Automated detection of Bornean white-bearded gibbon (*Hylobates albibarbis*)

vocalisations using an open-source framework for deep learning

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Passive acoustic monitoring is a promising tool for monitoring at-risk populations of vocal

species, yet extracting relevant information from large acoustic datasets can be time-

consuming, creating a bottleneck at the point of analysis. To address this, we adapted an open-

source framework for deep learning in bioacoustics to automatically detect Bornean white-

bearded gibbon (*Hylobates albibarbis*) "great call" vocalisations in a long-term acoustic dataset

from a rainforest location in Borneo. We describe the steps involved in developing this solution,

including collecting audio recordings, developing training and testing datasets, training neural

network models, and evaluating model performance. Our best model performed at a

satisfactory level (F score = 0.87), identifying 98% of the highest-quality calls from 90 hours of

manually-annotated audio recordings and greatly reduced analysis times when compared to a

human observer. We found no significant difference in the temporal distribution of great call

detections between the manual annotations and the model's output. Future work should seek

to apply our model to long-term acoustic datasets to understand spatiotemporal variations in

H. albibarbis' calling activity. Overall, we present a roadmap for applying deep learning to

identify the vocalisations of species of interest which can be adapted for monitoring other

endangered vocalising species.

I. INTRODUCTION

Ever-increasing anthropogenic pressures on the environment, such as habitat loss, have led to widespread population declines in many animal species (Bender et al. 1998). However, for many species, data on population trends are often sparse (Jetz et al. 2019), leading to an increased demand for wildlife population monitoring programs to inform conservation responses (Verma et al. 2016). To help achieve this, conservation scientists and ecologists have turned to developing technologies to automate data collection, enabling the rapid accumulation of large volumes of data (Piel and Wich 2021). While this has allowed for unprecedented insight, it can also make practical aspects of ecological inference challenging (Borowiec et al. 2022).

Manual extraction of relevant information from large datasets can be time consuming, resulting in a bottleneck at the point of analysis (Norouzzadeh et al. 2018). This bottleneck is evident in data generated as part of passive acoustic monitoring (PAM) programs, which involve the use of autonomous acoustic sensors to collect sound recordings in the field (Acevedo and Villanueva-Rivera 2010). Advancements in recording device design and cost, as well as improved data storage options have made the task of capturing many hours of acoustic data relatively straightforward (Morgan and Braasch 2021; Piel and Wich 2021). Data must then be browsed to identify relevant signals of interest, such as species-specific vocalisations, often by manually listening to each recording in full or visually inspecting the data in spectrogram form (a time-frequency pictorial representation of an audio signal), or both (van Kuijk et al. 2023; Clink et al. 2023). PAM can provide a step-change in standardized population monitoring of

vocal species at high temporal resolution and simultaneous large spatial scales, which would be

impossible to achieve with 'traditional' methods relying on manual data collection (Sugai et al.

2019). However, such PAM programs often capture datasets so large that they cannot be

studied manually in full in a reasonable timeframe, so automating this limiting data-processing

step is critical (Morgan and Braasch 2021; Clink et al. 2023).

Machine learning has proven to be an effective solution for fast and accurate analysis of

acoustic data, including the automated detection of signals of interest (Stowell 2022; Miller et

al. 2023). There are many options available for this task, including artificial neural networks

(ANNs) (Mielke and Zuberbühler 2013), Gaussian mixture models (GMMs) (Heinicke et al. 2015)

and support vector machines (SVMs) (Noda et al. 2016), among others (reviewed in Knight et al.

2017). These each have associated advantages, and due to the diversity of potential signal types

and acoustic environments, no single method is optimal in all situations (Clink et al. 2023).

However, it is worth noting that ANNs demonstrate comparatively strong adaptability and

proficiency in understanding complex patterns in data (Haykin 2009; Bengio et al. 2016). ANNs

can be arranged in various architectures, each suited for different tasks. Early work applying

ANNs to animal sound made use of the multi-layer perception (MLP) architecture, with

manually selected summary features such as syllable duration, peak frequency, etc., used to

inform the network's predictions (Stowell 2022). While MLPs have been effective in classifying a

wide variety of terrestrial and marine animal calls, the structure of non-speech acoustic events

can be highly variable (Kong et al. 2017), and reducing the data to a series of manually assigned

summary features can restrict the wealth of information available to train a network,

potentially limiting its effectiveness (Stowell 2022).

Deep neural networks, such as Convolutional Neural Network (CNN) architectures, rely

on feature sets that are not manually selected, but instead learned during the training process

(Morgan and Braasch 2021). CNNs are particularly effective for processing visual

representations of audio, such as spectrograms, leveraging their ability to learn patterns that

occur both spatially and temporally in data. This allows CNNs to learn local features regardless

of their spatial position within an image (Knight et al. 2017). CNNs are therefore ideal

candidates for the automated detection of signals within bioacoustic data, where instances of

relevant features within a spectrogram are not predefined or readily identifiable (Stowell 2022).

They have been used to analyse vocalisations from a variety of taxa, including insects (Hibino et

al. 2021), fish (Guyot et al. 2021), anurans (Colonna et al. 2016), birds (Narasimhan et al. 2017),

marine mammals (Miller et al. 2023), bats (Mac Aodha et al. 2018), and other terrestrial

mammals (Bjorck et al. 2019), including primates (Wood et al. 2023).

The potential of CNNs is far from fully realized however (Rammer and Seidl 2019), and

there are relatively few examples of CNNs being used to answer well-defined research

questions in ecology, as is so with other deep learning approaches (Dufourq et al. 2021).

Additionally, there are few guidelines on how to approach key steps such as model tuning and

performance assessments (Knight et al. 2017; Patterson and Gibson 2017; Stowell 2022).

Further case studies reporting successful applications will advance the development of best

practices for overcoming these challenges (Dufourq et al. 2021).

