1   Revealing the evolutionary history and contemporary population structure of Pacific salmon in the
2   Fraser River through genome resequencing
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Abstract

The Fraser River once supported massive salmon returns, but now years with half of the recorded historical maximum are considered good. There is substantial interest from surrounding communities, governments, and other groups to increase salmon returns for both human use and for functional ecosystems. To help generate resources for this endeavour, we resequenced hundreds of genomes at moderate coverage (~16x) of Chinook (*Oncorhynchus tshawytscha*), coho (*O. kisutch*), and sockeye salmon (*O. nerka*) from the Fraser River. The resequenced genomes are an important resource that can give us new insights. In this study, we found evidence that Chinook salmon have 1.5-2x more polymorphic loci than coho or sockeye salmon. Using principal component analysis (PCA) and admixture analysis, we also identified genetic groups similar to those previously identified with only a few microsatellite markers. As the higher density data supports these previous genetic groups, it suggests that the identity of these groups is not overly sensitive to the number of genetic markers or when the groups were sampled. With the increased resolution from resequenced genomes, we were able to further identify factors influencing these genetic groups, including isolation-by-distance, migration barriers, recolonization from different glacial refugia, and environmental factors like precipitation. We were also able to identify 20 potentially adaptive loci among the genetic groups by analyzing runs of homozygosity. All of the resequenced genomes have been submitted to a public database where they can be used as a reference for the contemporary genomics of Fraser River salmon.

Article Summary

Concerns over Fraser River salmon declines led us to generate resources to better understand the genetics of these salmon. Our findings were similar to previous studies that examined a very small fraction of the genetic markers examined here. We expanded upon previous studies by identifying possible influences on the genetics of Fraser River salmon. These included environmental factors, historical differences, and a lack of connectivity in the river caused by the Fraser Canyon.

Introduction

The Fraser River in Canada may have existed in some form for over 66 million years (Tribe 2005). The modern flow of the river is much more recent as there is evidence of a major change in flow around 760,000 years ago (Andrews *et al.* 2012). During the last glaciation period (the last in the Pleistocene), the entire region where the modern river is located may have been covered by the Cordilleran Ice Sheet, which reached its estimated maximum around 19,000 years ago (Clark *et al.* 2009). The retreat of the Cordilleran Ice Sheet began as early as 18,000 years ago and there were ice-free regions in British Columbia (where the Fraser River is located) as early as 17,700 years ago (Warner *et al.* 1982; Darvill *et al.* 2018). Some regions of British Columbia may have even been ice-free during the glacial maximum (e.g., (Byun *et al.* 1999; McPhail 2007; Stewart *et al.* 2009)). The Cordilleran Ice Sheet fully retreated by around 11,000 years ago, and its only remnants are the glaciers that still exist in modern times (Clague 2017). These massive changes in the Fraser River environment likely caused local extinctions of plants and animals from the river.

Based on fossil evidence, there were salmonids in the Fraser River before it was covered by glaciers (Harington 1996). From paleogeographic evidence mentioned above and this fossil evidence, we can infer that modern salmon in the river must have re-colonized after glaciers receded. There is

similar evidence of colonization of other species to this region, such as different plants, deer, and other
ray-finned fishes (Soltis *et al.* 1997; Bernatchez and Wilson 1998; Hewitt 2000; McPhail 2007; Latch
et al. 2009; Beatty and Provan 2010). Colonization or re-colonization of the 65-66 species of ray-
finned fishes native to British Columbia was thought to have started around 13,000 years ago (McPhail
2007). However, fossil evidence indicates that kokanee (*O. nerka*) could have re-colonized the interior
of British Columbia, possibly via the Columbia River, as early as 18,000 years ago (Harington 1996).
This is consistent with a large isolated and unique kokanee population in the upper Columbia River and
Fraser River system that remains today (Beacham and Withler 2017; Christensen *et al.* 2020).

      While there is a large body of paleogeographic research and a fossil record describing glaciation
and species colonization of the Fraser River after glaciers retreated, we often do not have information
on specific locations or populations. We can turn to genetic studies to complement what we understand
from these other two fields of study. We expect to observe genetic legacies of re-colonization events in
modern populations (Hewitt 1996, 2000). These legacies may influence modern populations, and
understanding them will be important for evidence-based conservation and management. Several
researchers have suggested that Fraser River salmon have experienced recent genetic bottlenecks that
could be the outcome of re-colonization events (Wehrhahn and Powell 1987; Wood *et al.* 1994;
Rondeau *et al.* 2023).

      Generally, there were two to three Fraser River (e.g., lower, middle, and upper) Chinook (*O.
tshawytscha*), sockeye (*O. nerka*), or coho salmon (*O. kisutch*) genetic groups identified from previous
studies (Wood *et al.* 1994; Small *et al.* 1998; Teel *et al.* 2000; Beacham *et al.* 2003; Beacham and
Withler 2017). This excludes additional groups identified from the Thompson River tributary that were
also commonly identified (Small *et al.* 1998; Withler *et al.* 2000; Beacham *et al.* 2003; Beacham and
Withler 2017). While these different genetic groups could have originated from different re-
colonization events or from different combinations of re-colonization events, they could also have
originated as a result of other factors. These factors include: limited gene flow between the groups
(e.g., due to barriers like the Fraser Canyon, a phenomenon sometimes referred to as isolation-by-
resistance (McRae 2006)), because of stochastic mechanisms like isolation-by-distance, adaptation to
environmental conditions (sometimes referred to as isolation-by-environment (Weber *et al.* 2017)),
ecotype differentiation in sockeye salmon (Beacham and Withler 2017), or through a combination of
these mechanisms. If these genetic groups were a result of different re-colonization events, it is also
possible that the re-colonizing populations came from different glacial refugia (Wood *et al.* 1994; Small
et al. 1998; Teel *et al.* 2000; Withler *et al.* 2000), which would increase their overall distinctiveness.

      Wehrhahn and Powell (1987) note that the Fraser Canyon might limit gene flow between lower
and upper Fraser River salmon. There are several identified velocity barriers in the Fraser Canyon that
could influence salmon migration and particularly limit migration of smaller salmon (Wright 2022).
This was dramatically emphasized between 1913-1914 when construction left a massive rockslide in
the Fraser Canyon that partially blocked some salmon species, and completely blocked pink salmon
(*O. gorbuscha*) (Kew 1992; Grant and Pestal 2009; Pess *et al.* 2012). This single event reduced salmon
returns to around a quarter of historic returns to the Fraser River during peak years (Pacific Salmon
Commission – psc.org). If the Fraser Canyon acts as a barrier, we might expect unique genetic
characteristics above and below the Fraser Canyon for all the species that must traverse it.

      Upper and lower Fraser River salmon are also separated by large distances and because they
have natal homing with variable straying (reviewed in (Quinn 1993; Keefer and Caudill 2013; Bett *et
al.* 2017)), we expect population structure to be influenced by isolation-by-distance (e.g., (Wright 1943;
Weber *et al.* 2017; Aguillon *et al.* 2017)). In general, we would expect isolation-by-distance to increase
the longer temporally that salmon are isolated from each other and with increased distance. We also
expect to observe nearby groups to have higher genetic similarity than those further away for all

species. An exception to this expectation would be salmon with different life history types (e.g., odd
and even-year pink salmon (Christensen *et al.* 2021)).
The expectation of increased genetic distance with increased geographic distance will not be the
main driver of differentiation if genetic groups were largely influenced by separate re-colonization
events. If re-colonization occurred by different genetic groups (e.g., populations from different glacial
refugia), we might expect genetic distance to be unrelated to geographic distance in the absence of
strong gene-flow or long periods of time. Patterns of re-colonization from different genetic groups
would also be expected to be variable among species.
Adaptation to the environment is of particular interest for the conservation and management of
a species because salmon from an area could be uniquely suited to that region. The headwaters of the
Fraser River flow north down the Rocky Mountain Trench, which leads to the Interior Plateau,
followed by the Fraser Canyon, and finally the river outlets through a major delta in the metropolis of
Vancouver, British Columbia (Reynoldson *et al.* 2005). These different regions have variable climates,
elevations, and river velocities (Reynoldson *et al.* 2005). They are also separated by great distances.
Genetic signatures correlated with environmental factors could be from neutral factors such as
isolation-by-distance or from non-neutral factors such as local adaptation or other forms of selection.
Researchers have not yet found a reliable method for distinguishing among the different mechanisms
with genetic data alone (reviewed in (Bierne *et al.* 2013; Lotterhos and Whitlock 2015; Ahrens *et al.*
2018; Saravanan *et al.* 2020)), and other experiments need to be used to firmly establish particular
genetic adaptations.
In this study, we analyze 954 resequenced genomes from three species of salmon (317 sockeye,
360 Chinook, and 277 coho salmon), mainly from the Fraser River, to understand differences among
the species, identify genetic groups among the collections of samples within each species, and to
characterize genetic adaptation to different environments. Data related to genetic structure and local
adaptation are particularly important to conservation and management. By using data from multiple
species, we compare among species to identify common patterns. This work provides a foundational
dataset that will be openly available and of significant use to the salmonid research community for
years to come.
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Materials and Methods

Sampling and whole genome resequencing

Chinook, coho, and sockeye salmon were sampled from Fraser River tributaries and lakes or
were from other bodies of water from previous studies (Figure 1, Figure S1, File S1, (Christensen *et al.*
2020; Rondeau *et al.* 2023)). Some of these sampling locations include hatchery sources (File S1).
Most modern hatchery stocks originate from local sources (e.g. (Heard 2012)), and they should reflect
the current Fraser River population if not the historical population. The samples were collected during
various years (File S1). For coho salmon, there were 125 salmon collected at the Big Bar Landslide
between 2019-2021. Spawning locations were unknown for these samples. These were included to
identify potentially unknown genetic structure since it is difficult to reach upper Fraser River sites
during coho salmon spawning.
Samples were collected by Fisheries and Oceans Canada personnel in compliance with the
Canadian Council on Animal Care Guidelines, and under the authority of the Fisheries and Oceans
Canada Pacific Region Animal Care Committee (Ex.7.1). Samples were taken either as operculum-
clips or as scales. Both were desiccated and stored on Whatman paper.

