In defense of the Type I functional response:
The frequency and population-dynamic effects of feeding on multiple prey at a time

Mark Novak[1]*, Kyle E. Coblentz[2] & John P. DeLong[2]

[1]Department of Integrative Biology, Oregon State University,
Corvallis, Oregon, 97331 USA
[2]School of Biological Sciences, University of Nebraska - Lincoln,
Lincoln, Nebraska 68588 USA

RUNNING TITLE: Multi-prey functional response

*Corresponding author: mark.novak@oregonstate.edu

1 Abstract

Ecologists differ in the degree to which they consider the linear Type I functional response to be an unrealistic versus sufficient representation of predator feeding rates. Empiricists tend to consider it unsuitably non-mechanistic and theoreticians tend to consider it necessarily simple. Holling's original rectilinear Type I is dismissed by satisfying neither desire, with most compromising on the smoothly saturating Type II response for which searching and handling are assumed to be mutually exclusive activities. We derive a "multiple-prey-at-a-time" functional response reflecting predators that can continue to search when handling an arbitrary number of already-captured prey. The multi-prey model clarifies the empirical relevance of Holling's two Type I forms and the conditions under which linearity can be a mechanistically-reasoned description of predator feeding rates, even when handling times are long. We find information-theoretic support for the linear Type I and multi-prey responses in 26% of 2,598 compiled empirical datasets, and find evidence that larger predator-prey body-mass ratios permit predators to search while handling greater numbers of prey. Incorporating the multi-prey response into the Rosenzweig-MacArthur population-dynamics model reveals that a non-exclusivity of searching and handling can lead to coexistence states and dynamics that are not anticipated by theory built on linear Type I or Type II responses. In particular, it can lead to bistable fixed-point and limit-cycle dynamics with long-term crawl-by transients between them under conditions where abundance ratios reflect top-heavy food webs and the functional response is effectively linear. We conclude that Type I responses should not be considered empirically unrealistic and that more bounded conclusions should be drawn in theory presuming the linear Type I to be appropriate.

KEYWORDS: *type 0 functional response, consumer-resource cycles, long transients, alternative states, predator-prey body-mass ratio, top-heavy food webs*

Introduction

The way that predator feeding rates respond to changes in prey abundance, their functional response, is key to determining how species affect each other's populations (Murdoch & Oaten, 1975). The challenge of empirically understanding and appropriately modeling functional responses is therefore central to myriad lines of ecological research that extend even to the projection of Earth's rapidly changing climate (DeLong, 2021; Rohr *et al.*, 2023).

The simplest functional response model, the Holling Type I response, describes feeding rates as increasing linearly with prey abundance. Interpreted to represent an analytically-tractable first-order approximation to all other prey-dependent functional response forms (Lotka, 1925; Volterra, 1926), its simplicity has caused the Type I response to become foundational to theory across Ecology's many sub-disciplines. Nonetheless, there is a common and persistent belief among empirically-minded ecologists that the Type I response is unrealistic and artifactual. Indeed, it is typically dismissed *a priori* from both empirical and theoretical efforts to "mechanistically" characterize predator feeding rates (e.g., Baudrot *et al.*, 2016; Kalinkat *et al.*, 2023). This dismissal is similarly levied at the rectilinear model (e.g., Koen-Alonso, 2007) in which feeding rates increase linearly with prey abundance up to an abrupt maximum and by which Holling (1959) originally depicted the Type I response (Denny, 2014). Supportive evidence comes from syntheses that indicate the linear Type I to be rare, with feeding rates more consistent with smoothly saturating Type II responses being by far the more frequently inferred (Dunn & Hovel, 2020; Jeschke *et al.*, 2004).

Countering justifications for the continued use of the linear Type I response in theory relate to the challenge of extrapolating the inferences of mostly small-scale experiments to natural field conditions (see also DeLong, 2021; Griffen, 2021; Jeschke *et al.*, 2004; Li *et al.*, 2018; Novak & Stouffer, 2021b; Novak *et al.*, 2017; Uiterwaal *et al.*, 2018). For example, prey abundances in the field may vary relatively little, making linearity a sufficiently good approximation for how species affect each other (Wootton & Emmerson, 2005). Further, prey abundances in nature are often much lower than those used in experiments to elicit predator saturation (Coblentz *et al.*, 2023), which may consequently be rare in nature (but see Jeschke, 2007). Functional responses could thus be *effectively* linear in their dependence on prey abundances even for predator-prey interactions having very long handling times (e.g., Novak, 2010).

Here, our goal is to offer a further way of resolving ecologists' views on the linear and recti-linear models by considering an additional reason for functional responses to exhibit effectively linear prey dependence. This reason is not one of experimental design or variation in species abundances per se, but rather is attributable to the mechanics of predator-prey biology: the ability of predator individuals to handle and search for more than just one prey individual at a time (i.e. the non-exclusivity of handling and searching). Although it is straightforward to show how the linear Type I can emerge when handling times are inconsequential, and although func-tional response forms that could result from a non-exclusivity of handling and searching have been considered before (Jeschke *et al.*, 2002; 2004; Mills, 1982; Sjöberg, 1980; Stouffer & Novak, 2021), we contend that the empirical relevance and potential prevalence of such "multiple-prey-

at-a-time" feeding (henceforth multi-prey feeding) are not sufficiently understood due to an inappropriately literal interpretation of the "handling time" concept (see also Jeschke *et al.*, 2002; 2004). Likewise, the potential implications of multi-prey feeding for predator-prey coexistence and population dynamics have not, to our knowledge, been assessed.

We begin by providing a derivation of a simple multi-prey functional response model for a single predator population feeding on a single prey species that relaxes the assumption of searching and handling being exclusive activities. This derivation helps clarify the empirical relevance of Holling's linear and rectilinear Type I forms and the conditions under which these can be good descriptions of feeding rates in the presence of long handling times (Jeschke *et al.*, 2004). We then fit the multi-prey model to a large number of datasets assembled in a new version of the FoRAGE compilation (Uiterwaal *et al.*, 2022) to quantify the potential prevalence of multi-prey feeding and to test the hypothesis that larger predator-prey body-mass ratios permit predators to handle and search for more prey at a time. Finally, we incorporate the multi-prey response into the Rosenzweig & MacArthur (1963) "paradox of enrichment" population-dynamic model to assess its potential influence on predator-prey coexistence and dynamics. With our statistical analyses demonstrating that many datasets are indeed consistent with linear and effectively-linear multi-prey feeding, and that larger predator-prey body-mass ratios are indeed more conducive to multi-prey feeding, our mathematical analyses demonstrate that even small increases in the number of prey that a predator can handle at a time can lead to dynamics that are not anticipated by theory assuming linear Type I and Type II response forms.

