagenome/metatranscriptome assembled genomes uncovered the main cysteine-degradation pathway mediated by cysteine desulfidase *cyuA* and revealed that some

25   bacteria potentially also use cysteine as a carbon source in sulfate-based H₂S production. Our findings call

26   for a revised view of microbial H₂S production in engineered systems.

27   **Key word:** amino acid fermentation; cysteine desulfidase; cyuA; hydrogen sulfide; yhaM

Introduction

Hydrogen sulfide (H_2S) is a volatile sulfur compound present in natural environments (e.g., hydrothermal

vents, oxygen minimal zones, marine sediments, hot springs, lakes (1–4); but can also build up in

engineered systems (e.g., sewage treatment plants, aquaculture systems, oil and gas processing facilities,

landfills (5–7)). While some of this H_2S is of abiotic origin, microorganisms also play a key role in its

production. In both natural and engineered environments, the main group of microbial H_2S producers is

considered to be sulfate reducing microbes (SRM) that use sulfate (SO_4^{2-}) as a terminal electron acceptor

for the oxidation of organic or inorganic compounds in order to obtain energy (8). However, H_2S can also

be produced through sulfur disproportionation, where other inorganic sulfur compounds (e.g. thiosulfate,

sulfite) are simultaneously oxidized and reduced (9). Microbes can also degrade and desulfurize more

complex sulfur (S)-rich compounds, such as cysteine, to produce H_2S (4,9,10). While microbial sulfate

reduction has been well characterized in many environments, the mechanisms and significance of organic

compound degradation to produce H_2S remain poorly understood.

In microorganisms, cysteine degradation is mainly catalyzed by two different types of enzymes: L-

cysteine desulfhydrases and L-cysteine desulfidases, both of which catalyze the decomposition of

cysteine into pyruvate, ammonium, and hydrogen sulfide. Enzymes enabling cysteine degradation and

H_2S production can mediate this process as either a primary or a secondary activity. For example, of the

four abundant genes coding for cysteine degrading enzymes in the human gut, only *cyuA* (also called

yhaM) is associated with cysteine desulfidase activity, while the three other enzymes (*mgl*, *malY*, and

cysK) have other primary functions, but can still use cysteine as a substrate to produce H_2S as a secondary

activity (11–13). While most of the existing knowledge on the degradation of cysteine and other S-rich

amino acids as an H_2S source comes from medical sciences (14–16), this process has also been described

in nature. In freshwater lake sediments, H_2S can for example be produced through degradation of methionine and cysteine under anaerobic conditions (4) and the abundance of proteolytic bacteria able to break down proteins and degrade the resulting amino acids has been found to correlate strongly with hydrogen sulfide concentration, unlike the abundance of SRM (10). Recently, cysteine degradation was also described in the oxic zone of a freshwater lake, being carried out by members of Gammaproteobacteria and Bacteroidetes (17).

In nature, the production of H_2S is generally in balance with its consumption, as several microbial groups use H_2S it as electron donor. However, in engineered systems, the prevailing conditions and processes might promote the accumulation of H_2S if subsequent consumption processes are inhibited. When accumulating, H_2S rapidly causes serious problems, being flammable, corrosive, and very toxic for eukaryotes even at low micromolar concentrations (18–21) . While the control and prevention of H_2S-producing bacterial activity has been suggested as a strategy to avoid H_2S accumulation (22), the current research in engineered environments is mainly focused on the dissimilatory sulfate reduction pathway, neglecting non-sulfate-based H_2S sources (5,23,24). At the same time, recent studies have found H_2S production to increase with protein content of sludge from a freshwater recirculating aquaculture system (RAS) and from wastewater treatment plant (25,26). High H_2S production rates also coincide with the presence of the bacteria phylum Fusobacteria (27), which includes the cysteine-degrading, H_2S producing genus *Fusobacterium* (15,28,29). These findings suggest the presence of non-sulfate based H_2S production also in engineered systems.

In this study, we examined the importance, the role, and the bacteria responsible for cysteine degradation and H_2S production in samples collected from an engineered environment, where both sulfate and S-rich amino acids are typically present (marine recirculating aquaculture systems). We performed laboratory reactor experiments and measured H_2S concentration produced from cysteine (non-sulfate substrate) and for comparison also from sulfate. Cysteine degradation was further verified by measuring dynamic change in concentrations and coupled stoichiometry of ammonium and volatile fatty acids (VFAs). To shed further light on the organisms involved, we combined 16S rRNA gene amplicon sequencing and in-

depth metagenome and metatranscriptome characterization to identify cysteine-degrading H$_2$S producers and reconstruct their cysteine-degradation pathways.

Materials and methods

Sampling

The starting material for enrichment reactors was collected from a commercial salmon facility in Northern Jutland (Denmark) in May 2022. Three types of starting material were collected; activated sludge that represents more degraded organic matter (Aqua-Sample-1), organic material accumulating in the biofilters (biofilter backwash) that represents fresh organic matter (Aqua-Sample-2), and suspended carriers with biofilm from the biofilters (Aqua-Sample-3). For each starting material, three samples of 2 ml were collected and stored at -20°C for further analysis of initial microbial community composition.

Enrichment of potential H$_2$S-producers

To examine the different routes for H$_2$S production, two different enrichment conditions were used: media with cysteine to enrich cysteine-degrading H$_2$S producers and media with sulfate to enrich traditional *dsrA/B*-carrying H$_2$S producers (Supplementary Table S1). For each enrichment condition and type of starting material, three 2-liter anaerobic batch reactors were filled with 100 ml of starting material (corresponding to total chemical oxygen demand of appr. 0.5 g O$_2$/l) and 900 ml of corresponding media. Reactors were sealed with screw caps fitted with two ports for sampling purposes and were subsequently flushed with N$_2$ for 15 min to create anaerobic condition. The enrichment reactors were operated for 4 weeks at 15°C with continuous magnetic stirring at 150 rpm. Every week, the magnetic stirring was stopped for 15 minutes to let enriched biomass settle, after which 800 ml of media was removed, and the reactor was filled with 800 ml of fresh media. At 1, 5, 7, 10, 14, 21, and 28 days of incubation, 50 ml of biomass sample was collected from stirred reactors for analysis of total dissolved sulfide (TDS) and sulfate (only from sulfate treatment reactors). Samples for microbial community/metagenomic analysis were collected from all reactors on days 0, 14, 21, and 28, and subsequently stored frozen at –20°C. Cysteine reactors were kept running until day 76 for a final sampling occasion.

H$_2$S production rates and pathways of cysteine-enriched microbial communities

To confirm the production of H$_2$S via cysteine degradation, we conducted a series of 48-hour experiments using biomass collected from the cysteine-enriched reactors. For the first two experiments, enriched biomass was harvested from Aqua-sample-1 and 2 cysteine-enriched reactors at days 28, 35, and 76 by gentle centrifugation (4,000 rpm for 15 min at 4°C), resuspended in cysteine media ("cysteine-fed"; n=3) (10% v/v) and transferred into 1-liter glass bottles. Incubation bottles were sealed with screw caps fitted with two ports for sampling purposes and were subsequently flushed with N$_2$ for 15 min to create anaerobic condition. The incubations were maintained for 48 hours at 15°C with stirring at 150 rpm. A biomass sample of 5 ml was collected in the beginning (0 h) and after 24 and 48 hours for analysis of TDS production and after 24 hours for RNA sequencing (only for experiment 1). In experiment 3, cysteine-enriched biomass from Aqua-sample-2 was collected in the same way, but the biomass was resuspended in cysteine ("cysteine-fed"; n=3) or sulfate media ("sulfate-fed"; n=3) and divided into three 250 ml bottles to create a replicated setup. A biomass sample was collected in the beginning (0h) and after 24 and 48 hours for measuring concentrations of TDS, total ammonia nitrogen (TAN), VFAs (formate, acetate, butyrate, propionate, and valerate), and sulfate (only sulfate-fed bottles) and after 48 hours for RNA. In experiment 4, Aqua-sample-2 biomass was used and resuspended to either cysteine media or media with cysteine and sulfate (Supplementary Table S1). TDS and RNA samples were collected for experiment 3.

Chemical analysis

Immediately after each sampling, pH, and temperature were measured from the collected biomass sample using a portable meter (Hach HQ40d, Hach Lange, Germany). For measuring total sulfide concentration, aliquots of 0.1 to 2 ml were fixed with ZnAc (10% v/v) and analyzed as in (27) using a spectrophotometer (DR 3900, Hach Lange). The concentration of H$_2$S was calculated from TDS using temperature, salinity, and pH according to (30). For measuring VFAs, sulfate, and TAN concentrations, biomass samples were first centrifuged at 4,500 rpm for 15 min at 4°C. Supernatants were then retrieved and filtered through 0.2 µm syringe filters (Filtropur S, SARSTEDT, Germany), and kept at +4°C until analysis. Samples for VFA

determination were further fixed by adding 1% v/v of 4 M sulfuric acid (H_2SO_4, Merck Millipore,

Germany). The concentrations of VFAs and sulfate were determined using a 930 Compact IC Flex 1 with

a Metrosep A supp 7 -250/4.0 column coupled with an 887 Professional UV/VIS detector (Metrohm,

Sweden), using 0.1 M H_2SO_4 as suppressor and 3.6 mM Na_2CO_3 as eluent and TAN concentration was

determined by 930 Compact IC Flex system (Metrohm, Switzerland) with 887 Professional UV/VIS

Detector (Metrohm, Switzerland).