Gibbons (family Hylobatidae) are ideal candidates for the automated detection of

species-specific vocalisations. They engage in loud, highly stereotyped song bouts which are

largely confined to a few-hour window before and after sunrise (Cheyne et al. 2008). During a

particular calling bout, they usually emit multiple calls, which facilitates the generation of

abundant training data (Clink et al. 2023). Since the great call is performed largely by mated

females, it is often used as an indicator of a gibbon family group, allowing for group density and

spatial distribution estimates to be derived from great call densities, assuming that estimates of

female calling rates are available (Cheyne et al. 2016). Furthermore, gibbons reside exclusively

in tropical forests, which are often visually challenging and inaccessible so studying their

populations using visual methods, such as line transect and camera trap surveys, is typically

very difficult (Vu and Tran 2019). For these reasons, gibbons are model organisms for

developing and testing guidelines for automated detection. So far, this has been done for a

handful of species, including the Hainan gibbon (*Nomascus hainanus*) (Dufourq et al. 2021),

Western black-crested gibbon (*Nomascus concolor*) (Zhou et al. 2023), Northern grey gibbon

(*Hylobates funereus*) (Clink et al. 2023), and southern yellow-cheeked crested gibbon

(*Nomascus gabriellae*) (Clink et al. 2024).

Here, we apply a variation of a pre-defined CNN architecture, DenseNet (Huang et al.

2016), to identify female great call vocalisations of the endangered Bornean white-bearded

gibbon (*Hylobates albibarbis*) from a long-term acoustic dataset. We train a detector with high

precision that minimizes false-positive rates (i.e., the rate of detections incorrectly labelled as

gibbon great calls) for application to large acoustic datasets recorded by PAM arrays. This can

then be used to facilitate accurate population monitoring of wild gibbons on an ever-greater

spatiotemporal scale and applied as a case study for developing automated detectors for other

endangered vocalising species.

II.   METHODS

A. Data collection

The long-term acoustic dataset used in this study derives from the Mungku Baru

Education and Research Forest (MBERF), a ~50 km^2 area of tropical rainforest in Central

Kalimantan Province, Indonesia. The MBERF lies in the centre of the wider Rungan Landscape,

which covers approximately 1,500 km^2 between the Kahayan and Rungan rivers north of the

provincial capital of Palangka Raya. This represents the largest area of continuous unprotected

lowland rainforest remaining on the island of Borneo (Purnama and Afitah 2021). There is an

estimated population of 4,000 white bearded gibbons in the Rungan Landscape and an

estimated density of 2.79 groups per km^2 in the MBERF, making the region significantly

important for the conservation of the species (Buckley et al. 2018).

Eight autonomous recording units (ARUs) (Song Meter SM4, Wildlife Acoustics,

Maynard, Massachusetts) were deployed here by WME in July 2018 (see Figure 1). These were

placed on trees, 5 meters above the ground, in a dispersed grid with approximately 1,200

meters between each device. The MBERF contains a mosaic of different forest types, and the

array was designed to capture this heterogeneity, with three ARUs situated in 'kerangas'

(heath) forest, three in 'low pole' peat swamp forest, and two in 'mixed swamp' forest, with the

latter representing a transition habitat between the former two (Buckley et al. 2018).

FIGURE 1. Map of the Mungku Baru Education and Research Forest (Buckley et al. 2018),

showing the distribution of different habitat types over the survey area and the location of the

ARUs.

The ARUs were set to record from 4 am to 6 pm (local time) daily to capture the full pre-

dawn and diurnal period of ape calling. These were set to default settings (sensitivity of -35 ± 4

dB (0 dB = 1V/pa@1kHz), dynamic range of 14 to 100 dB SPL at 0 dB gain, microphone gain of

16 dB, and inbuilt preamplifier gain of 26 dB) and recorded on two channels with a sampling

rate of 24 kHz. Audio was captured in 16-bit Waveform Audio File Format (WAV) and saved as

one-hour files. Memory cards and batteries were changed every two weeks.

B. Manual annotation

Manually-annotated training and testing datasets were required to develop the

automated detector. To create these samples, recordings between 4-10 am were selected from

a single day every four weeks from a randomly selected device for each habitat. This covers the

temporal period in which most *H. albibarbis* great calls occur (Cheyne et al. 2008) and ensured

that a variety of potential sound environments (i.e., capturing spatial and temporal variation)

were included as training inputs to the model, improving its ability to generalize over a wider

range of applications. The resultant subset contained 300 hours of recordings, covering 50 days

spanning from October 2018 to December 2019.

The selected sound files were loaded into the sound analysis software Raven Pro 1.6 (K

Lisa Yang Center for Conservation Bioacoustics 2024) and visualized as spectrograms using a

3462-sample Hann window with a 90% overlap and a 4096-sample Discrete Fourier

Transformation. With assistance from a team of undergraduate interns (see

Acknowledgements), each recording was listened to in full and visually scanned to identify

instances of great calls. These were defined as vocalisations containing introductory, climax,

and descending phrases (see Figure 2). A selection was created for each instance by drawing a

box around the call in the spectrogram, providing information about its time-frequency

boundaries.

FIGURE 2. A spectrogram image of a female *H. albibarbis* great call, created using Raven Pro 1.6.

The box represents the time and frequency boundaries of a manually annotated selection. The

time boundaries span from the start of the first introductory (a) note to the end of the last

descending (c) note. The frequency boundaries span from the lowest frequency descending (c)

note to the highest frequency climax (b) note.

160   Each selection was then annotated based on its completeness and quality. A selection

was marked as "clear" when the entirety of the call could be heard and was visually clear in the

spectrogram, "faint" when the whole call could be heard but was not fully shown in the

spectrogram (and vice versa), and very faint when the call was only partially seen and heard

(i.e., part of the great call was not captured in the recording). Each selection was reviewed and

edited where needed by AFO to prevent inter-observer bias. In total, 1,611 great calls were

annotated.

167   The manually-annotated data were then randomly split, with 70% allocated for training

168   (210 hours, 1,089 calls) and 30% allocated for testing (90 hours, 522 calls). Due to the

169   ambiguous nature of "very faint" calls, they were removed from the training process and used

170   solely for testing thereafter. This prevented misleading information, i.e., non-target events,

171 being fed into the positive class weightings. Models were trained using "clear" and "faint"

172 instances (729 calls), as well as only "clear" instances (522 calls).