A functional response for multi-prey feeding

Holling's Type II response

The multi-prey model may be understood most easily by a contrast to Holling's Type II model (a.k.a. the disc equation). There are several ways to derive it (Garay, 2019), but the most common approach takes the perspective of a single predator individual that can either be searching or handling a single prey individual at any point in time: In the time T_S that a predator spends searching it will encounter prey at a rate proportional to their abundance N, thus the number of prey eaten is $N_e = aNT_S$ where a is the attack rate. Rearranging we have $T_S = N_e/aN$. With a handling time h for each prey, the length of time spent handling all eaten prey will be $T_H = hN_e$. Given the presumed mutual exclusivity of the two activities, $T_S = T - T_H$ where T is the total time available. Substituting the second and third equations into the fourth, it follows that $N_e = aNT/(1 + ahN)$. We arrive at the predator individual's feeding rate by dividing by T, presuming steady-state predator behavior and constant prey abundances.

An alternative derivation on which we build to derive the multi-prey model considers a temporal snapshot of a predator population composed of many identical and independent individuals (see also Real, 1977). Assuming constant prey abundance and steady-state conditions, the rate at which searching individuals P_S become handling individuals P_H must equal the rate at which handling individuals become searching individuals such that $aNP_S = \frac{1}{h}P_H$, visually represented as

$$P_S \; \overset{\overset{\textstyle N}{\searrow} a}{\underset{N_e \nearrow 1/h}{\rightleftarrows}} \; P_H \,.$$

105 Given the mutual exclusivity of searching and handling, $P_S = P - P_H$, where P is the total

106 number of predators. Substituting this second equation into the first, it follows that the number

107 of handling predators $P_H = ahNP/(1 + ahN)$. Since eaten prey are generated at rate $\frac{1}{h}P_H$

108 by these predators as they revert back to searching, we obtain Holling's Type II (per-predator)

109 model by multiplying the proportion of all predators that are handling, P_H/P, by $\frac{1}{h}$.

The multi-prey response

The derivation of the multi-prey response follows the same logic but assumes that searching

and handling are not mutually exclusive activities until an arbitrary count of n prey individuals

are being handled. With constant prey abundance and steady-state conditions as before, we

assume that predators continue to handle each prey with handling time h and that predators

handling less than n prey continue to search for and encounter prey at rate aN. The rate at

which searching individuals P_S become P_{H_1} individuals handling one prey is then equal to the

rate at which they revert back to being searching individuals with no prey, thus $P_{H_1} = ahNP_S$.

Likewise, the rate at which P_{H_1} individuals become P_{H_2} individuals handling two prey must

equal the rate these revert back to handling just one prey, thus $P_{H_2} = ahNP_{H_1} = (ahN)^2 P_S$.

That is,

$$
P_S \;\underset{N_e \;1/h}{\overset{N\; a}{\rightleftharpoons}}\; P_{H_1} \;\underset{N_e \;1/h}{\overset{N\; a}{\rightleftharpoons}}\; P_{H_2} \;\underset{N_e \;1/h}{\overset{N\; a}{\rightleftharpoons}}\; \cdots \;\underset{N_e \;1/h}{\overset{N\; a}{\rightleftharpoons}}\; P_{H_n} \; .
$$

122 Generalizing by induction, the number of predators P_{H_i} handling i prey will be $(ahN)^i P_S$ for

123 $i \in \{1, 2, 3, \ldots, n\}$. The proportion of predators handling i prey at any point in time will then

be

$$\frac{P_{H_i}}{P} = \frac{(ahN)^i P_S}{P_S + P_{H_1} + \ldots + P_{H_n}} = \frac{(ahN)^i}{1 + \sum\limits_{i=1}^{n}(ahN)^i} \tag{1}$$

(Fig. S.1). With each of these groups generating eaten prey at rate $\frac{1}{h}P_{H_i}$, the per predator

feeding rate of the population is obtained by a summation across all groups, giving

$$f(N) = \frac{\frac{1}{h}\sum\limits_{i=1}^{n}(ahN)^i}{1 + \sum\limits_{i=1}^{n}(ahN)^i} \tag{2}$$

(Fig. 1). This is the multi-prey model for integer values of n. However, because the geometric

series $\sum_{i=1}^{n} x^i = x(1 - x^n)/(1 - x)$ for $x \neq 1$, we can also write the model more generally for

arbitrary values of n as

$$f(N) = \frac{aN(1 - (ahN)^n)}{1 - (ahN)^{n+1}} \tag{3}$$

to reflect predator populations capable of searching while handling a non-integer (e.g., average)

number of prey individuals.

We note that Sjöberg (1980) derived equivalent formulations in Michaelis-Menten enzyme-

kinematics form with parameters having correspondingly different statistical properties (Novak

& Stouffer, 2021a; Rohr *et al.*, 2022). We also note that despite the appearance of two sum-

mations in eqn. 2 and the unusual appearance of subtractions in eqn. 3 (see *Supplementary*

Materials), the model has only three parameters.

Relevance of the Type I response

The conditions under which Holling's Type I and II models can be good descriptions of predator

feeding rates are clarified by observing that both multi-prey formulations simplify to the Type

II when $n = 1$ and approach the rectilinear Type I as n increases (Fig. 1). Further, the linear

Type I is obtained when $n = \infty$ (Fig. 1) because the infinite power series $\sum_{i=1}^{\infty} x^i = x/(1-x)$ for

$|x| < 1$. Incorporating this infinite power series into eqn. 1 shows that the expected proportion

of predators handling prey at any given time will be ahN. Importantly, this proportion differs

from that which would be inferred to emerge by letting $h \to 0$ in the way the linear Type

I is typically obtained (e.g., Rohr *et al.*, 2022). In other words, the multi-prey model shows

that handling times need not be inconsequential for the functional response to appear linear

(Jeschke *et al.*, 2004). Rather, the Type I can be a very good approximation of feeding rates

so long as n is high and not all predators are handling prey (i.e. $ahN < 1$), which requires that

prey abundances remain less than $1/ah$. (For comparison, note that $1/ah$ is equivalent to the

half-saturation parameter of the Michaelis-Menten version of the Type II response for which

it reflects the prey abundance at which 50% of predators will be handling prey (i.e. the per

predator feeding rate is at half its maximum of $1/h$).)

Empirical support for multi-prey feeding

The multi-prey model shows that a spectrum of functional response forms can exist between

the extremes of the Type I and Type II when handling and searching are not assumed to be

mutually exclusive (Fig. 1). This motivated us to test two empirical hypotheses using the large

number of functional response studies that exist in the literature. The first hypothesis was that

prior syntheses indicating the linear Type I response to be rare (Dunn & Hovel, 2020; Jeschke

et al., 2004) were biased against the Type I despite its potential appropriateness. That is,

Figure 1: The potential forms of the multi-prey response. The multi-prey model diverges from the Holling Type II model (for which $n = 1$) and approaches Holling's rectilinear Type I model as the number n of prey individuals that a predator can handle while continuing to search increases. When $n = \infty$ it reduces to the linear Type I model which can remain a biologically appropriate description of predator feeding rates so long as $ahN < 1$ (indicated by non-dashed region of the black line). *Parameter values*: attack rate $a = 0.1$ and handling time $h = 4$.

feeding rates may have had response shapes between the Type II and rectilinear Type I (close

to the linear Type I at low prey abundances) but were classified as Type II due to the lack of a

sufficiently simple rectilinear-approaching model in prior analyses. The second hypothesis was

due to Sjöberg (1980) who motivated parameter n by considering it to be a measure of food

particle size relative to a zooplankter's gut capacity, with low n reflecting capacity for few large

prey and high n reflecting capacity for many small prey. We thus expected predator-prey pairs

with larger body-mass ratios to exhibit larger estimates of n when their functional responses

were assumed to follow the multi-prey model.

