Antarctic geothermal soils exhibit an absence of regional habitat generalist microorganisms

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Abstract

Active geothermal systems are relatively rare in Antarctica and represent

metaphorical islands ideal to test microbial dispersal. In this study, we tested the

macroecological concept that high dispersal rates result in communities being dominated by

either habitat generalists or specialists by investigating the microbial communities on four

geographically separated geothermal sites on three Antarctic volcanoes (Mts. Erebus,

Melbourne and Rittman). We found that the microbial communities at higher temperature

sites (Tramway Ridge on Erebus and Rittmann) were unique from each other and were

dominated by a variety of novel *Archaea* from class *Nitrososphaeria*, while lower

temperature sites (Western Crater on Erebus and Melbourne) had characteristically

mesophilic communities (*Planctomycetes, Acidobacteriota,* etc) that were highly similar. We

found that 97% of the detected microbial taxa were regional habitat specialists, with no

generalists, with community assembly driven by high dispersal rates and drift (25 and 30% of

community assembly, respectively), not environmental selection.

Introduction

In ecology, the niche concept is commonly used to group organisms based on the

range of different habitats they live in. Habitat generalists (which tend to be metabolic

generalists (1–4)) are adapted to thrive in a broader range of different ecosystems compared

to habitat specialists, although some organisms may be specialized to a habitat in one area but

have a broader distribution beyond the scope of the current study (4). Habitat generalists that

have a global distribution are usually called "cosmopolitan" to distinguish them from other

habitat generalists (5). In most environments, the diversity of habitat specialists is greater

than generalists, but habitat generalists tend to be the most abundant organisms; this is true

both for microorganisms (1,2,6,7) and macroorganisms (8,9). Exceptions to this trend are

few, but in macroecology, high rates of dispersal between communities can lead to habitat

specialists outcompeting habitat generalists, although the opposite has also been observed

(10). In contrast, two reported exceptions to this trend in microbial ecology have been found

under situations of severe environmental selection and dispersal limitation (e.g., strong

salinity gradients (11) or the mammalian gut (12)). In microbial ecology, it is unknown which

direction high dispersal rates would push the community: would it result in habitat specialists

finding their optimal niches and outcompeting all generalists, resulting in no generalists? Or,

would it lead to an enrichment in habitat generalists that are capable of filling any niche into

which they are dispersed, given their strong ability to survive dispersal (13,14)? It is

challenging to answer this question, however, given the complexity of most systems (many

routes of dispersal, ecological disturbances, the influence of macroorganisms, etc.).

Antarctic geothermal regions offer a unique opportunity to resolve the outcome of

high dispersal rates on the proportions of habitat generalists and specialists in a microbial

community for several reasons. First, Antarctica is one of the windiest places on earth (15),

which is likely to result in strong dispersal of microorganisms between sites within

Antarctica, as has been found in the McMurdo Dry Valleys (16) and other areas of Antarctica

(17,18). Second, aerial dispersal is the only possible route of dispersal between Antarctic

geothermal regions, given the absence of aquifers and bird traffic connecting regions, which

reduces complexity. Third, Antarctic geothermal regions have relatively simple communities,

with little to no macroscopic life present (19). Finally, Antarctic geothermal regions serve as

enrichment zones for thermophiles, given the vast amount of ice and snow separating regions

and the extremely cold air conditions in Antarctica, resulting in little regional influence of

non-thermophiles on the microbial communities present. Interestingly, past studies of

thermophiles, mainly in hot spring environments, have found high niche specialization for

temperature or pH, with environmental selection being the primary factor influencing

community assembly and low dispersal rates between sites (20–24). It is unknown whether

the strong winds in Antarctica are capable of overcoming the generally low dispersal rates for

thermophiles at Antarctic geothermal regions.

The southernmost active geothermal features in the world are located in Antarctica,

surrounded by vast swathes of ice and snow: Mt. Erebus, Mt. Melbourne, Mt. Rittmann, and

Deception Island (19). Within these geothermal areas, the interplay between cold air (as low

as -60℃) and hot soils results in a large variety of unique features such as ice caves, ice

chimneys, and ice hummocks. Of these regions, Mt. Erebus is the tallest (3794m) and most

volcanically active. Two of its largest geothermal sites are Tramway Ridge and Western

Crater, which have differing maximal temperatures (65℃ vs 50℃), geothermal features

(many active fumaroles in bare soil producing large amounts of steam vs. warm soil mostly

covered by ice hummocks with less steam), physicochemistry (e.g., low pH vs. high pH), and

productivity levels (visible phototrophic mats and moss beds vs. primarily chemotrophic

driven) (25). Mt. Melbourne (2733 m tall) is located 358 km northeast of Mt. Erebus and only

has two geothermal fields, with maximal temperatures of 50℃, limited fumarolic activity,

and some moss beds (26). Geothermal activity on Mt. Rittmann (2600m tall), 450 km

northeast of Mt. Erebus and 103 km north of Mt. Melbourne, consists of one steep wall with

actively steaming fumaroles and hot soil (maximum 65℃) that is relatively shallow due to

the steep aspect (27). Given the strong maritime influence, presence of vascular plants, and

off-continent location of Deception Island north of the Antarctic circle, we excluded this site

from consideration in our study.

The microbial communities present at these unique geothermal regions have been

poorly incompletely characterized, especially at Mts. Rittmann and Melbourne. On Mt.

Erebus, culture-independent microbial community analyses have shown that cosmopolitan,

thermophilic microorganisms dominate the surface community of Tramway Ridge, while

novel and endemic members of poorly understood archaeal and bacterial groups are dominant

>2cm below the surface (5,28,29). Our study comparing the communities at Tramway Ridge

and Western Crater along thermal transects found distinct microbial communities, primarily

driven by strong differences in pH and water content between the two sites (25). Studies of

life on Mt. Melbourne have focused on macroscopic organisms (26,30,31) or were culture-

based (31–33), with the lone metagenomic dataset showing a community comprised of

Proteobacteria, *Chloroflexota*, and candidate phylum AD3 (now *Pseudomonadota*,

Chloroflexota, and *Dormibacteraeota* respectively) (34). Similarly, only culture-dependent

studies of the microorganisms at Mt. Rittmann have been conducted to-date (27,31,35,36).

These high-elevation, Antarctic geothermal sites have been theorized to have acted as refugia

for terrestrial Antarctic life during highly glaciated periods (37).

In this study, we used a combination of 16S rRNA gene amplicon sequencing and

metagenomics to resolve the proportion of habitat generalists at these sites (i.e., do we see a

large group of microorganisms common to all sites?), and determine the ecological factors

shaping community assembly. Given that the microbial communities at these sites are largely

uncharacterized, we began by describing the taxonomy and functionality of the microbes

present at these sites before answering our ecological question. We hypothesized that the

lower temperature sites (Western Crater on Erebus and Melbourne) would have different

microbial communities than the higher temperature sites (Tramway Ridge on Erebus and

Rittmann), and that there would be a high proportion of habitat generalists across sites, driven

by high rates of aerial dispersal.

Methods

Site Description and Sample Collection

Soil samples were collected from sites on Mt. Erebus, Mt. Rittmann, and Mt.

