

1 **Galápagos yellow warblers in low- and high-traffic environments differ in**  
2 **song plasticity in response to noise.**

3

4 Leon Hohl <sup>\*1</sup>, Alper Yelimlieş <sup>\*1,2</sup>, Çağlar Akçay <sup>3</sup>, Sonia Kleindorfer <sup>1,2,4</sup>

5

6 <sup>\*</sup>co-first authors

7 <sup>1</sup> Konrad Lorenz Research Center for Behavior and Cognition, University of Vienna, Grünau im

8 Almtal, 4645 Upper Austria, Austria

9 <sup>2</sup> Department of Behavioral and Cognitive Biology, University of Vienna, 1030 Vienna, Austria

10 <sup>3</sup> School of Life Sciences, Anglia Ruskin University, Cambridge, UK

11 <sup>4</sup> College of Science and Engineering, Flinders University, Adelaide 5001, Australia

12

13 **ORCID:**

14 LH: 0009-0001-4485-7685

15 AY: 0000-0003-2319-6141

16 CA: 0000-0003-0635-9586

17 SK: 0000-0001-5130-3122

18

19 **Corresponding Author:**

20 Çağlar Akçay

21 School of Life Sciences, Anglia Ruskin University, East Road, Cambridge, UK CB1 1PT

22 Email: [caglar.akcay@aru.ac.uk](mailto:caglar.akcay@aru.ac.uk)

23 Phone: +44 (0)1245 493131

24 **Galápagos yellow warblers in low- and high-traffic environments differ in**  
25 **song plasticity in response to noise.**

26 **Abstract**

27 Acoustic communication between animals is increasingly disrupted by noise in human-  
28 altered environments. This can make signals less effective and create selection pressure to  
29 increase their detectability. A good example of this is bird song. Song is a signal used in  
30 agonistic interactions between territorial rivals, and birds may modify their song and singing  
31 behaviour in response to noise. However, if these modifications are still ineffective, this can lead  
32 to increased conflict between rivals. Here, we asked whether experimental traffic noise induces  
33 immediate changes in acoustic characteristics of song and aggressive behaviour in populations  
34 that differ greatly in traffic noise exposure. We conducted simulated territorial intrusions on  
35 Galápagos yellow warblers (*Setophaga petechia aureola*) living on Santa Cruz (high traffic) and  
36 Floreana (low traffic) islands. We assessed the focal birds' physical response levels and recorded  
37 their vocalisations in response to playback of conspecific song (control) and conspecific song  
38 coupled with traffic noise (noise treatment). Although we found that physical response levels did  
39 not differ between islands or treatment types, birds on each island adjusted their songs differently  
40 in response to noise. In particular, birds increased the minimum frequency of their songs in  
41 response to the noise treatment, and the increase was more pronounced on Floreana Island (low  
42 traffic) than on Santa Cruz (high traffic). In contrast, song duration increased in response to noise  
43 treatment only on the high-traffic island. Our results suggest behavioural flexibility in birdsong  
44 in response to traffic noise, which appears to depend at least in part on prior experience with  
45 traffic noise.

46 **Keywords:** Anthropogenic noise, Birdsong, Aggression, Galápagos, Yellow Warbler

47

48

49 **Highlights**

- 50 ● Galápagos yellow warblers on different islands modified their song structure in an  
51 experience-dependent way when exposed to a simulated intruder and traffic noise.
- 52 ● In Santa Cruz (higher traffic) birds increased the duration of their songs in response to  
53 experimental noise while in Floreana (lower traffic) they did not.
- 54 ● Santa Cruz birds sang with higher minimum frequencies in general, but Floreana birds  
55 increased minimum frequencies in response to noise more than the Santa Cruz birds
- 56 ● Birds on neither island changed their physical aggression towards the simulated intruder  
57 dependent on experimental noise.