C. Automated detection

The development and testing of the automated detector utilized Koogu (version 0.7.2),

an open-source framework for deep learning from bioacoustic data (Madhusudhana 2023).

Koogu offers a variety of functions for deep learning, including 1) preparing audio for use as

inputs to machine learning models, 2) training models, 3) assessing model performance, and 4)

using trained models for the automated analysis of large datasets. For a full workflow

describing how the following steps were implemented within Koogu, see Supplementary

Material A.

1. Data preparation

All audio files were first down-sampled to 4,500 Hz to reduce the overall file size and

improve the efficiency of downstream computations (cf. Miller et al. 2023). The resulting

Nyquist frequency (2,250 Hz) is above the highest frequency within the great call selections

(2,077 Hz), and so no relevant information was lost in this process. The recordings were then

split into consecutive 28s segments (longer than the longest manually-annotated great call at

27.7s) with a hop size of 1s, leaving an overlap of 27s between clips. The waveform of each

segment was normalized by scaling the amplitudes to occur in the range -1.0 to 1.0.

The resulting start and end times of each segment were then compared to those from

manual annotations. Segments that fully contained the temporal extents of an annotated great

call were considered as positive inputs while segments with partial overlap were excluded from

training as these could resemble non-great call events and so lead to uncertainties during the

training process. Segments without temporal overlap were considered as negative inputs (i.e.,

background noise).

Spectrograms for both the positive and negative classes were then computed with an

analysis window of 0.192s and a 75% overlap. The bandwidth was also restricted to between

200 and 2200 Hz to exclude noise outside of the target frequency range. This resulted in input

spectrograms with a shape of 384 x 580 (height x width) pixels. To address the imbalance

between positive and negative classes, the maximum number of training inputs for each class

was reduced to 10,000. This utilized all the positive class inputs while randomly subsampling

inputs from the negative class. Following this, there were 5,763 "clear" and 1,253 "faint"

positive class spectrograms as well as 10,000 negative class spectrograms.

2. Data augmentation

Despite manually-annotated calls showing a high variance in background noise, call

duration and note length, for example, data augmentation was applied to further improve input

variance. To do this, several pre-defined augmentations supported by Koogu were applied both

on waveforms before conversion into spectrograms, and on the spectrograms themselves.

These were performed at each epoch, i.e., each time the model passed through the entire

training dataset, during the feeding of inputs into the model. This meant that the same original

sample could have different levels of augmentations between epochs.

Firstly, Gaussian noise (Schlüter and Grill 2015) was added to 25% of the training input's

waveforms at each epoch to simulate varying levels of background noise. The amount of noise

added randomly varied from -20dB to -30dB below the peak dB of the input signal.

The spectrogram was then smeared and squished along the time axis (cf.

Madhusudhana 2023). These augmentations were independently added to 20% of the input

spectrograms each epoch, at a magnitude of -1,1 (smearing backwards and forwards by one

frame of the spectrogram) and -2,2 (stretching and squishing over up to 2 frames of the

spectrogram). This process essentially blurred inputs along the time axis, as while the target

calls were contained within a standard frequency range, the duration of the signals was highly

variable.

Finally, Koogu's "AlterDistance" augmentation was applied to 25% of the spectrogram

inputs. This aimed to mimic the effect of increasing or reducing the distance between the

calling gibbon and the receiver, by attenuating or amplifying higher frequencies while keeping

lower frequencies relatively unchanged. This was applied by a random factor between -5dB

(attenuation) and 5dB (amplification).

3. Network parameters and training

The DenseNet architecture was chosen as the base CNN architecture for this study as it

has been shown to achieve comparatively high accuracy with fewer parameters, making it

efficient in terms of computational resources (Huang et al. 2016). Early variations of the model

suffered from overfitting, occurring when the model learns noise or random fluctuations in the

training data rather than the underlying pattern itself. This occurs when the model is too

complex relative to its intended task. Bearing this in mind, the standard DenseNet architecture

was adapted to a "quasi-DenseNet" architecture (Madhusudhana et al. 2021) which reduces

the number of connections within each dense block, limiting model size and complexity. To

limit the model's complexity further and improve computational efficiency, bottleneck layers

were also added (cf. Huang et al. 2016). Finally, batch normalization was enabled to improve

model convergence. For the final model architecture, see Figure 3.

FIGURE 3. Flowchart showing the final model architecture. The final model had a growth rate of

12, began with a 16-filter pre-convolutional layer, contained 3 quasi-dense blocks with 4, 6, and

6 layers respectively, and finished on a 32-node dense layer. Average pooling layers downscaled

the inputs by a factor of 3x3 (height x width) in the transition blocks between quasi-dense

blocks. Global Average pooling was used to reduce the spatial dimensions of outputs of the

final block to the 32-node feature vectors.

Training inputs were then divided further, with 15% randomly selected as a validation

set to evaluate the model's performance throughout the training process. Dropout layers were

added (Srivastava et al. 2014) at a rate of 5% to further reduce overfitting and improve

generalization. The models were then trained over 80 epochs using the Adam optimizer

(Kingma and Ba 2014) with a minibatch size of 24. The learning rate was initially set at 0.01 and

then reduced successively by a factor of 10 at epochs 20, and 40.

4. Testing

Trained models were then applied to the test dataset to provide a preliminary

assessment of model performance and establish a desirable detector threshold value. To do

this, each test segment was assigned a confidence score by the model between 0 (lowest) and 1

(highest) indicating how likely it was to contain a great call. Performance scores were outputted

for thresholds at an interval of 0.01, calculating the number of true positives (TP), false

positives (FP) and false negatives (FN). If there was a 100% overlap between a segment and an

annotated great call, and the confidence score was above the threshold it was marked as a TP.

If there was no- or partial overlap and the score was above the threshold it was marked as an

FP. If there was full overlap but the confidence score was below the threshold, then it would be

marked as a FN.

These quantities were used to compute recall, Eq. (1), precision, Eq. (2), and F score, Eq.