We used the FoRAGE database of published functional response datasets to assess these

hypotheses (Uiterwaal *et al.*, 2022). Our v4 update contains 3013 different datasets representing

1015 unique consumer-resource pairs (i.e. not just predator and prey species, though we continue

to refer to them as such for simplicity). For our analyses, we excluded datasets having a sample size less than 15 observations as well as structured experimental studies that implemented less than 4 different treatment levels of prey abundance. (See *Supplemental Materials* for additional details.) Our model-fitting procedure followed the approach used by Stouffer & Novak (2021) and Novak & Stouffer (2021b), assuming one of two statistical models for each dataset: a Poisson likelihood for observational (field) studies and when eaten prey were replaced during the course of the experiment, and a binomial likelihood when eaten prey were not replaced. Experimental data available in the form of treatment-specific means and uncertainties were analyzed by a parametric bootstrapping procedure in which new datasets were created assuming either a treatment-specific Poisson or binomial process as dictated by the study's replacement of prey. In cases where measures of the uncertainty around non-zero means were not available, we interpolated them based on the global log-log-linear relationship between means and standard errors across all datasets following Uiterwaal *et al.* (2018); for zero means, we interpolated missing uncertainty values assuming a linear within-dataset relationship. Unlike in Stouffer & Novak (2021) and Novak & Stouffer (2021b), we bootstrapped data available in non-summarized form as well, using a non-parametric resampling procedure that maintained within-treatment sample sizes for treatment-structured datasets. While replacement data were bootstrapped 1000 times, non-replacement data were bootstrapped only 100 times due to the substantially-higher computational burden of having to, in these cases, numerically integrate the multi-prey model for which no sufficiently-simple Lambert W solution could be obtained.

Frequency of multi-prey feeding

We used the Bayesian Information Criterion (BIC) to test our first hypothesis, counting the number of datasets whose bootstrapped mean BIC score supported the multi-prey model over the other two models by more than two units (ΔBIC > 2). Our choice to use BIC was motivated both by its purpose of selecting the generative model (rather than the best out-of-sample predictive model, as per AIC) and by its generally stronger penalization of parametrically-complex models (thereby favoring simpler models, relative to AIC). Conclusions regarding evidence in support of the multi-prey model were thereby made more conservative.

Considering only those datasets that could differentiate among the performance of the three models, the results of this first analysis were that 20 of 203 (9.9%) replacement datasets and 551 of 2395 (23%) non-replacement datasets identified the multi-prey model as the sole best-performing model (Fig. 2a-2b). An additional 37 (18%) replacement and 459 (19%) non-replacement datasets identified the multi-prey model as performing equivalently well to either the Type I or Type II model. Although the linear Type I was the least frequently sole-supported model (15 of replacement and 94 of non-replacement datasets), it was supported by datasets representing all four of the most common predator taxonomic groups that constituted 90% of all datasets in FoRAGE (insects, arachnids, crustaceans, and fishes; Fig. S.2).

Effect of predator-prey body-mass ratio on n

To test the second hypothesis, we excluded datasets for which the linear Type I had alone performed best and regressed the remaining datasets' bootstrapped median point estimates of

n against their study's predator-prey body-mass ratio ($ppmr$), these having been compiled in

FoRAGE for most datasets. Although almost 90% of these datasets had estimates of $n \leq 8$

(Fig. S.3), both variables exhibited variation over several orders of magnitude. We therefore

performed linear least-squares regression using $\log_2(n)$ and $\log_{10}(ppmr)$.

Our analysis supported the hypothesis that predator-prey pairs with larger body-mass ratios

tend to exhibit larger estimates of n ($log_2(n) = 0.64 + 0.16 \cdot log_{10}(ppmr)$, $p < 0.01$, Table S.1),

but the predictive utility of this relationship was extremely poor ($R^2 = 0.02$, Fig. 2c). To assess

the sensitivity of this result to variation among datasets, we performed additional regressions

that restricted the considered datasets to (i) those having estimates of $n > 1$ (Fig. 2c, Table

S.1), (ii) those with sample sizes exceeding the median sample size of all datasets (Fig. S.4,

Table S.2), and (iii) the four most common predator taxonomic groups (insects, arachnids,

crustaceans, and fishes), including for this last regression a two-way interaction term between

predator group identity and predator-prey body-mass ratio (Fig. 2d, Table S.3). Each of these

analyses evidenced statistically clear, albeit predictively poor, positive relationships between n

and predator-prey body-mass ratios for all predators in general and for each predator group

individually as well.

Population-dynamic effects of multi-prey feeding

Given the empirical evidence that multi-prey feeding may indeed be common and a viable way

to describe functional responses, we next investigated its potential consequences for predator-

prey dynamics. Our goal was to understand how assuming either a linear Type I or a Type II

Figure 2: Empirical support for multi-prey feeding. Figs. 2a and 2b depict Venn diagrams categorizing the datasets of FoRAGE by their support for one or more of the three models as evaluated using a cut-off of 2 BIC units. Figs. 2c and 2d depict the observed relationship between estimates of n and the body-mass ratio of the studies' predator-prey pairs, excluding datasets for which the Type I model alone performed best. Regression lines in Fig. 2c reflect all considered datasets or only those with estimates of $n > 1$ (Table S.1). Regression lines in Fig. 2d reflect the identity of the four most common predator groups ($n \geq 1$, Table S.3).

response could lead to incorrect conclusions regarding these dynamics. We used the well-studied

Rosenzweig & MacArthur (1963) "paradox of enrichment" model to achieve this goal, employing

both graphical (i.e. isocline) analysis and simulations.