58

59

60

61

62

63

64

65

66

67

68

69

70 Interference from noise during signal transmission has important implications for animal  
71 communication (Bradbury & Vehrencamp, 2011). Given the rapidly increasing scale of human  
72 activities, animals need to produce and perceive signals in an ever-noisier world (Brumm, 2013;  
73 Halfwerk & Slabbekoorn, 2015). This is especially prevalent in the acoustic domain, where loud  
74 noises caused by traffic, construction, or other activities may mask animal vocalisations with  
75 implications for lower reproductive success and/or survival (Bowen et al., 2020; De Jong et al.,  
76 2018; Grade & Sieving, 2016).

77         The effect of anthropogenic noise on vocalisations has been most widely studied in  
78 songbirds (Brumm & Zollinger, 2013; Duquette et al., 2021). Bird song is a culturally  
79 transmitted acoustic signal that functions in both inter-sexual communication such as mate  
80 attraction as well as communication within each sex, such as territory defence against same-sex  
81 competitors (Catchpole & Slater, 2008). signaling in territorial disputes can be mutually  
82 beneficial because both parties can assess their rivals without the cost of fighting, which may  
83 result in injury or death (Logue et al., 2010; Maynard Smith & Harper, 1988). When  
84 anthropogenic noise masks animal signals, however, assessment of rivals via signals is impaired,  
85 increasing the probability that the opponent will need to resort to a conflict involving physical  
86 aggression. In support of this scenario, there is evidence from several species of songbirds that  
87 high levels of acoustic noise are correlated with increased physical aggression (Akçay et al.,  
88 2020; Davies & Sewall, 2016; Phillips & Derryberry, 2018). Of course, increased aggressiveness  
89 in urban environments could also be explained by other variables such as resource availability  
90 (Foltz et al., 2015), but experimental studies have shown that at least in some cases noise can  
91 lead to increased territorial aggression in birds (Grabarczyk & Gill, 2019; Önsal et al., 2022; but  
92 see Kleist et al., 2016).

93           To improve transmission of signals in noise, songbirds may use one of several  
94 adjustments to their signals and signaling behaviours (Brumm & Zollinger, 2013; Kunc &  
95 Schmidt, 2021). One common adjustment is to increase the amplitude of the signal, known as the  
96 Lombard effect (Brumm & Zollinger, 2011; Hage et al., 2013; Kunc et al., 2022). Alternatively,  
97 they can increase the redundancy of the signal through repetition or by singing for longer  
98 (Brumm & Slater, 2006; Ríos-Chelén et al., 2013). Increasing redundancy would be potentially  
99 more effective in response to intermittent noise (as opposed to constant noise) as there would be  
100 a possibility of at least partially avoiding masking during signal production.

101           Because anthropogenic noise generally occupies lower frequency bandwidths, another  
102 possibility to reduce the effects of masking is to change the spectral parameters of the  
103 vocalisations (Roca et al., 2016; Slabbekoorn & Peet, 2003). One possibility is to raise the  
104 frequency of vocalisations so that there is less overlap with low-frequency noise. Such changes  
105 in frequency may come about through either ontogenic effects (Moseley et al., 2018; but see Liu  
106 et al., 2021), or through immediate and short-term plasticity (Bermúdez-Cuamatzin et al., 2011;  
107 Verzijden et al., 2010). It is worth noting that the changes in frequency may partly come about as  
108 a by-product of the Lombard effect as singing louder can lead to higher frequencies (Nemeth &  
109 Brumm, 2010).