Melbourne, Antarctica, between 2010 and 2019. Sampling sites with coordinates are given in

Figure 1. On Mt. Erebus, two sites were sampled; the first was Tramway Ridge (TR), which

is the largest geothermal site on the Mt. Erebus summit and lies within Antarctic Specially

Protected Area (ASPA) No. 175. TR is relatively protected from the wind, has large numbers

of actively steaming fumaroles, producing gases enriched in CO_2 and CO (74), with hot soil

reaching 65°C in places, and has extensive moss and lichen mats covering 16% of the site

(75). The second site sampled on Mt. Erebus was Western Crater (WC), which is more

exposed to the wind and has very few active fumaroles and no visible moss or lichen mats;

soil temperatures only reach 50°C in ice-free grounds. On the Mt. Melbourne summit,

samples were collected on Cryptogram Ridge (also protected by ASPA No. 175), which,

similar to WC, is comprised of ice-free hot soil spots where temperatures reach 50°C. There

are some patches of moss on Melbourne, and the few active fumaroles on Melbourne are

covered by ice hummocks, generating CO_2 and methane-rich gases (30,75). On Mt. Rittmann,

samples were collected from the single geothermally active area on the mountain, which is

also within ASPA No. 175. This site is particularly steep (~40 degree slope), with ice free hot

soil reaching 63°C and many actively steaming fumaroles with gases of unknown

composition, as well as a few moss patches (75).

Soil temperatures were measured prior to sampling using a Checktemp 1C electronic

temperature sensor (Hanna Instruments, Rhode Island, United States). Mt. Erebus samples

were collected along temperature transects, as detailed in (25). Samples were collected using

a sterile spatula after first discarding the top centimetre of soil, except for one sample each at

Rittmann and Melbourne where the top mat-like layer of soil was taken as the sample

(Ritt3M-3 and Mel1M-2). Samples were placed into sterile 50 mL tubes and frozen for

transport to the University of Waikato (Hamilton, New Zealand) for analysis.

Geochemistry

Soil samples were analysed for geochemistry using a variety of analytical techniques.

pH and electroconductivity (EC) were measured by adding MilliQ water to soil samples at a

1:2.5 soil:water ratio. pH and EC measurements were made using a Thermo Scientific Orio 4-

Star Benchtop pH/Conductivity Meter (Thermo Fisher Scientific) using a 3-point calibration.

For gravimetric water content (GWC) measurements, 3 g of soil sample was dried at 105°C

until the sample weight remained unchanged. GWC was then calculated as the percentage of

weight loss during drying (5). For elemental composition (B, Na, Mg, P, K, Ca, V, Cr, Fe,

Mn, Co, Ni, Cu, Zn, Ga, As, Se, Sr., Ag, Cd, Ba, Tl, Pb, and U), dried samples (1g) were

prepared and analysed as described in (25). For total carbon and nitrogen measurements,

dried samples were also ground as for elemental analysis, then sent to Analytica (Hamilton,

New Zealand) for analysis. For total organic carbon (TOC) and nitrogen (TON)

measurements, samples were prepared similar to (76). Briefly, 1g of samples were weighed

out into acidified glass jars with cardboard disks in lids; jars and lids were washed as for the

ICP-MS protocol. 1 mL 6N HCl was added to samples and dried in an oven in a fume hood

for 24 hours at 95°C. 100 mg aliquots of samples were analysed for TOC and TON using an

automatic CN analyser.

DNA Extraction, Amplification, and Sequencing

For 16S rRNA gene amplification as well as metagenome sequencing, genomic DNA

was extracted from 0.5 to 0.9 g of each soil sample using a modified CTAB/bead beating

method (25). To have enough DNA from low biomass samples for metagenomes, multiple

replicates (up to 10) of samples were extracted and pooled. Genomic DNA was quantified

using a Qubit 2.0 Fluorometer (Thermo Fisher Scientific, Massachusetts, United States).

The 16S rRNA gene (V4-V5 region) was amplified using PCR with the 515F (Parada)

and 926R (Quince) primers (77,78), adapted for Ion Torrent sequencing using fusion primers

with a unique tag (79). PCR reaction conditions, library prep, and 16S rRNA gene

sequencing are the same as described in Noell, S. E. *et al.* (2022) (25).

Four samples were chosen for shotgun metagenome sequencing: Ritt1-3, Ritt3M-1

(mat sample from Rittmann), WC-50, and Mel1-3; the mat sample was included to improve

the odds of recovering cyanobacterial genomes. A fifth sample (from Tramway Ridge) was

previously extracted and sequenced using Illumina Hi-Seq; the two surface metagenome

samples from this previous project (SRR6519253 and SRR6519256) were downloaded from

the SRA using SRAtoolkit v3.0.1 and combined into one larger sequence file. For the four

new metagenomic DNA samples, extracted DNA was sent to GENEWIZ Genomics Center

(Suzhou, China). Next generation sequencing library preparations were constructed following

the manufacturer's protocol (NEBNext® Ultra™ DNA Library Prep Kit for Illumina®). For

each sample, 1 µg genomic DNA was randomly fragmented to <500 bp by sonication

(Covaris S220). The fragments were treated with End Prep Enzyme Mix for end repairing, 5'

Phosphorylation and dA-tailing in one reaction, followed by a T-A ligation to add adaptors to

both ends. Size selection of Adaptor-ligated DNA was then performed using AxyPrep Mag

PCR Clean-up (Axygen), and fragments of ~410 bp (with the approximate insert size of 350

bp) were recovered. Each sample was then amplified by PCR for 8 cycles using P5 and P7

primers, with both primers carrying sequences which can anneal with flowcell to perform

bridge PCR and P7 primer carrying a six-base index allowing for multiplexing. The PCR

products were cleaned up using AxyPrep Mag PCR Clean-up (Axygen), validated using an

Agilent 2100 Bioanalyzer (Agilent Technologies, Palo Alto, CA, USA), and quantified by

Qubit2.0 Fluorometer (Invitrogen, Carlsbad, CA, USA). Then libraries with different indexes

were multiplexed and loaded on an Illumina HiSeq instrument according to manufacturer's

instructions (Illumina, San Diego, CA, USA). Sequencing was carried out using a 2x150

paired-end (PE) configuration; image analysis and base calling were conducted by the HiSeq

Control Software (HCS) + RTA 2.7 (Illumina) on the HiSeq instrument.

Sequence Quality Control and Taxonomic Assignment

For 16S rRNA gene amplicon sequences, the raw reads from the Ion Torrent PGM

were processed using the DADA2 pipeline with recommended Ion Torrent parameters (80).

Chimeras were removed using the function removeBimeraDenovo, method "consensus."

Taxonomy assignment was performed using assignSpecies with the naïve Bayesian classifier

and SILVA database version 138.1. Eukaryotic, mitochondrial, and chloroplast sequences

were removed and an unrooted phylogenetic tree built using the neighbour-joining method

with the decipher (81) and phangorn (82) packages in R, generating final amplicon sequence

variants (ASVs).

Unfortunately, two samples (one from Tramway Ridge, TR2-64, one from Western

Crater, WC-10) produced no sequence data. As these two samples were sequenced in a

previous sequencing run, we investigated whether sequence data for these two samples from

the previous run could be safely used with the second sequence run samples. To control for

batch effects, we tested whether other samples from the two runs had similar microbial

communities and found that there was no pronounced batch effect for all samples between the

two runs (Figure S1). Thus, we decided to use sequence data for TR2-64 and WC-10 from the

first sequencing run alongside the second run; we denote these samples as TR2-64-1 and

WC-10-1 to distinguish them from samples in the second run (TR1-32-2, WC-20-2, etc.).

For metagenome sequences, demultiplexing was performed by bcl2fastq 2.17. Raw

data was filtered as follows: (1) Discard pair-end reads with adapter; (2) Discard pair-end

reads when the content of N bases is more than 10% in either read; (3) Discard pair-end reads

when the ration of bases of low quality (Q<20) is more than 0.5 in either read. All four

metagenome samples were quality checked using FastQC v0.11.5 (83) and the reports

visualized using MultiQC (84). Illumina adaptors were trimmed using trimmomatic v0.39

(85) and sequences were trimmed for quality using the following command:

trimmomatic PE -phred33 <input files> ILLUMINACLIP:NexteraPE-PE.fa:2:30:10:2:True

MINLEN:70 SLIDINGWINDOW:4:15

In light of the large amount of duplicate sequences observed in some samples

(especially the Rittmann sample) and recent evidence that suggests deduplication is beneficial

(86), we used fastp v0.23.2 (87) to deduplicate all metagenome sequences using the following

command:

fastp –in1 <input file forward reads> --out1 <output forward reads> --in2 <input reverse

reads> --out2 <output reverse reads> -A -D -G -Q -L –html <summary file>

Reads with ambiguous bases were removed using prinseq-lite v0.20.4 (88) using the -

ns_max_n 0 command.