The model describes the growth rates of the prey N and predator P populations as

$$\frac{dN}{dt} = rN\left(1 - \frac{N}{K}\right) - f(N)P \tag{4a}$$

$$\frac{dP}{dt} = ef(N)P - mP , \tag{4b}$$

where r and K are the prey's intrinsic growth rate and carrying capacity, $f(N)$ is the functional

response, and e and m are the predator's conversion efficiency and mortality rate. Logistic

prey growth and Holling's Type II response have become the component parts of the canonical

Rosenzweig-MacArthur model for which enrichment in the form of an increasing carrying ca-

pacity causes the populations' dynamics to transition from a regime of monotonically-damped

stable coexistence to damped oscillations to sustained limit cycles (Rosenzweig, 1971). Other

prey growth and Type II-like functional response forms affect a similar destabilization sequence

(e.g., Freedman, 1976; May, 1972; Rosenzweig, 1971; Seo & Wolkowicz, 2018). The location

of the Hopf bifurcation between asymptotic stability and limit cycles is visually discerned in

the model's P vs. N phase plane (Fig. 3) as the point where the vertical N^* predator iso-

cline intersects the parabolic P^* prey isocline at its maximum (Rosenzweig, 1969; Rosenzweig

& MacArthur, 1963). That is, the coexistence steady state entails a globally-stable fixed point

when the isoclines intersect to the right of the maximum and entails a locally-unstable fixed

point with a globally-stable limit cycle when they intersect to the left of the maximum (Seo

& Wolkowicz, 2018). Graphically, increasing K destabilizes dynamics by stretching the prey

isocline, moving its maximum to the right while the position of the vertical predator isocline

remains unchanged. In contrast, when logistic growth and a linear Type I are assumed, the

prey isocline is a linearly-decreasing function of prey abundance (Fig. 3) and predator-prey

coexistence entails a globally-stable fixed point for all levels of enrichment.

Graphical analysis

For our analysis we insert the multi-prey response (eqn. 3) for $f(N)$ in eqn. 4. Solving $dP/dt = 0$

for the N^* predator isocline then requires solving

$$\frac{m}{e} = f(N^*) \implies N^* = \frac{m\left(1 - (ahN^*)^{n+1}\right)}{ae\left(1 - (ahN^*)^n\right)} . \tag{5}$$

This leads to a solution for N^* that is independent of the predator's abundance (i.e. remains

vertical in the P vs. N phase plane) but is unwieldy for $n > 2$ (see *Supplementary Materials*).

Nonetheless, it represents a generalization of the predator isocline obtained for the Rosenzweig-

MacArthur model with $n = 1$, $N^* = \frac{m}{a(e-mh)}$, and converges on $N^* = m/ae$ as $n \to \infty$ when

$ahN^* < 1$, just as obtained assuming the Type I. In fact, N^* transitions smoothly from the

former to the latter as n increases (Fig. 3) because eqn. 5 is a monotonically declining function

of n for $ahN^* < 1$.

Solving $dN/dt = 0$ for the P^* prey isocline leads to the solution

$$P^* = \frac{rN}{f(N)}\left(1 - \frac{N}{K}\right) = \frac{r(K - N)\left(1 - (ahN)^{n+1}\right)}{aK\left(1 - (ahN)^n\right)} . \tag{6}$$

270 This too represents a generalization of the Rosenzweig-MacArthur model's prey isocline, $P^* =$

Figure 3: Predator and prey isoclines of the Rosenzweig-MacArthur model modified to include the multi-prey response correspond to those observed with the linear Type I and Type II responses when $n = \infty$ and $n = 1$ respectively. As the number n of prey that a predator can handling while searching increases, the prey abundance at which the predator's growth rate is zero (i.e. the vertical predator isocline, N^*) decreases from its value under the Type II response ($m/a(e - mh)$) and converges rapidly on the value expected under the Type I response (m/ae). In contrast, predator abundances at which the prey's growth rate is zero, P^*, converge on those expected under the Type I response only at low prey abundances to affect a second region of asymptotically stable dynamics; the "hump" does not flatten as it would if the handling time were presumed to be inconsequential (i.e. $h = 0$). Limit cycles occur when the predator and prey isoclines intersect on the left flank of the hump. With increasing n, the inflection point between the low-prey region of stability and limit cycles approaches the prey abundance where all predators are busy handling predators under the rectilinear model, $1/ah$ (indicated by non-dashed region of the black prey isocline). *Other parameter values*: attack rate $a = 0.02$, handling time $h = 2$, prey growth rate $r = 0.5$, prey carrying capacity $K = 100$, conversion efficiency $e = 0.25$, predator mortality rate $m = 0.08$.

$r(K - N)(1 + ahN)/aK$, which is itself a generalization of the isocline $P^* = r(K - N)/aK$

obtained with the Type I as $n \to \infty$. Between these the prey isocline under the multi-prey

response transitions from a parabolic dependence on the prey's abundance to having a second

region within which it is a declining function of prey abundance (Fig. 3). This second region has

a slope of $-r/aK$ at its origin regardless of n and is limited to low prey abundances of $N < 1/ah$;

as n increases, the region's upper extent approaches the prey abundance at which all predators

are busy handling prey under the rectilinear model. That is, for $1 < n < \infty$ the "hump" shape

does not flatten out as it does when one assumes handling times to become negligible. Rather,

the prey isocline exhibits two regions of negative prey dependence (where $\frac{dP^*}{dN} < 0$) that flank

an intermediate region of positive prey dependence (where $\frac{dP^*}{dN} > 0$).

Implications for coexistence and dynamics

The emergence of a second prey abundance region where the slope of the prey isocline is neg-

ative means that a second asymptotically-stable coexistence equilibrium — one having a high

predator-to-prey abundance ratio — is possible should the two isoclines intersect within it. The

fact that this may occur is discerned by noting that N^* (eqn. 5) is independent of r and K,

and that P^* (eqn. 6) is independent of m and e; the positions of the two isoclines are thus

independent except via the functional response parameters a, h, and n. In fact, because N^*

decreases while the upper limit of the low prey abundance region of P^* increases towards $1/ah$

as n increases, it is readily possible — conditional on the values of the other parameters — to

observe a stable state at $n = 1$ to first transition to limit cycles and then return to fixed-point

stability as n alone is increased. This is illustrated by Fig. 4 in the context of the paradox of

enrichment for values of K between approximately 75 and 115. Multi-prey feeding may thus be

seen as another potential mechanism with which to resolve the paradox (Roy & Chattopadhyay,

2007). Indeed, in addition to rescuing predators from deterministic extinction at low levels of

enrichment where a single-prey-at-a-time predator could not persist ($20 < K < 40$ in Fig. 4), sufficiently large values of n can preclude the occurrence of limit cycles altogether ($n > 9$ in Fig. 4).

Figure 4: The destabilization with enrichment that is seen under the classic Rosenzweig-MacArthur model (where $n = 1$) is altered when predators can search for and handle multiple prey at a time ($n > 1$). At low prey carrying capacities ($K < 40$), multi-prey feeding rescues predators from deterministic extinction. At intermediate carrying capacities ($40 < K < 110$), low levels of multi-prey feeding destabilize dynamics by causing perturbation responses to transition from a transient regime of monotonic damping to one of damped oscillations or from damped oscillations to a persistent limit cycle regime. Further increases in multi-prey feeding can have a qualitatively stabilizing influence on dynamics, with sufficiently high n precluding a transition to limit cycles altogether so long as perturbations are sufficiently small. Large perturbations, on the other hand, will cause a transition to an alternative stable state consisting of limit cycle dynamics (see Fig. 5). *Other parameter values* as in Fig. 3.