110           Whether animals can cope with noise by changing their signaling may depend on  
111 previous experience with noise (Courter et al., 2020; Gallego-Abenza et al., 2020; LaZerte et al.,  
112 2016). There are only a few studies that have directly examined whether previous experience  
113 with noise affects short-term changes in song in response to acute noise. In one study, LaZerte  
114 and colleagues (2016) found that black-capped chickadees (*Poecile atricapillus*) shifted their  
115 song frequencies to higher frequencies in response to noise only if they already lived in noisier  
116 territories. By contrast, birds in quieter territories shifted their song frequencies downward. In a

117 similar vein, urban, but not rural, white-crowned sparrows (*Zonotrichia leucophrys*) produced  
118 songs with lower maximum frequencies and bandwidths when exposed to experimental noise  
119 (Gentry et al., 2017). Given that urban birds live in noisier territories this again suggests a role  
120 for experience. In contrast to these results, other studies found that in at least some species,  
121 individuals living in rural or quieter areas can also respond to experimental noise with immediate  
122 plasticity in vocalisations (Bermúdez-Cuamatzin et al., 2011; Potvin & Mulder, 2013; Verzijden  
123 et al., 2010). Here we test whether immediate acoustic and behavioural plasticity in the face of  
124 experimental noise is dependent on previous experience with chronic noise in two island  
125 populations of a common songbird in the Galápagos archipelago, the Galápagos yellow warbler  
126 (*Setophaga petechia aureola*).

127         While long considered a natural laboratory to study natural selection (B. R. Grant &  
128 Grant, 1979; P. R. Grant & Grant, 2014), the Galápagos Archipelago is not spared the human  
129 footprint and is now increasingly also a source of research into urban ecological frameworks that  
130 explore human-wildlife conflict and the evolutionary dynamics of endemic island systems. In  
131 Darwin's finches, there is a growing body of research into the effects of urbanisation on altered  
132 foraging behaviour (De León et al., 2019; Rivkin et al., 2021), nesting success (Harvey et al.,  
133 2021), boldness (Gotanda, 2020), song phenotypes and aggressive behaviour (Colombelli-Négrel  
134 et al., 2023). Given the large growth in the human population on the Galápagos Islands (Kricher  
135 & Loughlin, 2022; INEC 2016), there is a concomitant increase in the number of trucks and cars,  
136 and as a result higher levels of noise as well as more roadkill (García-Carrasco et al., 2020).  
137 Between 1980 and 2013, there was a 57-fold increase in traffic on Santa Cruz Island (23 vehicles  
138 in 1980 and 1326 vehicles in 2013; Márquez 2000; García-Carrasco et al. 2020). The endemic  
139 Galápagos yellow warbler is widely distributed across the archipelago, both on inhabited and  
140 uninhabited islands, and is the most commonly killed bird (> 70%) among roadkill on the

141 heavily trafficked main road of Santa Cruz Island (García-Carrasco et al., 2020). Being  
142 commonly found across the human disturbance gradient, the Galápagos yellow warbler has the  
143 potential to be an important study system for investigating the impacts of human activities on  
144 island birds.

145 Here we aim to measure behavioural and song structure plasticity in response to a  
146 simulated intruder under traffic noise in the Galápagos yellow warblers and to compare the  
147 magnitude of response between two islands (Santa Cruz and Floreana) that differ in human  
148 population size and traffic. If the birds show behavioural and song structure plasticity in their  
149 response to an intruder song in a noisy environment, we expect that individuals will modify their  
150 songs and singing behaviours to increase detectability (increase in frequency, longer duration,  
151 more songs). Crucially, if the ability to modify song structure depends on experience with noise  
152 (Gentry et al., 2017; LaZerte et al., 2016), more modifications will be made on the island with  
153 more traffic exposure (Santa Cruz). Finally, we also expected that if birds do not modify their  
154 song or modifications are ineffective in dealing with experimentally increased noise, they will  
155 increase physical aggression by more closely approaching the simulated intruders and engaging  
156 in more search behaviours for the intruder.