(3), at each threshold (where P = precision and R = recall).

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$$R = TP/(TP + FN) \tag{1}$$

265
$$P = TP/(TP + FP) \tag{2}$$

266
$$F = 2(P \times R)/(P + R) \tag{3}$$

The optimal threshold was then selected using maximum F score, as is a good indicator of

overall model performance (Clink et al. 2023).

Once a threshold had been decided, the model was re-run on the test dataset audio files

to produce detections and analyse the models' output. Neighbouring segments for which

scores were above the identified threshold were grouped to form a single detection. The score

of each detection was then set as the maximum of its component segment scores, and its start

and end times were set to the start time of the first segment and the end time of the last

segment respectively. These detections were then outputted in the format of Raven Pro

selection tables.

D. Post-processing

Initial inspection of the models' outputs showed that they produced clusters of

detections for each great call, overestimating the number of calls within the data. Detections

that overlapped with the first detection in each cluster were grouped together (see

Supplementary Material B). These groups were then filtered to retain only the highest-scoring

detection(s) within each group to minimize duplicate detections for single great call events.

Where the start times of the remaining selections were the same, only one detection was

retained. There was no further judication between remaining detections where their scores

were tied, as this could indicate that two great calls overlapped or occurred close in time.

Finally, the non-post-processed and post-processed model outputs of the best

performing model were evaluated against the manually-annotated test dataset. To do this, we

compared the distribution of total detections every 5 minutes across the 4-10 am period using a

Kolmogorov-Smirnov test. By comparing the non-post-processed and post-processed model

outputs, this also served to validate the effectiveness of the post-processing stage.

III.   RESULTS

A. Preliminary assessment

The preliminary assessment of the best performing model, trained on only "clear" calls,

showed precision ranging from 0.19 to 0.97 and recall ranging from 0.93 to 0.23 at thresholds

from 0.01 to 0.99 (Figure 4). This gave a maximum F score of 0.75 at a threshold of 0.36

(precision: 0.80, recall: 0.70).

As the aim was to optimize the model for "clear" and "faint" calls, the testing was re-run

excluding "very faint" calls. In this case, precision ranged from 0.12 to 0.94 and recall ranged

from 0.99 to 0.33 at thresholds from 0.01 to 0.99 (Figure 4). The maximum F score was

improved to 0.78 at a threshold of 0.78 (precision: 0.80, recall: 0.76). To maximize performance

for "clear" and "faint" calls while minimizing false positives, a threshold of 0.78 was therefore

chosen to evaluate the model's performance on the test dataset.

FIGURE 4. Precision-recall curves of the best performing model tested against all calls (left) and

both "clear" and "faint" calls (right).

B. Comparison with manual annotations

After re-running the model selected in the preliminary assessment on the test dataset at

the desired threshold and processing the output, it produced 535 detections. These were then

visually analysed in Raven Pro and compared to the manually-annotated dataset to discern the

occurrences of TPs, FPs, and FNs. In this case, a TP instance was defined as any model detection

that overlapped with a great call. The model was found to have produced 511 TPs and 24 FPs,

missing a further 133 calls (FNs). This gives a precision of 0.96, a recall of 0.79, and an F score of

0.87.

Upon closer inspection of the post-processed model output, 86 of the TP detections

were the result of repeated detections for singular great call events. Additionally, 22 FNs were

classed as TPs before the post-processing stage. These were created as a result of inadvertently

grouping detections where two great calls overlapped in time or were adjacent to one another.

Overall, the non-post-processed model identified 409 (78%) of the 522 manually-

annotated calls. This included 98% of all "clear" calls, 73% of "faint" calls, and 44% of "very

faint" calls. Out of the FN instances before post-processing, 77% were "very faint" calls with

only 0.05% representing 6 missed "clear" annotations. A further 38 great calls were identified

which had been missed in the manual annotation stage, including 6 "clear", 8 "faint" and 24

"very faint" calls.

The distribution of total detections every 5 minutes across the 4-10 am period for both

the non-post-processed and post-processed model outputs were compared against the

manually-annotated test dataset (Figure 5). Kolmogorov–Smirnov tests indicated no significant

difference (D = 0.036, p > 0.05 and D = 0.045, p > 0.05 respectively). Also, we found no

significant difference between pre-processed and post-processed data (D = 0.020, p > 0.05).

FIGURE 5. Histogram showing the number of great calls detected every 5 minutes between

04:00 and 10:00 for the manually annotated test dataset, as well as the output for the non-

post-processed model, and the post-processed model when ran on the test dataset.

IV.   DISCUSSION

Our best performing model was effective at detecting high-quality *H. albibarbis* great

calls with a low rate of false positives. The best performing model (F score: 0.87) exceeded

previously reported SVM models for detecting gibbon vocalisations e.g., *Hylobates funereus*

detector, F score: 0.78 (Clink et al. 2023) and was largely comparable to other CNN models for

gibbon great calls, e.g., *Nomascus hainanus* detector, F score: 0.91 (Dufourq et al. 2021),

Nomascus concolor DenseNet detector, F score: 0.92 (Zhou et al. 2023).

We found that the best performing model detected 38 great calls that had been missed

during the manual annotation stage, amounting to 6.8% of the total number of target events.

Human listening is subject to error (Brauer et al. 2016; Knight et al. 2017), and this study relied

on multiple human observers with differing levels of training to construct the manually-

annotated dataset. While it has been shown that human observers with less experience may

perform worse than some automated detectors (Jennings et al. 2008), the consensus from

multiple observers may have reduced the level of human error (Drake et al. 2016).

Furthermore, signals with low signal to noise ratio (SNR) are difficult for both humans and

machine-learning algorithms to detect (Knight et al. 2017). Precision and recall are often

measured relative to manually-annotated datasets, but these are not always perfect. With this

in mind, it has been recommended to view the manually-annotated dataset as the output from

an alternative detector rather than a ground-truth set (Knight et al. 2017).