DNA and RNA extractions

DNA was extracted using the DNeasy PowerLyzer™PowerSoil DNA Isolation Kit (Qiagen, Germany)

according to the manufacturer's instructions and DNA concentration in extracts was measured with

Qubit™ dsDNA High Sensitivity (HS) assay kits and Qubit 2.0 Fluorometer (Thermo Fisher Scientific,

Waltham Massachusetts, US). RNA was extracted using the RNeasy kit (Qiagen, Germany) according to

the manufacturer's instructions and the concentration was measured with Qubit™ RNA High Sensitivity

(HS) assay kits and Qubit 2.0 Fluorometer.

Microbial community composition in the enrichment reactors

Microbial community composition was studied by 16S rRNA amplicon sequencing of Aqua-Sample-1

and 2 at day 14, 21, and 28 (sulfate-enriched condition) and 14, 21, 28, and 76 (cysteine-enriched

condition) as described in (31). For comparison we also included the starting material used for these

incubations in the analysis. Aqua-Sample-3 was discarded from further DNA/RNA-based analysis due to

difficulties obtaining representative samples for the biofilm biomass. The detailed step by step protocol

for the two-steps PCR procedure can be found in the protocols.io repository

(dx.doi.org/10.17504/protocols.io.468gzhw). Hypervariable regions V3-V4 of the 16S rRNA gene were

amplified with primers 341F (5′-CCTACGGGNGGCWGCAG-3′) and 805F (5′-

GACTACNVGGGTATCTAATCC-3′), resulting in an average amplicons size of 460 bp. The amplicon

libraries were sequenced with Illumina Miseq v3 sequencing chemistry and 2X300 cycles. In total

700,805 reads were obtained from the 18 samples. Raw sequencing reads were processed with QIIME 2

(Release 2022.8/ Version 2022.8.3; (32)). Raw sequences were demultiplexed using the q2-demux plugin

and DADA2 (via q2-dada2) was used for trimming, denoising and dereplicating to produce amplicon

sequence variants (ASV) (33). Taxonomy was assigned to ASVs using a Bayesian Naïve classifier pre-

trained against the Silva database (v138.1; (34)). Files generated in QIIME 2 were transferred to R

(version 4.2.1; (35)) for statistical analysis and visualization using packages "phyloseq" (36) and

"ggplot2" (37) after singletons and unassigned reads were removed.

Whole metagenome sequencing, binning and metatranscriptomic analyses

Shotgun metagenome sequencing was carried out for four samples from day 28 (Aqua-Sample-1 and

Aqua-Sample-2 from cysteine and from sulfate-enriched conditions) and two samples from day 76 (only

cysteine-enriched conditions). Metagenomic shotgun libraries were prepared using the MagicPrep system

(Tecan Genomics) according to the manufacturer's instructions. Short read Illumina NovaSeq 6000

sequencing on the SP flow cell was performed at the SciLifeLab National Genomics Infrastructure in

Uppsala (Sweden). Assembly, functional and taxonomic assignment of genes, and bins were obtained by

using the SqueezeMeta automatic pipeline (version 1.6.0 (38)). In brief, all metagenomes were pooled and

co-assembled with MEGAHIT (39) in order to obtain a single reference assembly. Functional and

taxonomic annotation was obtained using the DIAMOND software (40) with three different databases: i)

GenBank nr for taxonomic assignment, ii) eggNOG v4.5 for COG/NOG annotation (41), and iii) the

KEGG database for KEGG Orthology (KO) annotation (42). A final gene classification was carried out

using the PFAM HMM database and the HMMER3 software (43). Coverage and abundance of genes and

contigs were estimated with Bowtie2 (44), where custom scripts from the SqueezeMeta pipeline

subsequently computed and normalized the results. Binning was performed with CONCOCT (45) and

MetaBAT2 (46), and the resulting MAGs were combined into a single non-redundant set with DAS Tool

(47). Bin refinement was performed outside SqueezeMeta using RefineM (48) to identify and remove

outlier contigs with abnormal tetranucleotide signals, GC patterns and coverage. Contigs were also

classified from their taxonomic assignments against the latest reference database provided by RefineM

(gtdb_r95_protein_db.2020-07-30.faa) and if the affiliation was different from the bin, the contigs were

removed. GTDB-tk v2.3.0, data release R214 (49), was used with default settings to assign taxonomy to

each MAG. CheckM (50) was used to provide information about completeness, contamination, and strain

heterogeneity. MAGs surpassing a 70% completeness threshold with less than 10% contamination were

kept for further analysis. For metatranscriptomic analysis, RNA was extracted from the samples collected

in the H_2S-production experiments 1, 3, and 4. Metatranscriptomic libraries were prepared using the

Illumina Stranded Total RNA library preparation kit with ribosomal depletion using Ribo-Zero Plus

(Illumina Inc.). Short read Illumina NovaSeq 6000 sequencing on an SP flow cell of 2X150 bp was

performed at the SNP/SEQ facility of the National Genomics Infrastructure in Uppsala. Only

metagenomics reads were used for assembly and binning, while both the metagenomic and

metatranscriptomic reads were mapped back to the assembly for abundance estimation. This was achieved

by adding the flags "noassembly" and "nobinning" to the metatranscriptomic libraries in samples file used

as an input for SqueezeMeta. Metagenome and metatranscriptome data were further analyzed and

visualized using R (version 4.2.1; (35)) and package "SQMtools" (51). The number of transcripts per

million (TPM) has been recalculated based on the total number of reads assigned to the BIN rather than

by sample before creating gene expression heatmaps. Differences in gene expression between cysteine

and sulfate condition were analyzed with DESeq2 (52,53).

Results

H_2S production in the enrichment reactors and incubation experiments

During the 4-weeks enrichment period, reactors enriched with cysteine started producing H_2S earlier and

faster as compared to sulfate-enriched reactors: H_2S concentrations of 51 - 82 mg H_2S/l were reached

after 5 days, while sulfate-enriched reactors reached this concentration only after 14 - 28 days (Fig. 1). In

cysteine-enriched Aqua-Sample-2 reactor, H_2S concentration decreased between days 14 and 21, most

likely due to oxygen contamination, which led to the consumption of the accumulated H_2S by sulfide-

oxidizing bacteria. H_2S production had recovered by day 28 (Fig. 1).

In the 48-hour experiments with cysteine-enriched biomass fed with either cysteine or sulfate media, we

observed an increase in H_2S concentrations and production rate of 9.99 ± 7.86 mg H_2S/l/d across Aqua-

Sample-1 and Aqua-Sample-2 material when using cysteine media (Table 1). At the same time, both TAN

and acetate concentrations increased, showing that both compounds were being produced in the cysteine-

fed incubations. Formate concentration decreased fast in both cysteine and sulfate-fed incubations, while

there were no significant changes in the concentrations of other VFAs (Supplementary Table S2). For

sulfate-fed incubations, no production of H_2S, ammonium, or acetate (Table 1), or consumption of sulfate

(data not shown) was observed. The one sample fed with a combination of cysteine and sulfate had a

significant increase in H_2S and consequently a high H_2S production rate of 23.9 mg H_2S/l/d.

Microbial community composition in the enrichment reactors

In the original material collected from recirculating aquaculture system, a large proportion of the

microbial community comprised of families outside the 25 most abundant ones (Fig. 2). Of the 25 most

abundant families, *Flavobacteriaceae*, *Candidatus Campbellbacteria*, *Rhodobacteraceae*, and

Rubritaleaceae were detected in the original material. The microbial communities in the enriched reactors

had little resemblance to the original material: e.g. families of the phylum Campylobacterota

(*Arcobacteraceae*, *Sulfurimonadaceae*, and *Sulfurospirillaceae*) were abundant in both cysteine (relative

abundance between 6% to 62%) and sulfate-enriched reactors (24% to 67%) but were not present in the

original material. In the cysteine-enriched reactors, families *Dethiosulfatibacteraceae*, *Fusobacteriaceae*,

and *Vibrionaceae* increased in relative abundance during the experiment, from 31%-55% on day 28 to 67-

87% on day 76. In the sulfate-enriched reactors, the relative abundance of families affiliated with phylum

Desulfobacterota (*Desulfuromonadaceae*, *Desulfocapsaceae*, *Desulfobacteraceae*, and

Geopsychrobacteraceae) increased during the experiment, from 7% on day 14 to 27.5% on day 28.