## 157 **Methods**

### 158 *Study species and sites*

159 The Galápagos yellow warbler is an abundant songbird that occurs on all vegetated  
160 islands and across all habitats of the Galápagos Archipelago, Ecuador (Chaves et al., 2012;  
161 Snow, 1966). It is a small (~12 g) songbird that is year around territorial (Snow, 1966). Both  
162 sexes sing distinctive songs (Yelimlies et al. unpublished data) to defend the territory. Males  
163 have a reddish-brown crown and a chest streaking of the same colour against a yellow plumage,

164 whereas females are uniform yellow (Salgado-Ortiz et al., 2008), making the sexes  
165 distinguishable using binoculars in the field.

166 We conducted the experiment in the lowlands and highlands on Santa Cruz Island and  
167 Floreana Island. All study areas have previously been part of long-term annual nesting  
168 monitoring of Darwin's finch breeding biology monitored by our research group since 2000  
169 (Common et al., 2022; Kleindorfer et al., 2021). Santa Cruz highland study plots were located in  
170 the Los Gemelos area (-0.625982, -90.384829), and the lowland study plots in the El Barranco  
171 area (0.739068, -90.301467). The Floreana Island highland study plots were located in the Cerro  
172 Pajas (-1.299829, -90.455674) and Asilo de la Paz (-1.313595, -90.454935) areas, and the  
173 lowland sites near the Loberia area (-1.282974, -90.49208). Both islands have permanent human  
174 settlements but differ in human population size and number of vehicles (INEC 2016; Galápagos  
175 Conservancy 2023). Santa Cruz has a population of over 15,000 humans (INEC 2016) and over  
176 1,000 vehicles (Ovidio et al. 2009; García-Carrasco et al. 2020) compared with Floreana which  
177 has a human population of ~100 (Galápagos Conservancy 2023) and ~10 vehicles (Hohl et al.  
178 personal observation).

### 179 ***Number of vehicles***

180 To estimate relative traffic density between the two islands, we placed passive sound  
181 recorders developed by the New Zealand Department of Conservation along the main road in the  
182 highlands at Los Gemelos on Santa Cruz (28 and 29 Jan) and Cerro Pajas on Floreana Island (4  
183 and 5 Feb). The number of passing vehicles was counted by listening to the recordings at x8  
184 speed on headphones of recordings made between 0600 and 0800 and 1000 and 1200. We  
185 measured dB (A-weighted) in the field with a sound level meter (Voltcraft SL-100) standing



186 roadside to measure the maximum dBA at 3 m from passing vehicles on 28 Jan on Santa Cruz  
187 Island.

### 188 *Active territories*

189 We carried out playback trials at 38 (22 Floreana, 16 Santa Cruz) active yellow warbler  
190 territories in January and March 2022 on Santa Cruz, and February 2022 on Floreana. Focal  
191 birds were identified from 3-7 visits to the nest territories across three weeks. Sites were  
192 systematically searched every two days for the presence of male and female pairs and active  
193 nests (Lawson et al., 2023). A male or a male and female pair was considered resident if they  
194 were active on the territory over a three-day period as evidenced by nest building and/or males  
195 showing mate-guarding (following a female) and/or females exhibiting nest defence behaviours  
196 (alarm-calling, perch-switching, wing-flicking, circle-flight), which have been used to assign  
197 partulid nesting status in similar studies (Hobson & Sealy, 1989; Mitra, 1999). As in Lawson et  
198 al. (2023), if these conditions were met, the territory was included in the experiment.