Metagenome analysis, assembly, and binning

We assessed the taxonomy of the quality trimmed metagenomes at the read level

using singleM v0.13.2 (https://github.com/wwood/singlem) with the singleM pipe command;

we saw no differences when we used the raw, non-quality trimmed sequences or the quality

trimmed sequences (data not shown). We used Nonpareil v3.401 (38,89) to assess the fraction

of community diversity sampled by the metagenome.

For metagenome assembled genome (MAG) assembly, binning, and processing, we

used the metaWRAP pipeline v1.3.2 (90). To minimize the amount of ambiguous bases in the

assemblies, we used the non-scaffolded contigs for binning; these non-scaffolded contigs

produced bins that were similar in number and quality to the scaffolded contigs (for

scaffolded vs non-scaffolded, the bins produced had 94.1% vs 94.0% average completeness,

1.5% vs 1.6% average contamination, 0.63% vs. 0.64% average GC, 41309 vs. 40696

average N_{50}). We used prodigal v2.6.3 in metagenome mode (91,92) to predict coding genes

from the non-scaffolded assemblies; these predicted coding sequences were used for the

metagenome functional analysis. For binning contigs in the metaWRAP pipeline, we used

metaBAT2 (93), MaxBin2 (94), and CONCOCT (95). For the bin_refinement module of

metaWRAP, we used targets of 80% and 5% for completeness and contamination,

respectively. Bins were then re-assembled and cleaned with MAGpurify v2.1.2 (96) using the

phylo-markers, clade-markers, gc-content, known-contam, and clean-bin modules to produce

the final MAGs. The final statistics for these MAGs were produced using checkM v1.0.12

(97) with the lineage_wf command. We used GTDB-Tk v2.1.1 (98) for taxonomy assignment

of MAGs (GTDB database R214) and both Prodigal and Prokka v1.11 (99) for predicting

coding gene regions. We predicted optimal growth rate of MAGs using gRodon (100) and

optimal growth temperature using Tome v1.0.0 (101). Finally, we used pfam_scan

(https://github.com/aziele/pfam_scan) to predict pfam domains in MAGs (pfam release 35.0).

We calculated the presence/absence of MAGs across sites by utilising MinHash

sketching techniques implemented with sourmash (102). Specifically, we created sourmash

signatures of all MAGs and quality filtered metagenomic read sets from all four sites using a

k-mer size of 21 and scaled to 1000. These settings allowed us to query for the presence of a

MAG within a site with greater sensitivity than by using a higher k-mer or scale value. We

selected these metrics to minimize the false negative of not finding a MAG that is present,

though we acknowledge this possibly increases the false positive, something recently found

by other groups (103). Finally, we included abundance weighting in the signature creation

step of the MAG and metagenome signatures so we could measure relative abundances of

MAGs present. To query presence/absence, we used sourmash *gather* on each MAG against

each metagenome to calculate containment scores. These scores are a proxy for Average

Nucleotide Identity where a containment score of 0.2 is equivalent to an ANI of 0.95 for a

signature created with a k-mer of 21. We therefore concluded that a MAG was present within

a metagenome if its score was greater than 0.2, to account for the varying completeness levels

of the MAGs and to again minimize the false negative rate. We measured MAG abundances

within each site using the abundance weighted percent match between the MAG and the

metagenome. This correlates to the proportion of metagenomic reads that would map to the

MAG, a proxy for abundance. The final set of MAGs was deduplicated for all analyses using

a 95% ANI cutoff. Where duplicate MAGs were found, the MAG from the site with its

highest abundance was retained.

Analysis of 16S rRNA Gene Sequence Data

Using 16S rRNA gene sequences, microbial community structure was assessed using

PcoA of center log-ratio transformed (CLR) sequence abundance on a Unifrac distance

matrix using the vegdist function. Statistical differences between sites were evaluated first by

testing samples did not differ significantly in their dispersion (using an analysis of

multivariate homogeneity, PERMDISP), then by using a permutational analysis of variance

(PERMANOVA) within the R package vegan v2.6-2 (104). Alpha diversity was also assessed

within vegan via the Shannon index. Venn diagrams were plotted using the R package

ggVennDiagram v1.2.0 (105); a custom modification to the ggVennDiagram function was

used to color the sections of the Venn Diagram by the abundance of ASVs in those sections.

Heatmaps were plotted using the R package pheatmap v1.0.12 (106). The random forest

differential abundance analysis was conducted using the R package microeco v0.13.0 with

taxa level set to "all" (107).

To identify physicochemical parameters that correlated with microbial community

structure, the mantel test function within microeco was used with default parameters. The

factors that had a non-adjusted two-sided p-value < 0.05 were plotted using a distance-based

redundancy analysis (dbRDA). To test for significant differences between physicochemical

parameters, the cal_diff (method = anova) and plot_diff functions within the trans_env

function of microeco were used. To analyse clustering of sites by physicochemical

parameters, principal coordinates analysis (PcoA) of physicochemical parameters in a

Euclidian distance matrix was performed using the vegdist function in the microbiome

package v.1.12.0 (108).

Metagenome Functional Analysis

We estimated the metabolic potential of each site using a method described by

Greening and colleagues (109). In short, we used DIAMOND v. 2.1.6 (query cover > 80%) to

search quality filtered assembled ORFs of the metagenomes against custom protein databases

of representative metabolic marker genes designed by the authors. The metabolic marker

genes include multiple pathways or energy harvesting systems, namely: sulfur cycling (AsrA,

FCC, Sqr, DsrA, Sor, SoxB), succinate oxidation (SdhA), selenium cycling (YgfK),

rhodopsin (RHO), reductive dehalogenation (RdhA), phototrophy (PsaA, PsbA, energy-

converting microbial rhodopsin), oxidative phosphorylation (AtpA), nitrogen cycling

(AmoA, HzsA, NifH, NarG, NapA, NirS, NirK, NrfA, NosZ, NxrA, NorB), NADH oxidation

(NuoF), methane cycling (McrA, MmoA, PmoA), isoprene oxidation (IsoA), iron cycling

(Cyc2, MtrB, OmcB), hydrogen cycling (catalytic subunit of [NiFe]-hydrogenases, the

catalytic domain of [FeFe]-hydrogenases, and Fe-hydrogenases), fumarate reduction (FrdA),

carbon monoxide oxidation (CoxL, CooS), carbon fixation (RbcL, AcsB, AclB, Mcr, HbsT,

HbsC), arsenic cycling (ARO, ArsC), and aerobic respiration (CoxA, CcoN, CyoA, CydA).

We filtered results based on different identity thresholds: PsaA (80%), PsbA and IsoA, AtpA,

ARO, YgfK (70%), HbsT (75%), group 4 [NiFe]-hydrogenases, [FeFe]-hydrogenases, CoxL,

AmoA, NxrA, NuoF (60%), and all others with an identity threshold of 50%. We normalized

the relative abundance of present markers to RPKM.

RPKM is recommended for relative abundance comparisons within metagenomic

datasets because it normalizes the data based on both sequence depth (per million reads) and

sequence length (in kilobases) (110). We used a three-step procedure to calculate the relative

abundance of each metabolic marker in the community. First, we used quality controlled

reads to normalize coverage of ORFs to RPKM using CoverM v0.6.1 contig

(https://github.com/wwood/CoverM) (with parameters: -p bwa-mem -min-read-aligned-

percent 0.75 -min-read-percent-identity 0.95 --min-covered-fraction 0 -m rpkm). In the

second step, we aligned high-quality unassembled reads to 14 universal single-copy

ribosomal marker genes used in SingleM v.0.12.1 (https://wwood.github.io/singlem/) and

PhyloSift (111) using DIAMOND (query cover > 80%, bitscore > 40) and normalized to

RPKM as above. Finally, we calculated the average gene copy number of each marker in the

community by dividing the read count to the gene (in RPKM) by the mean of the read counts

of the 14 universal single-copy ribosomal marker genes (in RPKM). Finally, we visualized

the data using ggplot2.