Metagenomes from enrichment cultures

In order to identify potential bacterial families involved in the production of hydrogen sulfide, we selected

a large number of genes that could be involved in the production of hydrogen sulfide via cysteine

degradation (*cyuA, mgl, malY, cysK, cysM, sseA, metC, tnaA, dcyD, cbs, cse*), dissimilatory sulfate

reduction (*dsrA,B*), and thiosulfate reduction (*phsA/psrA*) (See Supplementary Table S3 for detailed

information on selection of these genes). We then examined their abundance in the metagenomes (Fig. 3)

and their taxonomic assignment (Fig. 4). We found genes encoding enzymes that may have either cysteine

degradation as secondary activity (*mgl, malY, cysK, metC, tnaA, cbs, cse, sseA*) and genes exclusively

associated with cysteine degradation (c*yuA, dcyD*) to be present under both sulfate-enriched and cysteine-

enriched conditions (Fig. 3). Some of the cysteine-degrading genes (c*ysK, sseA, tnaA, metC*) were found

to be present in equal abundances under both conditions, while the abundances of *cbs* (0.03 to 2.14 cpm)

and *cse* (0.04 to 2.48 cpm) were very low in all the reactors. Genes *cysM* and *dcyD*, as well as the genes

associated with dissimilatory sulfate reduction (*dsrA, B),* were more abundant in the sulfate-enriched than

in cysteine-enriched reactors. Three genes associated with cysteine degradation were more abundant in

cysteine-enriched reactors: *mgl* (sulfate day 28: 11.15 ± 0.18 cpm, cysteine day 28: 22.45 ± 11.65 cpm,

cysteine day 76: 49.59 ± 14.38 cpm), c*yuA* (11.84 ± 2.35 cpm, 37.23 ± 9.14 cpm, 76: 57.61 ± 3.61 cpm)

and *malY* (18.70 ± 4.74, 39.4 ± 0.19, 37.3 ± 1.31 cpm). The genes associated with thiosulfate reduction

(*phsA*/*psrA*) and oxidation of hydrogen sulfide (*sqr*) were present in all reactors without any significant

difference between the treatments.

Further examining the 8 most abundant microbial families carrying the three most abundant H_2S-

producing genes in cysteine-enriched reactors (*cyuA, mgl,* and *malY)* (Fig. 4)*, Fusobacteriaceae* was

found to be the only one associated with all three genes. Bacterial families specifically associated with

cyuA varied according to the enrichment condition. In the sulfate-enriched reactors, the largest share of

sequences for c*yuA* was assigned to the families *Campylobacteraceae* and unclassified Bacteria while the

rest of the sequences were assigned to families with low relative abundance. In the cysteine-enriched

reactors, the largest share of *cyuA* sequences was assigned to the families *Campylobacteraeae*,

Dethiosulfatibacteraceae, *Vibrionaceae*, *Fusobacteriaceae*, and *Desulfovibrionaceae*. The total

abundance of families associated with *malY* was generally higher in cysteine-enriched reactors as

compared with sulfate-enriched incubations. In both enrichment conditions, a large portion of the

sequences could be assigned to families *Thiovulaceae* and *Campylobacteraceae*. In the cysteine-enriched

reactors, gene sequences associated to *Fusobacteriaceae* emerged and alleles associated with families

outside the 10 top ones were less abundant than in the sulfate enriched reactors. The total abundance of

families associated with *mgl* was generally also higher in cysteine-enriched reactors as compared to

sulfate-enriched ones, except for the second cysteine-enriched replicate reactor on day 28. In the sulfate-

enriched reactors, most of the gene sequences associated with *mgl* were assigned to unclassified

Bacteroidetes and families outside the 10 most abundant ones, while in the cysteine-enriched reactors, the

largest share of gene sequences was associated with *Dethiosulfatibacteraceae*, *Clostridiaceae*,

Fusobacteriaceae, and unclassified Firmicutes.

Genomes encoding H₂S-producing pathways

Overall, 71 good quality MAGs (>70% completeness and <10% contamination) and 35 high quality

MAGs (>90% completeness and <5% contamination) were reconstructed (Supplementary Fig. S1) from

the metagenomic data. MAGs affiliated with *Dethiosulfatibacteraceae*, *Fusobacteriaceae*, and

Vibrionaceae were selected for in-depth analysis, because of their quantitative increase in cysteine-

enriched condition (Fig. 2), but also because of the prevalence of cysteine degrading genes in their

genomes (Fig. 4) and the broad representation of these MAGs across the different cysteine reactors

(Supplementary Table S4). A MAG associated with *Desulfovibrionaceae* was also included in this

pathway mapping, as it was found to be present in cysteine-enriched conditions (Supplementary Fig. S2)

and carried the *cyuA* gene (Fig. 4). All four MAGs encode the capacity for at least one cysteine degrading

gene (*cyuA, mgl, malY, cysK, sseA, metC, tnaA,* or *dcyD*).

Dethiosulfatibacteraceae MAG65b carried five different genes associated with potential cysteine

degradation activity (*cyuA, mgl, cysK, sseA,* and *tnaA*). *Psychrilyobacter* MAG139 also carried five

different genes with cysteine degradation potential (*cyuA, mgl, malY, cysK,* and *sseA)*, while Vibrio

MAG65 carried four genes with cysteine degradation potential (*cyuA, mgl, malY,* and *cysK*).

Desulfovibrionaceae MAG259 carried all the genes for dissimilatory sulfate reduction (*sat, aprAB,* and

dsrAB) and four different genes associated with cysteine degradation potential (*cyuA, cysK, metC,* and

dcyD) (Fig. 5).

Metagenome assembled genomes

Dethiosulfatibacteraceae MAG65b encoded a few amino acid or peptide transporters (Fig. 6A). We

detected branched-chain amino acid transporters (*LivKMGF*), the specific cysteine transporter (*cyuP,* also

called *yhaO*), and an oligopeptide transporter (*OppABCDF*). In addition to genes involved in cysteine

degradation (Fig. 5), MAG65b carried other amino acids degrading enzymes (threonine dehydratase

[EC:4.3.1.19], L and D-serine dehydratase [EC:4.3.1.17 4.3.1.18], and alanine dehydrogenase

[EC:1.4.1.1]) producing pyruvate and oxobutanoate, which could be fermented to acetate (phosphate

acetyltransferase [EC:2.3.1.8], acetate kinase [EC:2.7.2.1], and propionate CoA-transferase [EC:2.8.3.1]),

ethanol (alcohol dehydrogenase [EC:1.1.1.1]), or lactate (acryloyl-CoA pathway). The genome also

encoded the complete pathway for glycolysis and gluconeogenesis (glucose formation from pyruvate).

Overall, we detected 23 peptidase families that could be involved in catabolism of peptides and amino

acids.

Fusobacteriaceae MAG139 encoded several amino acid and peptide transporters (Fig. 6B). We detected

an S-methylcysteine transporter (*YxeMNO*), a cystine transporter (*TcyC*), a branched-chain amino acid

transporter (*LivK*), a D-methionine transporter (*MetQIN*), and an oligopeptide transporter (*OppABCDF*).

In addition to genes involved in cysteine degradation (Fig. 5), MAG139 encoded other amino acid-

degrading enzymes (threonine dehydratase [EC:4.3.1.19], L-serine dehydratase [EC:4.3.1.17], and L-

alanine dehydrogenase [EC:1.4.1.1]) producing pyruvate and oxobutanoate, which can then be fermented

to acetate (phosphate acetyltransferase [EC:2.3.1.8], acetate kinase [EC:2.7.2.1], propionate CoA-

transferase [EC:2.8.3.1], and succinyl-CoA:acetate CoA-transferase [EC:2.8.3.18]), ethanol (alcohol

dehydrogenase [EC:1.1.1.1]), or lactate (D-lactate dehydrogenase [EC:4.3.1.18]). Produced acetyl-CoA

could also feed the TCA cycle (tricarboxylic acid cycle), even if the key enzyme succinate dehydrogenase

was missing from the MAG. It could be that the gene was not recruited during the creation of MAG139,

although several copies of succinate dehydrogenase associated to *Fusobacteriaceae* family and even an

unclassified *Psychrilyobacter* were found in the metagenome samples (Supplementary Data). The genome

also encoded the complete glycolysis and gluconeogenesis pathways. Overall, we detected 34 peptidase

families that could be involved for catabolism of peptides and amino acids.

Vibrionaceae MAG65 encoded several amino acid and peptide transporters (Fig. 6C). We detected the

general L-amino acid transporters (*AapJQMP*), ABC-type polar-amino-acid transporter (*ArtIMQP*), D-

methionine transporter (*MetQIN*), oligopeptide transporter (*OppABCDF*), and the specific cysteine

transporter (*cyuP*). In addition to genes involved in cysteine degradation (Fig. 5), MAG65 encoded other

amino acid degrading enzymes (threonine dehydratase [EC:4.3.1.19], L-serine dehydratase [EC:4.3.1.17],

and D-alanine transaminase [EC:2.6.1.21]) to produce pyruvate and oxobutanoate which could be further

fermented to acetate (phosphate acetyltransferase [EC:2.3.1.8], acetate kinase [EC:2.7.2.1],

acylphosphatase [EC:3.6.1.7], and propionate CoA-transferase [EC:2.8.3.1]), ethanol (alcohol

dehydrogenase [EC:1.1.1.1]), or lactate (L and D-lactate dehydrogenase [EC:4.3.1.17 4.3.1.18]). Acetyl-

CoA produced could also be fed into the TCA cycle, even if 2-oxoglutarate ferredoxin oxidoreductase was

missing, probably due to it not being recruited during the creation of the MAG. The genome was also

found to encode the complete pathway for glycolysis and gluconeogenesis. Overall, we detected 31

peptidase families that could be involved in catabolism of peptides and amino acids.

