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5  # LAGOS-US LANDSAT: Remotely sensed water quality estimates for
6  U.S. lakes over 4 ha from 1984 to 2020

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## Abstract

Broad-scale, long-term studies of water quality (WQ) are critical to understanding global-scale pressures on inland waters, yet they are rare. This data product, LAGOS-US LANDSAT, addresses this gap by providing remote sensing-derived WQ estimates from machine learning models trained on in situ data of six essential WQ variables for 136,977 lakes in the continental US from 1984-2020. The dataset includes: (a) 45,867,023 sets of whole-lake water reflectances for six individual bands and 15 band ratios; (b) 740,627 matchups with in situ data for lake WQ data for chlorophyll, Secchi depth, true color, dissolved organic carbon, total suspended solids, and turbidity; and, (c) predictions from each reflectance set for all six WQ variables across the 37 yr period. Variance explained for the predictions ranged from 20.7% for TSS to 63.7% for Secchi. Data extraction from individual scenes was quality-controlled based on cloud-cover and pixel quality, and we tested and validated key parts of the workflow to inform future water quality studies using the Landsat platform.

## Background & Summary

Contemporary regional, continental, and global water quality (WQ) assessments of inland waters are limited by the large resources and effort required to conduct in situ sampling across broad geographic extents. When investments are made in such sampling, it is generally for a snapshot in time or for only a narrow spatial footprint. Consequently, an even larger gap is the lack of historic monitoring of lakes in most of the world is lacking, particularly the global south and inaccessible regions such as boreal and arctic zones. Therefore, approaches are needed to fill these massive data gaps to understand better and forecast WQ of the world's lakes in response to global changes in the coming decades. Recently, there has been a growing interest in leveraging long-standing remote sensing platforms to obtain current and historic WQ[1,2] estimates using spaceborne sensors[1,2]. Decades of research have developed and validated methods that enhance the use of these sensors for inland aquatic studies, such as through the integration of in situ sampling with remote sensing (RS) data to extract information relevant to aquatic studies, use of band ratios for prediction, use of machine learning models, removal of non-water pixels, and cross-sensor harmonization[3–8]. Complementing these advances, there is a growing availability of open-access in situ datasets [9,10] as well as existing match-up datasets [7]. Here, we describe an approach that leverages these advances to simultaneously predict six WQ variables from RS data in lakes of the conterminous US from 1984 to 2020. Our unprecedented data product provides at least one set of estimates for the six WQ measurements in 136,977 lakes (with a median of 121,812 lakes with at least one observation for any given year).

We include ground-truthed estimates for six variables that represent components of WQ related to lake productivity and the light environment of the water column, both of which are influenced by inputs from the surrounding watershed. Chlorophyll *a* (CHL; ug/L) is the biomass of suspended phytoplankton, corrected for pheophytin, in the pelagic zone of a lake, and is a key measure of lake productivity. Secchi depth (Secchi; meters) is a metric of water clarity measured using a disk lowered into the water column until it visually disappears. True water color (true color) is determined from water samples filtered to remove particulates, such as phytoplankton and sediment and is measured with a visual comparator in platinum cobalt units (PCU) that correspond to the colored fractions of dissolved organic material in water. Dissolved organic carbon (DOC; mg/L) measured from filtered water, includes both colored and non-colored fractions of organic carbon. Total suspended solids (TSS; mg/L) include silt, clay, sand,

algae, and plant material suspended in water. Turbidity, another measure of water clarity, measures light
scattering due to suspended particles and is measured in Nephelometric Turbidity Units (NTU).
Combined, these variables provide a multi-faceted snapshot of the general WQ status of inland waters
related to common eutrophication pressures and the biological functioning of lakes, both of which are
related to external and internal forcing by land-water interactions, climate change, hydrologic forcing, and
foodweb dynamics. While CHL, true color, DOC, TSS, turbidity represent different components of
dissolved and/or suspended particulate matter in the water column, Secchi is a more generalized measure
that integrates all these components. Full definitions, source information, and method descriptions for
these six variables are described in the documentation of methods for the LAGOS-US LIMNO data
product we used for our in situ WQ data[10].

Although multiple remote sensing platforms have the potential to fill these important gaps in WQ
data, in our data product, we use the Landsat Surface Reflectance (Landsat SR; *Landsat Collection 1*
*Level-2 Surface Reflectance Science Product courtesy of the U.S. Geological Survey*) data product
because of numerous advantages it has for broad-scale studies of inland waters. Critically, it has a long
time period for data collection (since mid-1980's), global coverage, moderate spatial resolution, and
comprehensive and standard image processing. Specifically, LAGOS-US LANDSAT uses data from the
launch of Landsat 5 in 1984 until our data extraction process began in mid-2021, a time interval that also
encompasses Landsat 7 and 8. Previous studies, particularly conducted on US lakes, have already made
significant progress in advancing the needed methods in data integration and modeling of WQ to leverage
this platform for broad-scaled studies such as ours[1,7,11–13] (**Table 1**). For example, AquaSat[7] for the
conterminous US, is an open-access dataset of match-ups, i.e., surface reflectances from Landsat SR and
in situ WQ observations collected within 1 day or less. In addition, predictive models to estimate specific
WQ components, such as Secchi depth[11] and a generalized trophic state index[13], have also been created at
the US national scale. These databases are supported by well-documented and fully reproducible
methods that enable other researchers to both use their data and build new applications off them. Other
broad-scale data products have more more limited reusability, such as ground-truth Secchi predictions
provided as 5-year averages for over 10,000 Chinese lakes across the Landsat platform period[14].
Additional global datasets that focus more on addressing scientific questions rather than producing
reusable data products have been published for Secchi[15] and for algal bloom metrics[15–18].