Phylogenomic Tree Analysis

For the *Cyanobacteria* phylogenomic tree, we used GtoTree v1.7.00 (112), with our 9

cyanobacterial MAGs along with GTDB species cluster representatives from the same three

Cyanobacterial families, resulting in 217 members of *Nostocaceae*, 38 for *Leptolyngbyaceae*,

and 55 for *Obscuribacteraceae* (accessed on 13 October 2023). We used the *Cyanobacteria*-

specific single copy gene (SCG) set available in the GtoTree package. The SCG threshold

was lowered to 35% to keep the *Obscuribacteraceae* genomes. We used four members of the

Terrabacteria supergroup (*Streptomyces coelicolor*, *Clostridium botulinum*, *Chloroflexus*

aurantiacus, and *Meiothermus ruber*) as the outgroups to root the tree. The tree was

visualized and edited using the website Interactive Tree of Life

(https://itol.embl.de/upload.cgi).

Community Assembly Analysis

To compare the abundance of cosmopolitan and site-specific microorganisms in the

sites/samples, ASVs were categorized by the number of sites they were present at, with the

threshold for presence at a site being greater than 0 relative abundance within either the

sample or site. MAGs were classified as present or absent as described above.

To calculate niche breadth for ASVs, we first eliminated ASVs only found at one site

(1878 ASVs reduced to 832 ASVs), equivalent to a 5% sample occupancy cutoff (stricter

than (13,59) but not as conservative as (113)). Next, we used the R package MicroNiche

v1.0.0 (42), with the default settings for the limit of quantification (LOQ) threshold (q = 1.6).

Levin's B_N was used to detect generalist/specialists across the four sample sites, while

Hurlbert's B_N was used to detect generalist/specialists with regards to temperature or pH. An

adjusted p-value cut-off of 0.05 was used, and thresholds for generalist/specialist were

determined from the null model distribution produced by the package during nice breadth

calculations. Additionally, we calculated social niche breadth scores for ASVs using the

script calculate_snb.py v0.2 with default parameters (43).

To measure the strength of different ecological factors on shaping community

assembly in the 16S rRNA gene sequence data, we used the following commands within the

microeco package with default settings: trans_nullmodel, cal_ses_betamntd, cal_rcbray, and

cal_process; this approach is based on (44). To generate a distance-decay plot to look at the

influence of geographic distance on the phylogenetic differences between ASVs, we

calculated the geographical difference between sampling sites using the distm function

(distVincentyEllipsoid option) in the geosphere v1.5-18 R package (114). To calculate

biological differences in the microbial communities, we chose to use the Chao index, given

the differences in the sample sizes in our data set (115), implemented with the vegdist

function in the vegan R package.

MAG Functional Analysis

Functional annotation of MAGs was conducted using METABOLIC-G (116), using

the predicted proteins from each MAG as predicted by Prodigal (as above) with default

parameters. Specifically, we used the full KOFAM database and set the cut-off value to

assign the presence of a module to 0.75. We categorized the metabolic categories used by

METABOLIC into either Pathway (sets of genes that take a metabolite from start to end-

point) or Category (contains multiple distinct, not inter-connected pathways). We counted a

Pathway as being present in a MAG if at least 50% of the genes in the pathway were present,

and a Category as being present if there was at least one gene in that Category present.

Statistical Analyses

All statistical analyses were conducted in R (117), with plots produced in ggplot2

v3.3.6(118) and edited for aesthetics in Inkscape (https://www.inkscape.org). Data analysis

was preformed using tidyverse v1.3.1 (119). The ggpubr v0.6.0 package in R was used to add

the results of a one-way ANOVA test (when comparing all four sites and when data was

normally distributed), a one-way Krustal-Wallis test (when comparing all four sites for non-

normally distributed data), or a two-sided Mann-Whitney test (when comparing active vs.

passive site types) to plots (120).

Results

Prokaryotic Community Diversity and Composition

To characterize the microbial communities inhabiting Antarctic geothermal sites, we

extracted DNA from 21 soil samples from three geothermal areas in Victoria Land,

Antarctica: Mt. Rittmann, Mt. Melbourne, and two sites on Mt. Erebus, Tramway Ridge (TR)

and Western Crater (WC), for a total of four sampling sites (see Figure 1 for maps and site

descriptions). Two of these four sites, Melbourne and WC, have little to no fumarolic activity

(i.e., fumarolic passive) and lower maximal soil temperatures (maximum 50℃). The other

two sites, Rittmann and TR, have many actively steaming fumaroles (i.e., fumarolic active)

and higher maximal soil temperatures (maximum 65℃). Thus, we categorized WC and

Melbourne as "passive" sites and TR and Rittmann as "active" sites. We used 16S rRNA

gene amplicon sequencing to study the microbial communities present in these samples,

generating 1,871 amplicon sequence variants (ASVs) from 926,649 reads (Table S1). We

also conducted metagenome sequencing on four samples, one from each site, with almost all

metagenomes reaching sequencing saturation (nonpareil C value of 1 for all except TR,

which was 0.93; Figure S2B). Metagenome diversity, based on nonpareil diversity indices,

were 17.8, 15.6, 15.3, and 15.2 for TR, WC, Melbourne, and Rittmann, respectively, all very

low compared to other habitat microbiomes (17.5 for human-associated, 19 for lake water)

(38). From these metagenomes, we binned 140 metagenome assembled genomes (MAGs): 42

from Melbourne, 15 from Rittmann, 23 from TR, and 60 from WC, all with completeness

>80% (except for two that dropped to 77% during cleaning) and contamination <5% (Table

S2).

Overall, we found that the microbial communities at WC and Melbourne (hereafter,

passive sites) were more closely related to each other than to TR and Rittmann (hereafter,

active sites) (Figure 2). This was most apparent from the 16S rRNA samples, where passive

sites had significantly higher alpha diversity than the active sites (Figure 2A), clustered very

closely together in a principal coordinates analysis (PCoA) (Figure 2B), and were more

similar to each other based on abundance patterns both at the phylum (Figure S3A) and class

level (Figure S3B). MAGs from passive sites also clustered together based on similarities in

abundance profiles, separate from active sites (Figure S4). However, a pairwise

PERMANOVA test using the 16S rRNA gene sequence data showed no significant

difference between the microbial communities at any of the sites except between TR and WC

($p = 0.05$), which could be due to the small number of samples analysed at Melbourne and

Rittmann. TR and Rittmann also clustered separately from each other in the PCoA plot,

indicating distinct microbial communities at the two active sites (Figure 2B).

Based on the 16S rRNA gene sequence-based and metagenome-based community

composition of the sites, we found that differences between site types were primarily

attributable to greater abundances of members of *Crenarchaeota* (class *Nitrososphaeria*) at

active sites, while passive sites had significantly greater abundances of phylum

Planctomycetota (family *Gemmataceae*) and class *Acidobacteriae* within *Acidobacteriota*

(Figure 2C-D, Figure S3). When looking for distinguishing taxa at each of the four sites, we

found that TR was the easiest to distinguish from all other sites based on greater abundances

of phylum *Bacteroidota* and the archaeal family *Nitrososphaeraceae* present (Figure 2E).

Interestingly, both the Rittmann and WC metagenomes had much greater relative abundances

of *Nitrososphaeria* compared to the corresponding 16S rRNA gene sequence-based

community profile (85% vs. 20% for Rittmann, 25% vs. 5% for WC for metagenome vs. 16S

rRNA, respectively), which could indicate primer selection against this group.