Desulfovibrionaceae MAG259 encoded different peptide transporters (Fig. 6D). We detected the general

L-amino acid transporter (*AapJQMP*), branched-chain amino acid transporter (*LivKHMGF*), L-glutamine

transporter (*GlnHPQ*), and the specific cysteine transporter (*cyuP*). In addition to genes involved in

cysteine degradation (Fig. 5), MAG259 encoded other amino acid degrading enzymes (threonine 3-

dehydrogenase [EC:1.1.1.103], L-serine dehydratase [EC:4.3.1.17], alanine dehydrogenase [EC:1.4.1.1],

and D-alanine transaminase [EC:2.6.1.21]) producing pyruvate and oxobutanoate, which could be

fermented to acetate (phosphate acetyltransferase [EC:2.3.1.8], acetate kinase [EC:2.7.2.1], and acetyl-

CoA synthetase [EC:6.2.1.1]), ethanol (alcohol dehydrogenase [EC:1.1.1.1]) or lactate (L-Lactate

dehydrogenase [EC:1.1.1.27]). Acetyl-CoA produced could also feed into the TCA cycle. The genome

also encoded the complete pathway for glycolysis, gluconeogenesis, and dissimilatory sulfate reduction.

Overall, we detected 21 peptidase families that could be involved for catabolism of peptides and amino

acids.

Gene expression

Of the four selected MAGs, only *Fusobacteriaceae* MAG139 was represented by too few transcripts to

enable meaningful expression analyses (Supplementary Table S6). The metatranscriptomes show that in

all MAGs, *cyuA* is the only gene expressed for cysteine degradation, and this gene was upregulated

significantly and transcribed in all cysteine-fed incubations as compared to sulfate-fed incubations (Fig.

7). In *Dethiosulfatibacteraceae* MAG65b, several genes related to cysteine transport and degradation,

pyruvate metabolism, and gluconeogenesis were upregulated when cysteine was available as sulfur source

as compared to when sulfate was used as sulfur source (Fig. 7A). Thiosulfate/3-mercaptopyruvate

sulfurtransferase ([EC:2.8.1.1 2.8.1.2]), *sseA,* exhibited a substantial TPM increase in response to

cysteine, but the expression was not significantly different between sulfate or cysteine-fed incubations.

The gene expression for acetate kinase [EC 2.7.2.1] and succinate/propionate CoA-transferase

[EC:2.8.3.1], both related to pyruvate metabolism and transformation of acetyl-coa to acetate, also

increased significantly in response to cysteine. In contrast, tryptophanase [EC:4.1.99.1], formate C-

acetyltransferase [EC:2.3.1.54], pyruvate-ferredoxin/flavodoxin oxidoreductase [EC:1.2.7.1 1.2.7.-], 2-

oxoglutarate/2-oxoacid ferredoxin oxidoreductase subunit alpha/beta [EC:1.2.7.3 1.2.7.11], 2,3-

bisphosphoglycerate-independent phosphoglycerate mutase [EC:5.4.2.12], phosphoglycerate kinase

[EC:2.7.2.3], glyceraldehyde 3-phosphate dehydrogenase [EC:1.2.1.12], and

triosephosphate isomerase (TIM) [EC:5.3.1.1] were all significantly upregulated when sulfate was

available (Fig. 7A).

In *Vibrio* MAG65, most genes related to cysteine transport and degradation, pyruvate metabolism, and

gluconeogenesis were upregulated in response to cysteine as compared to sulfate-fed incubations (Fig.

7B).

However, some upregulated genes exhibited only moderate expression with no significant difference

between cysteine and sulfate-fed incubations. For example, *cysK* and *mgl* associated to cysteine

degradation and D-lactate dehydrogenase exhibited a low expression in both sulfate and cysteine-fed

incubations. The expression of cysteine importer *cyuP*, phosphoglycerate kinase [EC:2.7.2.3], fructose-

1,6-bisphosphatase I [EC:3.1.3.11], and glucose-6-phosphate isomerase [EC:5.3.1.9] was higher but not

significantly different between cysteine and sulfate-fed incubations. The rest of the genes selected

exhibited a significantly higher expression in cysteine-fed incubations as compared to those that were

sulfate-fed. Only *malY* (cysteine-S-conjugate beta-lyase [EC:4.4.1.13]) was significantly downregulated

in response to cysteine.

In *Desulfovibrionaceae* MAG259, several genes were upregulated when fed cysteine (cysteine only and

cysteine and sulfate) as compared to sulfate. Formate C-acetyltransferase [EC:2.3.1.54], pyruvate-

ferredoxin/flavodoxin oxidoreductase [EC:1.2.7.1 1.2.7.-], phosphate acetyltransferase [EC:2.3.1.8],

acetate kinase [EC:2.7.2.1], and L-lactate dehydrogenase [EC:1.1.1.27] associated with pyruvate

metabolism, were significantly upregulated. Acetyl-CoA synthetase [EC:6.2.1.1] was the only gene that

was upregulated, with also a clear increase of expression in sulfate-fed incubations. as compared to

cysteine-fed incubations. Other genes did not show any clear expression pattern.

Discussion

In engineered systems, H_2S often accumulates due to the imbalance between H_2S production and

consumption, causing e.g., sewer corrosion, health hazards, and aquaculture production losses (54,55). So

far, H_2S has been considered to originate from sulfate reduction, while evidence from natural ecosystems

suggests that microbial degradation of S-rich amino acids can also be an important H_2S source. In this

study, we examined the microbial production of H_2S in a sulfate-free medium, based exclusively on the S-

rich amino acid cysteine and compared the production and microbiology to a "traditional H_2S source" via

the i.e., dissimilatory sulfate reduction pathway.

First and foremost, we could demonstrate microbial H_2S production from both cysteine and sulfate. The

role of cysteine as a significant source of H_2S was further supported by an increase of acetate and

ammonium concentrations in incubation experiments with cysteine. Ammonium is a direct product of the

cysteine degradation reaction, originating from the amino group of cysteine released during degradation

(56,57). Acetate is not produced directly from cysteine degradation but can be derived from pyruvate

produced by cysteine degradation if the bacteria carry the acetate synthesis pathway (58). Conversely,

acetate concentration decreased in the incubations with sulfate, attributed to the consumption of acetate

by SRM in dissimilatory sulfate reduction (59). We also observed differences in H_2S production profiles

between the two enrichment conditions, as cysteine led to almost immediate H_2S production, while the

process started slower in sulfate-enriched reactors. One putative explanation for this lag could be that the

more preferred electron acceptors i.e. oxygen, nitrate, and fumarate (60) needed to be fully depleted to

allow sulfate reduction. This could also explain why the relative abundance of known SRM

(*Desulfuromonadaceae*, *Desulfocapsaceae*, *Desulfobacteraceae*, and *Geopsychrobacteraceae*) was still

rather low on day 14 in sulfate-enriched reactors (Fig. 2).

We identified three bacterial families significantly increasing their abundance under cysteine-enriched

conditions: *Dethiosulfatibacteraceae*, *Fusobacteriaceae*, and *Vibrionaceae*. Metagenomic analysis

showed that each of them carried at least one of the three main genes encoding for cysteine degrading

enzymes (*cyuA, mgl*, and *malY)* and that the relative abundance of these genes increased in response to

cysteine-enrichment. In human gut and oral microbiomes, these same three genes have been identified as

the key enzymes/genes for cysteine degradation in diverse bacterial taxa (13,61–64). Our results suggest

that these genes and families are the principal ones involved in cysteine based H_2S production in

engineered environments, but potentially also in natural environment, as member of these families have

found in lake water column and sediment (65,66), where also cysteine degradation activity has been

found (4,67). The rest of our analysis and discussion will be focused on the aforementioned families. We

also found a significant share of sequences encoding for one the genes (*cyuA*) to be assigned to family

Desulfovibrionaceae, and included it in the analysis, even if it was not abundant in the overall microbial

community (Fig. 2).

We could successfully reconstruct MAGs for each of the four families: *Dethiosulfatibacteraceae*,

Fusobacteriaceae, *Vibrionaceae*, and *Desulfovibrionaceae*. *Dethiosulfatibacteraceae* is a bacterial family

that is commonly found in marine sediments but has also been observed in engineered environments with

potential for H_2S production (e.g., offshore subsea oil storage or wastewater treatment) (68,69), and in

this study, in a recirculating aquaculture system. The members of this family are known to carry the

capacity for thiosulfate reduction and could also use a large range of other compounds such as lactate,

pyruvate, glutamate, peptone and amino acids like serine, threonine, histidine, lysine, or cysteine for

fermentation (65). We could reconstruct two *Dethiosulfatibacteraceae* MAGs (MAG21 and 65b)

(Supplementary Fig. S1), but as MAG21 was not abundant nor active, it was not analyzed further

(Supplementary Tables S5 and S6). In MAG65b, *cyuA* was the only cysteine degrading gene highly

transcribed and upregulated in cysteine-fed reactors in experiments carried out with the cysteine-enriched

biomass. As *cyuA* is a cysteine desulfidase, with primary function to degrade cysteine, we suggest this

enzyme to be responsible for H_2S production from cysteine in MAG65b. We also found MAG65b to carry

sseA, a gene that has two functions: thiosulfate reduction and cysteine degradation. As

Dethiosulfatibacteraceae is known for capacity to perform thiosulfate-reduction, we expected to find

thiosulfate reductase genes (e.g., thiosulfate reductase (quinone) genes (phsABC)) but did not detect any

such gene other than sseA. In our experiments, we detected an increase of sseA transcription in cysteine-

fed reactors, but this increase was not significant when compared to sulfate-fed reactors. It is still possible

that thiosulfate was produced and then reduced to H_2S by *sseA* in our reactors, which would explain why

this gene was transcribed under both cysteine and sulfate amended conditions. More in-depth functional

studies will need to be carried out for *Dethiosulfatibacteraceae* in order to understand how *sseA* reacts

and how it is involved in cysteine and thiosulfate metabolism. Our metagenome analysis confirmed that

our MAG65b *Dethiosulfatibacteraceae* could use cysteine either for gluconeogenesis or fermentation of

pyruvate to acetate, as all pathways and key enzymes were present (Fig. 6A). Moreover,

metatranscriptome analysis revealed acetate kinase and succinate/propionate CoA-transferase to be

upregulated with a relatively high transcription in cysteine-fed compared to sulfate-fed reactors (Fig. 7A),

indicating that cysteine could has been mostly used for fermentation purposes and ATP production

(70,71). This acetate production explains the observed increase of acetate concentration in the 48h-

experiment when cysteine was added (Table 1).