### 199 *Playback stimuli*

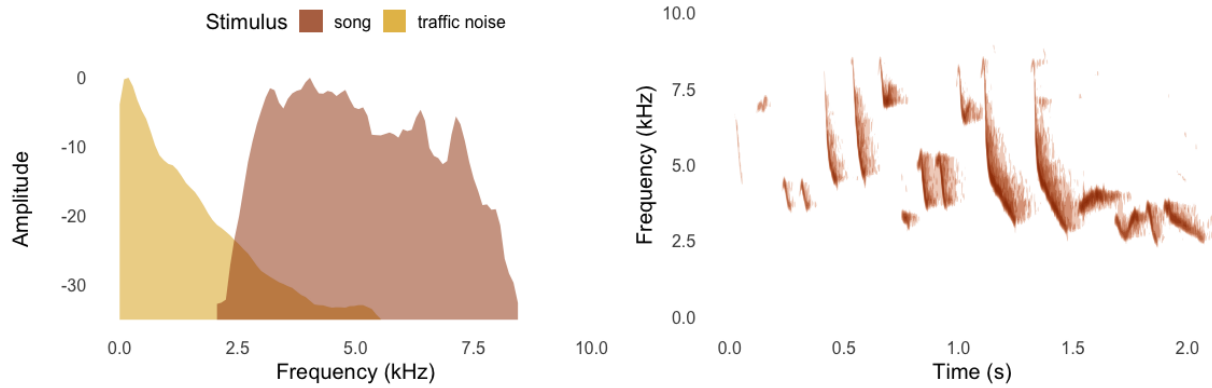
200 We prepared 15 song and 10 traffic noise stimulus files with our field recordings using  
201 the software Audacity. The 15 yellow warbler song stimuli (9 from Floreana, 6 from Santa Cruz)  
202 were created from recordings made in January and February 2022 using a Zoom H5 (Zoom  
203 Corporation, Japan) recorder with a Sennheiser MKE 600 (Sennheiser electronic GmbH & Co.  
204 KG, Germany) directional microphone. Recordings were made at a 48 kHz sampling rate and 24-  
205 bit resolution. We selected songs with a high signal-to-noise ratio from the recordings, filtered  
206 frequencies below 1500 Hz, and normalised the peak amplitude of each song. The song stimuli  
207 consisted of 2 or 3 renditions of a single song type presented at a rate of 6 songs per minute.

208

209 We generated traffic noise stimulus files using recordings of vehicles passing on the main  
210 road through Santa Cruz Island during January 2022 by standing on the roadside and pointing the  
211 microphone towards the road (using the same equipment and settings above). We edited the  
212 traffic noise recordings to first clip unique traffic events (a vehicle passing). Then, we merged on  
213 average 5.5 (SD = 1.2) events per minute to create a stimulus tape of traffic noise. In other  
214 words, our noise stimulus simulated a traffic of 5.5 vehicles per minute passing by. This noise  
215 stimulus simulates the typical intermittent and varying levels of noise (which increases and  
216 decreases as a vehicle passes) that birds experience. Figure 1 shows the average frequency  
217 spectra for the noise and the song stimuli. Each playback stimulus consisted of 4 minutes: 1-  
218 minute baseline (silence), 1-minute song (song or song coupled with traffic noise, depending on  
219 treatment), 1-minute silence or traffic noise, and 1-minute song (song or song coupled with  
220 traffic noise). The song stimulus for each subject was a local song from a non-neighbouring  
221 conspecific recorded at least 200 m away from the same island. Both stimuli (song and noise)  
222 were broadcast at 85 dBA max (measured at 1 m from the speaker with Voltcraft SL-100,  
223 Conrad Electronic, Germany).

224

225



226

227 **Figure 1.** On the left: mean frequency spectra for one example traffic noise stimulus file and one  
228 song stimulus file, showing the overlap between the two. On the right: spectrogram of the same  
229 yellow warbler song stimulus.

### 230 *Procedure and response measures*

231 Each male was tested twice, once with song only and once with song plus traffic noise,  
232 the two trials being 1 day apart except for the 3 birds that were tested 2 to 5 days apart due to  
233 rainy weather. 17 of the subjects received the noise treatment before the song treatment and for  
234 the rest the order was reversed. For each trial, we placed two JBL Clip 4 Bluetooth speakers  
235 (JBL Inc., USA) at 0.5 m and 1.5 m above ground (usually clipped on a branch). The lower  
236 speaker was used to broadcast the traffic noise on the traffic noise trials and the upper one was  
237 used for the song. After placing the speakers, two observers were positioned 10 m apart and  
238 waited until the male was seen within 20 m to start the playback. Once the male was located  
239 within range, one observer started recording with a Zoom H5 recorder with a Sennheiser MKE  
240 600 microphone and started the 4-minute playback stimulus via a smartphone. During the trials,  
241 starting with the 1-minute baseline period, both observers narrated the movements of the focal  
242 bird and recorded songs. From the narrations, we extracted the closest approach distance, number