Genomic DNA was extracted from tissue, after an overnight incubation in 95% ethanol, using a
Quick-DNA Kit (Zymo Research). Other genomic DNA samples were previously extracted by
Fisheries and Oceans Canada using automated BioSprint extractions, as per manufacturer's instructions
(Qiagen). DNA samples were then sent to the Michael Smith Genome Science Centre (Vancouver, BC)
for library preparation and whole genome resequencing.
Whole genome sequencing libraries were prepared by shearing the DNA samples individually
using a Covaris LE220 (duty cycle: 20%, PIP: 450, cycles per burst: 200, time per run: 90 s; with pulse
spin after 45 s). Individual libraries were then constructed using the MGIEasy PCR-Free DNA Library
Prep Set (MGI Tech Co.). The indexed libraries were then pooled and sequenced on a MGISeq-G400
sequencer (paired-end 200 bp). Data used in this study were also taken from previous studies
(Christensen *et al.* 2020; Rondeau *et al.* 2023) that used Illumina sequencing technology.
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SNP calling and filtering
Reads from each individual were aligned to their respective species reference genome assembly
(Chinook salmon: GCF_018296145.1 (Christensen *et al.* 2018), coho salmon: GCF_002021735.2
(Rondeau *et al.* 2023), sockeye salmon: unreleased version 2 (Christensen *et al.* 2020)) with BWA
(version 0.7.17, parameter -M) (Li and Durbin 2009, 2010; Li 2013). Reads were sorted with
SAMtools (version 1.12, default parameters) (Danecek *et al.* 2021). Picard (version 2.26.3, default
parameters) (Broad Institute 2019) was used to add read group information and mark reads that were
suspected PCR duplicates. GATK (version 3.8) (McKenna *et al.* 2010; Van der Auwera *et al.* 2013)
was then used to call nucleotide variants for each individual (parameters: -T HaplotypeCaller, --
genotyping_mode DISCOVERY, --emitRefConfidence GVCF) and then combined (parameters: -T
GenotypeGVCFs, --max_alternate_alleles 3). Truth and training SNP datasets (File S2) were then used
to recalibrate nucleotide variant scores and filter variants using GATK (parameters -T
ApplyRecalibration, --mode SNP –ts_filter_level 99.5). The truth SNPs came from multiple studies
(Brieuc *et al.* 2014; Meek *et al.* 2016; Nichols *et al.* 2016; Larson *et al.* 2017; Veale and Russello 2017;
Rondeau *et al.* 2023), and the training SNPs are described in File S2.
Additional filters were used to remove indels, SNPs with more or fewer than two alleles, or that
were missing genotypes in more than 10% of the individuals. SNPs were removed if their mean depths
were outside a range of 8-100x, and if they had less than 0.01 minor allele frequency (MAF). All
filtering was performed using VCFtools (version 0.1.15) (Danecek *et al.* 2011). The MAF would
eliminate alleles found in fewer than 6-8 heterozygous individuals (or 3-4 homozygous individuals),
depending on the species and its respective sample size. This threshold was chosen to reduce
sequencing errors, but to still keep all but the rarest variants. The MAF filter was not used for the
SMC++ analysis. Linkage disequilibrium was evaluated and used to filter SNPs in some analyses, as
noted below. The prune add-on for BCFtools (version 1.9) (Danecek *et al.* 2021) was used to filter
based on linkage disequilibrium (parameters: +prune, -w 20kb, -l 0.4, -n 2). The filters used for all
analyses are shown in Table 1.
195

Mapping variants among species
To compare analyses among species, variants were mapped from the coho and sockeye salmon
genomes to the Chinook salmon genome using a pipeline, *MapVCF2NewGenome* (see *Data*
Availability). This pipeline was also used to map sockeye salmon variants, which were on scaffolds, to
a chromosome level assembly (this assembly was submitted to the NCBI as version 2 of the reference
genome assembly – now available as GCA_034236695.1). The sockeye salmon chromosome mapped
version was needed for several analyses.

203
Coverage, relatedness metrics, and runs of homozygosity (ROH)

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Coverage was assessed using a python script, *VCFStats* (see *Data Availability*). This script
finds the average depth of all SNPs per individual. It also counts the number and type of genotypes per
individual. This information was used to remove individuals, using VCFtools, that had average
depth/coverage less than 8x, as a preliminary PCA analysis showed clustering of individuals based on
coverage rather than geography if they were below this threshold. These individuals are not discussed
and were not analyzed further. Individuals were filtered with an average depth less than 15x depending
on the analysis (noted for each analysis). This was because the frequency of heterozygous genotypes
dropped when coverage was below 15x and some analyses were sensitive to this issue (Table 1).
Genotype information and coverage was visualized in R (R Core Team 2022) using ggplot2 (Wickham
2016).
The related package (version 0.8, parameters of coancestry: ritland = 1) (Pew *et al.* 2015) was
used to calculate relatedness (Ritland 1996) in R. Values below zero (e.g., when markers are not
shared) and above one (e.g., when rare markers are shared) are expected with this type of estimation
(Ritland 1996). There were many more private alleles in coastal populations of Chinook and coho
salmon, which may have made them appear much more related. The linkage disequilibrium filtered
SNP datasets were converted to the related format using a python script, *VCF2Relate* (see *Data
Availability*). The reshape2 (Wickham 2007) and pheatmap (Kolde 2019) packages were used to
visualize relatedness. Related individuals were included in most analyses, but we tested what would
happen if related individuals were removed from PCA analyses. A relatedness score of 0.15 (an
arbitrary value greater than what might be expected for first cousins, but less than for half-sibs –
depending on dataset) was used to filter all but one individual from pairs that might be related. Most
samples were taken as adults and related individuals were not expected to be common within or among
sampling collections.
Runs of homozygosity were identified using PLINK (version 1.9, parameters: --homozyg –
double-id, --allow-extra-chr, --homozyg-snp 25, --homozyg-kb, --homozyg-density, --homozyg-gap, --
homozyg-window-het 1, --homozyg-window-snp 50, --homozyg-window-threshold 0.05, --homozyg-
window-missing 5) (Chang *et al.* 2015; "PLINK 1.9"). This was to compare species levels of runs of
homozygosity. All individuals were included in these analyses. No consistent relationship was
observed between coverage and runs of homozygosity (e.g., runs of homozygosity might be expected
to increase as coverage decreases since there are fewer heterozygous genotypes as noted above). We
plotted each individual when comparing runs of homozygosity among species to determine if coverage
or relatedness might influence species comparisons.

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Population structure and clustering analyses

PCA and admixture analyses were used to identify distinct genetic groups for each species. The
admixture software requires independent variants, and PCAs can be sensitive to LD blocks, which can
cause clustering that is unrelated to population structure. For this reason, the LD filtered SNPs were
used in these analyses. To perform the admixture analyses, the LD filtered data was converted from
VCF format to a suitable format using PLINK (version 1.9, parameters: --double-id, --allow-extra-chr).
We converted the chromosome names to numbers using Unix commands. The admixture analysis was
performed using ADMIXTURE (version 1.3.0, parameters: --cv) (Alexander *et al.* 2009). Cluster
values 1-20 were tested and we accepted the value with the lowest cross-validation score. Admixture
ancestry values were visualized in R for individuals and in QGIS (QGIS Development Team 2022) as
an average score per sampling site.

PCA were used to assess the groupings produced from the admixture analyses. PLINK
(parameters: --pca, --double-id, --allow-extra-chr) was used to perform the PCA, and they were
visualized in R using ggplot2, reshape2, and ggrepel (Slowikowski 2021). PCA were tried with all the
individuals, individuals with ≥15x coverage, and with individuals that were highly related removed.
The outputs of these three approaches were compared to determine the influence of coverage and
relatedness on clustering and the groups produced from the admixture analyses. We also verified that
samples among the new and previous datasets clustered together if they were from the same sampling
site.
257

Environmental variable PCA

To determine whether population structure was associated with environmental factors, we
clustered sampling sites by environmental factors using PCA in R (parameters: prcomp, scale=T).
Each sampling site was assigned to an admixture group if the average ancestry value was ≥ 0.7 for a
particular site. Environmental factors were taken from the WorldClim version 2.1 dataset (Fick and
Hijmans 2017). Elevation for each site was either estimated from Google maps, mapcarta, or from data
downloaded from the Federal Geospatial Platform (maps.canada.ca) and viewed in QGIS. Distance to
the ocean was estimated as the river distance with QGIS or Google maps.
266

Historical estimates of effective population size

To estimate effective population size through time, we used the program SMC++ (parameters: -
c 1000000) (Terhorst *et al.* 2017). The only mutation rate estimate for these species available at the
time of writing was for coho salmon (8e-9 from (Rougemont *et al.* 2020)). A correction was applied to
this mutation rate for Chinook and sockeye salmon based on the ratio of total SNPs from these species
to the total SNPs from coho salmon. For Chinook salmon, this ratio was 1.7274 for the individuals
subset from a similar geographic range and for the same number of individuals (mutation rate of
1.382e-8). For sockeye salmon, this was 0.79785342 for the individuals subset, with a mutation rate of
6.38283e-9. Figures were plotted in R using ggplot2, and ocean surface temperature data was taken
from previous studies (Zachos *et al.* 2008; Hansen *et al.* 2013).
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Genetic diversity

The percent of heterozygous genotypes per individual was calculated by dividing the number of
heterozygous genotypes (see *Coverage, relatedness metrics, and runs of homozygosity* section) by all
genotypes (including missing data to standardize against all variants) and multiplying by 100. If
missing genotypes were excluded from the calculation, values changed by 0.09% on average, and the
difference ranged from 0.03%-1.4%. Nucleotide diversity (Pi) and the number of polymorphic loci
was calculated using the Stacks populations module (version 2.54, default parameters) (Catchen *et al.*
2011, 2013).
The percent of heterozygous genotypes per individual and nucleotide diversity per sampling site
were both calculated from the subset of individuals with ≥15x coverage because these analyses were
sensitive to coverage (i.e., as coverage increased, so did the percent of heterozygous genotypes until
≥15x coverage). The number of sampling sites were included in the reported percent of heterozygous
genotypes per individual to show if the reported value could be sensitive to relatedness (e.g., fewer
locations could be more sensitive to highly related individuals from a single location). Since the
nucleotide diversity was plotted per sampling site using the inverse distance weighting interpolation
analysis in QGIS, other nearby sites can also be used to determine if relatedness influenced nucleotide
diversity regionally.

Admixture group private alleles

A python script was used to identify private alleles among admixture groups of each species (*PrivateAllele* see *Data Availability*). Alleles were identified that were unique to each group if they were present for a specified number of individuals in that group (parameter: -min 3). Individuals were assigned to an admixture group if they had ancestry values ≥ 0.7 for a particular group. Salmon with ancestry values below 0.7 were excluded from these analyses. The private allele counts were visualized in R using ggplot and the gridExtra library (Auguie 2017).

The number of private alleles per individual among admixture groups was identified using another python script, *PrivateAllelePerInd* (see *Data Availability*). This metric identifies if there are individuals or sampling sites that were more responsible for the number of private alleles within a genetic group. The minimum number of individuals with the private allele was set to one. While decreasing this parameter to one could increase the chance that SNP calling errors are included, it also allows us to get an idea for the number of private alleles from a group that are within the genome of each individual. We expect sequencing errors to be evenly distributed among sampling locations and similar among individuals.

Potentially adaptive variants

The rehh library (Gautier and Vitalis 2012; Gautier *et al.* 2017) in R was used to identify extended haplotype homozygosity within and among admixture groups (assignments were made for individuals with ≥ 0.7 ancestry values). To perform the rehh analyses, we used SHAPEIT5 (Hofmeister *et al.* 2023) (version 5.1.1, parameter: phase_common_static) to first phase the genotypes of the MAF 0.01 filtered variants. The output was then converted to VCF format using BCFtools, and the different admixture groups were separated using VCFtools. The rehh command *data2haplohh* was then used to read in each admixture group VCF file, and the *scan_hh* command (parameters, polarized=FALSE, interpolate=FALSE) was used to calculate extended haplotype homozygosity. The within population metric iHS was calculated using the *ihh2ihs* command (default parameters), and the pairwise population metric Rsb was calculated using the *ines2rsb* command (default parameters). The *calc_candidate_regions* function (parameters: threshold=10, window_size=10000, min_n_extr_mrk=8) was used to identify regions with significant differences in extended haplotype homozygosity (the Bonferroni correction p-value threshold was between 8.8-9.13 after -log10 transformation depending on species). A higher threshold was used to identify only the strongest candidates as this is only a preliminary study. We also examined overlapping 10 kbp windows among all three species to identify candidates of potential convergent evolution.