Unlike *Dethiosulfatibacteraceae*, members of the family *Fusobacteriaceae* (genus *Fusobacterium*) have

been shown to produce H_2S via the cysteine degradation pathway in human bodies (15,28,29), with

methionine gamma-lyase (mgl) as a key enzyme (13,61). The presence of *Fusobacterium* has also been

found to coincide with H_2S production from aquaculture organic waste (27). In our cysteine-enriched

samples, the key cysteine degrading genes sequences were all associated with the *Fusobacteriaceae*

family (Fig. 4), and most of these gene sequences were annotated to the *Fusobacterium* genus (see

Supplementary Data). These findings highlight the importance of this family for non-sulfate based H_2S

production in the presence of S-rich amino acids and absence of oxygen. Unfortunately, we could not

reconstruct any good quality MAGs associated with *Fusobacterium*. However, we did reconstruct a

Fusobacteriaceae MAG139 with 96% identity to *Psychrilyobacter sp02342785* (GTDB). Previously, the

Psychrilyobacter genus has been found to use a wide range of amino acids and peptides for growth in

both marine sediments (72) and abalone gut (73). Indeed, we found this MAG to carry an array of

peptidases, different transporters for either oligopeptide or amino acids, and different genes for the

fermentation of amino acids. It also carried four genes involved in cysteine degradation. Although we did

not obtain enough metatranscriptomic reads from the *Psychrilyobacter* MAG to enable further analysis

(Supplementary Tables S5 and S6), the high abundance of *Fusobacteriaceae* in the cysteine-enriched

reactors and the fact that they harbor key cysteine degrading genes combined with earlier findings in the

medical field, strongly suggests that *Psychrilyobacter* and/or the broader *Fusobacteriaceae* family play an

important role in the production of H_2S from cysteine also in engineered ecosystems.

The family *Vibrionaceae* is mainly known for its many pathogens capable of causing human and animal

disease (74,75), but this diverse genus also includes other members native to a wide range of aquatic

habitats (e.g., sediment, column water, deep sea, or as symbionts (66,76)). The *Vibrionaceae* family is

known to be capable of cysteine degradation (77), and at least some members, such as *Vibrio cholerae,*

are using the putative cystathionine B-synthase (*cbs*) for H_2S-producing cysteine degradation (78). We did

not detect any cbs genes in the metagenomes, but found the genes *cyuA, malY, cysK, sseA, metC*, and *tnaA*

associated with the *Vibrionaceae* family. We could reconstruct two MAGs (MAG30, 65) associated to the

genus *Vibrio*. MAG30 carried all the cysteine degrading genes listed above but was not abundant and

showed little presence and expression in the metatranscriptomes, suggesting it not being active

(Supplementary Tables S5 and S6). In contrast, the second *Vibrio* MAG (MAG65), containing the genes

cyuA, mgl, malY, and *cysK,* appeared to be extremely active based on our community-wide expression

profiling. The *cyuPA* (also known as *yhaOM*) operon is known to be the major anaerobic cysteine

transporter and cysteine desulfidase in *Escherichia coli* and *Salmonella enterica* (63). We found these

genes to be highly expressed in our study (Fig. 7B), but only *cyuA* was upregulated significantly in

cysteine-fed reactors as compared to sulfate-fed reactors, suggesting it to be the main enzyme responsible

for cysteine degradation and H_2S production in our *Vibrio* MAG65. To the best of our knowledge, this is

the first time that *cyuA* has been identified as the key gene for H_2S production in a *Vibrio sp*. Previously,

some members of the *Vibrio* genus have been found to ferment glucose to pyruvate and subsequently

transform this metabolite further to acetate (79). We found Vibrio MAG65 to carry such a complete

pathway for pyruvate fermentation to acetate and genes related to this fermentation were all upregulated

and highly transcribed in cysteine-fed reactors compared to sulfate-fed reactors (Fig. 7B). This inferred

acetate production can explain the observed increase of acetate concentration in the 48h-experiment when

cysteine was added (Table 1). We also observed the transcription of all genes for gluconeogenesis, a

majority being significantly upregulated in cysteine-fed reactors as compared to sulfate-fed treatments,

suggesting that pyruvate produced from cysteine was also transformed into glucose as an energy-reserve

and that cysteine was not only used for sulfur and ATP but also as a carbon source in our Vibrio MAG65.

In addition to the three above-mentioned families, we also found the family *Desulfovibrionaceae* to be of

interest. *Desulfovibrionaceae* is a well-known SRM, which is commonly found in marine environments

but also in engineered H_2S-rich environments (e.g., aquaculture, wastewater) (6,79,80). However, we did

not detect the *Desulfovibrionaceae* family in the sulfate-enriched reactors but only in cysteine-enriched

reactors (Supplementary Fig. S2). This could be explained by the media used in this study. The media

contained a combination of formate, acetate, and propionate, as it was formulated following previous

knowledge on optimal SRM enrichment media. However, some genera within the *Desulfovibrionaceae*

lack the capacity to use acetate or propionate to fuel dissimilatory sulfate reduction (84,85), meaning that

they would be restricted to use formate for this purpose. Hence when formate would have been depleted

in our reactors, *Desulfovibrionaceae* would not be able to survive in sulfate media, while in cysteine-

enriched media, they could potentially have used cysteine as an alternative carbon source for fermentation

and growth, as previously suggested for *Desulfovibrionaceae* (81–83). Indeed, our metagenome and

metatranscriptome analysis confirm the potential of *Desulfovibrionaceae* MAG259 to function as a

cysteine degrader and fermenter in addition to being a sulfate reducing bacteria. Of the four cysteine

degrading genes seen in this population, only *cyuA* was highly transcribed and upregulated in cysteine-fed

reactors compared to sulfate-fed reactors (Fig. 7C), strongly suggesting that *cyuA* was also the key

enzyme for cysteine degradation in our *Desulfovibrionaceae* MAG259. Pyruvate produced by *cyuA* could

have been converted into acetyl-coA and then acetate for energy acquisition and ATP production

explaining the upregulation of formate C-acetyltransferase, pyruvate-ferredoxin/flavodoxin

oxidoreductase, phosphate acetyltransferase, and acetate kinase in cysteine-fed incubations as compared

to sulfate-fed ones. Analogous to the previous MAGs, this acetate production can explain observed

increases in acetate concentration during the 48h-experiment when cysteine was added (Table 1).

Pyruvate could also be used by *Desulfovibrionaceae* MAG259 to produce lactate or formate by either L-

lactate dehydrogenase or formate C-acetyltransferase. Both lactate and formate could then be used for

dissimilatory sulfate reduction if sulfate is available (86). Finally, pyruvate could also be transformed into

glucose via gluconeogenesis, as described for the previous MAG. Therefore, in our MAG259, cysteine

could have a central place in the metabolism of the cell as either a carbon or energy source or as a source

of organic matter to drive dissimilatory sulfate reduction.

By combining production measurements, amplicon sequencing, and metagenomic and metatranscriptomic

analysis, we demonstrate that cysteine degradation can be an important H_2S source with four microbial

families centrally involved in this process. *cyuA* seems to be the main gene responsible for cysteine

degradation, and while there is previous evidence on the ability of two of these families

(*Fusobacteriaceae* and *Vibrionaceae*) to carry out H_2S production through cysteine degradation, we also

identified one new family (*Dethiosulfatibacteraceae*) to be involved in H_2S production from cysteine and

one family (*Desulfovibrionaceae*) to have the potential to engage in either SRM or cysteine degradation.

We also showed that cysteine can be used as a carbon and energy source by some bacteria to fuel general

cellular anabolic and catabolic needs.

While known to take place in nature, cysteine degradation has not been recognized as a significant H_2S

source in engineered environments. Previous studies assigned it as insignificant compared to sulfate

reduction in sewage biofilms (80), but those biofilms were on the other hand enriched with sulfate and

might have overlooked the forementioned process. While in some engineered systems, the supply and

inputs of cysteine and other S-rich amino acids is limited, environments such as sewage, biogas facilities,

and aquaculture systems are by design rich in proteins, which could enable and enhance non-sulfate based

H_2S production. In line with this, recent evidence shows a positive relationship between H_SS production

and sludge protein content (25,26). Overall, our findings encourage further research on H_2S and sulfur

cycling in engineered environments and call for such work to be expanded and also cover cysteine-fueled

other non-sulfate-based H_2S production pathways in order to gain a complete understanding of the factors

controlling H_2S production. We can then use this information to develop effective management strategies.

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Conflicts of interest

The authors declare no conflicts of interest.