243 of flights, and crosses to the speaker (defined as flights within a 1m radius of the conspecific  
244 playback speaker). These measures are commonly used to quantify physical aggressive responses  
245 in songbirds (Falls & Brooks, 1975; Searcy et al., 2006).

#### 246 *Song analysis*

247 From the trial recordings, we extracted high-quality songs that did not overlap with other  
248 vocalisations using the software Raven Pro 1.6.4 (K. Lisa Yang Center for Conservation  
249 Bioacoustics, 2022). The resulting dataset consisted of 490 songs (Santa Cruz = 225, Floreana =  
250 265) from 36 birds (Santa Cruz = 15, Floreana = 21). We then measured the minimum and  
251 maximum frequency of the selected songs -20 dB relative to the peak frequency (i.e. the  
252 threshold method, Brumm et al., 2017); and calculated bandwidth as their difference using the  
253 freq\_detec function from the warbleR package in R (Araya-Salas & Vidaurre, 2017; R Core  
254 Team, 2023). We also measured the peak frequency and the duration of the songs within Raven  
255 Pro (for the latter we used the duration 90% function which measures the duration in each  
256 selection that is between the times dividing the selection to the 5th and 95th percentile of the  
257 energy in the selection).

#### 258 *Statistical analysis*

259 All data were analysed using R. Because our measures of physical approach were all correlated  
260 with each other (minimum distance, crosses, flights) we used a principal component analysis  
261 (PCA) using the package psych in R (Revelle, 2017) for these variables (see Table 1 for the  
262 correlation matrix and coefficients). The first component (PC1) of the PCA (unrotated,  
263 correlation matrix) explained 76% of the variance and had high factor loadings for behavioural

264 response associated with physical aggression so it was taken as the aggression scores. A high  
265 aggression score (PC1) indicates a close approach and many flights and crosses over the speaker.

266 To test our hypothesis that behavioural response differs when exposed to an intruder in a  
267 noisy environment, we created a linear mixed model (LMM) using the aggression scores (PC1)  
268 as the dependent variable, island and treatment as the fixed factors, and the individual ID as the  
269 random factor. We also analysed the number of songs and the song structure parameters using  
270 separate LMMs with the same fixed and random factors. All models were created in R using  
271 package nlme (Pinheiro et al., 2022).

### 272 ***Ethical note***

273 This study complies with all current Austrian laws and regulations and was supported by  
274 Animal Experiment License Number 66.006/0026-WF/V/3b/2014 issued by the Austrian Federal  
275 Ministry for Science and Research (EU Standard, equivalent to the Animal Ethics Board). The  
276 subjects were not captured during the experiment and each trial only took 3 minutes. The birds  
277 returned to their normal activities within a few minutes.

### 278 **Results**

#### 279 ***Traffic conditions between islands***

280 There was significantly more vehicle traffic on Santa Cruz Island ( $62 \pm 7.8$  per hr) than  
281 on Floreana Island ( $1.8 \pm 0.3$  per hr) ( $t = 7.679$ ,  $df = 6$ ,  $P < 0.001$ ). The average amplitude of 27  
282 vehicles recorded at 3 m was  $84.7 \pm 0.7$  dBA max, with a minimum of 79.9 and a maximum of  
283 95.4.