Because the use of Landsat SR for broadscale WQ predictions is still relatively new as well as
being computationally intensive, it is important to test which workflow variations to ultimately improve
ground-truthed predictions of WQ because seemingly minor decisions can take months to implement
when conducted on hundreds of thousands of images across space and time. Based on a synthesis of
methods available to date, we identified seven major steps in the workflow for consideration: retrieval of
whole-lake reflectances versus single pixels or buffers around a single pixel; the use of band ratios
calculated on a pixel-specific basis versus calculations of ratios after extraction of band values for all
target pixels; reflectance quality such as atmospherically-overcorrected pixels; in situ WQ data
transformations; individual bands and ratio combinations; model type; and the optimal matchup window
length for predictive models. We present the results from our testing of these steps to guide future studies
in the use of Landsat SR for inland water studies.

Our study was designed to build on prior studies to advance the use of Landsat SR for broad-scale
studies of inland waters by making our WQ predictions and the intermediate data publicly available. Our
advancements include: (1) testing computationally expensive components of the workflow to determine
when and where they are needed to improve predictions; (2) creating a single data product that includes

all components into a single data product: reflectances, match-ups with in situ data, ground-truthed
predictions, and data quality flags; and, (3) creating and testing the validity of using Landsat SR for a
broader range of WQ variables[19].
Our final data product strives to meet all the FAIR data principles of findability, accessibility,
interoperability, and reusability[20] and includes extensive validation of predictions, methods, and
recommendations for specific processing to maximize the predictive success of estimates of WQ from the
Landsat Surface Reflectance product. Our WQ estimates of the six variables include multiple dates within
the year over a maximum 37-year time series from March 1984 - December 2020 derived from over 45
million retrievals. Another key advantage of our data product is its interoperability with an existing US
lake research platform (LAGOS-US) that provides millions of data points on other critical features of
lakes that are related to WQ including delineated watersheds, regional characteristics, land use, lake depth
and many other variables[21–23]. Our data and methods are extensively documented to enable users to use
the data in many ways and to reproduce results in the future.
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## Methods

We followed established approaches to leverage the Landsat SR data product including using the
Google Earth Engine (TM) platform (GEE) to extract reflectances for over 100,000 lakes. However,
because many specific methods have not been thoroughly evaluated, we also tested some key components
of a workflow that uses in situ data to ground truth remotely sensed water quality estimates that we
consider critical to help future researchers using the Landsat SR data product in GEE. Our overall
approach is graphically displayed in Figure 1, which displays the four main steps described in detail next.
151

*(1) Process Landsat SR scenes*
We used the Tier 1, Level-2 Landsat Surface Reflectance product (Collection 1), which employs
the Landsat Ecosystem Disturbance Adaptive Processing System (LEDAPS)[24] to atmospherically correct
scenes from Landsat 5 (TM) and 7 (ETM+) and the Land Surface Reflectance Code (LaSRC)[25] to
atmospherically correct scenes from Landsat 8 (OLI). We extracted surface reflectance values for all
137,465 LAGOS-US lakes $\geq$ 4 hectares in surface area[21], from 1984-2020 (1985-2020 complete years)
using GEE (**Figure 1,** step 1). We selected 4 ha as our minimum lake size because any smaller area
results in large numbers of lakes with pixels that have mixed water and land components, making it
challenging to capture true water-only pixels. Because lakes were run as separate instances and the default
average concurrent tasks allowed on GEE at the time of execution was two, the total runtime for all lakes
took over a year to complete (March 2021 to April 2022). Extractions for the full time series of
reflectances for the largest lakes in our dataset required GEE tasks that ran over 0.5 hours (e.g., Lake
Okeechobee, FL at 1,715 $km^2$ had a runtime of 32 minutes), which added substantial processing time.
For each lake and for each sensor overpass and band, we extracted the median, minimum, and
standard deviation of individual pixel reflectance values, and band ratios calculated at the per-pixel level
before a lake-wide median was extracted. We removed pixels coded as clouds, cloud shadows, snow/ice,
or fill data, retaining only clear-sky pixels over water using the CFMASK-derived pixel quality
attributes[26] in the Landsat Pixel Quality Assessment Band. To enable historical time series analyses across
the full study period we harmonized bands across the TM, ETM+, and OLI sensors, for which we used
complete handoff following the launch of a new sensor, using published surface reflectance
transformations[6] prior to reflectance extraction.