Author contributions

AN and SLA designed the study with input from SB. AN carried out the experimental work and chemical

analysis with support from SLA. AN performed amplicon sequencing and prepared samples for

metagenomic and metatranscriptomic analysis with support from SB. AN and FPS performed

bioinformatics. All authors contributed to data interpretation. AN prepared the first version of the

manuscript with major contributions from SLA and additional inputs from FPS and SB in shaping the

final manuscript.

Supplementary material

Supplementary material is available at The ISME Journal online.

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Data availability

Amplicon sequencing data sets, metagenome-assembled genomes, and metagenome and

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References

1. Berner RA. Electrode studies of hydrogen sulfide in marine sediments. *Geochim Cosmochim Acta*.

1963;27:563–75. https://doi.org/10.1016/0016-7037(63)90013-9

2. Gaill F. Aspects of life development at deep sea hydrothermal vents. *FASEB J*. 1993;7:558–65.

https://doi.org/10.1096/fasebj.7.6.8472894

3. Schunck H, Lavik G, Desai DK, Großkopf T, Kalvelage T, Löscher CR, et al. Giant Hydrogen

Sulfide Plume in the Oxygen Minimum Zone off Peru Supports Chemolithoautotrophy. *PLoS One*.

2013;8:e68661. https://doi.org/10.1371/journal.pone.0068661

4. Zinder SH, Brock TD. Methane, Carbon Dioxide, and Hydrogen Sulfide Production from the

Terminal Methiol Group of Methionine by Anaerobic Lake Sediments. *Appl Environ Microbiol*.

1978;35:344–52. https://doi.org/10.1128/aem.35.2.344-352.1978

5. Marriott RA, Pirzadeh P, Marrugo-Hernandez JJ, Raval S. Hydrogen sulfide formation in oil and

gas. *Can J Chem*. 2016;94:406–13. https://doi.org/10.1139/cjc-2015-0425

6. Rojas-Tirado P, Aalto SL, Åtland Å, Letelier-Gordo C. Biofilters are potential hotspots for H_2S

production in brackish and marine water RAS. *Aquaculture*. 2021;536:736490.

https://doi.org/10.1016/j.aquaculture.2021.736490

7. Sharma KR, Yuan Z, de Haas D, Hamilton G, Corrie S, Keller J. Dynamics and dynamic

modelling of H_2S production in sewer systems. *Water Res*. 2008;42:2527–38.

https://doi.org/10.1016/j.watres.2008.02.013

8. Muyzer G, Stams AJM. The ecology and biotechnology of sulphate-reducing bacteria. *Nat Rev*

Microbiol. 2008;6:441–54. https://doi.org/10.1038/nrmicro1892

9. Wasmund K, Mußmann M, Loy A. The life sulfuric: microbial ecology of sulfur cycling in marine

sediments. *Environ Microbiol Rep*. 2017;9:323–44. https://doi.org/10.1111/1758-2229.12538

10. Dunnette DA, Chynoweth DP, Mancy KH. The source of hydrogen sulfide in anoxic sediment.

Water Res. 1985;19:875–84. https://doi.org/10.1016/0043-1354(85)90146-0

11. Bruinenberg PG, De Roo G, Limsowtin G. Purification and Characterization of Cystathionine

(gamma)-Lyase from *Lactococcus lactis* subsp. *cremoris* SK11: Possible Role in Flavor

Compound Formation during Cheese Maturation. *Appl Environ Microbiol*. 1997;63:561–6.

https://doi.org/10.1128/aem.63.2.561-566.1997

12. Awano N, Wada M, Mori H, Nakamori S, Takagi H. Identification and Functional Analysis of

Escherichia coli Cysteine Desulfhydrases. *Appl Environ Microbiol*. 2005;71:4149–52.

https://doi.org/10.1128/AEM.71.7.4149-4152.2005

13. Suwabe K, Yoshida Y, Nagano K, Yoshimura F. Identification of an l-methionine γ-lyase involved

in the production of hydrogen sulfide from l-cysteine in *Fusobacterium nucleatum* subsp.

nucleatum ATCC 25586. *MicroSoc*. 2011;157:2992–3000. https://doi.org/10.1099/mic.0.051813-0

14. Blachier F, Beaumont M, Kim E. Cysteine-derived hydrogen sulfide and gut health: a matter of

endogenous or bacterial origin. *Curr Opin Clin Nutr Metab Care*. 2019;22:68–75.

https://doi.org/10.1097/mco.0000000000000526

15. Braccia DJ, Jiang X, Pop M, Hall AB. The Capacity to Produce Hydrogen Sulfide (H_2S) via

Cysteine Degradation is Ubiquitous in the Human Gut Microbiome. *Front Microbiol*. 2021;12.

https://doi.org/10.3389/fmicb.2021.705583

16. Hale VL, Jeraldo P, Mundy M, Yao J, Keeney G, Scott N, et al. Synthesis of multi-omic data and

community metabolic models reveals insights into the role of hydrogen sulfide in colon cancer.

Methods. 2018;149:59–68. https://doi.org/10.1016/j.ymeth.2018.04.024

17. Tran PQ, Bachand SC, Hotvedt JC, Kieft K, McDaniel EA, McMahon KD, et al. Physiological

and genomic evidence of cysteine degradation and aerobic hydrogen sulfide production in

freshwater bacteria. *mSystems*. 2023;8:e00201-23. https://doi.org/10.1128/msystems.00201-23

18. Cooper CE, Brown GC. The inhibition of mitochondrial cytochrome oxidase by the gases carbon

monoxide, nitric oxide, hydrogen cyanide and hydrogen sulfide: chemical mechanism and

physiological significance. *J Bioenerg Biomembr*. 2008;40:533. https://doi.org/10.1007/s10863-

008-9166-6

19. Rumbeiha W, Whitley E, Anantharam P, Kim DS, Kanthasamy A. Acute hydrogen sulfide–induced

neuropathology and neurological sequelae: challenges for translational neuroprotective research.

Ann N Y Acad Sci. 2016;1378:5-16. https://doi.org/10.1111/nyas.13148

20. Li P, Li M, Liu Z, Zhao Y, Qian X, Huang P. Effect of high temperature and sulfur vapor on the

flammability limit of hydrogen sulfide. *J Clean Prod*. 2022;337:130579.

https://doi.org/10.1016/j.jclepro.2022.130579

21. Joseph AP, Keller J, Bustamante H, Bond PL. Surface neutralization and H_2S oxidation at early

stages of sewer corrosion: Influence of temperature, relative humidity and H_2S concentration.

Water Res. 2012;46:4235–45. https://doi.org/10.1016/j.watres.2012.05.011

22. Zhang L, De Schryver P, De Gusseme B, De Muynck W, Boon N, Verstraete W. Chemical and

biological technologies for hydrogen sulfide emission control in sewer systems: A review. *Water*

Res. 2008;42:1–12. https://doi.org/10.1016/j.watres.2007.07.013

23. Zhang L, Qiu YY, Sharma KR, Shi T, Song Y, Sun J, et al. Hydrogen sulfide control in sewer

systems: A critical review of recent progress. *Water Res*. 2023;240:120046.

24. Jasmin MY, Syukri F, Kamarudin MS, Karim M. Potential of bioremediation in treating

aquaculture sludge: Review article. *Aquaculture*. 2020;519:734905.

25. Choudhury A, Lepine C, Good C. Methane and Hydrogen Sulfide Production from the Anaerobic

Digestion of Fish Sludge from Recirculating Aquaculture Systems: Effect of Varying Initial Solid

Concentrations. *Fermentation*. 2023;9:94. https://doi.org/10.3390/fermentation9020094

26. Forouzanmehr F, Solon K, Maisonnave V, Daniel O, Volcke EIP, Gillot S, et al. Sulfur

transformations during two-stage anaerobic digestion and intermediate thermal hydrolysis. *Sci*

Total Environ. 2022;810:151247. https://doi.org/10.1016/j.scitotenv.2021.151247

27. Letelier-Gordo CO, Aalto SL, Suurnäkki S, Pedersen PB. Increased sulfate availability in saline

water promotes hydrogen sulfide production in fish organic waste. *Aquac Eng*. 2020;89:102062.

https://doi.org/10.1016/j.aquaeng.2020.102062

28. Carda-Diéguez M, Rosier BT, Lloret S, Llena C, Mira A. The tongue biofilm metatranscriptome

identifies metabolic pathways associated with the presence or absence of halitosis. *NPJ Biofilms*

Microbiomes. 2022;8:1–10. https://doi.org/10.1038/s41522-022-00364-2

29. Hampelska K, Jaworska MM, Babalska ZŁ, Karpiński TM. The Role of Oral Microbiota in Intra-

Oral Halitosis. *J Clin Med*. 2020;9:2484. https://doi.org/10.3390/jcm9082484

30. Eaton AD, Clesceri LS, Greenberg AE, Franson MAH. Standard Methods for the Examination of

Water and Wastewater. 20th ed. American Public Health Association: Washington, DC, USA,

1998.

31. Vass M, Székely AJ, Lindström ES, Langenheder S. Using null models to compare bacterial and

microeukaryotic metacommunity assembly under shifting environmental conditions. *Sci Rep*.

2020;10:2455. https://doi.org/10.1038/s41598-020-59182-1

32. Bolyen E, Rideout JR, Dillon MR, Bokulich NA, Abnet CC, Al-Ghalith GA, et al. Reproducible,

interactive, scalable and extensible microbiome data science using QIIME 2. *Nat Biotechnol*.