Notably, however, the just-described high-predator low-prey steady state is only a locally stable fixed point and coexists with a stable limit cycle that surrounds it (Figs. 4 and 5). The high-predator low-prey state thus exhibits bi-stability. The consequences of this bi-stability are that predator-prey interactions with multi-prey feeding are destined to exhibit (i) transitions to persistent limit cycles when subjected to large perturbations that send abundances beyond the domain of attraction of the fixed-point steady state (Fig. 5a,c), and (ii) transient dynamics that are prone to damped oscillations (rather than monotonic damping) in response to small perturbations within the domain of attraction. These transient oscillations occur for substantially lower levels of enrichment than is the case for single prey-at-a-time predators (Fig. 4). Moreover, their temporal duration can be exceedingly long (Fig. 5b) because the limit cycle acts akin to a crawl-by attractor (Hastings $et\ al.$, 2018) that impinges upon the steady state's local resilience. Thus, when subjected to continual perturbations in an explicitly stochastic setting (Barraquand $et\ al.$, 2017), the system can readily transition between the stable fixed-point attractor and the stable limit cycle attractor that surrounds it (Fig. 6), resulting in dynamical epochs of irregular duration that appear characteristic of many empirical time-series (Blasius $et\ al.$, 2020; Rubin $et\ al.$, 2023). Multi-prey feeding does not, therefore, provide a robust resolution to the paradox of enrichment but rather leads to a richer range of population dynamics and coexistence states than can result from the Type I and Type II responses alone.

Figure 5: Because of the system's bi-stability at high predator-to-prey abundance ratios, even small differences in the size of a perturbation to the steady state can affect a large change in the duration of the system's transient response (compare panels a and b with c and d) and can even cause the system to become entrained in a stable limit cycle (illustrated in panels e and f). The only difference between each of the above panel rows is that the predator's initial population size $P(0)$ is perturbed away from its P^* steady state as: (a, b) $P(0) = P^* - 6$; (c, d) $P(0) = P^* - 7.0645$; and (e, f) $P(0) = P^* - 7.065$. For all cases $N(0) = N^*$. *Parameter values* as Fig. 3 with $n = 10$.

Figure 6: When subjected to continually-occurring stochastic perturbations, the high-predator low-prey coexistence state can exhibit time periods during which its dynamics are influenced primarily by the stable fixed-point attractor and time periods during which dynamics are primarily influenced by the alternative stable limit cycle attractor, switching between these on an irregular basis. Simulation implemented using an Itô integral process as $dN = rN(1 - N/K) - f(N)P \, dt + \sigma N dW$ and $dP = ef(N) - mP \, dt - \sigma P dW$, with $f(N)$ as in eqn. 3 and Gaussian white environmental noise $dW(t)$ of volatility $\sigma = 0.04$ (*cf.* Barraquand, 2023). *Other parameter values* and initial population sizes as in Fig. 5c-d.

Discussion

Our study was motivated by the apparent disconnect that exists between the way that many empirically-minded ecologists perceive the Type I functional response model and the way that many modelers and theory-minded ecologists justify its use in their representations of consumer-resource interactions. While the former are prone to dismiss the Type I as being overly simplistic and hence unsuitable for describing predator feeding rates, the latter are prone to rely on and justify its sufficiency for the sake of computational ease and analytically-tractable insight. Since the potential for predators to feed on multiple prey at a time (i.e. the non-exclusivity of handling and searching activities) has been little considered by either group, we set out to address three

aspects of this disconnect: (i) deriving a multiple-prey-at-a-time functional response model that mechanistically connects the linear and piecewise-linear Type I forms to the more empirically palatable Type II model, (ii) assessing the extent to which published datasets provide support for the Type I, II, and multi-prey models, and (iii) investigating how multi-prey feeding and the linearity it can impose on functional responses can alter our understanding of predator-prey coexistence. Because they bear additional insight with which to elaborate on the circumstances under which linear functional responses may be empirically relevant, we structure the discussion of our work by considering the latter two aspects first.

Empirical support

Our information-theoretic analysis of the datasets compiled in FoRAGE demonstrates that both the linear Type I and the multi-prey model are viable descriptions (*sensu* Skalski & Gilliam, 2001) of the feeding rates that predators have exhibited in many single-prey experiments (Figs. 2a-2b). This result is consistent with handling and searching being non-exclusive activities for a substantial number of predator-prey pairs. Although this contrasts with the prior syntheses of Jeschke *et al.* (2004) and Dunn & Hovel (2020), these (*i*) did not consider models capable of response forms in between the strictly linear Type I and Type II forms and (*ii*) either relied on the conclusions reached by each studies' original authors (who used varied model-fitting and comparison approaches) or visually assessed functional response forms from plotted data. One might argue that many of the datasets providing sole support to the Type I in our analysis came from experiments using prey abundances that were insufficient to elicit saturation (see

also Coblentz *et al.*, 2023), but the point can be made that, from an information-theoretic perspective, the linear Type I performed best across the range of prey abundances that the original authors considered empirically reasonable (and logistically feasible). The even greater number of datasets that provided sole support to the multi-prey model, along with the result that many of the point estimates for parameter n (the maximum number of prey eaten at a time) were sufficiently large to affect a response approaching a piecewise-linear Type I (*cf.* Figs. 1 and 2c), indicates that feeding rates exhibited effectively linear responses for many predator-prey interactions having long handling times as well. Moreover, the statistically-clear positive relationships we observed in our subsequent regression analyses of n and predator-prey body-mass ratios (Figs. 2c-2d) confirm Sjöberg's hypothesis regarding a proximate reason for this effective linearity, indicating that linearity in feeding rates is more likely for larger predators feeding on small prey because handling is less preclusive of searching for these interactions.

Unfortunately, the amount of variation in n that was explained by body-mass ratio alone was extremely low, making the relationship of little predictive utility relative to several other body-mass relationships (e.g., Brose *et al.*, 2006; Coblentz *et al.*, 2023; Hatton *et al.*, 2015; Rall *et al.*, 2012). That said, the relationship's low explanatory power is not unsurprising given that none of the experiments in FoRAGE was designed with the multi-prey model in mind. In particular, and although most estimates of n were of a seemingly reasonable magnitude (Fig. S.3), we caution against giving too much credence to the very large-valued estimates we observed. This is for two primary reasons. First, given that a given dataset's ability to distinguish between

possible values of n diminishes rapidly as n increases (Fig. 1), datasets exhibiting saturation at high prey abundances but having few or no observations near the inflection point of $1/ah$ will have been sensitive to issues of parameter identifiability. Low identifiability will have caused an inflation of estimates despite our effort to guard against it by removing datasets with fewer than 4 prey abundance levels. Second, given that initiating experiments with predator individuals having empty guts is a common protocol (Griffen, 2021; Li *et al.*, 2018), many experiments will have strictly violated the assumption of predator behavior being at steady state. This will also have inflated estimates of n by causing transient rates of prey ingestion to exceed rates of handling completion (i.e. $aN > 1/h$) to affect faster-than-steady-state feeding, especially at prey abundances below $1/ah$. We therefore suggest that the very large estimates of n observed in our analyses be better interpreted as qualitative (rather than quantitative) support for the non-exclusivity of searching and handling and encourage future experiments and analyses with additional covariate predictors to better understand the biological sources of variation in n.