As part of our data product, we created the data table, *LAGOS_US_LANDSAT_compiledRS*,
which contains output from the steps described in Figure 1a to aid in later steps and to make data
available to other users for their modeling purposes in the future. Each row in the data table represents a
unique lake-scene combination and includes: median lake-wide values for the 6 bands available for all
sensors, pixel-specific ratios for all possible band combinations (n=15), the percent cloud cover of each
scene, the total number of pixels extracted, and a 3BDA-like "KIVU" algorithm applied to the individual
pixels[27]. We also created data flags for: (1) any pixel or any band with negative reflectances; (2) any lake
median reflectance that is negative; (3) the percent pixels retrieved for a lake; and (4) duplicate scenes on
the same day for a lake (rare, but present due to overlapping overpasses). Our final data product
represents 136,977 lakes (99.6% of LAGOS-US lakes $\geq$ 4 hectares) that had at least one set of Landsat
reflectances and provides 45,867,023 reflectance sets of lake and overpass combinations through time.
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*(2) Build a matchup training dataset for six water quality variables*
Our matchup training dataset **(Figure 1, step 2)** is based on the LAGOS-US LIMNO data
product[10] that contains in situ data extracted primarily from the Water Quality Portal and three sampling
years from the USEPA National Lakes Assessment. We conducted quality control analyses on each of the
six WQ variable training set and made the following data decisions. First, for CHL, we removed negative
values and extremely large values > 800 ug/L that are rarely measured in pelagic zones of lakes. Second,
if a lake had multiple measurements for a given WQ variable on a calendar day, the median value was
calculated prior to matching with remote sensing data. Third, because the few observations for true color
> 100 PCU appeared to skew the distribution, we removed them from the training set, recognizing that
our resulting models would likely not be able to capture the most extreme high range of true color.
Nevertheless, because lakes are considered tannic if they have values > 20 PCU[28] and highly colored
lakes were rare, we made the decision to apply a data limit in order to produce more accurate training
models. The ranges of data for all other variables were less skewed and were not trimmed.
After merging the remote sensing dataset with LIMNO on a calendar date basis (with a maximum
matchup window of +/- 7 days), we had 723,206 training matchups of lake-by-date combinations with at
least one value for any of the six WQ variables across 13,756 LAGOS-US lakes. The number of training
data matchups (including multiple values by date per lake) by WQ variable were: CHL = 238,248; Secchi
= 666,060, DOC = 8,269; true color = 37,839; TSS = 52,677; turbidity = 68,573. Matchup lakes were
widely distributed across the conterminous US with spatially high densities of observations related to
regional lake density and monitoring program specifications **(Figure 2)**. Despite the availability of
substantial matchup training data over time, the likely effect of monitoring intensity combined with the
data retrieval process followed to construct the LIMNO module[10] is the number of training matchups
increased across the study period with a peak around 2008  **(Figure 3)**. The complete set of matchup data
is provided in the data table *LAGOS_US_LANDSAT_matchups*.
209
*(3) Test key components of the workflow for each water quality variable*
The major methodological steps that required testing were related to image and pixel processing,
modeling, and matchup windows **(Figure 1, step 3)**. Using the matchup training dataset, we conducted
several exploratory assessments to test key workflow components that were used to train the models for
each of the 6 water quality variables. Note that because decisions were interrelated, we did not perform all
analyses in a simple linear order. All training models used 5-fold cross-validation, reserving 20% of the
data for validation. We conducted most of these tests on all water quality variables. However, due to the

intense computational nature of some workflow components, particularly those conducted in GEE, for
some tests, we selected one to three water quality variables. For example, using a randomly selected lake
matchup training dataset of 500 lakes with in situ DOC data, we ran a set of 40 training models that
varied the image and pixel processing steps, transformations, and matchup windows. In addition to
identifying the best-performing model type, we tested the influence on prediction quality for the
following model inputs (**See Figure 1, step 3**): 5 types of pixel retrievals; 2 types of band ratio
calculations; 2 types of scene quality indicators; and inclusion of all bands or all bands plus all ratios.
Because many components that we tested have not been analyzed in any published studies to date, the
outcomes of these tests are informative not only for our final model selection but also for future studies
using the Landsat SR platform for broadscale water quality studies.
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<u>Key outcomes of the workflow tests</u>
Based on all analyses described above we summarize our main recommendations for building remote-
sensing models across hundreds of thousands of lakes using the Landsat platform for these six variables
(Table 2):
- **The machine learning model** performed best. We tested three types of training models that have
been used by other studies using the CHL training dataset: linear models with all bands plus band
ratios; the three-band KIVU retrieval algorithm[29] that was previously found to be relatively
effective for inland water chlorophyll retrieval[27]; and random forest regression models using all
bands plus band ratios. We found that the random forest regression training model, applied using
the 'randomForest' package in R[30] far outperformed the other two models with over two times
reduction in error compared to the other two models (**Figure 4**). We found similar gains in
performance for Secchi data (data not shown). Based on this test, all of the remaining tests were
performed only using the machine learning model, a decision supported by other recent studies
documenting the superior performance of machine learning models[11]. *Tested for CHL.*
- **Pixel-retrieval calculated at the whole-lake scale** performed best. The lake center with a buffer
performed close to the whole-lake estimates while the lake center calculated using the Chebyshev
point was consistently better than the lake centroid (**Figure 5a**). However, because these
calculations were performed within GEE, they were computationally slow, preventing us from
making comparisons for the entire dataset. *Tested for CHL, Secchi, and DOC subsets.*
- **Calculating band ratios at the pixel level** was always better, i.e., with lower prediction error,
than first calculating lake-wide or buffer-wide band averages and then calculating ratios from
those (Figure 5a). Because this step must also be conducted within GEE, it was critical for
consideration early in the workflow. *Tested on all WQ variables.*
- **Models that removed scenes based on two quality flags** related to cloud cover and negative
reflectances, performed best for all variables except DOC (Table 2; Figure 5a). We speculate that
the DOC models retained these scenes because of the low sample size (the DOC training dataset
was the smallest of all of the WQ variables). However, the differences in this test were not large,
so when sample sizes are low, retaining some scenes with poorer quality may be better because
sample size is increased. In general, these poor-quality images were relatively rare, so the gain in
predictive performance was not large across many lakes. *Tested on all WQ variables.*
- **Log-transformed in situ data** resulted in better-performing models than their non-transformed
versions (Table 2, final column). Because the distributions of most of our WQ variables were left-
skewed and more data-rich in the lower ranges, the models performed better when the