2019;37:852–7. https://doi.org/10.1038/s41587-019-0209-9

33. Callahan BJ, McMurdie PJ, Rosen MJ, Han AW, Johnson AJA, Holmes SP. DADA2: High-

resolution sample inference from Illumina amplicon data. *Nat Methods*. 2016;13:581–3.

https://doi.org/10.1038/nmeth.3869

34. Quast C, Pruesse E, Yilmaz P, Gerken J, Schweer T, Yarza P, et al. The SILVA ribosomal RNA

gene database project: improved data processing and web-based tools. *Nucleic Acids Res*.

2013;41:D590–6. https://doi.org/10.1093/nar/gks1219

35. R Core Team. R: A Language and Environment for Statistical Computing. R foundation for

Statistical Computing: Vienna, Austria. 2023. https://www.r-project.org/

36. McMurdie PJ, Holmes S. phyloseq: An R Package for Reproducible Interactive Analysis and

Graphics of Microbiome Census Data. *PLoS One*. 2013;8:e61217.

https://doi.org/10.1371/journal.pone.0061217

37. Wickham H. ggplot2. *Wiley Interdiscip Rev Comput Stat*. 2011;3:180–5.

https://doi.org/10.1002/wics.147

38. Tamames J, Puente-Sánchez F. SqueezeMeta, A Highly Portable, Fully Automatic Metagenomic

Analysis Pipeline. *Front Microbiol*. 2019;9. https://doi.org/10.3389/fmicb.2018.03349

39. Li D, Liu CM, Luo R, Sadakane K, Lam TW. MEGAHIT: an ultra-fast single-node solution for

large and complex metagenomics assembly via succinct de Bruijn graph. *Bioinformatics*.

2015;31:1674–6. https://doi.org/10.1093/bioinformatics/btv033

40. Buchfink B, Xie C, Huson DH. Fast and sensitive protein alignment using DIAMOND. *Nat*

Methods. 2015;12:59–60. https://doi.org/10.1038/nmeth.3176

41. Huerta-Cepas J, Szklarczyk D, Forslund K, Cook H, Heller D, Walter MC, et al. eggNOG 4.5: a

hierarchical orthology framework with improved functional annotations for eukaryotic,

prokaryotic   and   viral   sequences.   *Nucleic   Acids   Res*.   2016;44:286–93.

https://doi.org/10.1093/nar/gkv1248

42. Kanehisa M, Goto S. KEGG: Kyoto Encyclopedia of Genes and Genomes. *Nucleic Acids Res*.

2000;28:27–30. https://doi.org/10.1093/nar/28.1.27

43. Finn RD, Bateman A, Clements J, Coggill P, Eberhardt RY, Eddy SR, et al. Pfam: the protein

families database. *Nucleic Acids Res*. 2014;42:222–30. https://doi.org/10.1093/nar/gkt1223

44. Langmead B, Salzberg SL. Fast gapped-read alignment with Bowtie 2. *Nat Methods*. 2012;9:357–

9. https://doi.org/10.1038/nmeth.1923

45. Alneberg J, Bjarnason BS, de Bruijn I, Schirmer M, Quick J, Ijaz UZ, et al. Binning metagenomic

contigs   by   coverage   and   composition.   *Nat   Methods*.   2014;11:1144–6.

https://doi.org/10.1038/nmeth.3103

46. Kang DD, Li F, Kirton E, Thomas A, Egan R, An H, et al. MetaBAT 2: an adaptive binning

algorithm for robust and efficient genome reconstruction from metagenome assemblies. PeerJ.

2019;7:e7359.

47. Parks DH, Rinke C, Chuvochina M, Chaumeil PA, Woodcroft BJ, Evans PN, et al. Recovery of

nearly 8,000 metagenome-assembled genomes substantially expands the tree of life. *Nat*

Microbiol. 2017;2:1533–42. https://doi.org/10.1038/s41564-017-0012-7

48. Chaumeil PA, Mussig AJ, Hugenholtz P, Parks DH. GTDB-Tk v2: memory friendly classification
with the genome taxonomy database. *Bioinformatics*. 2022;38:5315–6.
https://doi.org/10.1093/bioinformatics/btac672

49. Parks DH, Imelfort M, Skennerton CT, Hugenholtz P, Tyson GW. CheckM: assessing the quality
of microbial genomes recovered from isolates, single cells, and metagenomes. *Genome Res*.
2015;25:1043–55. https://doi.org/10.1101/gr.186072.114

50. Tiwari RR. Occupational health hazards in sewage and sanitary workers. *Indian J Occup Environ*
Med. 2008;12:112. https://doi.org/10.4103/0019-5278.44691

51. Puente-Sánchez F, García-García N, Tamames J. SQMtools: automated processing and visual
analysis of 'omics data with R and anvi'o. BMC Bioinformatics. 2020;21(1):358.

52. Love MI, Huber W, Anders S. Moderated estimation of fold change and dispersion for RNA-seq
data with DESeq2. Genome Biol. 2014;15(12):550.

53. Quinn TP, Crowley TM, Richardson MF. Benchmarking differential expression analysis tools for
RNA-Seq: normalization-based vs. log-ratio transformation-based methods. BMC Bioinformatics.
2018;19(1):274.

54. Boyd CE. Chapter 6 - General Relationship Between Water Quality and Aquaculture Performance
in Ponds. In: Jeney G, editor. Fish Diseases. Academic Press, 2017. pp 147–166.

55. Tchong SI, Xu H, White RH. L -Cysteine Desulfidase: An [4Fe-4S] Enzyme Isolated from
Methanocaldococcus jannaschii That Catalyzes the Breakdown of L -Cysteine into Pyruvate,
Ammonia, and Sulfide. *Biochemistry*. 2005;44:1659–70. https://doi.org/10.1021/bi0484769

56. Kredich NM, Foote LJ, Keenan BS. The Stoichiometry and Kinetics of the Inducible Cysteine
Desulfhydrase from *Salmonella typhimurium. J Biol Chem*. 1973;248:6187–96.
https://doi.org/10.1016/S0021-9258(19)43526-6

57. Göbbels L, Poehlein A, Dumnitch A, Egelkamp R, Kröger C, Haerdter J, et al. Cysteine: an
overlooked energy and carbon source. *Sci Rep*. 2021;11:2139. https://doi.org/10.1038/s41598-021-
81103-z

58. Sørensen J, Christensen D, Jørgensen BB. Volatile Fatty Acids and Hydrogen as Substrates for
Sulfate-Reducing Bacteria in Anaerobic Marine Sediment. *Appl Environ Microbiol*. 1981;42:5–11.
https://doi.org/10.1128/aem.42.1.5-11.1981

59. Malik S, Kishore S, Dhasmana A, Kumari P, Mitra T, Chaudhary V, et al. A Perspective Review on
Microbial Fuel Cells in Treatment and Product Recovery from Wastewater. *Water*. 2023;15:316.
https://doi.org/10.3390/w15020316

60. Chen YW, Camacho MI, Chen Y, Bhat AH, Chang C, Peluso EA, et al. Genetic Determinants of
Hydrogen Sulfide Biosynthesis in *Fusobacterium nucleatum* Are Required for Bacterial Fitness,
Antibiotic   Sensitivity,   and   Virulence.   *mBio*.   2022;13:e01936-22.
https://doi.org/10.1128/mbio.01936-22

61. Zhou Y, Imlay JA. *Escherichia coli* Uses a Dedicated Importer and Desulfidase To Ferment
Cysteine. *mBio*. 2022;13:e02965-21. https://doi.org/10.1128/mbio.02965-21

62. Loddeke M, Schneider B, Oguri T, Mehta I, Xuan Z, Reitzer L. Anaerobic Cysteine Degradation
and Potential Metabolic Coordination in *Salmonella enterica* and *Escherichia coli*. *J Bacteriol*.
2017;199:e00117-17. https://doi.org/10.1128/jb.00117-17

63. Li K, Xin Y, Xuan G, Zhao R, Liu H, Xia Y, et al. *Escherichia coli* Uses Separate Enzymes to
Produce H_2S and Reactive Sulfane Sulfur From L-cysteine. *Front Microbiol*. 2019;10.
https://doi.org/10.3389/fmicb.2019.00298

64. Li W, Liu J, Zhen Y, Lin M, Sui X, Zhao W, et al. Simultaneous removal of nitrite and organics in
a biofilm-enhanced high-salt wastewater treatment system via mixotrophic denitrification coupled
with   sulfate   reduction.   *J   Water   Process   Eng*.   2021;40:101976.
https://doi.org/10.1016/j.jwpe.2021.101976

65. Vigneron A, Cruaud P, Ducellier F, Head IM, Tsesmetzis N. Syntrophic Hydrocarbon Degradation
in a Decommissioned Off-Shore Subsea Oil Storage Structure. *Microorganisms*. 2021;9:356.
https://doi.org/10.3390/microorganisms9020356

66. Takii S, Hanada S, Tamaki H, Ueno Y, Sekiguchi Y, Ibe A, et al. *Dethiosulfatibacter aminovorans*

gen. nov., sp. nov., a novel thiosulfate-reducing bacterium isolated from coastal marine sediment

via sulfate-reducing enrichment with Casamino acids. *Int J Syst Evol Micr*. 2007;57:2320–6.

https://doi.org/10.1099/ijs.0.64882-0

67. Ingram-Smith C, Martin SR, Smith KS. Acetate kinase: not just a bacterial enzyme. *Trends*

Microbiol. 2006;14:249–53. https://doi.org/10.1016/j.tim.2006.04.001

68. Schäfer T, Schönheit P. Pyruvate metabolism of the hyperthermophilic archaebacterium

Pyrococcus furiosus. *Arch Microbiol*. 1991;155:366–77. https://doi.org/10.1007/BF00243457

69. Pelikan C, Wasmund K, Glombitza C, Hausmann B, Herbold CW, Flieder M, et al. Anaerobic

bacterial degradation of protein and lipid macromolecules in subarctic marine sediment. *ISME J*.