We used our complete set of non-dereplicated MAGs, categorized by the site the

MAG was binned from, to examine genomic characteristics of microbes from each site. We

found that MAGs binned from active sites were distinguishable by their significantly smaller

average genome size (average: 3.1 Mbp for active, 4.8 Mbp for passive) and significantly

higher predicted optimal growth temperature (OGT) (average: 50.8℃ for active, 38.0℃ for

passive); MAGs from Melbourne had the largest average genome size and the lowest

predicted OGT of the four sites (Figure 3A-B). We did not see significant differences in

%GC, coding density, or predicted maximal growth rate between MAGs from the two site

types or the four sites (Figure S5A-B).

When we explored the physicochemical parameters that could be influencing the

microbial communities at these sites, we first found that the three volcanoes clustered

separately from each other in the physicochemical data (Figure S6A), with elemental

composition playing a large role in this separation (Figure S6B), reflecting the unique

geological origin of each volcano. When we looked for correlations between physicochemical

parameters and biological communities (16S rRNA gene sequence-based), although many

parameters were significantly correlated (Figure S6C), only magnesium remained significant

($p = 0.033$) after Benjamini and Hochberg adjustment. Values for selected parameters are

presented in Figure S6D and all physicochemical parameter values are given in Table S3.

Exploration of Cyanobacterial Diversity and Functions

We used the *Cyanobacteria* group as a case study to explore the influence of

physicochemistry on the distribution and functioning of these important primary producers

and to investigate the global uniqueness of these microorganisms. We binned an additional

five cyanobacterial MAGs from the metagenome of Ritt3M-1, a cyanobacterial mat sample

from Rittmann; this was necessary for our analysis as there were no cyanobacterial MAGs

recovered from the other Rittmann metagenome sample. In total, we retrieved 50

cyanobacterial ASVs and 9 cyanobacterial MAGs from our data set.

Overall, we found organisms from three families (*Nostocaceae*, *Leptolyngbyaceae*,

and *Obscuribacteraceae*), but could only identify a few instances of clear correlations

between physicochemical parameters and the distribution and functioning of these groups

(Figure 4A, B). Supporting earlier findings at the entire microbial community scale,

Tramway Ridge had a distinct *Cyanobacteria* community, with species from

Leptolyngbyaceae and the genus *Fischerella* (family *Nostocaceae*) dominant and an absence

of *Obscuribacteraceae*. *Leptolyngbyaceae* was positively correlated with gravimetric water

content (GWC) and negatively correlated with nickel (Figure 4C). In general, *Nostocaceae*

was the most widely distributed and had a positive relationship with nickel and sulfur and

was negatively correlated with GWC (Figure 4C). Only MAGs from *Nostocaceae* had the

capacity to fix nitrogen (Figure S7), which matches previously measured nitrogen fixation by

this group at Melbourne (26).We found that *Obscuribacteraceae* MAGs were non-

photosynthetic, as previously observed for this group (39), with pathways for ethanol

fermentation and fatty acid degradation instead (Figure S7B).

To illuminate potential evolutionary relationships between the *Cyanobacteria* MAGs

from this study and globally distributed representatives of these families from GTDB (GTDB

species cluster representatives, 217 *Nostocaceae*, 38 *Leptolyngbyaceae*, and 55

Obscuribacteraceae), we used a phylogenomic approach (Figure S8). We observed that all

MAGs from our study clustered with globally distributed members of these families, from the

Mojave Desert to China to the continental subsurface. The short branch lengths we observed

for our MAGs also confirming the cosmopolitan distribution of these families, as has been

found previously at Tramway Ridge (29). These results are dependent on the completeness of

the database used and should not be treated as conclusive evidence.

Functional Comparison of Sites

We analyzed the functional potential of the microbial communities at the four sites

using the community metagenome sequences, post-assembly, looking at the abundances of 52

conserved marker genes that characterize various key metabolic pathways (Figure 3C). All

sites had a high abundance of marker genes for aerobic organotrophic respiration (NuoF,

SdhA, AtpA, and CoxA), as well as low abundances of hydrogenases involved in

atmospheric H_2 scavenging (Group 1h and 1l NiFe hydrogenases) (40).

Several key differences were notable between sites. First, TR was distinguishable by

an enrichment in the number of genes involved in sulfur (SoxB, Sqr, and FCC) and nitrogen

(NosZ, NxrA, NirS, NirK, NorB, and NarG) metabolisms (Figure 3C; see Figure S5D for the

same heatmap in log scale for distinguishing lower abundance gene categories). These

differences are expected given the relatively large concentrations of TN and sulfur at TR

(Figure S6). A notable exception to the suite of nitrogen metabolising genes detected at TR is

the lack of AMO (ammonia monooxygenase) encoding genes, an exception which has

recently been accounted for (29). Melbourne and WC were highly similar to each other,

except for the presence of methane monooxygenase genes at Melbourne and not WC (Figure

3C). Several MAGs in the genus *Methylocella* were found at Melbourne, confirming the

potential for methane metabolism there, which aligns with large measured concentrations of

methane at Melbourne (41) (Table S2). Genes for methane monooxygenase genes were also

detected in the samples from Rittmann, but no *Methylocella* MAGs. Rittmann was

distinguishable by an enrichment in marker genes for anaerobic metabolism (FrdA) and CO

oxidation (CoxL) (Figure 3C). The gas composition of Rittmann fumaroles has yet to be

studied, but based on these results, is likely to be enriched in CO and methane.

Habitat Specialization and Community Assembly Factors

Following our taxonomic and functional analysis of these Antarctic geothermal sites,

we asked whether there was a large community of shared microorganisms between these sites

(habitat generalists) and, if so, whether dispersal between sites was an important factor in

structuring these communities. To do this, we first classified 16S rRNA gene sequence ASVs

and MAGs by the number of sites or (for ASVs) samples at which they were present, after

confirming that all samples were similarly sampled using rarefaction (Figure S2). Based on

this classification, we observed that most ASVs (90%) and MAGs (77%) in our study were

only found at either one or two sites/samples (Figure S9A-C, F, I). However, these

site/sample-specific ASVs and MAGs comprised only a small fraction of the total read count

at all samples. Instead, it was the ASVs and MAGs that were found across all four sites (36

ASVs, 3% of all ASVs; 2 MAGs, 2% of all MAGs) that were generally the most abundant

(>30% of all reads for both ASVs and MAGs) (Figure S9A, B, D, G, J). Melbourne and WC

shared the most ASVs and MAGs (120 ASVs, 9% of all ASVs; 36 MAGs, 32% of all MAGs)

(Figure S9A, B), consistent with their similar overall microbial community composition. TR

shared very few ASVs with any other site (Figure S9A, B).

Next, we asked whether these common microorganisms were adapted to thrive at all

of the sites (habitat generalists), or if their presence at all sites was a product of stochastic

processes. We used a robust implementation of Levin's niche breadth (42) with our 16S

rRNA gene sequence data (reduced to 832 ASVs by eliminating ASVs only present in one

sample) to identify specialists and generalists. Levin's niche breadth (B_N) is calculated by:

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$$Levins' B_N = \frac{1}{R \sum_{i=1} p_i^2}$$

where R is the number of differing environments and p_i is what proportion of all instances of

taxon j is found in environment i. B_N values close to 1 indicate broad habitat occupancy,

while values close to $1/R$ indicate habitat specialization. In our data set, we defined each of

our four sites as different environments and used ASV read counts for the taxon proportions.