transformations resulted in greater granularity at the lower ends of the data distributions. *Tested on all WQ variables.*

- **Using all bands and all band ratios** outperformed RGB-only models. Although we present results only for DOC (**Figure 5a**) because machine learning models are not penalized for including more predictors, we felt that this result applies to all WQ variables. The variable importance plots for the final models for all WQ variables demonstrate that although the non-RGB bands were never the top predictor, ratios containing one of these bands were often the second or third most important variable (see Figure 6). *Tested on all WQ variables.*
- **The optimal matchup window ranged from 1 to 7 days and should always be tested.** We found the optimal matchup window differed by WQ variable and was likely due to the inherent daily variability in the WQ variable (**Table 2; Figure 5b**). For example, true color is a more temporally stable variable for many lakes than CHL, TSS, and turbidity which can all respond readily to large precipitation or wind events day to day. Consequently, the optimal matchup window for the latter 3 variables was 1 day but was 7 days for Secchi, true color, and DOC. Nevertheless, we encourage users to test this important part of the workflow in future applications. *All WQ variables.*

## *(4) Build separate predictive models for each water quality variable*

The final model type and all workflow steps followed to build separate random forest regression models for each WQ variable are described fully in **Table 2** with data decisions described in each row. Final models were based on in situ training data that were $\log_{10}$-transformed (except for Secchi). Variance explained of predicted versus in situ data, a valuable model diagnostic, ranged from a low of 20.7 % for TSS to a high of 63.7 % for Secchi, ranging between 42-52 % for the remaining WQ variables. Thus, except for TSS, the variance explained is quite high and certainly suitable for broad-scaled water quality studies. All random forest models included all bands and ratios. We found that the top predictors were always a 2-band ratio based on variable importance factors calculated for each WQ model (**Figure 6**) like other studies demonstrating the importance of the blue/red ratio and blue/green ratios for Secchi and CHL. However, fewer studies have also included the SWIR bands, which were also quite important. Interestingly, the top predictor for true color was the Green/Red ratio, which was also the second highest ratio for closely related DOC, although, for DOC, the top predictor was Blue/Red, which was also top for Secchi and CHL. These results make ecological sense since DOC is related to both true color (which is the colored fraction of DOC) as well as CHL which is known to track with the non-colored fraction of DOC. Turbidity and TSS have similar dominant predictors, although the contributions of the other wavelengths and ratios differ suggesting they reflect distinct WQ features.

An important component of our data product is the inclusion of quality-control flags that allow the user to decide which data to use based on the requirements of their study. We created 5 data flags in our final dataset: Negative reflectance of any band, negative reflectance of the median of any band; the % of maximum lake pixels with quality reflectances (in other words, if 95% of the lake was covered by a cloud that day and so only 5% of pixels were retrieved, we flagged this row to be of suspect quality); whether there is a date in which two satellite overpasses resulted in two predictions (because this will affect queries and sample sizes, we indicate whether this has occurred due to overlapping rows in Landsat); and the flag *QAQC_recommend*, which selects the highest quality predictions based on the above flags. Therefore, users can easily query the dataset to select predictions that meet the following