Due to the apparent likelihood of great calls to be missed by manual annotation (Knight

et al. 2017), it is unrealistic to presume that all target calls were identified in the 210-hour

training dataset. This could pose a problem when training the model if any of the randomly

selected unannotated time periods for the negative class contained target events, potentially

increasing the rate of false negatives. Dufourq et al. (2021) found that better results were

obtained by specifically including negative class segments with typical ambient noise, such as

other species' vocalisations, which could potentially confuse the classifier. This method of 'hand

picking' the negative class could reduce the false negative rate and the false positive rate by

negatively labelling potentially confusing information. The low false-positive rate reported in

this study, as well as the low false-negative rate for "clear" calls, suggests that our method of

randomly selecting the negative class was sufficient for our aims. Where reducing the false-

negative rate has greater importance, such as when identifying infrequent vocalisations, a more

thorough approach could be preferable. Nonetheless, it is important to recognize the trade-off

between ensuring that the negative class contains as little erroneous information as possible

and the time required to construct an adequate training dataset.

In this study, we categorized selected calls according to their quality (cf. Cañas et al.

2023), although we acknowledge that the distinction of "clear", "faint" and "very faint" great-

call selections was not based on p measurements. These categories cannot be translated

directly to distance; however, in most cases, it is likely that "clear" calls were recorded from

gibbons singing closer to the ARUs. Overall, our best performing model identified 98% of all

"clear" calls, and only missed 6 from the manually-annotated dataset. This was comparable to

the human observers, who also missed 6 "clear" calls. The model performed worse than the

human detector at detecting "faint" and "very faint" calls, however, picking up 73% and 44% of

the manually-annotated instances, respectively. With this in mind, the results suggest that

detection likelihood is affected by the caller-ARU distance (Spillmann et al. 2015). This supports

the suggestion by Jahn et al. (2017) that the difference in recall between an automated

detector and a human listener is caused by the former having a smaller detection radius. Future

studies should aim to apply relationships between signal strength and the distance of the

source from the receiver to estimate call detection probability over distance. This will help to

determine the effective area being sampled by PAM studies.

Although the model did not have perfect recall, the importance of detecting all great

calls within a recording will depend on the research question. With regards to calling activity

over time, the distribution of calling frequency detected by the model was not significantly

different to that of the manual annotations. Therefore, our model can be used to reliably

estimate spatiotemporal variations in *H. albibarbis* calling activity. Through analysing recordings

from multiple habitats, our model could be applied to understand the relative importance of

forest subtypes for the species, as singing behaviour is density-dependent (Cheyne et al. 2008),

with less singing activity at lower group densities. This could operate on a continuous long-term

time frame which would be hard to achieve when relying on human observers in the field.

For an in-depth understanding of gibbon group abundance and density, further

information is necessary. One method is to use individuals as the sampling unit (Buckland 2006)

by analysing their call structure (Clink et al. 2023), or by localizing vocalisations using estimates

of direction to the source from multiple ARUs (Stevenson et al. 2015). This may prove highly

effective at estimating gibbon abundance and density over the short term yet could prove too

complex over the course of hundreds, or thousands of hours of audio. An alternative method is

to estimate vocalisation density per unit time, apply an estimation of vocalisation rate, and

then convert vocalisation density into group density (Marques et al. 2013). This does require a

knowledge of the area covered by each ARU, yet, for this task, Marques et al. (2013) notes that

automated detectors need not perform extraordinarily well so long as true positive and false

positive rates are characterized accurately. This method does not require for the effective area

of ARUs to overlap, so in theory, could monitor a larger area with the same resources.

While post-processing greatly reduced the number of repeated detections for single

great call events, these would still account for many false positives if only one true positive per

great call is allowed. The post-processing protocol did not seek to adjudicate between

detections if there was a tie in the highest scoring instances within a group, as in some cases

this represented two great call events close in time. In fact, 22 false negatives derived from

instances when the grouping of detections failed to take this into account, and limiting the

number of detections per group to one would have increased the false negative rate further. An

alternative approach would be an improved post-processing stage to better interpret the

output of the model. In audio classification, CNNs inspect audio recordings as image-like

segments and so are unable to use broader-scale contextual information, such as whether the

current point in a recording is preceded by a target call (Wang et al. 2022). It has therefore

been proposed to combine CNNs with other machine learning techniques, such as Hidden

Markov Models (HMMs) or deep learning techniques, such as Recurrent Neural Networks

(RNNs). Postprocessing the output of a CNN with a HMM or an RNN has been shown to improve

the F score (Madhusudhana et al. 2021; Wang et al. 2022). For instance, Wang et al. (2022)

showed that the application of a combined Convolutional Recurrent Neural Network (CRNN)

reduced error rates arising from overestimation of gibbon calls (49-54%) to 0.5%. It was noted,

however, that while post-processing the model output with a HMM performed second best, it

required much less computational power. Future work should therefore consider both of these

options when attempting to improve on our post-processing method.

A key aim of this study was to help address the analysis bottleneck evident in PAM data

by improving on the time taken to manually analyse recordings. During the batch processing

stage, it took 19 seconds for our trained model to process each hour of test recordings. This

greatly improved on a human processing speed of minimum 1 hour per hour of audio for this

study, varying depending on the level of observer experience. One caveat is the significant

amount of time required to construct a manually-annotated dataset to train and test a CNN

when compared to other machine learning approaches (Stowell 2022). Despite our study

showing how data augmentation can be effective where training data is limited, careful

consideration should be taken when under time-pressure if no training datasets are already

available. In some cases, it may be better to adopt an approach that requires less data to

develop, such as an SVM or a GMM (Clink et al. 2023).

Finally, Stowell (2022) noted that it was increasingly common to evaluate deep learning

models on test sets specifically designed to differ in some respects from the training data, such

as location, signal to noise ratio, or by season. In this case, the model was designed with

application to the MBERF bioacoustic dataset in mind, and so it is appropriate that the test data

came from the same location as the training data. Test inputs spanned across all times of the

year and were recorded from three different habitats. This ensured a variety of potential sound

environments were included in the testing stage. However, for applications in other locations,

especially outside the Rungan Forest Landscape, it would be advantageous to first test the

model on recordings captured elsewhere within *H. albibarbis'* range.