In contrast to the theoretical distribution of niche breadth scores (Figure S10A),

surprisingly, we did not find any habitat generalists, with 95% of ASVs instead identified as

specialists and the rest as intermediate (Figure 5A). Because two of our sample sets (TR and

WC) were collected along temperature gradients, we re-ran the analysis with only high

temperature samples (>35°C), but again found 96% of ASVs were specialists; only one ASV

(ASV_5, GAL15) was returned as a generalist (Figure S10B). This absence of habitat

generalists held true both with a relaxed limit of quantification (LOQ) for including ASVs in

the niche breadth analysis, as well as when we randomly subsampled within each site to bring

the sample number from all sites to the same amount (data not shown). A recently developed

social niche breadth metric (SNB(43)) identifies specialists or generalists based on the

compositional similarity of the communities that microbes inhabit. Using this metric, we

found a distinct skew towards specialization (lower SNB score) in our data set, with a much

longer tail on the specialist side both in raw SNB scores and in modified z-scores at all

taxonomic ranks, supporting our findings of high habitat specialization (Figure S10).

A refinement to Levins' niche breadth adds in a measure of the resource availability

across environments, called Hulbert's niche breadth (42). This allows measurement of habitat

specialism or generalism in reference to environmental gradients; in our case, we chose to

examine temperature and pH, as these were identified as having the strongest correlations

with community composition (Figure S6C). Again, we found a strong skew towards

specialization for temperature and pH (95% specialists for temperature, 73% for pH) (Figure

5B, C). However, there were two ASVs identified as generalists for temperature alone (ASV

5, phylum GAL15) or both temperature and pH (ASV 42, phylum *Acidobacteriota*, genus

Bryobacter). When examining the abundance of these ASVs across temperature and pH, they

clearly had a broader distribution across different temperature/pH values than highly

specialized ASVs (Figure S10E). However, their distribution within sites is more variable

(Figure S10E).

Next, we explored what ecological factors might be driving the lack of habitat

generalists in our data set. First, we used a common method, beta nearest taxon index and

Bray-Curtis-based Raup-Crick, to calculate contribution of different factors to community

assembly in the 16S rRNA gene sequence data (44–46). We found that, across all sites,

ecological drift was the strongest factor (30.5%), with homogenous dispersal (i.e., similar

community structure caused by high dispersal rates) was the second strongest factor (25.5%)

(Figure 5D). On the other hand, 20% of the community assembly could be explained by

dispersal limitation (i.e., dissimilar community structure caused by low dispersal rates)

(Figure 5D). Environmental selection (either variable selection, i.e. dissimilar community

structure caused by different environmental conditions, or homogenous selection, i.e. similar

community structure caused by similar environmental conditions) was the weakest factor

(11.4% variable selection, 12.9% homogenous selection) (Figure 5D). Similar results were

found when comparing only active sites or only passive sites, although homogeneous

environmental selection was much stronger within passive sites than active (Figure 5D). The

importance of ecological drift and dispersal in structuring these communities was again

apparent when we examined the distance-decay plot for the 16S rRNA gene sequence data,

finding a relatively flat slope of the linear regression line (m = -0.0003), which points to

strong dispersal between sites (Figure 5E). This shallow slope held true when Active and

Passive site samples were considered separately as well (m = -0.0007 and -0.0003 for Active

and Passive, respectively) (Figure 5F, G). Slopes for similarly large-scale studies have shown

much larger slopes (e.g., (47–49)).

Functional Imprint of Habitat Specialization

Finally, we explored the impact of high levels of habitat specialization on the

microbial communities from these sites by exploring the genomic and functional profiles of

MAGs based on site distribution. Generally, habitat generalists tend to be metabolic

generalists (1,2); given the absence of habitat generalists in our data set, we predicted that the

number of sites that a microbe is present at should have little to no correlation with genome

size, number of unique protein domains (used as a proxy for metabolic breadth), or predicted

metabolic pathways present. Rather, we would expect that most microorganisms would likely

be metabolic specialists.

Indeed, we found very few significant differences (two-sided Wilcox test, Benjamini

& Hochberg-adjusted p-value < 0.05) between different MAG site occupancy categories in

genome size, number of unique pfam domains, or predicted metabolic pathways/categories

present (Figure 6A-B). We did find that MAGs occupying three sites had significantly more

metabolic pathways/categories than MAGs only occupying one site. Additionally, we

observed significantly larger genome sizes and numbers of unique pfam domains in MAGs

present at both WC and Melbourne compared with those occupying only WC (Figure 6B).

Correlations between other genome characteristics and site occupancy were non-significant

(Figure S11A-B). When we explored the ASVs that were identified previously as generalists

for temperature and pH, ASVs 5 and 42, we were able to match ASV 5 to the MAG

Ritt_bin.6.strict_final with 100% identity. This MAG had some hallmarks of generalism,

with 20.9 predicted metabolic pathways/categories but a genome size of only 3.4 Mbp. Two

MAGs (Erebus-WC_bin.32.strict_final and Mel_bin.5.orig_final) were classified to the

Bryobacteraceae family, as was ASV 42; neither MAG contained any 16S sequence, so we

cannot be sure if they match ASV 42. Still, both MAGs have genome sizes and numbers of

predicted metabolic pathways/categories on the upper end of the distribution (6.2 Mbp and

10.1 Mbp, respectively for genome size; 18.2 and 26.2, respectively, for metabolic

pathways/categories).

581   We investigated the different kind of metabolisms present in MAGs based on site

occupancy patterns and found few patterns (Figure 6C, Table S4). We did observe a trend

towards MAGs present at two or three sites being more likely to have the capacity for nitrile

hydration (6%, 17%, or 11% of MAGs present at one, two, or three sites, respectively) and

nitrate reduction (6%, 13%, or 11% of MAGs present at one, two, or three sites, respectively)

(Figure 6C). Nitrile hydration provides microorganisms with another avenue of acquiring

carbon and nitrogen needed for growth (50). Also, MAGs at one or two sites were more

likely to have the capacity for selenate reduction compared to those at three or four sites (9%,

12%, or 0% of MAGs present at one, two, or three and four sites, respectively) (Figure 6C).

Notably, the two MAGs that were present at all four sites lack many pathways/categories that

were found in other MAGs, such as ethanol fermentation, nitric and nitrous oxide reduction,

urea utilization, hydrogenases, and halogenated compound utilization, among others (Figure

6C), indicating that presence at all four sites does not require metabolic generalism.

594   As we initially hypothesized, there was no observable trend towards metabolic

specialization, rather, relatively even numbers of metabolic specialists and generalists across

sites and site occupancy patterns (Figure 6C). Rather, there were relatively even numbers of

metabolic generalists (many types of pathways, aerobic/anaerobic, multiple energy sources,

etc.) and metabolic specialists between the different site occupancy patterns. One example of

a metabolic generalist, the MAG with the second largest number of metabolic pathways

present, Ritt_bin.7.strict_final (*Gammaproteobacteria*, genus *Paraburkholderia*), appears to

be capable of, among other things, urea utilization, thiosulfate and sulfur oxidation, nitrite

reduction to ammonia, fatty acid degradation, ethanol fermentation, some types of complex

carbon degradation and C1 metabolism, carbon fixation (via RuBisCo), and aerobic CO

oxidation, yet was only present at Rittmann (Figure 6C). On the other hand, we observed

multiple metabolic generalist MAGs that also had broad habitat ranges, such as Erebus-

WC_bin.32.strict_final (*Acidobacteriota*, order *Bryobacterales*) which was found at all sites

except TR and is capable of both organic and inorganic metabolisms, and both aerobic and

anaerobic metabolism (among other things, sulfur oxidation, nitrile hydration, nitrite

reduction to ammonia, iron oxidation, fatty acid degradation, ethanol fermentation, some

level of complex carbon degradation, chlorite reduction, and aerobic CO oxidation) (Figure

6C).