304    quality criteria: > 10% of lake pixels retrieved; no negative median bands included; and duplicate days
305    removed and the scene with the highest % lake pixels selected as the single value.
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307    **Data Records**
308    The LAGOS-US LANDSAT data product[19] consists of the following four data tables that include all
309    components of our workflow described in Figure 1.
310
311    **LAGOS_US_LANDSAT_data_description.csv**
312    This table is the combined data dictionary for the three tables below that describes and defines variables
313    included in each of the four data tables. For each variable, we include the variable name, description, data
314    type, and column index.
315
**LAGOS_US_LANDSAT_compiledRS.rds**
This data table, in the form of an R Data Serialization (RDS) file, provides reflectance extraction output
generated from Google Earth Engine that was extracted from a single lake at a specific time. Each row is
a single LAGOS lake and Landsat scene combination. This data table contains 57 variables related to
whole-lake summaries of reflectance values calculated for lake water pixels on all individual bands (min,
max, median, standard deviation) and all 2-band ratios in addition to the KIVU calculation.  Ratios are
first calculated at the pixel level (referred to as pixel-wise in the data dictionary), and then the median is
calculated across all lake-water pixels. We also include scene identifier information from the Landsat
platform (satellite name, sensing time, scene ID, etc), quality estimates obtained from the Landsat
platform (e.g., % cloud cover of the scene, image quality indicators), earth-sun distance, solar angles, etc.
326
**LAGOS_US_LANDSAT_matchups.csv**
This data table includes all possible matchups within 0 to +/- 7 days between Landsat overpasses and in
situ data obtained from the LAGOS-US LIMNO[10] data module for the six WQ variables (CHL, Secchi,
True color, DOC, TSS, Turbidity). For reference, we include the WQ variables from individual dates and
the associated reflectance values and scene identifier information from the above dataset. If any date had
more than one WQ estimate, we took the median value, however, more than one estimate was extremely
rare, and the vast majority of data were from a single sampling event per day. We also include the
matchup window value for each WQ in situ value so that future users can select the matchup window.
Note that not all WQ variables are sampled on the same dates, so the sample sizes vary by WQ variable.
336
337    **LAGOS_US_LANDSAT_predictions_v1_QAQC.csv**
338    The predicted WQ data for all combinations of date-reflectance sets. As described above in step 4 of our
339    workflow, we include five data flags including the field QAQC_recommend in which we apply our
340    criteria for quality predictions that users can select for a highest-quality version of the dataset.
341
**Technical Validation**
The validity of using Landsat 5, 7, and/or 8 sensors for inferring water quality in lakes has been
supported by decades of studies conducted on individual and combination of sensors[3,11,31–35]. However,
challenges remain that careful methods development can address. For example, the spectral signals of
WQ variables are not always independent; for example, colored dissolved organic matter and chlorophyll
have overlapping absorption spectra[36]. Further, Landsat imagery may have lower accuracy for estimating

WQ values in oligotrophic relative to eutrophic waters[37]. Therefore, we used a series of analyses to
characterize the level of uncertainty in the LAGOS-US Landsat data and identify potential biases. In
addition to quantifying numerous fit statistics and conducting 5-step cross-validation that supports the use
of the six training models that we developed (Table 2), we also investigated the ecological interpretability
of the predicted water quality estimates.
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*Model prediction error*
In addition to examining the overall model fits that quantify model error across the whole dataset,
we examine quantiles of predictions to examine whether errors differ across the range of the matchup
dataset (**Figure 7**). We found that in most comparisons, 90% of the data were close to the 1:1 line and that
differences between predicted and in situ data are not that far off ecologically, especially when one
considers predictions of water quality were made in over 100,000 lakes across 34 years. The model with
the poorest performance was TSS (Table 2), which had the largest bias in predictions, with the median
TSS being underpredicted relative to in situ data. Plots of error by binned values of predicted values also
show that TSS has higher errors, particularly for high TSS lakes. We also found error by binned values
for DOC to also be particularly high for the highest DOC values (> 16 mg/L). A study of global lake
DOC estimates predicted the average DOC was 3.9 mg/L, with a maximum value of 27.0 mg/L[38].
Therefore, while our modeled DOC values are more accurate for values < 16 mg/L, predictions at rarely
observed higher values have more error.
We also examined model prediction error for an individual lake for which we had good temporal
data in the matchup dataset with a long time series of both *in situ* sampling and error-free imagery. First,
we looked at Secchi data for Lake Okeechobee, a well-studied lake in Florida that contained 10,158
matchups within the ±7 window with satellite overpasses with 2,989 Secchi measurements. In this
comparison, higher prediction errors occurred in concert with known issues with imagery data that we
have flagged for in our database. Secchi was overestimated when the proportion of lake pixels available
was low (< 10%; **Figure 8a**), an effect that is removed by eliminating such predictions from analysis.
Second, in the same set of data from Lake Okeechobee, we found that negative reflectance values also led
to prediction errors (**Figure 8 b-c).** Removing date-image combinations flagged with these errors results
in the majority of the predicted versus in situ data falling close to the 1:1 line (**Figure 8d).**
377
*Comparison of time trends for selected lakes for CHL and DOC only:*
Further validation of the LAGOS-US LANDSAT dataset comes from a lake which has
documented changes in water quality following a deliberate management intervention. The eutrophic
Lake McCarrons in Minnesota was treated in 2004 with aluminum salts to reduce internal phosphorus
loading, and, consequently, decreased algal growth[39]. The comparison between predicted and *in-situ* data
for CHL suggested first, a close relationship near the 1:1 line (Figure 9a). Second, the alignment between
the in situ data and predicted values improved when a narrower matchup window was applied (Figure 9b-
c). However, the most striking was that the timing of the dramatic decrease in the in situ CHL was
matched exactly in the predicted values from the remote-sensing models (**Figure 9d**). This lake thus
presents an excellent case study to examine the application of our data product to individual lake studies
beyond our primary goal of facilitating regional and continental-scale research.
Finally, we demonstrate how our remote-sensing-based prediction dataset can be used to extend
time series in lakes for which in situ data are missing. For example, DOC is not sampled as commonly as
other water quality variables and yet is an important regulator of the ecology and biogeochemistry of