Mechanistic approximations

The multi-prey model may be considered a mechanistic model in that its derivation and each of its parameters has at least one biologically-specific interpretation. However, it is also rather phenomenological in that it encodes only an essence of the biologically possible non-exclusivity of searching and handling processes. For example, the model's derivation assumes that both the attack rate and handling time remain constant and independent of the number of prey that predators are already handling (below the maximum number n). Although this assumption

may result in a very good approximation to feeding rates, it is unlikely to reflect biological reality particularly as the number of prey being handled by a given predator approaches n. In such circumstances either or both searching and handling process rates are likely to become dependent on the feeding rate and thereby prey abundance (see also Okuyama, 2010; Stouffer & Novak, 2021).

Functional responses where such dependence is important may be better and more mechanistically described by more flexible models (see also Novak & Stouffer, 2021a). Prominent among these is the extended Steady State Saturation model (SSS[1]) of Jeschke *et al.* (2004) in which handling and digestion are explicitly distinguished (see *Supplementary Materials*). In this four-parameter model, searching and handling are mutually exclusive, but searching and digestion are not because the predator's search effort depends on its gut fullness (i.e. hunger level) and is thus dictated by the digestion rate. A phenomenological shape parameter controls the non-linearity of the search-effort hunger-level relationship. For high values of this shape parameter (reflecting predators that search at their maximum rate even when their guts are quite full) and inconsequential handling times, the model approaches the rectilinear Type I, just like the multi-prey model at high n, while for consequential handling times it retains a saturating curvature at low prey abundances (see Figs. A1 and A2 of Jeschke *et al.*, 2004).

---

[1]We would be remiss not to point out that all functional response models of which we are aware assume steady state conditions at the behavioral foraging scale. The SSS model's name does not, therefore, reflect a limitation that is unique to it.

Population-dynamic effects

The population-dynamic consequences of the extended SSS model remain unstudied, but our analysis of the simpler multi-prey model reveals the relevance of it and other models for understanding how the effective linearity of multi-prey feeding can impact predator-prey dynamics. These other models are the arctangent and hyperbolic tangent functional responses because for these it has been more rigorously shown that two limit cycles — one stable and the other unstable — can co-occur with a locally-stable fixed point at low prey abundances (Seo & Kot, 2008; Seo & Wolkowicz, 2015; 2018), just as we observed for the multi-prey model. The key feature common to all three models is that they affect a prey isocline that *decreases* from a *finite*-valued origin at zero prey abundance. This differs from the Type II and other functional responses that exhibit saturating curvature at low prey abundance. For these the prey isocline *increases* from a finite-valued origin, the low-prey fixed point is unstable, and only the stable limit cycle is thus of relevance under logistic prey growth. It also differs from functional responses that accelerate at low prey abundances (e.g., the Type III) and from consumer-resource models more generally in which, for example, prey have a physical refuge, exhibit sublinear density-dependence, or experience density-independent immigration. For these the prey isocline decreases from an origin that approaches infinity and the low prey steady state is a stable fixed point around which limit cycles do not occur. We surmise that the effective linearity brought about by the non-exclusivity of searching and handling in the multi-prey model is (*i*) replicated by the more phenomenological arctangent and hyperbolic tangent models, and that (*ii*) it is the cause of the greater range

of dynamical outcomes that these functional responses affect as compared to responses having a more nonlinear form at low prey abundances.

The broader implication of our analysis is that population-dynamic theory that relies on the linear Type I may not be as globally relevant from a biological perspective as its mathematics would suggest. In particular, it shows that the stabilization which the Type I contributes to dynamics is dependent on perturbation magnitude. More specifically, the relevance of theory that relies on the linear Type I is limited to perturbations that are small enough to preclude the influence of the attracting stable limit cycle that will exist when the functional response is more realistically described as having a potentially unobserved maximum feeding rate.

Our consideration of the paradox of enrichment illustrates a specific example of this. When the functional response is assumed to be linear Type I, the fixed point is globally-stable and perturbations to it decay monotonically. In contrast, when the functional response is only effectively linear, as when multi-prey feeding occurs, the fixed point is only locally stable and perturbations can elicit cycles that may persist for many generations or even indefinitely. In fact, as indicated by Rubin *et al.* (2023) in their analysis of a stochastic implementation of the Rosenzweig-MacArthur model, the dynamics will additionally be influenced by the crawl-by inducing origin (dual extinction) and prey-only (carrying capacity) steady states that will extend the lifetime of long-term transients even further. This influence, too, will not be observed when a linear Type I is assumed because these steady states will rarely if ever be approached.

Relevance revisited

As discussed above (see *Relevance of Type I response*), the multi-prey model shows that handling times need not be inconsequential to observe an effectively linear functional response when the number of prey that a predator individual can handle at a time is relatively high and the maximum proportion of individuals in a predator population that are simultaneously handling prey remains sufficiently low. This is not to say that other factors and processes cannot cause functional responses to be far from linear, but within the confines of our work's assumptions the latter condition can be satisfied when prey abundances remain less than $1/ah$.

Our statistical and mathematical analyses add insight into where this condition on prey abundances is more likely to be met. Specifically, functional responses are more likely to be effectively linear when the predator-to-prey body-mass ratios are high (Fig. 2c), when predator-to-prey abundance ratios are high (Fig. 3), and thus, we predict, in top-heavy systems with high predator-to-prey biomass ratios. Top-heavy interactions and food webs more generally occur in all ecosystem types (McCauley *et al.*, 2018), but are more likely for ectothermic and invertebrate consumers, in aquatic habitats, among higher trophic levels, and in ecosystems of low total biomass (Brose *et al.*, 2006; Hatton *et al.*, 2015; Perkins *et al.*, 2022). The development of methods for gauging the non-linearity of functional responses in diverse field settings (e.g., Novak *et al.*, 2017; Uiterwaal & DeLong, 2024) will be useful for directly testing our prediction that these same systems should also exhibit more linear functional responses.

Importantly, our work also shows that predator-prey dynamics need not be destabilized by

food web top-heaviness. Rather, increases in top-heaviness can lead to greater food web stability — be it stable coexistence potential or perturbation resilience (Fig. S.5) — when multi-prey feeding occurs, provided that perturbations are small enough for population abundances to remain well within the local attractor of the stable fixed point (Fig. 5). This contrasts with existing theory on top-heavy food webs that has largely assumed Type II responses (McCauley *et al.*, 2018). Indeed, our analyses show that even small departures from mutual exclusivity can lead to qualitatively different coexistence states and dynamics than predicted by existing theory, including the possibility of long-term transients and the just-mentioned bi-stability of fixed-point and limit-cycle dynamics. Food web models that incorporate multi-prey feeding and how its prevalence may change with species- and system-level attributes will be useful for understanding just how much multi-prey feeding must occur within food webs as a whole to alter their structure and dynamics. A first step towards such food web models will be to extend the multi-prey model to multi-species formulations appropriate for generalist rather than single-prey-species predators.