Results

Coverage, relatedness metrics, and runs of homozygosity (ROH)

We identified ~6-13 million SNPs per species (Figure 2a). Chinook salmon had 1.58-2.07x more SNPs than either coho or sockeye salmon, respectively. Between 1-1.75% of all SNP loci were common among species, and only up to 0.1% were found to be polymorphic in all species (Figure 2b). Chinook salmon had a reduction in the total length of runs of homozygosity (ROH), consistent with a higher number of polymorphic loci (Figure 2c). Sockeye and coho salmon had on average ~3.5-3.6x longer total lengths of ROH than Chinook salmon. The length of ROH decreased as the fraction of heterozygous genotypes increased, but generally sockeye and coho salmon had longer ROH when

fractions of heterozygous genotypes were below 0.2 (Figure 2d). No consistent relationship among
species was observed between SNP coverage and the total length of ROH (Figure S2).
The average SNP coverage of the combined 954 salmon was 16x (Figure 3a). The average
coverage of Chinook salmon was 16.5x (n=360) , coho salmon 18.1x (n=277), and sockeye salmon
13.7x (n=317). The percent of heterozygous genotypes for each individual appeared to depend on the
SNP coverage until a depth of 15x (Figure 3a). This relationship was consistent in all species (Figure
3b). This issue did not appear to influence the scale of the number of SNPs among species. The
average number of heterozygous genotypes per salmon for those with 15x coverage or greater reflected
the trend observed for the number of SNPs per species. These values were 1,667,558 in coho salmon,
1,305,776 in sockeye salmon, and 2,636,167 in Chinook salmon (1.58-2.02x greater in Chinook
salmon).
Coastal samples (all sites downstream of the Thompson River or outside the Fraser River – see
Figure 1) appeared to be highly related in Chinook and coho salmon (Figure 3c and 3d). These high
relatedness scores were independent of the coverage or the technology used. In Chinook and coho
salmon, coastal samples had much higher counts of private alleles (see below), which is known to
increase relatedness metrics (Ritland 1996). Relatedness values were greatly reduced by sub-setting
individuals by coastal or non-coastal (data not shown).
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Population structure and clustering analyses

We focused on large-scale resolution of genetic groups in this study, but there was still
substantial genetic variation within each of the genetic groups (Figure S3). Each species had three
supported Fraser River admixture groups (Figures 4 and S3). All groups were based on geography and
were supported by PCA (Figure S4). These included a lower/coastal, mid, and upper Fraser River
admixture group.
All of the lower Fraser River (LFR) admixture groups included sample sites in the Fraser Valley
south of the Fraser Canyon (approximately where the river turns from a southern flow to a western
flow before its outlet to the Pacific Ocean) (Figure 4). In Chinook and coho salmon, the LFR
admixture groups also included nearby coastal sites. For coho salmon, this included locations from
Alaska to California.
We sampled fewer coho salmon north of the Fraser Valley than for the other species due to
accessibility issues during their spawning season, but those north of the valley clustered in a pattern
similar to Chinook salmon (Figure 4). All samples, excluding those from McKinley Creek (a tributary
of the Horsefly River), and some samples without known natal streams (sampled at the Big Bar
Landslide), form one admixture group. In Chinook salmon, a similar admixture group is formed from
the collection of sites west of the Horsefly River and from the Nechako River south. This geographic
region is known as the Interior Plateau. In sockeye salmon, the mid Fraser River (MFR) group is
composed of bodies of water from the Chilcotin and Quesnel River watersheds (Figure 4).
The upper Fraser River (UFR) coho and Chinook salmon admixture groups include most of the
Quesnel watershed locations, with the exception of the lower Cariboo River near the confluence with
the Fraser River (Figure 4). For coho salmon, this was from a single known sampling location and
potentially other locations from salmon that were collected at the Big Bar Landslide. The Chinook
salmon UFR group also included sampling sites above the Nechako River (in or near Robson Valley, a
part of the Rocky Mountain Trench). The UFR sockeye salmon admixture group is comprised of
Nechako River watershed sites.
The LFR admixture groups, in all species, have the greatest genetic differentiation (i.e., F_{st})
from the other admixture groups (Figure 4). In addition, the UFR admixture groups have higher

differentiation from the LFR admixture groups than the MFR groups (Figure 4). F_{st} values were
generally much lower between UFR and MFR admixture groups. While there were some significant
differences in the percent of heterozygous genotypes per individual among groups, these do not show a
consistent pattern (Figure 4). In Chinook salmon, the significant differences in heterozygous genotypes
per individual between the MFR admixture group and the other groups were driven by high values in
upper Chilcotin River samples (File S1).
392

Environmental factors influencing genetic structure

394   Sampling locations from different admixture groups clustered by environmental factors (Figure
395 5, Figure S5). Sites with predominately LFR ancestry values were in regions with higher annual
396 precipitation and higher mean annual temperatures (Figure S5). While there was some overlap between
397 MFR and UFR admixture groups, there was still distinct environmental factors between them overall.
398 This was the case for all three species.
399

Historical estimates of effective population size

Historical estimates of effective population size (N_e) are strikingly different among species
(Figure 6, Figure S6). Prior to 200,000 years before present, almost all Chinook and multiple coho
salmon sampled experienced drops in estimates of effective population size (Figure 6, Figure S6). This
coincides with a time when there was an increase in ocean surface temperatures (Figure 6d).
Between 120,000 and 150,000 years before present, coho and sockeye salmon both experienced
major declines in estimated effective population size (Figures 6 and 7). In sockeye salmon, there are
distinct times when these decreases occurred (Figure 7). For some sockeye salmon locations and for
most coho salmon sites, the decrease in effective population size occurred around the penultimate
glacial maximum, when ocean temperatures were at their lowest. Neither species recovered to previous
estimates of effective population size for over 100,000 years (Figures 6 and S6). This is consistent with
the decrease in polymorphic loci and increases of runs of homozygosity compared to Chinook salmon.
Interestingly, Chinook salmon effective population size increased during this same time frame. A
similar reversal among species was observed between 50,000 and 90,000 years before present, where
there were increases in estimates of effective population size for coho and sockeye salmon, but
decreases in Chinook salmon (Figure 6, Figure S6).
In all species, there were decreases in effective population size around the last glacial
maximum, 19,000 years before present (Figures 6 and 7, Figure S6). The timing of these decreases in
effective population size varied in time by species, admixture group, and sampling sites within
admixture groups (Figures 6 and 7, Figure S6). Most of the demographic histories have a step pattern
near or after the last glacial maximum. Some coastal Fraser River groups have earlier decreases in
effective population sizes than the other Fraser River groups. The exception being LFR sockeye
salmon sites, which have similar or more recent decreases in population size than MFR and UFR
locations (Figure S6).
424

Understanding the genetic diversity of Fraser River salmon

On average, total runs of homozygosity (ROH) were shorter in LFR locations than MFR and
UFR sites of Chinook and coho salmon (Figure 8). Lower ROH is a proxy for greater genetic diversity,
suggesting that the Chinook and coho salmon LFR admixture groups have higher diversity than MFR
and UFR groups. Sampling sites with the longest average total ROH were the Yakoun River (Haida
Gwaii) and Salmon River (tributary of the Thompson River) for Chinook and coho salmon,
respectively (Figure S7). Locations with the shortest average total ROH (i.e., highest diversity) for

Chinook and coho salmon, respectively, were the upper Chilcotin River (tributary of the Fraser River)
and the Deschutes River (tributary of the Columbia River) (Figure S7). ROH levels were most variable
in the Morkill and upper Chilcotin River locations for Chinook salmon. The individuals with the
shorter ROH in the upper Chilcotin River drove the average to the lowest for Chinook salmon.
There was a reverse trend for ROH in sockeye salmon compared to Chinook and coho salmon
(Figure 8). Most of the sockeye salmon sites with the longest ROH (i.e., the lowest diversity) were
from the LFR (i.e., Cultus Lake, Pitt River, and Widgeon Slough). Cultus Lake and Widgeon Slough
have small population sizes, often under 1000 salmon in recent generations (DFO 2020; Doutaz *et al.*
2023). The Harrison River, although not a part of this admixture group due to being below the 0.7
admixture threshold, had the lowest average ROH of all sockeye salmon collections (Figure S7). The
Harrison River, in the early 1980s, had around 300,000 spawning adults (Doutaz *et al.* 2023). Widgeon
Slough and possibly the Harrison River were also the only samples of the sea-type ecotype taken from
the Fraser River (Beacham and Withler 2017; Doutaz *et al.* 2023), all the other Fraser River samples
were likely lake-type. Also, some Harrison River salmon are sea and some are lake-type. Bodies of
water in the Yukon, Alaska, Russia, and the upper Columbia River had slightly lower average ROH (i.e,
more diverse) compared to the LFR, but higher than sites from the MFR and UFR (Figure 8 and Figure
S7).
Nucleotide diversity (π), which is a population-level metric rather than an individual-level
metric like ROH, had a more complex pattern. Nucleotide diversity ranged from 0.147 (Chinook
salmon – Yakoun River) to 0.22 (sockeye salmon – Pinchi Creek). All sockeye salmon sampling
locations had nucleotide diversity metrics above 0.19 except for Widgeon Slough, which was 0.17.
Cultus Lake might have had lower nucleotide diversity than Widgeon Slough based on its longer ROH,
but this location was excluded from this analysis since it had lower SNP coverage than 15x. Similarly,
most northern coastal and upper Columbia River samples were excluded from the nucleotide diversity
analysis because they had SNP coverage below the threshold. Most Fraser River salmon locations had
comparable nucleotide diversity to upper Columbia River Kokanee that were available for comparison.
Similar to the ROH results, the LFR salmon had lower diversity than the MFR and UFR sockeye
salmon.
Most Chinook salmon locations had π values between 0.18 and 0.20. However, the Yakoun
River (Haida Gwaii) had the lowest π of all sites for Chinook salmon (π = 0.147). Many of the
sampling locations with the highest nucleotide diversity were from the MFR (Figure 8). The Fraser
River Chinook salmon sampling sites had average nucleotide diversity values comparable or lower than
other coastal locations along the BC coast (Figure 8).
Coho salmon ranged in nucleotide diversity from 0.17 to 0.21, with the lowest collection site
being the Salmon River in the South Thompson River drainage and the highest being the Kawkawa
Creek below the Fraser Canyon (Figure 8). Fraser River coho salmon had comparable or higher levels
of nucleotide diversity as those from California to Alaska. The most extreme northern (Kwethluk
River) and southern sites (Klamath River) had the second and third lowest nucleotide diversity metrics
respectively.
The number of polymorphic loci was highest in Chinook salmon sampling sites, followed by
coho salmon and then sockeye salmon, which is consistent with the total number of SNPs identified in
each species (Figure S8, Figure 2). While the number of polymorphic loci was dependent on the
number of samples per location, it remained higher in Chinook salmon at comparable sample counts
per location (Figure S8). Even though Chinook salmon had the greatest number of polymorphic loci
per location, nucleotide diversity was often lower than for coho and sockeye salmon locations (Figure
S8). This is consistent with higher individual percent heterozygous genotypes in coho and sockeye
salmon (Figure 3).