2021;15:833–47. https://doi.org/10.1038/s41396-020-00817-6

70. Liu M, Wei G, Lai Q, Huang Z, Li M, Shao Z. Genomic and metabolic insights into the first host-

associated isolate of *Psychrilyobacter*. *Microbiol Spectr*. 2023;11:e03990-22.

https://doi.org/10.1128/spectrum.03990-22

71. Colwell RR, Grimes DJ. *Vibrio* diseases of marine fish populations. *Helgol Meeresunters*.

1984;37:265–87. https://doi.org/10.1007/BF01989311

72. Jones MK, Oliver JD. *Vibrio vulnificus*: Disease and Pathogenesis. *Infect Immun*. 2009;77:1723–

33. https://doi.org/10.1128/iai.01046-08

73. Dryselius R, Kurokawa K, Iida T. *Vibrionaceae*, a versatile bacterial family with evolutionarily

conserved variability. *Res Microbiol*. 2007;158:479–86.

https://doi.org/10.1016/j.resmic.2007.04.007

74. Thompson FL, Iida T, Swings J. Biodiversity of Vibrios. *Microbiol Mol Biol Rev*. 2004;68:403–31.

https://doi.org/10.1128/mmbr.68.3.403-431.2004

75. Gram L, Trolle G, Huss HH. Detection of specific spoilage bacteria from fish stored at low (0°C)

and high (20°C) temperatures. *Int J Food Microbiol*. 1987;4:65–72. https://doi.org/10.1016/0168-

1605(87)90060-2

76. Ma Y, Yang X, Wang H, Qin Z, Yi C, Shi C, et al. CBS-derived H₂S facilitates host colonization of

Vibrio cholerae by promoting the iron-dependent catalase activity of KatB. *PLoS Pathogens*.

2021;17:e1009763. https://doi.org/10.1371/journal.ppat.1009763

77. Rasheedkhan Regina V, Noorian P, Sim CBW, Constancias F, Kaliyamoorthy E, Booth SC, et al.

Loss of the Acetate Switch in *Vibrio vulnificus* Enhances Predation Defense against *Tetrahymena*

pyriformis. *Appl Environ Microbiol*. 2022;88:e01665-21. https://doi.org/10.1128/AEM.01665-21

78. Rurangwa E, Verdegem MCJ. Microorganisms in recirculating aquaculture systems and their

management. *Rev Aquac*. 2015;7:117–30. https://doi.org/10.1111/raq.12057

79. Chen Z, Chang Z, Qiao L, Wang J, Liu Y, Song X, et al. Effect of hydraulic retention time on

solid-phase   denitrification   reactor   in   recirculating   aquaculture   system.   *Aquaculture*.

2021;543:736928. https://doi.org/10.1016/j.aquaculture.2021.736928

80. Forsberg CW. Sulfide Production from Cysteine by *Desulfovibrio desulfuricans*. *Appl Environ*

Microbiol. 1980;39:453–5. https://doi.org/10.1128/aem.39.2.453-455.1980

81. Ueno A, Tamazawa S, Tamamura S, Murakami T, Kiyama T, Inomata H, et al. *Desulfovibrio*

subterraneus sp. nov., a mesophilic sulfate-reducing deltaproteobacterium isolated from a deep

siliceous   mudstone   formation.   *Int   J   Syst   Evol   Micr*.   2021;71:004683.

https://doi.org/10.1099/ijsem.0.004683

82. Galushko A, Kuever J. Desulfovibrionaceae. In: Bergey's Manual of Systematics of Archaea and

Bacteria.   John   Wiley   &   Sons,   Ltd;   2020.   pp.   1–13.

https://doi.org/10.1002/9781118960608.fbm00199.pub2

83. Postgate JR, Campbell LL. Classification of *Desulfovibrio* species, the nonsporulating sulfate-

reducing bacteria. *Bacteriol Rev*. 1966;30:732–8. https://doi.org/10.1128/br.30.4.732-738.1966

84. Liu P, Conrad R. *Syntrophobacteraceae*-affiliated species are major propionate-degrading sulfate

reducers in paddy soil. *Environ Microbiol*. 2017;19:1669–86. https://doi.org/10.1111/1462-

2920.13698

85. Park MJ, Kim YJ, Park M, Yu J, Namirimu T, Roh YR, et al. Establishment of Genome Based

Criteria for Classification of the Family Desulfovibrionaceae and Proposal of Two Novel Genera,

Alkalidesulfovibrio gen. nov. and *Salidesulfovibrio* gen. nov. *Front Microbiol*. 2022;13.

https://doi.org/10.3389/fmicb.2022.738205

86. Nielsen P, Raunkjær K, Norsker NH, Jensen N, Hvitved-Jacobsen T. Transformation of

Wastewater in Sewer Systems – A Review. *Wat Sci Tech*. 1992;25:17–31.

https://doi.org/10.2166/wst.1992.0111

 Figure legends

 Figure 1. Hydrogen sulfide concentrations (mg/l) under two different enrichment conditions (solid line =

 cysteine-enriched, dotted line = sulfate-enriched) from three samples, Aqua-Sample-1 in black, Aqua-

 Sample-2 in grey and Aqua-Sample-3 in blue.

 Figure 2. Relative abundances of the top 25 most abundant bacterial families in original material of

 Aqua-Samples 1 and 2 (raw), at day 14, 21, and 28 under cysteine and sulfate-enriched conditions, and at

 day 76 under cysteine-enriched condition. "Other" refers to the families outside of the top 25 most

 abundant ones.

 Figure 3. Abundance of genes in the metagenomes reported as coverage per million reads (cpm) with

 focus of those known to be associated with cysteine degradation in orange, dissimilatory sulfate reduction

 in green, thiosulfate reduction in blue, and sulfide oxidation gene in pink. Data are for sulfate-enriched

 reactors at day 28 (Sulfate-d28-1/2) and cysteine-enriched reactors at days 28 and 76 (Cysteine-d28-1/2,

 Cysteine-d76-1/2).

 Figure 4. The weight (in percentage of total reads) and taxonomic distribution of *cyuA* (cysteine

 desulfidase), *malY* (cystathionase), and *mgl* (methionine gamma-lyase) in metagenomes under sulfate-

 enrichment after 28 days (Sulfate-d28-1/2) and under cysteine-enrichment after 28 and 76 days (Cysteine-

 d28-1/2, Cysteine-d76-1/2). "Other" refers to the families outside of the 10 most abundant ones.

 Figure 5. Diagram showing the inferred metabolic pathways responsible for H_2S production from sulfur

 and organosulfur compound for the most abundant MAG of the predetermined target families. The

 presence or absence of the key genes related to the different metabolic features are shown by solid or

 open circles, respectively, and with different colors for each of the four MAGs. Dissimilatory sulfate

 reduction pathway is associated with the presence of *sat, aprAB,* and *dsrAB*; cysteine degradation activity

 is associated with *cyuA, mgl, malY, cysK, cysM, sseA, metC, tnaA, dcyD, cbs,* and *cse.*

 Figure 6. Central metabolism, pyruvate metabolism, and amino acids degradation pathway reconstructed

 for the four most abundant MAGs associated with the target families and qualified as at least good quality

 MAGs: A) *Dethiosulfatibacteraceae* MAG65b, (72.4% complete, 3.5% contamination) with no species

related from GTDB database, B) *Fusobacteriaceae* MAG139 (99.4% completion, 1.9% contamination)

classified as the same species of *Psychrilyobacter* sp002342785 from GTDB database, C) *Vibrionaceae*

MAG65 (77.3% complete and 0% contamination) with no species related from GTDB database, and D)

Desulfovibrionaceae MAG259 (99.4% complete and 9.5% contamination) with no species related from

GTDB database. FCA: formate C-acetyltransferase [EC:2.3.1.54]; PFFO: pyruvate-ferredoxin/flavodoxin

oxidoreductase [EC:1.2.7.1 1.2.7.-]; 2OOF: 2-oxoglutarate/2-oxoacid ferredoxin oxidoreductase

[EC:1.2.7.3 1.2.7.11].

Figure 7. Differences in the gene expression analyzed with DESeq2 and heatmap showing TPM of the

genes of interest annotated and grouped according to pathway as cysteine transporters (in yellow),

cysteine degradation (in green), pyruvate metabolism in (purple) and gluconeogenesis (in orange) for A)

Dethiosulfatibacteraceae MAG65b between sulfate or cysteine-fed reactors B) *Vibrionaceae* MAG65

between sulfate or cysteine-fed reactors, and C) *Desulfovibrionaceae* MAG259 between sulfate-fed,

cysteine-fed and sulfate+cysteine-fed reactors in experiments 1, 3, and 4. Statistically significant

differences (adjusted $p < 0.05$) in genes expression are marked with an asterisk (*).