inland waters, particularly related to carbon budgets[40]. Beaver Lake, Arkansas has a robust time series
beginning in 2002 for in situ DOC data that aligns well with predicted DOC (**Figure 10a**). With our
remote-sensing predictions, we can extend that time series back in time an additional 18 years (**Figure**
**10b**). These individual lake studies lend strong support for the potential for greatly expanding the
estimates of water quality for individual inland lakes across the US in lakes with incomplete records or
that have never been studied to date. Being able to estimate water quality for lakes across the entire US is
an urgent priority to ensure that people have equitable access to good water quality regardless of
socioeconomic status, race, ethnicity, and other factors. A recent study has shown that in situ estimates of
water quality from government agencies and community science programs are not equitably distributed
across all communities based on race and ethnicity[41]. Remotely sensed water quality predictions such as
ours can help fill this gap in sampling.
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**Usage Notes**
In addition to the data values, our prediction dataset includes multiple measures of quality that
provide users with the information needed to apply more lenient or stricter data quality filters: (1) if any
pixel in any band for a lake retrieval had a negative reflectance value, (2) if any band for the whole lake
retrieval had a negative median reflectance value, (3) the percent of pixels retrieved out of the maximum
number of pixels ever retrieved for the lake across the prediction time series, and (4) if the retrieval for a
lake shares the same calendar date as other retrievals due to overlapping scenes. The first two  flags
represent overcorrection over lake water pixels or other scene artifacts while the third provides an
estimate of the extent of the lake surface represented in the predictions. Of all predictions, 3,145,912
(6.9%) were flagged for any negative reflectance pixel while 410,065 (0.9%) were flagged for median
negative reflectance across the whole lake. Finally, because a lake can be within multiple Landsat row
and path combinations due to scene overlap, we additionally flag predictions that have multiple retrievals
for a single calendar date since users may want to treat these data differently than predictions for a single
day. We retrieved an average of 48.0% of all pixels within a lake polygon, with 37,529,125 lake-scene
combinations meeting or exceeding our 10 percent pixel threshold (81.8%). Finally, 4,877,886 lake
reflectance sets had shared calendar days for a lake due to scene overlap (1.1%). We recommend that
users select the following criteria as a starting point for a good quality dataset (found in the
QAQC_recommend variable with the TRUE value), which combines removal of median negative
reflectance sets and a threshold of least 10% of pixels being retrieved, alongside removing duplicate day
predictions. The dataset filtered in this way includes a total sample size of 35,794,821 lake and date
combinations and includes 78.0% of the total prediction sets. Another layer of quality control that users
may want to consider is to drop predictions for DOC that exceed 16 mg/L and TSS that exceed 80 mg/L
(Figure 7).
The location and density of predictions for the six WQ variables are well-distributed across the
conterminous US (**Figure 11**). Due to cloudiness that is higher in the eastern US compared to the western
US, the highest densities of prediction data occur in spatial banding patterns that result where Landsat
overpasses overlap, and are more likely to contain cloud-free scenes that are known to occur with the
Landsat platform[42]. A decrease in the spatial banding of the data can be seen from east to west and is
likely a function of clouds. Because our dataset began with all lakes $\geq$ 4 ha, it is close to a census
population of lakes, with the majority of lakes having at least 1 scene. Observations are also distributed
across time (**Figure 12**), with peak numbers of scenes occurring in April to May and then October

through November - again, likely a result of general patterns of clouds across the US. The number of
images extracted through time increases through time as well, with the largest retrievals per year
occurring after around 2015 that contain overlapping satellites that increase the potential for cloud-free
scenes. Nevertheless, over 70,000 lakes across the US contain at least one summer water quality estimate
across the full time period, resulting in an unprecedented time series dataset for lake WQ.

The ranges of the matchup training dataset include broad ranges of all six WQ variables (**Figure
13; Table 4).** The prediction dataset generally had narrower distributions of the data as shown by the
quantile plots that showed less spread in the data. However, given that we know that most lakes in the US
have never been sampled (Shuvo et al. 2023), we would not have expected the ranges of the matchup
training dataset to match the prediction dataset closely given the biases in water quality sampling
programs[43,44]. Examining sampling sizes through time is also valuable. The median number of
observations per lake is 257 (Table 4); the median lake-year combination with WQ observations is 37;
and the median number of lakes per year with observations is 121,812.

Although there is overlap in reflectances that these six WQ variables are related to (e.g,
overlapping absorption spectra for CDOM and chlorophyll documented by Olmanson et al. 2015), the
dominant factors that these WQ variables represent are distinct enough that they are predicted by different
bands and band ratios (Figure 6), and, despite very high sample sizes, they are not highly correlated with
the exception of Secchi depth which is highly correlated to everything except true color (Figure 14). In
fact, these correlations match expected ecological relationships. For example, Secchi and turbidity, both
measures of water clarity, are the most strongly correlated in the dataset (r=0.90). In contrast, the
correlation between true color, which measures dissolved fractions, and TSS, which measures suspended
particles, is the lowest among all comparisons (r=0.3).