V.   CONCLUSION

Our study demonstrates how an open-source deep-learning framework can be adapted

to produce a CNN capable of detecting *H. albibarbis* great calls, performing at a comparable

level to similar CNN approaches for gibbon great calls. Our model performed best on the

highest-quality calls and yielded a low false-positive rate, meeting the objectives of this study.

There was a much lower likelihood of successful detection for the lowest-quality calls, however,

and future studies should aim to estimate call detection probability over distance to determine

the effective area being sampled.

Further development of the post-processing stage could help to reduce the number of

repeat detections for each call. However, the current output can be used to estimate calling

rate over time. Further work should seek to apply this model to long-term acoustic datasets

over a variety of habitats to study spatial and temporal variation in gibbon calling activity.

Furthermore, in combination with future studies on sound propagation of gibbon vocalisations,

this represents an opportunity to monitor *H. albibarbis'* populations on an ever-greater

spatiotemporal scale. Our work presents some key considerations to inform decision-making

for such projects, and a full workflow script to visualize how these can be implemented in

developing an automated detector.

SUPPLEMENTARY MATERIAL

For the full Koogu workflow, see Supplementary Material A (SuppPub1.txt). For the post-

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and MS carried out building and analysis of the model; AFO drafted the manuscript to which all

other authors then contributed to produce the final version.

AUTHOUR DECLARATIONS

The authors declare no competing interests.

DATA AVAILABILITY

Recordings of great calls for both the training and testing datasets are openly available on

Zenodo at https://doi.org/10.5281/zenodo.10926304.

ETHICS APPROVAL

Ethical approval was provided by the University of Exeter (application ID: 1845574), BRIN

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REFERENCES

Acevedo, M.A. and Villanueva-Rivera, L.J. 2006. From the Field: Using Automated Digital

Recording Systems as Effective Tools for the Monitoring of Birds and Amphibians. *Wildlife*

Society Bulletin 34(1), pp. 211–214. doi: https://doi.org/10.2193/0091-

7648(2006)34[211:UADRSA]2.0.CO;2.

Mac Aodha, O., Gibb, R., Barlow, K., Browning, E., Firman, M., Freeman, R., Harder, B., Kinsey,

L., Mead, G.R., Newson, S.E., Pandourski, I., Parsons, S., Russ, J., Szodoray-Paradi, A., Szodoray-

Paradi, F., Tilova, E., Girolami, M., Brostow, G. and Jones, K. 2018. Bat detective—Deep learning

tools for bat acoustic signal detection. *PLOS Computational Biology* 14(3), p. e1005995. doi:

10.1371/journal.pcbi.1005995.

Bender, D.J., Contreras, T.A. and Fahrig, L. 1998. Habitat Loss and Population Decline: A Meta-

Analysis of the Patch Size Effect. *Ecology* 79(2), p. 517. Available at:

http://www.jstor.org/stable/176950?origin=crossref.

Bjorck, J., Rappazzo, B.H., Chen, D., Bernstein, R., Wrege, P.H. and Gomes, C.P. 2019. Automatic

Detection and Compression for Passive Acoustic Monitoring of the African Forest Elephant.

Proceedings of the AAAI Conference on Artificial Intelligence 33(01), pp. 476–484. doi:

10.1609/aaai.v33i01.3301476.

Borowiec, M.L., Dikow, R.B., Frandsen, P.B., McKeeken, A., Valentini, G. and White, A.E. 2022.

Deep learning as a tool for ecology and evolution. *Methods in Ecology and Evolution* 13(8), pp.

1640–1660. doi: 10.1111/2041-210X.13901.

Brauer, C.L., Donovan, T.M., Mickey, R.M., Katz, J. and Mitchell, B.R. 2016. A comparison of

acoustic monitoring methods for common anurans of the northeastern United States. *Wildlife*

Society Bulletin 40(1), pp. 140–149. doi: 10.1002/wsb.619.

Buckland, S.T. 2006. Point-Transect Surveys for Songbirds: Robust Methodologies. *The Auk*

123(2), pp. 345–357. doi: 10.1093/auk/123.2.345.

Buckley, B.J.W, Capilla, B.P., Maimunah, S., Adul, Armadyanto, Boyd, N., Cheyne, S.M., Iwan,

Husson, S.J., Santiano, Salahudin, Ferisa, A., Namaskari, N., van Veen, F. and Harrison, M.E.

2018. Biodiversity, Forest Structure & Conservation Importance of the Mungku Baru Education

Forest, Rungan, Central Kalimantan, Indonesia. *BNF Reports*.

Cañas, J.S., Toro-Gómez, M.P., Sugai, L.S.M., Restrepo, H.D.B., Rudas, J., Bautista, B.P., Toledo,

L.F., Dena, S., Domingos, A.H.R., de Souza, F.L., Neckel-Oliveira, S., da Rosa, A., Carvalho-Rocha,

V., Bernardy, J.V., Sugai, J.L.M.M., dos Santos, C.E., Bastos, R.P., Llusia, D. and Ulloa, J.S. 2023. A

dataset for benchmarking Neotropical anuran calls identification in passive acoustic monitoring.

Scientific Data 10(1), p. 771. doi: 10.1038/s41597-023-02666-2.

Cheyne, S., Gilhooly, L., Hamard, M.C., Höing, A., Houlihan, P.R., Kursani, Loken, B., Phillips, A.,

Rayadin, Y., Ripoll Capilla, B., Rowland, D., Sastramidjaja, W.J., Stephanie, S., Thompson, C.J.H.

and Zrust, M. 2016. Population mapping of gibbons in Kalimantan, Indonesia: correlates of

gibbon density and vegetation across the species' range. *Endangered Species Research* 30, pp.

133–143. doi: 10.3354/esr00734.

Cheyne, S.M., Thompson, C.J.H., Phillips, A.C., Hill, R.M.C. and Limin, S.H. 2008. Density and

population estimate of gibbons (Hylobates albibarbis) in the Sabangau catchment, Central

Kalimantan, Indonesia. *Primates* 49(1), pp. 50–56. doi: 10.1007/s10329-007-0063-0.