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Admixture group private alleles

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There were more private alleles identified in the LFR admixture groups of Chinook and coho
salmon than the other admixture groups, with 5-13% of all the identified SNPs being private to the LFR
group (Figure 9a). Sockeye salmon had a more even distribution of private alleles among admixture
groups than Chinook or coho salmon (Figure 9a). However, the sample size of the LFR sockeye
salmon group was the lowest of all comparisons, and this may have impacted the number of private
alleles identified. Also, sockeye salmon from the Harrison River (in the coastal region) had admixture
ancestry value below 0.7 and were excluded from this analysis. This site had greater genetic diversity
metrics than most other sockeye salmon locations (see previous section), and so the removal of this
river would be expected to reduce the observed private alleles from this region.
Analyses of individual private alleles from Chinook and coho salmon provided more evidence
that LFR individuals have more unique alleles than individuals from MFR and UFR groups (Figure 9b-
d). Each coastal individual had roughly 130,000 private alleles that were not found in the other
admixture groups. Other admixture groups, on average, have much fewer individual private alleles
compared to LFR salmon (~ 0.04% - 10%). The upper Chilcotin River and Spius River contributed
more private alleles to the Chinook salmon MFR admixture group than other rivers in that group
(Figure 9b). Likewise, the Kawkawa River contributed fewer private alleles to the coho salmon LFR
group than the other rivers (Figure 9d).
Sockeye salmon do not have a comparable admixture group with the large store of private
alleles relative to that observed in the LFR coho and Chinook salmon. The northern coast and upper
Columbia River kokanee admixture groups have the highest average individual private allele counts
(Figure 9c). Most sampling locations in the Fraser River have lower individual private alleles
compared to those outside the Fraser River. The main exceptions are samples from the Quesnel River
watershed, which have similar values as those outside of the Fraser River (Figure 9c). This is
consistent with a higher nucleotide diversity in the Quesnel River watershed than the Chilcotin River
watershed (see previous section).
507
Potentially adaptive variants

Twenty candidate adaptive loci were identified in pairwise comparisons among admixture
groups based on expanded haplotype homozygosity analyses (Figure S9, Table 2, File S3). No
significant loci were identified from within admixture group analyses. Most candidate loci were
identified among Chinook salmon admixture groups (14 out of 20). Twelve of these loci overlapped
with the boundary of a gene (Table 2). Two different homologs of the *trace amine-associated receptor*
13C (TAAR13C) gene overlapped with candidate loci identified in comparisons of admixture groups of
sockeye salmon (Figure S9, Table 2). One of these genes was identified from the comparison with the
LFR group and the MFR group. The other was identified from the comparison of the LFR group and
the UFR group.
To identify if there were potential adaptive variants shared among species, we searched for
overlapping regions with high -log10 p-values (we were searching for signals of convergent evolution
with this analysis). The overlap of 10 kb regions with the highest -log10 p-value was 3.6 among
species. The chance of having an overlap with this high of a value was expected by random at least
twice by chance given the distribution of p-values. These overlaps were not considered significant and
they are not shown or discussed below.
524

Discussion

The Fraser River drains a very large and diverse section of British Columbia from the Rocky Mountains to its outlet in the city of Vancouver (Reynoldson *et al.* 2005). This river has at times been the largest producer of salmon in North America (Milne 1964; Northcote and Atagi 1997), and remains important commercially (Henderson and Healey 1993) and culturally for its salmon fisheries. The massive declines in salmon, predominately sockeye and pink salmon from a 1913-1914 landslide caused by the construction of a railway, have never fully recovered (Pacific Salmon Commission – psc.org). There are many pressures potentially for why recovery has not occurred and why there has been continued declines in Fraser River salmon (Arbeider *et al.* 2020; DFO 2021; Doutaz *et al.* 2023). Understanding these pressures and their influence on salmon will be valuable for governance and conservation. Our goal is to provide the genetic tools that will facilitate this understanding and help with recovery.

While there have been many studies to examine population structure in the Fraser River (Small *et al.* 1998; Teel *et al.* 2000; Withler *et al.* 2000; Beacham *et al.* 2001, 2003, 2006b, 2006a, 2017; Nelson *et al.* 2001; Beacham and Withler 2017; Xuereb *et al.* 2022), this is the first to use whole genome resequencing data for a large distribution of multiple salmon species. The improved resolution of resequenced genomes allowed us to compare the genetic diversity among salmon species, calculate nucleotide diversity at the genome level, and evaluate historical influences on population structure. These types of analyses would suffer using fewer genetic markers. As an example, genetic and nucleotide diversity estimates are dependent on the nucleotide variants used to estimate them (e.g., polymorphic loci in one region might not be polymorphic in another region) and whole genome analyses allow unbiased estimates among locations. These genome sequences also us to capture a better snapshot in time that can be used in comparisons for future studies.

Species level differences of polymorphic loci

Genetic diversity can be understood at both a species and population level. We first discuss the diversity at a species level and will return to the population level later. From previous research, we expected different levels of polymorphic loci among Pacific salmon, with a generally lower genetic diversity of Pacific salmon compared to other species such as rainbow trout (*O. mykiss*) (Allendorf and Utter 1979). Several studies observed that sockeye salmon, and sometimes coho salmon, have the lowest levels of polymorphic loci or average heterozygosity (Utter *et al.* 1973; Allendorf and Utter 1979; Wood *et al.* 1994).

Our findings indicate that Chinook salmon have ~1.5-2x the number of polymorphic loci and a corresponding drop in the length of ROH (~27-29% of the ROH length) when compared to coho or sockeye salmon. We observed the same trend based on the total number of heterozygous genotypes among the species when only looking at individuals with higher than 15x coverage. While sequencing coverage likely impacts the number of polymorphic loci identified, the scale of the differences we observed, the greater geographic sampling distance covered by the species with fewer polymorphic loci, a similar trend in ROH, and a similar trend with heterozygous genotypes of individuals with at least 15x coverage supports that these findings are robust. If we assume that salmon species had similar levels of genetic diversity during speciation and that mutation rates are similar among species, coho and sockeye salmon likely had a greater reduction in standing genetic variation than Chinook salmon.

From modelling of effective population size through time, there is evidence that sockeye and
coho salmon experienced a large drop in effective population size around the penultimate glacial
maximum that Chinook salmon did not experience (Chinook salmon did experience an earlier drop in
effective population size, but the size rebounded after ~50,000 years). We must consider that we do not
have precise estimates of mutation rates for all of these species. Even without precise dates or effective
population size estimates, however, we note that the drop in effective population size did not rebound
to prior levels in sockeye and coho salmon as it did in most Chinook salmon sites. This could have
resulted in a loss of polymorphic loci in both sockeye and coho salmon not observed in Chinook
salmon.
While the SNP count, effective population size, and runs of homozygosity data are consistent
with this hypothesis, nucleotide diversity and private allele analyses appear to be inconsistent at first
glance. For example, why would sockeye and coho salmon have higher nucleotide diversity if Chinook
salmon have more polymorphic loci overall (sockeye salmon average π: 0.205, coho salmon: 0.197,
Chinook salmon: 0.189)? Also, why would the average private allele count be higher in the LFR coho
salmon admixture group than the Chinook salmon LFR group if the Chinook salmon have more
polymorphic loci?
First, we note that some coastal samples of sockeye salmon were removed due to low coverage
in these analyses, and these samples would have likely lowered the average nucleotide diversity score.
Technical artifacts like this and others (e.g., how nucleotide diversity was measured, sequence
coverage, or differences among genome assemblies) might help to explain a part of these discrepancies,
but they appear to be inadequate for all of the data. For example, at comparable SNP coverages,
sockeye salmon had higher proportions of heterozygous genotypes than coho or Chinook salmon even
though coho and Chinook salmon have more polymorphic loci overall.
The timing of re-colonization and the distribution of groups with unique genetic characteristics
could help to account for some of the discrepancies among genetic diversity metrics. In Chinook and
coho salmon, for example, the coastal genetic groups have the largest stores of private alleles and the
highest metrics of genetic diversity (comparable between species, even if slightly higher in coho
salmon). If there was a difference in the timing or the size of the groups that re-colonized the Fraser
River from these coastal genetic groups, we might expect the fraction of private alleles from a group to
reflect these differences.
Estimates of effective population size dip in Chinook salmon locations soon after the last glacial
maximum when compared to coho salmon. If we assume these dips are founding events and the times
are roughly accurate, we infer that Chinook salmon re-colonized the Fraser River earlier. The founding
events also appeared to have reduced the effective population size of Chinook salmon sites less than
they did for coho salmon, perhaps reflective of larger founding populations in Chinook salmon. With
this in mind, even with fewer polymorphic loci overall, the LFR coho salmon might appear to have
higher levels of private alleles per individual than Chinook salmon when comparing the LFR group to
the MFR and UFR groups because those groups have lower genetic diversity as a result of more recent
and smaller founding events. The MFR and UFR coho salmon groups could have retained comparable
nucleotide diversity metrics to Chinook salmon groups due to gene flow, or because not as much time
has passed for genetic drift to impact this metric as it has for Chinook salmon. Also, heterozygosity
excess (e.g., (Wang *et al.* 2016)), where allele frequency varies by chance more often between males
and females due to a low breeding population, is another viable explanation for how MFR or UFR coho
salmon might have comparable nucleotide diversity to Chinook salmon.
Nucleotide diversity might also be higher in a species with lower overall counts of polymorphic
loci and genetic diversity if there was recent admixture between diverse populations. In sockeye
salmon, there were at least two distinct demographic patterns identified based on when effective

population size declined around the penultimate glacial maximum (perhaps indicative of groups from
different glacial refugia as previously suggested (Wood *et al.* 1994; Beacham *et al.* 2006b)). In the
Fraser River, both demographic patterns were observed, with the MFR and UFR admixture groups
mostly having the earlier decline in effective population size (earlier than 150,000 years before
present), while the LFR locations had mostly the later (after 150,000 years before). Gene flow between
these groups could explain the higher nucleotide diversity in the MFR and UFR even though sockeye
salmon as a whole have fewer polymorphic loci.
Overall, the data point to a model where Chinook salmon have more polymorphic loci, but less
diversity at these loci than coho or sockeye salmon. Demographic modelling points to one explanation
for how this might have occurred, but other explanations are possible. For example, heterozygosity
excess is expected in populations with smaller effective population size (Robertson 1965; Wang *et al.*
2016).
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630 *Influences on the population structure of Fraser River salmon*
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Turning from the species level to the population level, researchers have known the basic
population structure of Fraser River salmon since the mid 1990's to the mid 2000's (Wood *et al.* 1994;
Small *et al.* 1998; Teel *et al.* 2000; Withler *et al.* 2000; Nelson *et al.* 2001; Beacham *et al.* 2003, 2006a,
2006b, 2017; Beacham and Withler 2017; Xuereb *et al.* 2022). This knowledge was based on relatively
few genetic markers but often many samples from a wide distribution. In the current work, we sampled
a moderate number of locations and individuals, but resequenced entire genomes. The basic population
structure identified in the Fraser River from resequencing genomes was similar to findings from earlier
studies using microsatellite and allozyme genetic markers (e.g., lower, mid, and upper groups), except
the Thompson River groups were missing since we did not sample the Thompson River drainage
enough to reconstruct them. We interpret the consistency among studies to mean that these genetic
clusters are stable through time (as sampling took place at various times, even within studies (Beacham
et al. 2003, 2006a)), the clusters are robust since they can be identified with only a few genetic
markers, and the entire genome is impacted by these groupings.
Understanding why we observe these genetic groups is important (e.g., (Nadeau *et al.* 2016;
Rougemont *et al.* 2023)). Are they a technical artifact from trying to cluster genomes influenced by
isolation-by-distance? Are they an artifact from the transfers between regions of the Fraser River
(Withler 1982)? Were they formed from different re-colonization events (or a combinations of events)?
Are they a consequence of environmental hurdles to gene-flow? Were they formed from adaptations to
specific environmental conditions?
If the groups are technical artifacts from isolation-by-distance, they may not be useful from a
conservation or management perspective because the groups were shaped by stochastic processes rather
than adaptations to different environments. If instead, groups were shaped by different re-colonization
events, they might reflect those colonization histories rather than the environments they occupy. If the
groups are a result of limited gene flow, we might expect the salmon to have specific adaptations to
environmental factors that influence gene flow (e.g., the ability to pass regions of the Fraser River such
as the Fraser Canyon), but not necessarily to other environmental conditions. These are only some
examples of why these groups might exist. The genetic groups could also be influenced by a
combination of these processes.
Previous researchers have suggested that isolation-by-distance is an important factor
influencing salmon genetics (Withler *et al.* 2000; Rougemont *et al.* 2020, 2023). Others have
suggested that different re-colonization events (e.g., from different glacial refugia) influence population