To examine relationships among the variables further, we also applied a common threshold for
eutrophication that separates oligo-mesotrophic lakes from eu-hypereutrophic lakes (CHL ≤ 7 ug/l) and
examined relationships among CHL, Secchi, and DOC. The patterns ( **Figure 15)** make ecological sense
based on our current understanding of key limnologic relationships [28,45]. First, DOC is positively related
to true color in both sets of lakes, but DOC at a given color value tends to be lower in oligo-mesotrophic
lakes than in eu-hypereutrophic lakes. These latter lakes have the highest DOC values in the dataset,
particularly where with true color > 20 suggesting high contributions from both non-colored DOC from
high algal growth and colored DOC from tannins that occur in high color waters. Second, Secchi versus
true color also diverges in the lakes in the two trophic classes, with a far clearer relationship between the
two variables in oligo-mesotrophic lakes compared to eu-hypereutrophic lakes suggesting a shift in
control of water clarity from color to algal biomass. Finally, the importance of Secchi and its relationships
with all other variables is demonstrated in **Figure 16**. Secchi is the most commonly sampled WQ variable
in US lakes[10], and its inclusion in nutrient modeling of lakes has been found to decrease prediction
error[46,47]. Being able to predict Secchi in an even broader range of lakes has great potential for improving
water quality modeling at broad scales.

To our knowledge, LAGOS-US LANDSAT is the first published dataset for nearly a census
population of lakes ≥ 4ha (over a hundred thousand lakes) at the continental scale that includes: ground-
truthed remote-sensing derived estimates of six major WQ variables; data on lakes as small as 4 ha, with
50% of observations coming from lakes < 9 ha; water pixel extractions at the whole-lake scale; all parts
of the workflow products in a single product; and interoperability with a major existing research platform,
LAGOS-US. This research platform for studying lakes of the US across both space and time includes
most of the necessary geospatial and temporal datasets for macroscale studies of lakes (e.g., watershed

delineations, land use/cover, geology, soils, climate, stream connections, and many others) and the
LANDSAT data module adds extensive and unprecedented temporal data for water quality up to 37 years.
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484  **Code Availability**
485  The entire code, including the code used for extracting reflectances in Google Earth Engine, is available
486  in Zenodo at https://zenodo.org/records/11110284
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499  **Author contributions**
500  PJH had the initial idea and developed and implemented the analytical approach for this dataset. All
501  authors contributed to making decisions about the modeling approach as well as outlining and writing the
502  article. PJH created all figures and tables except for Figure 1 and Table 1-2 (PAS) and Figure 12, 14-16,
503  and Table 4 (KEW).
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506  **Competing interests**
507  The authors declare no competing interests.
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# Figures

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515 description of all steps.

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524 LIMNO in situ observations for each of the six water quality variables. The map outlines are NEON
525 region borders.

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processing for a subset of the DOC dataset. (a) A plot of variance explained for each training model run
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for the 5 different ways that reflectance is summarized by lake (see text). 'Ratios' are ratios calculated on
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560 **Figure 6.** The model variable importance (mean decrease Gini) for all predictors in the random forest
561 regression models for each water quality variable. Orange bars are individual bands, and purple bars are
562 ratios between two bands.

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and Y axes is trimmed to the 90th percentile of in situ data (CHL = 53 ug/L; Secchi = 5.3 m; true color =
63 PCU; DOC = 13 mg/L; Turbidity = 19 NTU; TSS = 133 mg/L). The dashed red line is the 1:1 line.
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633
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635 (left plots) and from the prediction dataset (right plots). Black lines are median values.
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640 **Figure 14.** Correlation matrix of predicted values for the six water quality variables. Data points were
641 randomly selected single summer (June - September) predicted values for each year from 1984-2020 for
642 the 20,191 lakes with 37 years of data. Note, all WQ variables were from the same scene for a lake year.
643 Spearman rank correlations were based on $\log_{10}$ transformed data. All p values << 0.00001. Color is true
644 color.
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660 log$_{10}$ scale. Data points were randomly selected single summer (June - September) predictions for each
661 year from 1984-2020 for the 20,191 lakes with 37 years of data. Note, all WQ variables were from the
662 same scene for a lake year.
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# Figure Legends

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# Tables

**Table 1.** Synthesis of published studies that use Landsat for macroscale studies of lake water quality

| Characteristics of macroscale lake studies using Landsat | | | | | | | | OA data product includes | | | | |
|---|---|---|---|---|---|---|---|---|---|---|---|---|
| Country or spatial extent | Name of resource | Reference | Temporal extent (maximum possible) | Min lake size | Number of lakes | Aquatic variables of interest | Study type | Reflectances | Matchups | Ground-truthed predictions | Calculated values from reflectances | Methods documentation |
| Global | NA | Fang et al. 2022 | 1982-2018 | > 100 ha | 176,032 | Algal blooms (FAI) | Analysis | | | | ✔ | No (Only methods section; no code) |
| Global | Global Bloom Database | Hou et al. 2022 | 1982-2019 | > 10 ha | 248,243 | Algal blooms (FAI) | Analysis | | | | ✔ | No (Only methods section; no code) |
| Global | NA | Song et al. 2022 | 2019 | > 1 ha | NA | Secchi | Model development | | | | | No (Only methods section; no code) |
| Global | NA | Maciel et al. 2023 | 1984-2022 | NA | > 300 | Secchi | Model development | ✔ | | ✔ | | Partial (Docum. & code available, but not reproducible) |
| China | Annual inland water clarity dataset of China | Tao et al. 2022 | 1984-2018 (5-yr means) | > 1 ha | 10,814 | Secchi | Dataset | | | ✔ | | Partial |
| Continental US | AquaSat | Ross et al. 2019 | 1984-2019 | NA | Sites, not lakes | TSS, DOC, CDOM, CHL, Secchi | Dataset | ✔ | ✔ | | | ✔ |
| Continental US | LimnoSat-US | Topp et al. 2020 | 1984-2020 | > 10 ha | 56,792 | NA | Dataset | ✔ | | | | ✔ |
| Continental US | NA | Topp et al. 2021 | 1984-2018; 5-yr means | > 10 ha | 14,000 | Secchi | Analysis | ✔ | | ✔ | | ✔ |
| Continental US | NA | Topp et al. 2021 | 1984-2020 | > 10 ha | 26,000 | Spectral color | Analysis | ✔ | | | ✔ | ✔ |
| Continental US | LTS-US | Meyer et al. 2024 | 1984-2020 | > 10 ha | 55,662 | Trophic state index (TSI) | Dataset | | | ✔ | | ✔ |
| Continental US | LAGOS-US LANDSAT | This study | 1984-2020 | > 4 ha | 136,977 | Secchi, CHL, DOC, True color, TSS, Turbidity | Dataset | ✔ | ✔ | ✔ | | ✔ |