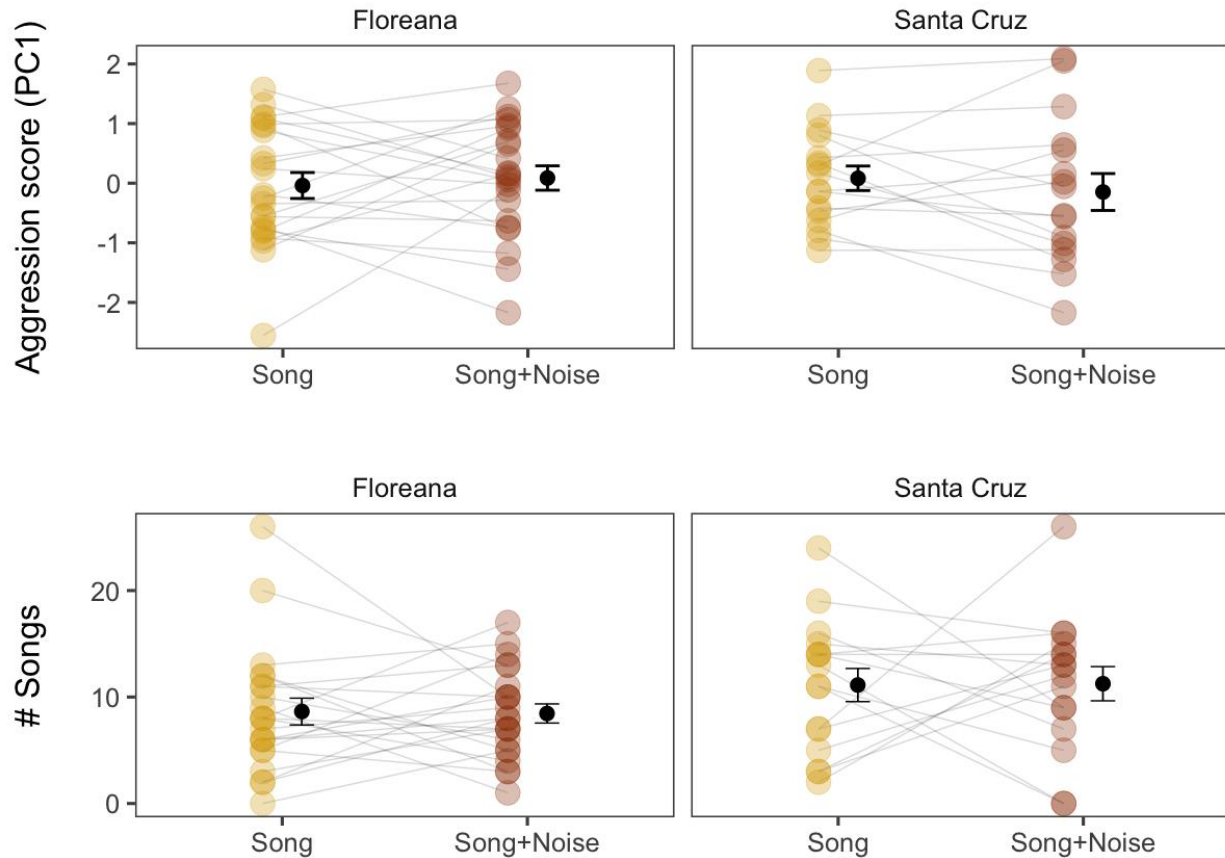
#### 284 ***Behavioural responses to playback***

285 Although all yellow warblers approached the speaker and displayed territorial defence  
286 behaviours during the playbacks (Table S2), there were no significant differences in the

287 aggression scores (PC1) depending on the treatment, island, or their interaction effect (Table 1,  
 288 Figure 2). Number of songs in response to playback also did not differ between the birds  
 289 depending on treatment, island, or their interaction (Table 1, Figure 2).