structure in Fraser River salmon (Wood *et al.* 1994; Small *et al.* 1998; Teel *et al.* 2000; Withler *et al.*
2000). Still others have suggested that barriers to gene-flow like the Fraser Canyon (specifically Hell's
Canyon) are important for population structure (Wehrhahn and Powell 1987). Finally, researchers have
also suggested genetic adaptation to environmental conditions are important (Small *et al.* 1998).
Analyzing multiple species can help with distinguishing which hypotheses are more supported (Nadeau
et al. 2016).
While we observed lower F_{st} between neighbouring admixture groups in all species, which
would be consistent with isolation-by-distance, we also observed that private alleles were more
common in the UFR group of Chinook salmon than in the MFR group. If populations were delineated
strictly by isolation-by-distance, we might expect UFR salmon to have the lowest count of private
alleles. Also, in all three species, sampling sites of an admixture group clustered together based on
environmental factors in a PCA. If admixture groups were the result of neutral genetic variation or
different colonization events alone, we would expect random sorting of groups or for it to be based on
distance between locations. This suggests that the environment is a major factor influencing genetic
variation in Fraser River salmon, but still does not exclude isolation-by-distance or separate
colonization events.
Indeed, from demographic modelling, separate colonization events appear to have been an
important source of sockeye salmon in the Fraser River and possibly for Chinook salmon. For sockeye
salmon, there are at least two unique demographic histories, with one common to LFR sites and the
other to MFR and UFR locations. These groups have unique demographic histories starting as early as
the penultimate glacial maximum. In Chinook salmon, LFR sampling locations have an earlier
decrease in effective population size that may be indicative of an earlier colonization history. Overall,
the variability in demographic histories of the species suggest a complex re-colonization history.
In all three species, there was a delineation between admixture and PCA groups between the
LFR group and the MFR group. The LFR groups had higher F_{st} values among groups, and 5-13% of
the SNPs we identified in Chinook and coho salmon were private alleles of the LFR group. If we also
consider the velocity barriers between these groups (Wright 2022), we have several pieces of evidence
to suggest that there is a substantial gene flow and migration barrier due to the Fraser Canyon. In
support of this hypothesis, we observed that the Chinook and coho salmon from the LFR groups had
higher nucleotide diversity and that longer runs of homozygosity in coho salmon appeared to be
demarcated near Hell's Canyon in the Fraser Canyon. If the Fraser Canyon is a migration barrier, we
would expect to observe these genetic delineations in all species and generally this is what was
observed.
In a study of Fraser River sockeye salmon, researchers observed possible adaptations to
environmental differences among populations (Eliason *et al.* 2011). They observed coastal populations
had significantly different cardiac morphology and performance from other groups, and that cardiac
morphology was correlated with migration difficulty. River temperature was also correlated with
cardiac performance. This study highlights the importance of the environment on population structure,
migration difficulty, and relates the genetic structure among admixture groups to phenotypic variation.
We note that in a comparison of extended haplotype homozygosity between the LFR coho salmon with
the MFR group, a potential adaptive locus was identified that overlapped with the *Ankyrin Repeat And*
SOCS Box Containing 2 (ASB2) gene. This gene is thought to be involved in heart development
(Yamak *et al.* 2020; Min *et al.* 2021).
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707   *Influence of adaptation on population structure of Fraser River salmon*
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When comparing genetic groups, we identified twenty candidate adaptive loci. No significant
loci overlapped among species, which would have been evidence for convergent evolution. Rather
than discuss all twenty regions, we will focus on what these loci reveal in general, and as an example
discuss the olfactory receptor gene TAAR13C. Potentially adaptive loci overlapped with two
TAAR13C genes and were identified in separate comparisons of the LFR sockeye salmon with MFR
and UFR groups.
In general, adaptive loci among the genetic groups reveal that there are different selective
pressures along the Fraser River drainage. While the environmental PCA gave us insight into how
genetic groups were organized based on environmental components, adaptive loci can reveal how these
and other elements shape the genomes of salmon through generations. In the type of analysis we used,
we did not need to know what these elements were. This means we can discover adaptation caused by
unknown and unexamined factors.
We may be able to formulate a hypothesis regarding a mode of adaptation in the case of the
TAAR13C genes. The olfactory receptor gene, TAAR13C, is directly involved in detecting putrescine,
cadaverine, and other diamines (typically associated with decomposing flesh) (Hussain *et al.* 2013;
Tessarolo *et al.* 2014; Liberles 2015; Gainetdinov *et al.* 2018). The TAAR13C gene has also
previously been found to be associated with sea age at maturity in Atlantic salmon (*Salmo salar*)
(Sinclair-Waters *et al.* 2022), and to possibly be under selection in another study of Atlantic salmon
(not peer reviewed at the time of writing) (Miettinen *et al.* 2023). Since the TAAR13C gene appears to
be involved in the timing of maturation, one hypothesis is that diamines could act as a signal for
maturation for the different genetic groups of sockeye salmon in the Fraser River. One source of
diamines to consider for this hypothesis comes from eggs. A study of Arctic charr (*Salvelinus alpinus*)
eggs and alevin revealed that during these developmental stages different amounts of putrescine and
cadaverine were produced during the spawning season (Srivastava *et al.* 1992). If diamines influenced
maturation of MFR and UFR groups of sockeye salmon, this information would be valuable in
conservation and management. Understanding the modes of adaptation of other potential adaptive loci
could likewise be useful for these purposes.
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Conclusion
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From analyzing hundreds of resequenced genomes of Chinook, sockeye, and coho salmon,
mostly from the Fraser River, we identified genetic groups that had previously been identified with
only a few microsatellites. With data from resequenced genomes, we were able to examine how these
groups might have formed. They appear to have been influenced by many factors, including isolation-
by-distance, migration barriers, separate glacial refugia in sockeye salmon, and by the diverse
environmental factors of the Fraser River drainage. We identified twenty potentially adaptive loci
among the different groups of salmon, which is indicative of unique patterns of selection in the various
regions of the river. Two of these loci overlapped with homologs of the TAAR13C gene. This gene
could be an important target for future studies, to investigate if there is a link between it and timing of
maturation in Fraser River sockeye salmon. Finally, by examining three species, we were able to
identify commonalities and differences. Patterns of historical effective population size were
dramatically different among the three species and could explain the variable genetic diversity currently
observed among them. In terms of information useful from a conservation and management
perspective, we have generated resequenced genomes that can be reused in future studies, determined
metrics for comparison, and identified loci that could impact the timing of maturation of salmon in the
Fraser River.

Data availability

Resequenced genomes from this study are available on the NCBI (coho salmon: PRJNA986075, sockeye salmon: PRJNA930425, and Chinook salmon: PRJNA694998 and PRJNA1090956). Truth SNP datasets used for SNP calling are available in File S2. SNP datasets are available in FigShare. Scripts are available at github.com/KrisChristensen (repositories: PrivateAllelePerInd, PrivateAllele, VCF2Relate, VCFstats, and MapVCF2NewGenome).

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References

Aguillon, S. M., J. W. Fitzpatrick, R. Bowman, S. J. Schoech, A. G. Clark *et al.*, 2017 Deconstructing isolation-by-distance: The genomic consequences of limited dispersal. PLOS Genetics 13: e1006911.

Ahrens, C. W., P. D. Rymer, A. Stow, J. Bragg, S. Dillon *et al.*, 2018 The search for loci under selection: trends, biases and progress. Molecular Ecology 27: 1342–1356.

Alexander, D. H., J. Novembre, and K. Lange, 2009 Fast model-based estimation of ancestry in unrelated individuals. Genome Res 19: 1655–1664.

Allendorf, F. W., and F. M. Utter, 1979 Population genetics, pp. 407–454 in *Fish physiology*, edited by W. S. Hoar, D. J. Randall, and J. R. Brett. Academic Press, New York, NY.

Andrews, G. D. M., J. K. Russell, S. R. Brown, and R. J. Enkin, 2012 Pleistocene reversal of the Fraser River, British Columbia. Geology 40: 111–114.

Arbeider, M., L. M. Ritchie, D. Braun, B. Jenewein, K. Rickards *et al.*, 2020 Interior Fraser Coho Salmon Recovery Potential Assessment: Fisheries and Oceans Canada Canadian Science Advisory Secretariat 2020/025, xi + 211 p. p.

Auguie, B., 2017 gridExtra: Miscellaneous Functions for "Grid" Graphics.

Beacham, T. D., J. R. Candy, K. J. Supernault, T. Ming, B. Deagle *et al.*, 2001 Evaluation and Application of Microsatellite and Major Histocompatibility Complex Variation for Stock Identification of Coho Salmon in British Columbia. Transactions of the American Fisheries Society 130: 1116–1149.

Beacham, T. D., K. L. Jonsen, J. Supernault, M. Wetklo, L. Deng *et al.*, 2006a Pacific Rim Population Structure of Chinook Salmon as Determined from Microsatellite Analysis. Transactions of the American Fisheries Society 135: 1604–1621.

Beacham, T. D., B. McIntosh, C. MacConnachie, K. M. Miller, R. E. Withler *et al.*, 2006b Pacific Rim Population Structure of Sockeye Salmon as Determined from Microsatellite Analysis. Transactions of the American Fisheries Society 135: 174–187.

Beacham, T. D., K. J. Supernault, M. Wetklo, B. Deagle, K. Labaree *et al.*, 2003 The geographic basis for population structure in Fraser River chinook salmon (Oncorhynchus tshawytscha). Fishery Bulletin 101: 229–242.

Beacham, T. D., C. Wallace, C. MacConnachie, K. Jonsen, B. McIntosh *et al.*, 2017 Population and individual identification of coho salmon in British Columbia through parentage-based tagging and genetic stock identification: an alternative to coded-wire tags. Can. J. Fish. Aquat. Sci. 74: 1391–1410.

Beacham, T. D., and R. E. Withler, 2017 Population structure of sea-type and lake-type sockeye salmon and kokanee in the Fraser River and Columbia River drainages. PLOS ONE 12: e0183713.

Beatty, G. E., and J. Provan, 2010 Refugial persistence and postglacial recolonization of North America by the cold-tolerant herbaceous plant Orthilia secunda. Molecular Ecology 19: 5009–5021.

Bernatchez, L., and C. C. Wilson, 1998 Comparative phylogeography of Nearctic and Palearctic fishes. Molecular Ecology 7: 431–452.

Bett, N. N., S. G. Hinch, N. J. Burnett, M. R. Donaldson, and S. M. Naman, 2017 Causes and Consequences of Straying into Small Populations of Pacific Salmon. Fisheries 42: 220–230.

Bierne, N., D. Roze, and J. J. Welch, 2013 Pervasive selection or is it…? why are FST outliers sometimes so frequent? Molecular Ecology 22: 2061–2064.

Brieuc, M. S. O., C. D. Waters, J. E. Seeb, and K. A. Naish, 2014 A dense linkage map for Chinook salmon (Oncorhynchus tshawytscha) reveals variable chromosomal divergence after an ancestral whole genome duplication event. G3 (Bethesda) 4: 447–460.