763 **Table 2.** LAGOS-US LANDSAT workflow decisions and fit statistics for the final model for each water quality predictive model.
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| | Workflow components tested and selected | | | | | | | | | Final model fit statistics | | | | |
| | Image and pixel processing by lake | | | | | Modeling steps | | Matchup training data | | | | | | |
| Variable | Pixel retrievals | Band ratio calculation | Scene quality: Negative reflectances | Scene quality: Scene clouds > 50% | In situ data transformed | # of bands and ratios | Training model type | Matchup window ( ± days) | Final matchups (N) | Variance explained (%) | RMSE | MAPE (%) | MAPE (%) Improvement of $\log_{10}$ model |
| CHL | Whole-lake | Pixel-specific | Removed | Removed | $\mathrm{Log}_{10}$ | All bands & ratios | Random forest | 1 | 43,755 | 44.1 | 0.43 | 23.3 | (+9.8) |
| Secchi | Whole-lake | Pixel-specific | Removed | Removed | none | All bands & ratios | Random forest | 7 | 586,368 | 63.7 | 1.23 | 22.8 | -- |
| True color | Whole-lake | Pixel-specific | Removed | Removed | $\mathrm{Log}_{10}$ | All bands & ratios | Random forest | 7 | 33,194 | 48.6 | 0.26 | 7.5 | (+12.7) |
| DOC | Whole-lake | Pixel-specific | Retained | Retained | $\mathrm{Log}_{10}$ | All bands & ratios | Random forest | 7 | 7,466 | 42.0 | 0.24 | 10.2 | (+6.1) |
| TSS | Whole-lake | Pixel-specific | Removed | Removed | $\mathrm{Log}_{10}$ | All bands & ratios | Random forest | 1 | 10,845 | 20.7 | 0.62 | 27.0 | (+16.3) |
| Turbidity | Whole-lake | Pixel-specific | Removed | Removed | $\mathrm{Log}_{10}$ | All bands & ratios | Random forest | 1 | 58,999 | 52.1 | 0.37 | 8.8 | (+20.9) |

**Table 3.** Statistical summary table of predictions for the six water quality variables, along with summaries of predictions per lake, and per year. Lake area was from LAGOS-US LOCUS [23].

| Variable | Units | min | q5 | q25 | q50 | q75 | q95 | max |
|---|---|---|---|---|---|---|---|---|
| *Water quality variables (37,251,456 predictions)* | | | | | | | | |
| **CHL** | µg/L | 0.04 | 3.1 | 5.7 | 8.7 | 14.0 | 29.1 | 299.9 |
| **True color** | PCU | 0.1 | 8.7 | 16.4 | 23.9 | 32.4 | 44.7 | 88.1 |
| **DOC** | mg/L | 0.5 | 2.8 | 4.3 | 5.7 | 7.1 | 9.7 | 67.7 |
| **Turbidity** | NTU | 0.05 | 1.2 | 2.4 | 4.4 | 9.3 | 21.2 | 197.8 |
| **Secchi** | meters | 0.02 | 0.7 | 1.2 | 1.9 | 2.7 | 4.3 | 59.3 |
| **TSS** | mg/L | 0.5 | 4.8 | 9.1 | 14.2 | 24.3 | 51.6 | 675.9 |
| *Lake-related (136,877 lakes)* | | | | | | | | |
| **Predictions per lake** | N | 1 | 32 | 160 | 255 | 361 | 568 | 1377 |
| **Lake–years with predictions\*** | N | 1 | 13 | 32 | 37 | 37 | 37 | 37 |
| **Lake water area** | ha | 4 | 4 | 6 | 9 | 19 | 116 | 338,950 |
| *Year-related (37 years)* | | | | | | | | |
| **Number of lakes per year** | N | 108,158 | 110,454 | 117,689 | 121,821 | 123,719 | 125,928 | 126,177 |
| **Number of predictions per year** | N | 534,405 | 604,508 | 938,970 | 1,021,304 | 1,066,120 | 1,317,022 | 1,455,641 |

\*The percentages of lakes with 37 and 30 + years with at least 1 summer obs (June-Sept) are, respectively, 14.8% and 51.6%

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