Conclusion

Natural history observations show that diverse types of predators are capable of (literally) handling and searching for prey simultaneously: sea otters capture several snails on a dive; crabs process mussels with their mouthparts while picking up more with their claws; spiders capture insects in their webs while processing others for later ingestion. Many more examples become apparent and potentially relevant to the context of functional responses when it is recognized

that the "handling time" parameter of most models represents not just the literal manipulation of prey that may be seen by an observer of the interaction but rather reflects processes that limit steady-state feeding rates (DeLong, 2021; Jeschke *et al.*, 2002).

The degree to which handling and searching actually represent mutually exclusive activities, and the degree to which each of the many processes potentially encapsulated by handling parameters measurably contributes to a predator's functional response, is nonetheless poorly discerned from observation alone. Knowing that handling times are short or long, or that searching and literal handling do or do not overlap, is neither sufficient to dismiss or assume a given functional response model on *a priori* grounds. This is because all models are phenomenological approximations of biological process at some level. This applies as much to predator-prey interactions studied in controlled experiments as it does to those studied in natural settings, and is particularly true in the context of building understanding and theory when extrapolating the former to the latter across Ecology's wide-ranging scales.

We thus draw two overarching conclusions: that neither of the Type I response forms should be dismissed by empiricists as irrelevant descriptions of predator feeding rates, and that modelers and theoreticians should be more cautious in reaching empirical conclusions of system dynamics when presuming the linear Type I response to be appropriate. In broader terms, our research demonstrates how disagreements between different perspectives can be addressed by identifying and easing the fundamental assumptions that underpin them, and how improved communication between empiricists and theoreticians will benefit Ecology as a whole (Grainger *et al.*, 2022).

Acknowledgments

MN thanks the OSU MathBio group for feedback, is indebted to Patrick DeLeenheer for setting him straight, thanks CJ Keist for technical assistance with OSU's Cosine High Performance Cluster, and was supported by NSF DEB-2129758.

Author contributions

MN conceived of the study, performed the analyses, and wrote the first draft. JPD compiled functional response datasets. KEC and JPD discussed the analyses and edited the manuscript.

Code and data availability

The FoRAGE compilation is available from the *Knowledge Network for Biocomplexity* (DeLong & Uiterwaal, 2018). All code and data are available at `https://github.com/marknovak/FR_n-prey-at-a-time` and `FigShare_url_posted_after_acceptance`.

References

Barraquand, F. (2023). No sensitivity to functional forms in the Rosenzweig-Macarthur model with strong environmental stochasticity. *Journal of Theoretical Biology*, 572, 111566.

Barraquand, F., Louca, S., Abbott, K. C., Cobbold, C. A., Cordoleani, F., DeAngelis, D. L., Elderd, B. D., Fox, J. W., Greenwood, P., Hilker, F. M., Murray, D. L., Stieha, C. R., Taylor, R. A., Vitense, K., Wolkowicz, G. S. K. & Tyson, R. C. (2017). Moving forward in circles: challenges and opportunities in modelling population cycles. *Ecology Letters*, 20, 1074–1092.

Baudrot, V., Perasso, A., Fritsch, C., Giraudoux, P. & Raoul, F. (2016). The adaptation of generalist predators' diet in a multi-prey context: insights from new functional responses. *Ecology*, 97, 1832–1841.

Blasius, B., Rudolf, L., Weithoff, G., Gaedke, U. & Fussmann, G. F. (2020). Long-term cyclic persistence in an experimental predator–prey system. *Nature*, 577, 226–230.

Brose, U., Jonsson, T., Berlow, E. L., Warren, P., Banasek-Richter, C., Bersier, L.-F., Blanchard, J. L., Brey, T., Carpenter, S. R., Blandenier, M.-F. C., Cushing, L., Dawah, H. A., Dell, T., Edwards, F., Harper-Smith, S., Jacob, U., Ledger, M. E., Martinez, N. D., Memmott, J., Mintenbeck, K., Pinnegar, J. K., Rall, B. C., Rayner, T. S., Reuman, D. C., Ruess, L., Ulrich, W., Williams, R. J., Woodward, G. & Cohen, J. E. (2006). Consumer-resource body-size relationships in natural food webs. *Ecology*, 87, 2411–2417.

Coblentz, K. E., Novak, M. & DeLong, J. P. (2023). Predator feeding rates may often be unsaturated under typical prey densities. *Ecology Letters*, 26, 302–312.

DeLong, J. P. (2021). *Predator ecology: Evolutionary ecology of the functional response.* Oxford University Press.

DeLong, J. P. & Uiterwaal, S. F. (2018). The FoRAGE (Functional Responses from Around the Globe in all Ecosystems) database: a compilation of functional responses for consumers and parasitoids. *Knowledge Network for Biocomplexity*, doi:10.5063/F17H1GTQ.

Denny, M. (2014). Buzz Holling and the functional response. *The Bulletin of the Ecological Society of America*, 95, 200–203.

Dunn, R. P. & Hovel, K. A. (2020). Predator type influences the frequency of functional responses to prey in marine habitats. *Biology letters*, 16, 20190758.

Freedman, H. (1976). Graphical stability, enrichment, and pest control by a natural enemy. *Mathematical Biosciences*, 31, 207–225.

Garay, J. (2019). Technical review on derivation methods for behavior dependent functional responses. *Community Ecology*, 20, 28–44

Grainger, T. N., Senthilnathan, A., Ke, P.-J., Barbour, M. A., Jones, N. T., DeLong, J. P., Otto, S. P., O'Connor, M. I., Coblentz, K. E., Goel, N., Sakarchi, J., Szojka, M. C., Levine, J. M. & Germain, R. M. (2022). An empiricist's guide to using ecological theory. *The American Naturalist*, 199, 1–20.

Griffen, B. D. (2021). Considerations when applying the consumer functional response measured under artificial conditions. *Frontiers in Ecology and Evolution*, 9, 461.

Hastings, A., Abbott, K. C., Cuddington, K., Francis, T., Gellner, G., Lai, Y.-C., Morozov, A., Petrovskii, S., Scranton, K. & Zeeman, M. L. (2018). Transient phenomena in ecology. *Science*, 361.

Hatton, I. A., McCann, K. S., Fryxell, J. M., Davies, T. J., Smerlak, M., Sinclair, A. R. E.

& Loreau, M. (2015). The predator-prey power law: Biomass scaling across terrestrial and aquatic biomes. *Science*, 349, aac6284.

Holling, C. S. (1959). Some characteristics of simple types of predation and parasitism. *The Canadian Entomologist*, 91, 385–398.