Broad Institute, 2019 Picard Toolkit. Broad Institute, GitHub Repository.

Byun, A. S., B. Koop, and T. E. Reimchen, 1999 Coastal Refugia and Postglacial Recolonization Routes: A Reply to Demboski, Stone, and Cook. Evolution 53: 2013–2015.

Catchen, J. M., A. Amores, P. Hohenlohe, W. Cresko, and J. H. Postlethwait, 2011 Stacks: Building and Genotyping Loci De Novo From Short-Read Sequences. G3 Genes|Genomes|Genetics 1: 171–182.

Catchen, J., P. A. Hohenlohe, S. Bassham, A. Amores, and W. A. Cresko, 2013 Stacks: an analysis tool set for population genomics. Mol Ecol 22: 3124–3140.

Chang, C. C., C. C. Chow, L. C. Tellier, S. Vattikuti, S. M. Purcell *et al.*, 2015 Second-generation PLINK: rising to the challenge of larger and richer datasets. Gigascience 4:.

Christensen, K. A., J. S. Leong, D. Sakhrani, C. A. Biagi, D. R. Minkley *et al.*, 2018 Chinook salmon (Oncorhynchus tshawytscha) genome and transcriptome. PLOS ONE 13: e0195461.

Christensen, K. A., E. B. Rondeau, D. R. Minkley, D. Sakhrani, C. A. Biagi *et al.*, 2020 The sockeye salmon genome, transcriptome, and analyses identifying population defining regions of the genome. PLOS ONE 15: e0240935.

Christensen, K. A., E. B. Rondeau, D. Sakhrani, C. A. Biagi, H. Johnson *et al.*, 2021 The pink salmon genome: Uncovering the genomic consequences of a two-year life cycle. PLOS ONE 16: e0255752.

Clague, J. J., 2017 Deglaciation of the Cordillera of Western Canada at the end of the Pleistocene. Cuadernos de Investigación Geográfica 43: 449–466.

Clark, P. U., A. S. Dyke, J. D. Shakun, A. E. Carlson, J. Clark *et al.*, 2009 The Last Glacial Maximum. Science 325: 710–714.

Danecek, P., A. Auton, G. Abecasis, C. A. Albers, E. Banks *et al.*, 2011 The variant call format and VCFtools. Bioinformatics 27: 2156–2158.

Danecek, P., J. K. Bonfield, J. Liddle, J. Marshall, V. Ohan *et al.*, 2021 Twelve years of SAMtools and BCFtools. GigaScience 10: giab008.

Darvill, C. M., B. Menounos, B. M. Goehring, O. B. Lian, and M. W. Caffee, 2018 Retreat of the Western Cordilleran Ice Sheet Margin During the Last Deglaciation. Geophysical Research Letters 45: 9710–9720.

DFO, 2020 Recovery Potential Assessment – Cultus Lake Sockeye Salmon (Oncorhynchus nerka) (2019): DFO Canadian Science Advisory Secretariat Science Advisory Report 2020/011.

DFO, 2021 Recovery Potential Assessment for 11 Designatable Units of Fraser River Chinook Salmon, Oncorhynchus tshawytscha, Part 2: Elements 12 to 22: Canadian Science Advisory Secretariat Science Advisory Report 2021/030.

Doutaz, D., A.-M. Huang, S. Decker, and T. Vivian, 2023 Recovery Potential Assessment for Fraser River Sockeye Salmon (Oncorhynchus nerka), Nine Designatable Units Part 2: Biology, Habitat, Threats, Mitigations and Allowable Harm - Elements 1-11, 14, 16-18, 22: Fisheries and Oceans Canada Canadian Science Advisory Secretariat 2023/003, xiii + 250 p.

Eliason, E. J., T. D. Clark, M. J. Hague, L. M. Hanson, Z. S. Gallagher *et al.*, 2011 Differences in Thermal Tolerance Among Sockeye Salmon Populations. Science 332: 109–112.

Fick, S. E., and R. J. Hijmans, 2017 WorldClim 2: new 1-km spatial resolution climate surfaces for global land areas. International Journal of Climatology 37: 4302–4315.

Gainetdinov, R. R., M. C. Hoener, and M. D. Berry, 2018 Trace Amines and Their Receptors (J. M. Witkin, Ed.). Pharmacol Rev 70: 549–620.

Gautier, M., A. Klassmann, and R. Vitalis, 2017 rehh 2.0: a reimplementation of the R package rehh to detect positive selection from haplotype structure. Mol Ecol Resour 17: 78–90.

Gautier, M., and R. Vitalis, 2012 rehh: an R package to detect footprints of selection in genome-wide SNP data from haplotype structure. Bioinformatics 28: 1176–1177.

Grant, S., and G. Pestal, 2009 CERTIFICATION UNIT PROFILE: FRASER RIVER PINK SALMON: Department of Fisheries and Oceans Canada Canadian Manuscript Report of Fisheries and Aquatic Sciences 2875, vii+36p. p.

Hansen, J., M. Sato, G. Russell, and P. Kharecha, 2013 Climate sensitivity, sea level and atmospheric carbon dioxide. Philosophical Transactions of the Royal Society A: Mathematical, Physical and Engineering Sciences 371: 20120294.

Harington, C. R., 1996 Quaternary Animals: Vertebrates of the Ice Age, pp. 261–273 in *Life in Stone: A Natural History of British Columbia's Fossils*, edited by R. Ludvigsen. University of British Columbia Press, Vancouver.

Heard, W. R., 2012 Overview of salmon stock enhancement in southeast Alaska and compatibility with maintenance of hatchery and wild stocks. Environ Biol Fish 94: 273–283.

Henderson, M. A., and M. C. Healey, 1993 Doubling sockeye salmon production in the Fraser River—Is this sustainable development? Environmental Management 17: 719–728.

Hewitt, G. M., 1996 Some genetic consequences of ice ages, and their role in divergence and speciation. Biological Journal of the Linnean Society 58: 247–276.

Hewitt, G., 2000 The genetic legacy of the Quaternary ice ages. Nature 405: 907–913.

Hofmeister, R. J., D. M. Ribeiro, S. Rubinacci, and O. Delaneau, 2023 Accurate rare variant phasing of whole-genome and whole-exome sequencing data in the UK Biobank. Nat Genet 55: 1243–1249.

Hussain, A., L. R. Saraiva, D. M. Ferrero, G. Ahuja, V. S. Krishna *et al.*, 2013 High-affinity olfactory receptor for the death-associated odor cadaverine. Proceedings of the National Academy of Sciences 110: 19579–19584.

Keefer, M. L., and C. C. Caudill, 2013 Homing and straying by anadromous salmonids: a review of mechanisms and rates. Rev Fish Biol Fisheries 24: 333–368.

Kew, M., 1992 Salmon Availability, Technology, and Cultural Adaptation in the Fraser River Watershed, in *A Complex Culture of the British Columbia Plateau*, edited by B. Hayden. UBC Press, Vancouver, BC.

Kolde, R., 2019 pheatmap: Pretty Heatmaps.

Larson, W. A., M. T. Limborg, G. J. McKinney, D. E. Schindler, J. E. Seeb *et al.*, 2017 Genomic islands of divergence linked to ecotypic variation in sockeye salmon. Molecular Ecology 26: 554–570.

Latch, E. K., J. R. Heffelfinger, J. A. Fike, and O. E. RHODES Jr, 2009 Species-wide phylogeography of North American mule deer (Odocoileus hemionus): cryptic glacial refugia and postglacial recolonization. Molecular Ecology 18: 1730–1745.

Li, H., 2013 Aligning sequence reads, clone sequences and assembly contigs with BWA-MEM. arXiv:1303.3997 [q-bio].

Li, H., and R. Durbin, 2010 Fast and accurate long-read alignment with Burrows-Wheeler transform. Bioinformatics 26: 589–595.

Li, H., and R. Durbin, 2009 Fast and accurate short read alignment with Burrows-Wheeler transform. Bioinformatics 25: 1754–1760.

Liberles, S. D., 2015 Trace amine-associated receptors: ligands, neural circuits, and behaviors. Current Opinion in Neurobiology 34: 1–7.

Lotterhos, K. E., and M. C. Whitlock, 2015 The relative power of genome scans to detect local adaptation depends on sampling design and statistical method. Molecular Ecology 24: 1031–1046.

McKenna, A., M. Hanna, E. Banks, A. Sivachenko, K. Cibulskis *et al.*, 2010 The Genome Analysis Toolkit: A MapReduce framework for analyzing next-generation DNA sequencing data. Genome Res 20: 1297–1303.

McPhail, J. D., 2007 *The Freshwater Fishes of British Columbia*. University of Alberta.

McRae, B. H., 2006 ISOLATION BY RESISTANCE. Evolution 60: 1551–1561.

Meek, M. H., M. R. Baerwald, M. R. Stephens, A. Goodbla, M. R. Miller *et al.*, 2016 Sequencing improves our ability to study threatened migratory species: Genetic population assignment in California's Central Valley Chinook salmon. Ecology and Evolution 6: 7706–7716.

Miettinen, A., J. Dannewitz, S. Palm, E. P. Ahi, A. Romakkaniemi *et al.*, 2023 Loci associated with maturation, migration and appetite control are linked with signals of fine-scale local selection in a large Atlantic salmon population. 2023.08.23.553800.

Milne, D. J., 1964 The Chinook and Coho Salmon Fisheries of British Columbia: Fisheries Research Board of Canada 142.

Min, K.-D., M. Asakura, M. Shirai, S. Yamazaki, S. Ito *et al.*, 2021 ASB2 is a novel E3 ligase of SMAD9 required for cardiogenesis. Sci Rep 11: 23056.

Nadeau, S., P. G. Meirmans, S. N. Aitken, K. Ritland, and N. Isabel, 2016 The challenge of separating signatures of local adaptation from those of isolation by distance and colonization history: The case of two white pines. Ecology and Evolution 6: 8649–8664.

Nelson, R., M. Small, T. Beacham, and J. Supernault, 2001 Population structure of Fraser River chinook salmon (Oncorhynchus tshawytscha): An analysis using microsatellite DNA markers. Fishery Bulletin 99: 94–107.

Nichols, K. M., C. C. Kozfkay, and S. R. Narum, 2016 Genomic signatures among Oncorhynchus nerka ecotypes to inform conservation and management of endangered Sockeye Salmon. Evol Appl 9: 1285–1300.

Northcote, T. G., and D. Y. Atagi, 1997 Pacific Salmon Abundance Trends in the Fraser River Watershed Compared with Other British Columbia Systems, pp. 199–219 in *Pacific Salmon & their Ecosystems: Status and Future Options*, edited by D. J. Stouder, P. A. Bisson, and R. J. Naiman. Springer US, Boston, MA.

Pess, G. R., R. Hilborn, K. Kloehn, and T. P. Quinn, 2012 The influence of population dynamics and environmental conditions on pink salmon (Oncorhynchus gorbuscha) recolonization after barrier removal in the Fraser River, British Columbia, Canada. Can. J. Fish. Aquat. Sci. 69: 970–982.

Pew, J., P. H. Muir, J. Wang, and T. R. Frasier, 2015 related: an R package for analysing pairwise relatedness from codominant molecular markers. Molecular Ecology Resources 15: 557–561.

PLINK 1.9 cog-genomics.org.

QGIS Development Team, 2022 QGIS_software.

Quinn, T. P., 1993 A review of homing and straying of wild and hatchery-produced salmon. Fisheries Research 18: 29–44.