Clink, D.J., Kim, J., Cross-Jaya, H., Ahmad, A.H., Hong, M., Sala, R., Birot, H., Agger, C., Vu, T.T.,

Thi, H.N. and Chi, T.N. 2024. Automated detection of gibbon calls from passive acoustic

monitoring data using convolutional neural networks in the "torch for R" ecosystem. *arXiv*

preprint arXiv:2407.09976.

Clink, D.J., Kier, I., Ahmad, A.H. and Klinck, H. 2023. A workflow for the automated detection

and classification of female gibbon calls from long-term acoustic recordings. *Frontiers in Ecology*

and Evolution 11. doi: 10.3389/fevo.2023.1071640.

Colonna, J., Peet, T., Ferreira, C.A., Jorge, A.M., Gomes, E.F. and Gama, J. 2016. Automatic

Classification of Anuran Sounds Using Convolutional Neural Networks. In: *Proceedings of the*

Ninth International C* Conference on Computer Science & Software Engineering - C3S2E '16.

New York, New York, USA: ACM Press, pp. 73–78. doi: 10.1145/2948992.2949016.

Drake, K.L., Frey, M., Hogan, D. and Hedley, R. 2016. Using digital recordings and sonogram

analysis to obtain counts of yellow rails. *Wildlife Society Bulletin* 40(2), pp. 346–354. doi:

10.1002/wsb.658.

Dufourq, E., Durbach, I., Hansford, J.P., Hoepfner, A., Ma, H., Bryant, J.V., Stender, C.S., Li, W.,

Liu, Z., Chen, Q., Zhou, Z. and Turvey, S.T. 2021. Automated detection of Hainan gibbon calls for

passive acoustic monitoring. *Remote Sensing in Ecology and Conservation* 7(3), pp. 475–487.

doi: 10.1002/rse2.201.

Goodfellow, I., Bengio, Y. and Courville, A. 2016. *Deep Learning*. Cambridge, Massachusetts: The

MIT Press.

Guyot, P., Alix, F., Guerin, T., Lambeaux, E. and Rotureau, A. 2021. Fish migration monitoring

from audio detection with CNNs. In: *Audio Mostly 2021*. New York, NY, USA: ACM, pp. 244–247.

doi: 10.1145/3478384.3478393.

Haykin, S.S. 2009. *Neural Networks: A Comprehensive Foundation*. 3rd ed. McMaster University,

Canada: Pearson Education.

Heinicke, S., Kalan, A.K., Wagner, O.J.J., Mundry, R., Lukashevich, H. and Kühl, H.S. 2015.

Assessing the performance of a semi-automated acoustic monitoring system for primates.

Methods in Ecology and Evolution 6(7), pp. 753–763. doi: 10.1111/2041-210X.12384.

Hibino, S., Suzuki, C. and Nishino, T. 2021. Classification of singing insect sounds with

convolutional neural network. *Acoustical Science and Technology* 42(6), p. E2152. doi:

10.1250/ast.42.354.

Huang, G., Liu, Z., van der Maaten, L. and Weinberger, K.Q. 2016. Densely Connected

Convolutional Networks.

Jahn, O., Ganchev, T.D., Marques, M.I. and Schuchmann, K.-L. 2017. Automated Sound

Recognition Provides Insights into the Behavioral Ecology of a Tropical Bird. *PLOS ONE* 12(1), p.

e0169041. doi: 10.1371/journal.pone.0169041.

Jennings, N., Parsons, S. and Pocock, M.J.O. 2008. Human vs. machine: identification of bat

species from their echolocation calls by humans and by artificial neural networks. *Canadian*

Journal of Zoology 86(5), pp. 371–377. doi: 10.1139/Z08-009.

Jetz, W., McGeoch, M.A., Guralnick, R., Ferrier, S., Beck, J., Costello, M.J., Fernandez, M., Geller,

G.N., Kiel, P., Merow, C., Meyer, C., Muller-Karger, F.E., Pereira, H.M., Regan, E., C., Schmeller,

D.S. and Turak, E. 2019. Essential biodiversity variables for mapping and monitoring species

populations. *Nature Ecology & Evolution* 3(4), pp. 539–551. doi: 10.1038/s41559-019-0826-1.

K Lisa Yang Center for Conservation Bioacoustics. 2024. Raven Pro: Interactive sound analysis

software. Ithaca, NY: The Cornell Lab of Ornithology. Available at:

http://ravensoundsoftware.com.

Kingma, D.P. and Ba, J. 2014. Adam: A Method for Stochastic Optimization.

Knight, E.C., Hannah, K.C., Foley, G.J., Scott, C.D., Brigham, R.M. and Bayne, E. 2017.

Recommendations for acoustic recognizer performance assessment with application to five

common automated signal recognition programs. *Avian Conservation and Ecology* 12(2), p.

art14. doi: 10.5751/ACE-01114-120214.

Kong, Q., Xu, Y. and Plumbley, M.D. 2017. Joint detection and classification convolutional neural

network on weakly labelled bird audio detection. In: *2017 25th European Signal Processing*

Conference (EUSIPCO). IEEE, pp. 1749–1753. doi: 10.23919/EUSIPCO.2017.8081509.

van Kuijk, S.M., O'Brien, S., Clink, D.J., Blake, J.G. and Di Fiore, A. 2023. Automated detection

and detection range of primate duets: a case study of the red titi monkey (Plecturocebus

discolor) using passive acoustic monitoring. *Frontiers in Ecology and Evolution* 11. doi:

10.3389/fevo.2023.1173722.

Madhusudhana. 2023. shyamblast/Koogu: v0.7.2.

Madhusudhana, S., Shiu, Y., Klinck, H., Fleishman, E., Liu, X., Nosal, E., Helble, T., Cholewiak, D.,

Gillespie, D., Širović, A. and Roch, M.A. 2021. Improve automatic detection of animal call

sequences with temporal context. *Journal of The Royal Society Interface* 18(180), p. 20210297.

doi: 10.1098/rsif.2021.0297.