One factor that seems to be largely negating the trends towards metabolic generalism

in MAGs with higher site occupancy is the presence in our dataset of a large number of

archaeal MAGs with broad habitat ranges (present at 3+ sites) with few predicted metabolic

pathways and very small genomes (mean archaeal genome size: 2 Mbp). For example,

Erebus-TR_bin.23.permissive_final (class *Nitrososphaeria*), present at all four sites, has a

limited metabolic repertoire (genes for aerobic CO oxidation, chlorite reduction, nitrite

reduction, and sulfur oxidation), with mostly aerobic pathways (Figure 6C). Similarly,

Ritt_bin.9.permissive_final (class *Nitrososphaeria*, order *Conexivisphaerales*), present at all

sites except Melbourne, is capable of the same set of metabolic pathways minus nitrite

reduction (Figure 6C).

We did observe that most MAGs only present at active sites (e.g., Rittmann only, TR

only, or Rittmann + TR only) were lacking in several metabolic categories that were found in

MAGs present in at least one passive site (Figure S11C). These pathways include methane

oxidation (unique to MAGs in Melbourne, matching the metagenome functional analysis

results), nitrile hydration, manganese cycling, nitrate reduction, urea utilization, and ethanol

fermentation (Figure S11C). These metabolic pathways do not match with measured,

corresponding nutrient concentrations (e.g., manganese and total nitrogen levels are

significantly greater at TR than WC or Melbourne (Figure S6D), despite manganese cycling

being enriched in MAGs present at WC and Melbourne compared to TR), so it is not clear

why these pathways are enriched in these MAGs.

Discussion

In this research, we have characterized the microbial communities inhabiting remote,

isolated Antarctic geothermal sites and investigated the impacts of high dispersal rates on the

abundance and distribution of habitat specialists and generalists in a patchy environment. We

found that lower temperature, passive sites (WC, and Melbourne) share very similar

microbial communities that were dominated by *Plancomycetota* and *Acidobacteriota*, while

the hotter, active sites (TR, and Rittmann), each harbored distinct communities with high

abundances of *Nitrososphaeria*. We also discovered an absence of habitat generalists in our

data set, with no microbes thriving at all four sites, driven by high dispersal between sites and

high ecological drift. It should be noted that "habitat generalists" here refers to generalists

across Antarctic geothermal sites, which is different from "globally cosmopolitan" microbes.

Both taxonomically (Figure 2, Figure S3, Figure S4, Figure S9A, B) and functionally

(Figure 3), the microbial communities at active and passive sites are distinct from each other.

The higher temperatures at active sites appear to be a major barrier preventing

microorganisms from passive sites colonizing active sites. This is apparent from the

significantly lower predicted optimal growth temperatures (OGT) of MAGs from passive vs.

active sites as well as that of MAGs from the Rittmann mat sample (32°C) vs. the Rittman

surface sample (60°C), which was collected directly underneath the Rittmann mat sample. It

is likely that the steep temperature gradient between the air (generally below -30°C) and hot

soil (60-65°C) is a large factor in facilitating microorganisms with lower temperature optima,

such as *Cyanobacteria* and *Bacteroidota*, to survive at hot spots at Rittmann and TR.

Additionally, the lack of methanotrophic activity at the active fumarolic sites is

understandable as thermophilic methanotrophy is only carried out by a limited number of

taxa at temperatures >55 °C, e.g. *Methylacidiphilum*, *Methylothermus* and putatively

Bathoarchaeota (51,52). MAGs present at both active sites and at least one passive site

tended to have higher predicted OGTs than all other MAGs (Figure 6B), which has been

observed before for thermophiles (i.e., there is a positive correlation between maximal

growth temperature and temperature growth range (21,53)).

TR and Rittmann have additional, unique physicochemical properties (nutrient

availability and soil depth) that appear to influence the microbial community composition

present. The extensive phototrophic mats and moss beds present at TR that are not found at

the other sites(30) likely provide the complex carbohydrates required for microorganisms

from *Bacteroidota* to thrive there, despite their low predicted OGT (35°C). The high levels of

total nitrogen (TN) at TR (likely, in part, due to the extensive moss beds observed) could

explain the presence of *Leptolyngbyaceae* there and its absence from other sites, given the

inability of MAGs from this family in our data set to fix nitrogen. Finally, Rittmann has

lower richness than other sites, both in terms of 16S rRNA gene sequence data (Figure 2A)

and number of MAGs binned (Rittmann was lowest at 15), perhaps due to the shallow layer

of topsoil present caused by the steep aspect. Shallow depth to bedrock has been previously

associated with lower bacterial counts and diversity (54,55). Rittmann has previously been

found to have a lower diversity of *Cyanobacteria* than Melbourne and TR (27).

We had hypothesized that high rates of dispersal between Antarctic geothermal sites

would result in a large number of shared microbes (habitat generalists) across all sites.

Although we did, indeed, find evidence of high dispersal between sites (25% of all variation

in our 16S rRNA gene sequence data can be explained by homogenous dispersal), we found

support for the opposite of our hypothesis: there were no habitat generalists between sites,

with the only shared microbial taxa present being clearly due to mass effects, not adaptation

to multiple sites. This lack of habitat generalists cannot be explained by environmental

selection, as variable environmental selection only accounted for 11% of community

differences between all sites and there were very few environmental parameters significantly

correlating with the overall biological community, and *Cyanobacteria* in particular (Figure

6C). Our results stand in stark contrast to past studies of thermophiles(22,45,56,57) and

habitat specialists(58,59) that have found environmental selection to be the primary factor

influencing distribution, not dispersal, although one study did find a large influence of

dispersal on the abundant biosphere of Tibetan hot springs (7). We note that many

thermophilic habitat generalists have been identified before, even along strong environmental

gradients, e.g. (21,60–62), so an absence of habitat generalists in our data set is surprising,

although many thermophiles tend to be habitat specialists for temperature (20,45,63).

Our findings provide key insights into how microbial communities assemble, with

high dispersal rates and ecological drift leading to a lack of habitat generalists. These results

are similar to past macroecological studies that have found specialist species are favored

when dispersal is high and conditions are stable (i.e., low short-term habitat disturbance) (64–

68). High dispersal rates allow specialist species to reach optimal niches and outcompete

generalist species. This does not preclude the existence of globally cosmopolitan microbes at

these sites, as has been postulated previously (5), given that cosmopolitan species may be

specialists for certain environments. Additionally, highly stable conditions like those

experienced at these sites (28), partially a result of their isolation and the absence of outside

disturbances, results in few opportunities for habitat generalists to take over niches vacated

by specialists during a disturbance. High environmental stability has been found in

macroecology to result in a selection towards K-strategists (lower growth rate, few offspring)

(68), which may be a contributing factor in the slow predicted growth rates of MAGs in our

data set (average across all four sites: 9.2 d^{-1}, fastest: 1.9 d^{-1}). The other factor influencing the

lack of habitat generalists in our data set is the large effect of ecological drift (30% of

community assembly), likely due to the small population size at these sites. In macroecology,

ecological drift can override niche effects in small populations, resulting in a co-occurrence

of two species even when one is a superior competitor (69).

The high level of habitat specialization at these sites may also be indicative of high

levels of endemism, as these two have been found to be correlated in macroecological studies

(70,71). Antarctica is home to an unusually high level of endemic organisms due to its

geographic isolation (72). Tramway Ridge has previously been found to harbor a number of

endemic microbes (29). Additionally, the most abundant organism across all sites was a

highly divergent member of the class *Nitrososphaeria* that lacks genes for ammonia

oxidation, despite this class being primarily known for its ammonia oxidation activities (73).

The functional potential of this MAG has previously been inferred, with the proposed name

Candidatus Australlarchaeum erebusii (29). Despite its high abundance, this MAG (Erebus-

TR_bin.23.permissive_final) had the smallest genome size (1.4 mbp) in our data set.

Culturing efforts are needed to better understand the mechanisms by which this novel

archaeon is capable of wide habitat range despite limited metabolic flexibility, especially

given that *Archaea* are generally much more dispersal limited than *Bacteria* (56). Our hope is that further microbial explorations at these unique sites, with broader sampling at more areas and further metagenomic, transcriptomic, and culturing work, will help uncover the wealth of novel microorganisms present.