R Core Team, 2022 R: A Language and Environment for Statistical Computing.

Reynoldson, T. B., J. Culp, R. Lowell, and J. S. Richardson, 2005 Fraser River Basin, pp. 696–732 in *Rivers of North America*, edited by A. C. Benke and C. E. Cushing. Academic Press, Burlington.

Ritland, K., 1996 Estimators for pairwise relatedness and individual inbreeding coefficients. Genetics Research 67: 175–185.

Robertson, A., 1965 The interpretation of genotypic ratios in domestic animal populations. Animal Science 7: 319–324.

Rondeau, E. B., K. A. Christensen, D. R. Minkley, J. S. Leong, M. T. T. Chan *et al.*, 2023 Population-size history inferences from the coho salmon (Oncorhynchus kisutch) genome. G3 Genes| Genomes|Genetics 13: jkad033.

Rougemont, Q., T. Leroy, E. B. Rondeau, B. Koop, and L. Bernatchez, 2023 Allele surfing causes maladaptation in a Pacific salmon of conservation concern. PLOS Genetics 19: e1010918.

Rougemont, Q., J.-S. Moore, T. Leroy, E. Normandeau, E. B. Rondeau *et al.*, 2020 Demographic history shaped geographical patterns of deleterious mutation load in a broadly distributed Pacific Salmon. PLOS Genetics 16: e1008348.

Saravanan, K. A., M. Panigrahi, H. Kumar, B. Bhushan, T. Dutt *et al.*, 2020 Selection signatures in livestock genome: A review of concepts, approaches and applications. Livestock Science 241: 104257.

Sinclair-Waters, M., T. Nome, J. Wang, S. Lien, M. P. Kent *et al.*, 2022 Dissecting the loci underlying maturation timing in Atlantic salmon using haplotype and multi-SNP based association methods. Heredity 129: 356–365.

Slowikowski, K., 2021 ggrepel: Automatically Position Non-Overlapping Text Labels with "ggplot2."

Small, M. P., T. D. Beacham, R. E. Withler, and R. J. Nelson, 1998 Discriminating coho salmon (Oncorhynchus kisutch) populations within the Fraser River, British Columbia, using microsatellite DNA markers. Molecular Ecology 7: 141–155.

Soltis, D. E., M. A. Gitzendanner, D. D. Strenge, and P. S. Soltis, 1997 Chloroplast DNA intraspecific phylogeography of plants from the Pacific Northwest of North America. Plant Syst. Evol. 206: 353–373.

Srivastava, R. K., J. A. Brown, and M. E. Brosnan, 1992 The presence of polyamines during embryonic development of arctic charr, *Salvelinus alpinus*. Comparative Biochemistry and Physiology Part B: Comparative Biochemistry 101: 153–157.

Stewart, J. R., A. M. Lister, I. Barnes, and L. Dalén, 2009 Refugia revisited: individualistic responses of species in space and time. Proceedings of the Royal Society B: Biological Sciences 277: 661–671.

Teel, D. J., G. B. Milner, G. A. Winans, and W. S. Grant, 2000 Genetic Population Structure and Origin of Life History Types in Chinook Salmon in British Columbia, Canada. Transactions of the American Fisheries Society 129: 194–209.

Terhorst, J., J. A. Kamm, and Y. S. Song, 2017 Robust and scalable inference of population history from hundreds of unphased whole genomes. Nat Genet 49: 303–309.

Tessarolo, J. A., M. J. Tabesh, M. Nesbitt, and W. S. Davidson, 2014 Genomic Organization and Evolution of the Trace Amine-Associated Receptor (TAAR) Repertoire in Atlantic Salmon (Salmo salar). G3 Genes|Genomes|Genetics 4: 1135–1141.

Tribe, S., 2005 Eocene paleo-physiography and drainage directions, southern Interior Plateau, British Columbia. Can. J. Earth Sci. 42: 215–230.

Utter, F. M., F. W. Allendorf, and H. O. Hodgins, 1973 Genetic Variability and Relationships in Pacific
 Salmon and Related Trout Based on Protein Variations. Systematic Biology 22: 257–270.

Van der Auwera, G. A., M. O. Carneiro, C. Hartl, R. Poplin, G. Del Angel *et al.,* 2013 From FastQ data
 to high confidence variant calls: the Genome Analysis Toolkit best practices pipeline. Curr
 Protoc Bioinformatics 43: 11.10.1-11.10.33.

Veale, A. J., and M. A. Russello, 2017 Genomic Changes Associated with Reproductive and Migratory
 Ecotypes in Sockeye Salmon (Oncorhynchus nerka). Genome Biol Evol 9: 2921–2939.

Wang, J., E. Santiago, and A. Caballero, 2016 Prediction and estimation of effective population size.
 Heredity 117: 193–206.

Warner, B. G., R. W. Mathewes, and J. J. Clague, 1982 Ice-Free Conditions on the Queen Charlotte
 Islands, British Columbia, at the Height of Late Wisconsin Glaciation. Science 218: 675–677.

Weber, J. N., G. S. Bradburd, Y. E. Stuart, W. E. Stutz, and D. I. Bolnick, 2017 Partitioning the effects
 of isolation by distance, environment, and physical barriers on genomic divergence between
 parapatric threespine stickleback. Evolution 71: 342–356.

Wehrhahn, C. F., and R. Powell, 1987 Electrophoretic Variation, Regional Differences, and Gene Flow
 in the Coho Salmon (Oncorhynchus kisutch) of Southern British Colombia. Can. J. Fish. Aquat.
 Sci. 44: 822–831.

Weir, B. S., and C. C. Cockerham, 1984 Estimating F-Statistics for the Analysis of Population
 Structure. Evolution 38: 1358–1370.

Wickham, H., 2016 *ggplot2: Elegant Graphics for Data Analysis*. Springer-Verlag New York.

Wickham, H., 2007 Reshaping Data with the reshape Package. Journal of Statistical Software 21: 1–20.

Withler, F. C., 1982 Transplanting Pacific salmon: Canada Department of Fisheries and Oceans
 Canadian technical report of fisheries and aquatic sciences 1488-5379 1079.

Withler, R. E., K. D. Le, R. J. Nelson, K. M. Miller, and T. D. Beacham, 2000 Intact genetic structure
 and high levels of genetic diversity in bottlenecked sockeye salmon (Oncorhynchus nerka)
 populations of the Fraser River, British Columbia, Canada. Can. J. Fish. Aquat. Sci. 57: 1985–
 1998.

Wood, C. C., B. E. Riddell, D. T. Rutherford, and R. E. Withler, 1994 Biochemical Genetic Survey of
Sockeye Salmon (Oncorhynchus nerka) in Canada. Can. J. Fish. Aquat. Sci. 51: 114–131.

Wright, M., 2022 A typology of hydraulic barriers to salmon migration in a bedrock river [Masters]:
Simon Fraser University, 57 p.

Wright, S., 1943 Isolation by Distance. Genetics 28: 114–138.

Xuereb, A., Q. Rougemont, X. Dallaire, J.-S. Moore, E. Normandeau *et al.*, 2022 Re-evaluating Coho
salmon (Oncorhynchus kisutch) conservation units in Canada using genomic data. Evolutionary
Applications 15: 1925–1944.

Yamak, A., D. Hu, N. Mittal, J. W. Buikema, S. Ditta *et al.*, 2020 Loss of Asb2 Impairs Cardiomyocyte
Differentiation and Leads to Congenital Double Outlet Right Ventricle. iScience 23: 100959.

Zachos, J. C., G. R. Dickens, and R. E. Zeebe, 2008 An early Cenozoic perspective on greenhouse
warming and carbon-cycle dynamics. Nature 451: 279–283.

Figure Legends

Figure 1. Fraser River salmon sampling locations. a) Each point on the map represents a sockeye
salmon sampling site (average of 6.5 samples per location). Each location is shown as the closest point
of the specified body of water to the Fraser River. The Fraser River and associated watersheds are
highlighted on the map. The Nechako River is highlighted as the top left watershed (darker blue, with
the river as a lighter blue), and the Thompson River is highlighted as the bottom right watershed
(green). Watershed and geographical data are from ced.org and naturalearthdata.com, respectively.
The map was generated using QGIS software. Information on sampling locations outside of the
boundaries of this map can be found in File S1. b) Same as a) for Chinook salmon samples (average of
7.8 samples per site). c) Same as a) for coho salmon samples (average of 6.9 samples per site –
excluding the Big Bar Landslide sampling site).

Figure 2. Polymorphic loci per species and runs of homozygosity. a) Counts of all SNPs identified
for each species with the same pipeline – differing only in score recalibration and reference genome
assemblies. The number of samples per species is shown below each column. b) Overlapping SNP
loci among species. SNP loci from sockeye and coho salmon were mapped to the Chinook salmon
genome assembly. The percent of loci that overlap are shown for each species relative to the number of
SNPs identified in that species. c) Boxplots of the total lengths of runs of homozygosity for each
species. d) A scatterplot of runs of homozygosity and the fraction of heterozygous genotypes per
individual. The sockeye sample on the far left is a doubled-haploid. Lines were fitted using the loess
method.

Figure 3. SNP coverage and relatedness metrics. a) A scatterplot of the relationship between the average SNP coverage and the percent of the genotypes that were heterozygous for each salmon (all species). The relationship between the two variables was modelled using the loess method. The vertical line shows 15x coverage, which was used as a threshold in other analyses. b) Same as a), but each salmon is identified by species. c) Heatmap of pairwise relatedness scores between each Chinook salmon. The salmon from a previous study are highlighted under the Illumina banner. A boxplot of SNP depth is displayed in an insert for the previous (Illumina) and current datasets (MGI). Samples from a Thompson River tributary are highlighted from the previous dataset. d) Same as c) for the coho salmon datasets.

Figure 4. Comparison of Fraser River salmon admixture groups. a) (left) Map of sockeye salmon admixture groups plotted in QGIS by overlaying the admixture ancestry raster plots from Figure S3. Each group is labelled (the colour of the group was randomly assigned). The northern coastal admixture group is outside the frame of this map. (right) Average percent of heterozygous genotypes for each of the Fraser River admixture groups. The different Fraser River admixture group names are lower Fraser River – LFR, mid Fraser River – MFR, and upper Fraser River – UFR. The number of samples (≥15x SNP coverage) and sampling sites per admixture group is shown. Weir and Cockerham's (Weir and Cockerham 1984) F_{st} are shown for comparisons between admixture groups. F_{st} comparisons between coastal (LFR) and interior clusters (MFR and UFR) are in bold text. b) Same as a) for Chinook salmon. The LFR group includes coastal sites outside the Fraser River. Significant differences in the percent of heterozygous genotypes among groups are shown with the *p-value* (two-tailed Welch's t-test). c) Same as b) for coho salmon.

Figure 5. PCA of sampling location environmental factors. a) A biplot of a PCA of environmental factors from each sockeye salmon sampling site. These included environmental factors from WorldClim version 2.1, elevation, and distance to the ocean. Environmental factors were simplified as temperature (including: annual mean temperature, minimum temperature of the coldest month, maximum temperature of the warmest month, and temperature range), precipitation (including: driest month precipitation, wettest month precipitation, and annual precipitation), elevation, and distance to the ocean. Figure S5 does not simplify these variables. Locations were coloured by their assignment to admixture genetic groups, but no genetic data was used to produce these PCA. Only sites that had average ancestry values ≥ 0.7 from the LFR, MFR, and UFR genetic groups are shown for simplicity. See Figure S5 for all locations. Each small symbol represents a sampling site, and the larger symbol represents the centre of the ellipse if drawn (ellipse level 0.4). b) Same as a), except for coho salmon. The LFR group includes coastal sites outside the Fraser River. c) Same as b), except for Chinook salmon.