Marques, T.A., Thomas, L., Martin, S.W., Mellinger, D.K., Ward, J.A., Moretti, D.J., Harris, D. and

Tyack, P.L. 2013. Estimating animal population density using passive acoustics. *Biological*

Reviews 88(2), pp. 287–309. doi: 10.1111/brv.12001.

Mielke, A. and Zuberbühler, K. 2013. A method for automated individual, species and call type

recognition in free-ranging animals. *Animal Behaviour* 86(2), pp. 475–482. doi:

10.1016/j.anbehav.2013.04.017.

Miller, B.S., Madhusudhana, S., Aulich, M.G. and Kelly, N. 2023. Deep learning algorithm

outperforms experienced human observer at detection of blue whale D-calls: a double-observer

analysis. *Remote Sensing in Ecology and Conservation* 9(1), pp. 104–116. doi: 10.1002/rse2.297.

Morgan, M.M. and Braasch, J. 2021. Long-term deep learning-facilitated environmental

acoustic monitoring in the Capital Region of New York State. *Ecological Informatics* 61, p.

101242. doi: 10.1016/j.ecoinf.2021.101242.

Narasimhan, R., Fern, X.Z. and Raich, R. 2017. Simultaneous segmentation and classification of

bird song using CNN. In: *2017 IEEE International Conference on Acoustics, Speech and Signal*

Processing (ICASSP). IEEE, pp. 146–150. doi: 10.1109/ICASSP.2017.7952135.

Noda, J.J., Travieso, C.M., Sanchez-Rodriguez, D., Dutta, M.K. and Singh, A. 2016. Using

bioacoustic signals and Support Vector Machine for automatic classification of insects. In: *2016*

3rd International Conference on Signal Processing and Integrated Networks (SPIN). IEEE, pp.

656–659. doi: 10.1109/SPIN.2016.7566778.

Norouzzadeh, M.S., Nguyen, A., Kosmala, M., Swanson, A., Palmer, M.S., Packer, C. and Clune, J.

2018. Automatically identifying, counting, and describing wild animals in camera-trap images

with deep learning. *Proceedings of the National Academy of Sciences* 115(25). doi:

10.1073/pnas.1719367115.

Patterson, J. and Gibson, A. 2017. *Deep Learning: A practitioners approach*. Loukides, M. and

McGovern, T. eds. O'Reilly Media.

Piel, A.K. and Wich, S.A. 2021. *Conservation Technology*. Wich, S. A. and Piel, A. K. eds. Oxford

University PressOxford. doi: 10.1093/oso/9780198850243.001.0001.

Purnama, A. and Afitah, I. 2021. Motivasi Masyarakat Terhadap Pengelolaan Khdtk Mungku

Baru, Palangka Raya. *Anterior Jurnal* 20(2), pp. 43–49. doi: 10.33084/anterior.v20i2.2162.

Rammer, W. and Seidl, R. 2019. Harnessing Deep Learning in Ecology: An Example Predicting

Bark Beetle Outbreaks. *Frontiers in Plant Science* 10. doi: 10.3389/fpls.2019.01327.

Schlüter, J. and Grill, T. 2015. Exploring Data Augmentation for Improved Singing Voice

Detection with Neural Networks. In: *16th International Society for Music Information Retrieval*

Conference.

Spillmann, B., van Noordwijk, M.A., Willems, E.P., Mitra Setia, T., Wipfli, U. and van Schaik, C.P.

2015. Validation of an acoustic location system to monitor Bornean orangutan (*Pongo*

pygmaeus wurmbii) long calls. *American Journal of Primatology* 77(7), pp. 767–776. doi:

10.1002/ajp.22398.

Srivastava, N., Hinton, G., Krizhevsky, A., Sutskever, I. and Salakhutdinov, R. 2014. Dropout: A

Simple Way to Prevent Neural Networks from Overfitting. *Journal of Machine Learning*

Research 15, pp. 1929–1958.

Stevenson, B.C., Borchers, D.L., Altwegg, R., Swift, R.J., Gillespie, D.M. and Measey, G.J. 2015. A

general framework for animal density estimation from acoustic detections across a fixed

microphone array. *Methods in Ecology and Evolution* 6(1), pp. 38–48. doi: 10.1111/2041-

210X.12291.

Stowell, D. 2022. Computational bioacoustics with deep learning: a review and roadmap. *PeerJ*

10, p. e13152. doi: 10.7717/peerj.13152.

Sugai, L.S.M., Silva, T.S.F., Ribeiro, J.W. and Llusia, D. 2019. Terrestrial Passive Acoustic

Monitoring: Review and Perspectives. *BioScience* 69(1), pp. 15–25. doi: 10.1093/biosci/biy147.

Verma, A., van der Wal, R. and Fischer, A. 2016. Imagining wildlife: New technologies and

animal censuses, maps and museums. *Geoforum* 75, pp. 75–86. doi:

10.1016/j.geoforum.2016.07.002.

Vu, T.T. and Tran, L.M. 2019. An Application of Autonomous Recorders for Gibbon Monitoring.

International Journal of Primatology 40(2), pp. 169–186. doi: 10.1007/s10764-018-0073-3.

Wang, Y., Ye, J. and Borchers, D.L. 2022. Automated call detection for acoustic surveys with

structured calls of varying length. *Methods in Ecology and Evolution* 13(7), pp. 1552–1567. doi:

10.1111/2041-210X.13873.

Wood, C.M., Barceinas Cruz, A. and Kahl, S. 2023. Pairing a user-friendly machine-learning

animal sound detector with passive acoustic surveys for occupancy modeling of an endangered

primate. *American Journal of Primatology* 85(8). doi: 10.1002/ajp.23507.

Zhou, X., Hu, K., Guan, Z., Yu, C., Wang, S., Fan, M., Sun, Y., Cao, Y., Wang, Y. and Miao, G. 2023.

Methods for processing and analyzing passive acoustic monitoring data: An example of song

recognition in western black-crested gibbons. *Ecological Indicators* 155, p. 110908. doi:

10.1016/j.ecolind.2023.110908.