Jeschke, J. M. (2007). When carnivores are "full and lazy". *Oecologia*, 152, 357–364.

Jeschke, J. M., Kopp, M. & Tollrian, R. (2002). Predator functional responses: Discriminating between handling and digesting prey. *Ecological Monographs*, 72, 95–112.

Jeschke, J. M., Kopp, M. & Tollrian, R. (2004). Consumer-food systems: why type I functional responses are exclusive to filter feeders. *Biological Reviews*, 79, 337–349.

Kalinkat, G., Rall, B. C., Uiterwaal, S. F. & Uszko, W. (2023). Empirical evidence of type III functional responses and why it remains rare. *Frontiers in Ecology and Evolution*, 11.

Koen-Alonso, M. (2007). A process-oriented approach to the multispecies functional response. In: *From Energetics to Ecosystems: The Dynamics and Structure of Ecological Systems* (eds. Rooney, N., McCann, K. S. & Noakes, D. L. G.). Springer, Dordrecht, pp. 2–32.

Li, Y., Rall, B. C. & Kalinkat, G. (2018). Experimental duration and predator satiation levels systematically affect functional response parameters. *Oikos*, 127, 590–598.

Lotka, A. J. (1925). *Elements of physical biology*. Williams & Wilkins.

May, R. M. (1972). Limit cycles in predator-prey communities. *Science*, 177, 900–902.

McCauley, D. J., Gellner, G., Martinez, N. D., Williams, R. J., Sandin, S. A., Micheli, F.,

Mumby, P. J. & McCann, K. S. (2018). On the prevalence and dynamics of inverted trophic

pyramids and otherwise top-heavy communities. *Ecology Letters*, 21, 439–454.

Mills, N. J. (1982). Satiation and the functional response: a test of a new model. *Ecological*

Entomology, 7, 305–315.

Murdoch, W. W. & Oaten, A. (1975). Predation and population stability. *Advances in Ecological*

Research, 9, 1–131.

Novak, M. (2010). Estimating interaction strengths in nature: experimental support for an

observational approach. *Ecology*, 91, 2394–2405.

Novak, M. & Stouffer, D. B. (2021a). Geometric complexity and the information-theoretic

comparison of functional-response models. *Frontiers in Ecology and Evolution*, 9, 776.

Novak, M. & Stouffer, D. B. (2021b). Systematic bias in studies of consumer functional responses.

Ecology Letters, 24, 580–593.

Novak, M. & Stouffer, D. B. (2024). Corrigendum: Geometric complexity and the information-

theoretic comparison of functional-response models. *Frontiers in Ecology and Evolution*, 12.

Novak, M., Wolf, C., Coblentz, K. E. & Shepard, I. D. (2017). Quantifying predator dependence

in the functional response of generalist predators. *Ecology Letters*, 20, 761–769.

Okuyama, T. (2010). Prey density-dependent handling time in a predator-prey model. *Community Ecology*, 11, 91–96.

Perkins, D. M., Hatton, I. A., Gauzens, B., Barnes, A. D., Ott, D., Rosenbaum, B., Vinagre, C. & Brose, U. (2022). Consistent predator-prey biomass scaling in complex food webs. *Nature Communications*, 13, 4990.

Rall, B. C., Brose, U., Hartvig, M., Kalinkat, G., Schwarzmüller, F., Vucic-Pestic, O. & Petchey, O. L. (2012). Universal temperature and body-mass scaling of feeding rates. *Philosophical Transactions of the Royal Society B: Biological Sciences*, 367, 2923–2934.

Real, L. A. (1977). The kinetics of functional response. *The American Naturalist*, 111, 289–300.

Rohr, T., Richardson, A. J., Lenton, A., Chamberlain, M. A. & Shadwick, E. H. (2023). Zooplankton grazing is the largest source of uncertainty for marine carbon cycling in CMIP6 models. *Communications Earth & Environment*, 4, 212.

Rohr, T., Richardson, A. J., Lenton, A. & Shadwick, E. (2022). Recommendations for the formulation of grazing in marine biogeochemical and ecosystem models. *Progress in Oceanography*, 208, 102878.

Rosenzweig, M. L. (1969). Why the prey curve has a hump. *American Naturalist*, 103, 81–87.

Rosenzweig, M. L. (1971). Paradox of enrichment: Destabilization of exploitation ecosystems in ecological time. *Science*, 171, 385–387.

Rosenzweig, M. L. & MacArthur, R. H. (1963). Graphical representation and stability conditions

of predator-prey interactions. *American Naturalist*, 97, 209–223.

Roy, S. & Chattopadhyay, J. (2007). The stability of ecosystems: A brief overview of the paradox

of enrichment. *Journal of Biosciences*, 32, 421–428.

Rubin, J. E., Earn, D. J. D., Greenwood, P. E., Parsons, T. L. & Abbott, K. C. (2023). Irregular

population cycles driven by environmental stochasticity and saddle crawlbys. *Oikos*, 2023,

e09290.

Seo, G. & Kot, M. (2008). A comparison of two predator–prey models with Holling's type I

functional response. *Mathematical biosciences*, 212, 161–179.

Seo, G. & Wolkowicz, G. S. K. (2015). Existence of Multiple Limit Cycles in a Predator-Prey

Model with $\arctan(ax)$ as Functional Response. *Communications in Mathematical Analysis*,

18, 64 – 68.

Seo, G. & Wolkowicz, G. S. K. (2018). Sensitivity of the dynamics of the general Rosenzweig–

Macarthur model to the mathematical form of the functional response: a bifurcation theory

approach. *Journal of Mathematical Biology*, 76, 1873–1906.

Sjöberg, S. (1980). Zooplankton feeding and queueing theory. *Ecological Modelling*, 10, 215–225.

Skalski, G. T. & Gilliam, J. F. (2001). Functional responses with predator interference: Viable

alternatives to the Holling Type II model. *Ecology*, 82, 3083–3092.

Stouffer, D. B. & Novak, M. (2021). Hidden layers of density dependence in consumer feeding rates. *Ecology Letters*, 24, 520–532.

Uiterwaal, S. F., Dell, A. I. & DeLong, J. P. (2018). Arena size modulates functional responses via behavioral mechanisms. *Behavioral Ecology*.

Uiterwaal, S. F. & DeLong, J. P. (2024). Foraging rates from metabarcoding: Predators have reduced functional responses in wild, diverse prey communities. *Ecology Letters*, 27, e14394.

Uiterwaal, S. F., Lagerstrom, I. T., Lyon, S. R. & DeLong, J. P. (2022). FoRAGE database: A compilation of functional responses for consumers and parasitoids. *Ecology*, 103, e3706.

Volterra, V. (1926). Fluctuations in the abundance of a species considered mathematically. *Nature*, 118, 558–560.

Wootton, J. T. & Emmerson, M. (2005). Measurement of interaction strength in nature. *Annual Review of Ecology, Evolution, and Systematics*, 36, 419–444.