Figure 6. Estimated historical effective population size of Fraser River salmon. a) Effective population size of Chinook salmon estimated using SMC++ and whole genome data. Only one sampling site from each Fraser River admixture group was retained for clarity. Figure S6 shows all sites. The estimated glacial maximum of around 19,000 years ago (Clark *et al.* 2009) is shown with a dashed vertical line. Other regions are highlighted when there were major changes in effective population size for more than one species. b) Same as a) for coho salmon. c) Same as a) for sockeye salmon. d) Estimates of ocean surface temperature (Zachos *et al.* 2008; Hansen *et al.* 2013).

Figure 7. Estimated historical effective population sizes of some admixture groups of sockeye salmon. a) Effective population size of sockeye sampling locations from the UFR admixture group

851 estimated using SMC++ and whole genome data. b) Same as a), but for sites from the MFR admixture
852 group. c) Same as a), but for locations from the northern coastal admixture group. d) Same as a), but
853 for locations from the upper Columbia River kokanee admixture group.
854
Figure 8. Genetic diversity of Fraser River salmon. a) (left) Map of sockeye salmon total length
(kb) of runs of homozygosity (ROH) averaged for each sampling site. This analysis included all
individuals. The colour scale is quantile-based. (middle-left) Barplot of total lengths of runs of
homozygosity for samples assigned to the different admixture groups with ancestry values ≥ 0.7.
(middle-right) Map of sockeye salmon nucleotide diversity (pi) of each sampling site excluding salmon
with less than 15x SNP coverage. The colour scale is quantile-based. (right) Nucleotide diversity
barplots of sampling sites with average admixture ancestry values ≥ 0.7 from individuals with at least
15x SNP coverage. b) Same as a) for Chinook salmon. The LFR group contains samples outside the
Fraser River. c) Same as b) for coho salmon.
864
Figure 9. Private allele analyses of Fraser River genetic groups. a) The count of private alleles of
Fraser River admixture groups (only individuals with ≥ 0.7 ancestry values were used). Only private
alleles identified in at least three individuals from a group were counted. Below the private allele count
is the percent that this represents of all the SNPs identified for each species. The circles are scaled by
this value. The number of samples per genetic group after these criteria were met are also included. b)
Boxplots, with jittered points added, of individual private allele counts of the different Chinook salmon
admixture groups. The individual private allele counts are the number of private alleles each individual
has from the private alleles common to the admixture group. This is a way to compare how much
individuals contribute to the identified private alleles of the group, and to be able to compare these
values to understand if sample size might influence group counts. The LFR group has individuals from
outside the Fraser River. c) Same as b) for sockeye salmon, except there were five total admixture
groups of sockeye salmon and the LFR is strictly Fraser River samples. d) Same as b) for coho salmon.
877

Tables

879
880 **Table 1. SNP filtering parameters used for each analysis.**

Analyses	Indels, >10% missing, depth	MAF 0.01	LD Filter	15x coverage	Relatedness
SMC++	all				
Rehh	all	all			
Coverage	all	all			
ROH	all	all			
Relatedness	all	all	all		
Admixture	all	all	all		
% Heterozygous	all	all		all	
Nucleotide diversity	all	all		all	
Polymorphic loci	all	all		all	
PCA	all	all	all	some*	some*

881 *See Figure S4 for all PCAs, including some with the 15x SNP coverage and relatedness filters.
882
883 **Table 2. Significant extended haplotype homozygosity among admixture groups.**

Species	Admixture	Admixture	Admixture	Nearest gene

	comparison LFR vs. MFR	comparison LFR vs. UFR	comparison MFR vs. UFR	
Chinook salmon	LG09:14.36-14.37	LG09:14.36-14.37		CSGALNACT1*
	LG10:5.22-5.23			ENDOD1*
	LG11:36.37-36.38			MACO2*
	LG21:43.68-43.69			CTNNA1
			LG01:54.33-54.34	HCA2
			LG07:46.47-46.48	SRGAP3*
			LG10:2.25-2.26	AIMP1B*
			LG13:37.80-37.81	EPHA7*
			LG13:37.98-37.99	LOC121838999
			LG18:25.30-25.31	METTL24*
			LG18:33.39-33.40	FOSL2
			LG25:28.04-28.05	LRPPRC*
			LG26:1.26-1.27	KLF12B*
			LG30:17.88-17.90	MINK1
Coho salmon	LG18:24.62-24.63			ASB2*
			LG10:24.15-24.16	PTPRD
Sockeye salmon	LG26:45.73-45.74			TAAR13C*
		LG09:23.72-23.73		TAAR13C*
			LG11:31.22-31.53	Two genes overlap
			LG04:28.48-31.57	Many genes

All locations are relative to the Chinook salmon genome assembly for comparison (see File S3 for species specific information). Format of location: Chinook salmon linkage group : start (distance in Mb) – end (distance in Mb). *Location overlaps with gene.

Supporting Information

Figure S1. Expanded maps and sampling locations. Slide 1) Sockeye salmon maps, Slide 2) Chinook salmon map, and Slide 3) Coho salmon maps.

Figure S2. Relationship between average SNP coverage and the total length of runs of homozygosity. a) The average SNP coverage per sockeye salmon sample relative to the total length of runs of homozygosity per sample. The line was determined using the loess function (local regression). b) Same as a) for Chinook salmon. c) Same as a) for coho salmon.

Figure S3. Admixture analyses. Slide 1) Average admixture ancestry values of the different admixture groups were plotted for sampling locations in QGIS using the inverse distance weighted interpolation method. The darker the sampling location, the greater the ancestry assignment to the specified admixture group. Slide 2) Plot of admixture ancestry values for clusters above and below the best supported k=5 for sockeye salmon. Slide 3) Plot of admixture ancestry values for clusters above and below the best supported k=3 for Chinook salmon. The arrows indicate that the name of a location was omitted because of space limitations. Slide 4) Plot of admixture ancestry values for clusters above and below the best supported k=3 for coho salmon. The arrows indicate that the name of a location was omitted because of space limitations.

Figure S4. PCA of each species with different filtering criteria. a) PCA of sockeye salmon. The first column is a PCA of all samples, the second column is a PCA with only one individual from pairs with high relatedness scores (≥ 0.15), and the third column is a PCA of samples with high SNP coverage (>15x coverage). The different admixture groups are noted in the PCA. Admixture numbers were arbitrary, so names were given based on geography for the first column PCA. Individuals with admixture ancestry values < 0.7, were given a different colour from the the main groups. b) Same as a) for Chinook salmon. c) Same as a) for coho salmon.

Figure S5. PCA of sampling locations based on environmental factors. Slide 1) Biplot of sockeye salmon sampling sites based on environmental variables from WorldClim version 2.1 and elevation. Sampling sites with an average ancestry assignment of less than 0.7 were assigned to intermediate genetic groups with the highest fraction named first (e.g., if a site had average ancestry fractions of 0.6 MFR and 0.4 LFR, it was assigned Mid-Coast). The Okanagan was an exception as it had high ancestry values for multiple genetic groups. Each sampling location is represented by a small point (unless there was only one site, in which case it has a larger symbol), larger points show the middle of ellipses, and ellipses were drawn at the 0.4 level. The average annual precipitation and distance to the ocean are shown on the right. Groups were combined if they shared a genetic group as the largest contributor to their ancestry values. Slide 2) Same as Slide 1, but for coho salmon, Slide 3) Same as Slide 1 but for Chinook salmon.

Figure S6. Estimates of historical effective population size for different salmon species and groups. Slide 1) Effective population size for all species and sampling sites. Some Chinook sampling locations are off the graph. These sites may have recent admixture, which can throw off these estimates. These sites were removed in all other figures. Slides 2-3) Fraser River admixture group comparisons of each species. Slides 4-5) Effective population size estimations of different groups of Chinook salmon. Individuals were grouped by admixture and region (Coastal: Northern BC Coast, Vancouver Island, Lower Fraser River; Mid Fraser: Mid-Upper Fraser River, Mid-Coast Fraser River, Mid Fraser River; Upper Fraser: Upper Fraser River, Upper Mid Fraser River). Slides 6-7) Effective population size estimations of different groups of coho salmon (Coastal: Mid Alaska, Northern BC/Southern Alaska, Southern BC, Coastal Fraser River, California-Oregon; Mid Fraser: Mid Fraser-Coastal Fraser, Mid Fraser; Upper Fraser River). Slides 8-9) Effective population size estimations of different groups of sockeye salmon (Coastal: Northern Coast, Coast-Mid Fraser, Coast-Coast Fraser, Coast Fraser; Mid Fraser: Mid-Coast Fraser, Mid-Upper Fraser, Mid Fraser; Upper Fraser: Upper Fraser – Columbia, Upper Fraser River; Columbia: Upper Columbia River Kokanee, Upper Columbia – Upper Fraser; Okanagan Lake).

Figure S7. Total runs of homozygosity for each sample location. Slide 1) Boxplot of total runs of homozygosity (kb) for all individuals from each sample location of Chinook salmon. Individuals were also highlighted by Admixture group (individuals with ancestry values < 0.7 were represented by mixes of admixture groups with the largest contributor first). The x-axis represents the sampling site (see File S1 for full name) and the y-axis represents the total length of runs of homozygosity within each individual genome. Slide 2) same as slide 1 except for coho salmon. Slide 3) same as slide 1 except for sockeye salmon.

Figure S8. Number of polymorphic loci for each sample location. a) The number of polymorphic loci per location is shown on the top for all species. The number of samples per location is shown on the bottom. b) The relationship between the number of samples per location and the number of

polymorphic loci identified for those locations. c) A boxplot of the number of polymorphic loci based
on admixture group. Locations with average admixture group ancestry values ≥ 0.7 only have one
label (i.e., lower Fraser River – LFR, mid Fraser River – MFR, and upper Fraser River – UFR). Those
with < 0.7 have two labels with the first having the largest ancestry value. Only locations with at least
four samples were used for this analysis. d) Relationship between the number of polymorphic loci and
nucleotide diversity for locations with at least four individuals per site. Each species has its own linear
regression line plotted.
962
963 **Figure S9. Extended haplotype homozygosity comparisons between LRF and MFR admixture**
964 **groups.** a) A manhattan plot of -log10 p-values from an Rsb analysis between LFR and MFR sockeye
965 salmon admixture groups. Sockeye salmon chromosomes were based on a draft genome assembly
966 submitted to the NCBI (now GCA_034236695.1). Significant peaks were indicated on the graph with
967 a vertical line and a number. b) Same as a) for Chinook salmon. c) Same as a) for coho salmon.
968
969 **File S1. Sample summary information in a spreadsheet format.**
970 **File S2. Truth SNPs used in GATK recalibration for each species (includes a readme file).**
971 **File S3. Significant extended haplotype homozygosity from all comparisons among groups.**

a Chinook salmon (Ne)
LFR
MFR
UFR

b Coho salmon (Ne)
UFR
MFR
LFR

c Sockeye salmon (Ne)
UFR
MFR
LFR

d Surface temperature (C°)
19,000

Years before present

present

Chinook salmon
Coho salmon
Sockeye salmon

a

UFR

Sockeye salmon (Ne)

b

MFR

Sockeye salmon (Ne)

c

North Coast

Sockeye salmon (Ne)

d

Upper Columbia kokanee

Sockeye salmon (Ne)

Years before present