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Data Availability

All R scripts used to analyze the data are available on Github at https://github.com/ThermophileResearchUnit/Antarctic_volcano_comparison_manuscript. 16S rRNA gene sequence data has been deposited at DDBJ/EMBL/GenBank under the accession KIGX00000000. The version described in this paper is the first version, KIGX01000000. Metagenome reads are available on SRA under BioProject ID PRJNA1113340.

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Figure Legends

Figure 1 Map and images of sampling sites. (A) Satellite image of Antarctica (top right inset, red circle indicates Victoria Land); main image shows Victoria Land and Ross Island. (B) Aerial picture of Mt. Erebus main crater. (C) Picture of the hot soil site sampled at Western Crater on Mt. Erebus. (D) Satellite image showing the summit of Mt. Erebus and the two sampling sites in this study, Tramway Ridge (-77.519444, 167.116389 DD) and Western Crater (-77.519167, 167.187778 DD). (E) Image of sampling site on Tramway Ridge. (F) Image of sampling on the side of Mt. Rittmann (-73.448445, 165.497444 DD). (G) Image of Mt. Rittmann. (H) Image of Mt. Melbourne. (I) Image of sampling at a hot soil site on Mt. Melbourne (-74.350016, 164.700139 DD). Satellite image in (A) is from the Antarctic Digital Database Map Viewer https://www.add.scar.org/, Open Source, using hillshade and bathymetry layers. Satellite image in (D) was purchased from DigitalGlobe Incorporated, Longmont CO USA (2019). All other images were taken by Ian McDonald and S. Craig Cary.

Figure 2 Diversity and composition of microbial communities across the four sites. (A) Alpha diversity of microbial communities based on 16S rRNA gene sequences across the four sites (Tramway Ridge (TR), Western Crater (WC), Melbourne, or Rittmann, n = 8, 6, 4, or 3 respectively) using the Shannon index of diversity. Paired significance test used an undirected Wilcox test. (B) Principal coordinates analysis (PCoA) of the microbial communities based on Unifrac similarities between 16S rRNA gene sequences. The percentages in brackets for each axis indicate the percentage of the variation in community composition explained along that axis. (C) Microbial community composition at the phylum level for both 16S rRNA gene sequencing samples (16S) and metagenome samples at the read level (Metagenome reads); singleM was used for determining taxonomy for the metagenome reads. For the 16S rRNA gene sequence data, the sequencing run used for

Tramway Ridge (TR) and Western Crater (WC) samples are denoted by the last digit, either 1

for the first (data from (25)) or second sequencing run. For this plot, phylum names for

metagenome reads were changed to match SILVA 16S rRNA gene taxonomy to allow for

ease of comparison between the two methods (*Thermoproteota* to *Crenarchaeota*,

Chloroflexota to *Chloroflexi*, CSP1-3 to GAL15, *Dormibacterota* to *Chloroflexi* (class AD3),

Eremiobacterota to WPS-2). (D, E) Results of a differential abundance random forest

analysis to identify taxa at any level that are most important for distinguishing between (D)

active and passive sites, or (E) all four sites in the 16S rRNA gene sequence data. The plots

on the left give the mean decrease in Gini value, which is a measure of the importance of that

taxa to distinguishing the sites. The plot on the right shows the relative abundance of those

taxa at the respective sites. A threshold of either 0.2 or 0.1 relative abundance was used for

plotting ASVs for the active vs. passive comparison or all four sites comparison, respectively.

Figure 3 Genomic and functional characteristics of MAGs and metagenomes from each site.

(A, B) Comparison of genome characteristics for MAGs recovered from different site

categories (A) or sites (B). (A) n = 32 or 79 for active or passive, respectively (*p*-values: two-

sided Wilcox test). (B) n = 19, 37, 42, or 13 for Tramway Ridge (TR), Western Crater (WC),

Melbourne, or Rittmann, respectively. p-values given for the pairwise comparisons are for

two-sided Wilcox tests. (C) Analysis of the functional potential of the microbial communities

at the four site samples with metagenomes, based on the abundance of marker genes for

different metabolic pathways. Abundance within community was calculated using RPKM

values, which were normalized by the abundance of single copy marker genes within each

metagenome. TR: Tramway Ridge. WC: Western Crater.

Figure 4 Analysis of Antarctic geothermal *Cyanobacteria* diversity and factors shaping it.

Abundance heatmaps of *Cyanobacteria* communities using (A) 16S rRNA gene sequence and

(B) metagenome assembled genome (MAG) data. For (A), relative abundances of

Cyanobacteria families are presented as relative to the abundance of other Cyanobacteria.

For (B), MAG abundances were marked as zero if the MAG was not present at the site. Then,

the abundances were normalized: $\log_{10}(\text{abundance} + 0.001) + 4$. (C) Correlation heatmap,

using the Pearson coefficient, of the cyanobacterial families to the relevant environmental

variables (selected for having Mantel test $p < 0.05$). GWC: gravitational water content; TN:

total nitrogen.

Figure 5 Ecological predictions for microorganisms at Antarctic geothermal sites. (A-C)

Frequency plots identifying the number of either specialists or generalists for (A) habitat

(using Levin's niche breadth, B_n), (B) temperature (using Hulbert's niche breadth, B_n), or (C)

pH (using Hulbert's niche breadth, B_n) in the 16S rRNA gene amplicon sequence data.

Arrows point to generalists where difficult to visualize. (D) Contribution of different

ecological factors to shaping the community assembly of microorganisms in the 16S rRNA

gene amplicon data. (E, F, G) Distance-decay plots for the 16S rRNA gene amplicon data

with a linear regression best fit model for (E) all samples, (F) only samples from Active sites,

or (G) only samples from Passive sites.

Figure 6 (A-B) Exploring evidence for genomic markers of metabolic generalism in

metagenome assembled genomes (MAGs) from the four sites. For (A), n = 35, 52, 27, or 2

for the categories 1 site, 2 sites, 3 sites, or 4 sites, respectively. p-values for ANOVA tests

conducted in (A) are, from top left clockwise order: 0.08, 0.07, 0.04, and 0.02. p-value shown

on plot is the result of an undirect Wilcox test, Benjamini-Hochburg corrected. No

differences for pfam domains remained significant after BH correction. For (B) and (C), n =

11, 2, or 7 for Tramway Ridge (TR) only, Rittmann only, or TR/Rittmann only MAGs,

respectively; n = 5, 3, 8, 15, or 2 for TR/Western Crater (WC) only, WC/Rittmann only,

TR/WC/Rittmann only, WC/Melbourne/Rittmann only, or MAGs at all four sites,

respectively; n = 10, 11, or 36 for WC only, Melbourne only, or WC/Melbourne only MAGs,

respectively. For (B), * represents a statistically significant ($p < 0.05$) result of an undirected

Wilcox test after Benjamini-Hochburg correction. See Methods for how Metabolic

Pathways/Categories were marked as present or absent in MAGs. (C) Heatmap showing

functional predictions for MAGs based on METABOLIC predictions, separated by site

occupancy (number and identity of sites occupied) and whether the metabolic predictions are

categories (containing multiple, non-sequential metabolic reactions) or pathways. MAGs that

are specifically mentioned in the text are indicated by an arrow. Pink boxes differentiate

different site occupancy patterns. X-axis MAGs are clustered first by site occupancy pattern,

then alphabetically by Phylum-Class. TR: Tramway Ridge. WC: Western Crater. RT:

Rittmann. ML: Melbourne.

C

B

A

Mt. Erebus

D

0.5 km

Western
Crater

Tramway
Ridge

E

I

H

Mt. Melbourne

100 km
50 mi

N

Mt. Rittmann

G

F