<i>Variable</i>	<i>Predictors</i>	<i>Estimates</i>	<i>CI</i>	<i>Statistic</i>	<i>p</i>
<b>Aggression score</b>	<b>(Intercept)</b>	-0.04	-0.48 – 0.40	-0.18	0.858
	<b>treatment [Song+Noise]</b>	0.13	-0.35 – 0.60	0.54	0.591
	<b>island [Santa Cruz]</b>	0.12	-0.56 – 0.80	0.36	0.719
	<b>treatment [Song+Noise] × island [Santa Cruz]</b>	-0.36	-1.08 – 0.37	-0.99	0.327
<b>Number of songs</b>	<b>(Intercept)</b>	8.64	6.19 – 11.08	7.16	<b>&lt;0.001</b>
	<b>treatment [Song+Noise]</b>	-0.18	-3.64 – 3.28	-0.11	0.916
	<b>island [Santa Cruz]</b>	2.49	-1.28 – 6.26	1.34	0.189
	<b>treatment [Song+Noise] × island [Santa Cruz]</b>	0.31	-5.03 – 5.64	0.12	0.908

290  
 291 **Table 1.** Effects of island type and treatment on Aggression Scores (PC1) and number of songs.  
 292 Results are shown for linear mixed effect models with island and treatment as fixed effects, and  
 293 territory ID as a random effect.



294

295 **Figure 2.** Behavioural response showing mean  $\pm$  SE for a) Aggression Scores (PC1), and b) the  
296 number of songs. No significant differences across the islands or treatment groups existed in any  
297 of the variables (see results).

298

299 ***Song structure***

300 Male yellow warblers changed their song structure when exposed to a simulated intruder  
301 with noise compared to without noise (see Table 2, Figure 2). In particular, males on Santa Cruz  
302 sang songs with higher minimum frequencies than on Floreana Island and birds on both islands  
303 increased the minimum frequency of their songs under experimental traffic noise. The latter  
304 effect, however, was more pronounced on Floreana as evidenced by a significant interaction

305 effect (Table 2, Figure 3). The maximum frequency and the bandwidth (the difference between  
 306 maximum and minimum frequency) did not differ between islands or treatments. The duration of  
 307 songs was similar across the two islands but yellow warblers on Santa Cruz sang longer songs in  
 308 song + noise treatment compared to the song-only treatment while birds on Floreana did not  
 309 adjust the duration of their song, as evidenced by a significant interaction effect of treatment and  
 310 island (Table 2, Figure 3). Finally, the peak frequency did not differ between islands and was  
 311 slightly but significantly increased in noise treatment.  
 312

<i>Variable</i>	<i>Predictors</i>	<i>Estimates</i>	<i>CI</i>	<i>Statistic</i>	<i>p</i>
<b>Minimum frequency</b>	<b>(Intercept)</b>	2468.62	2398.14 – 2539.11	68.83	<b>&lt;0.001</b>
	<b>treatment [Song + Noise]</b>	100.49	48.70 – 152.29	3.81	<b>&lt;0.001</b>
	<b>island [Santa Cruz]</b>	125.92	13.69 – 238.15	2.28	<b>0.029</b>
	<b>treatment [Song + Noise] × island [Santa Cruz]</b>	-82.44	-158.64 – -6.24	-2.13	<b>0.034</b>
<b>Maximum frequency</b>	<b>(Intercept)</b>	7151.61	6933.72 – 7369.49	64.51	<b>&lt;0.001</b>
	<b>treatment [Song + Noise]</b>	65.37	-96.06 – 226.80	0.80	0.427
	<b>island [Santa Cruz]</b>	-121.03	-467.92 – 225.85	-0.71	0.483
	<b>treatment [Song + Noise] × island [Santa Cruz]</b>	111.43	-126.07 – 348.94	0.92	0.357
<b>Bandwidth</b>	<b>(Intercept)</b>	4684.76	4464.13 – 4905.40	41.73	<b>&lt;0.001</b>
	<b>treatment [Song + Noise]</b>	-36.70	-192.48 – 119.08	-0.46	0.644

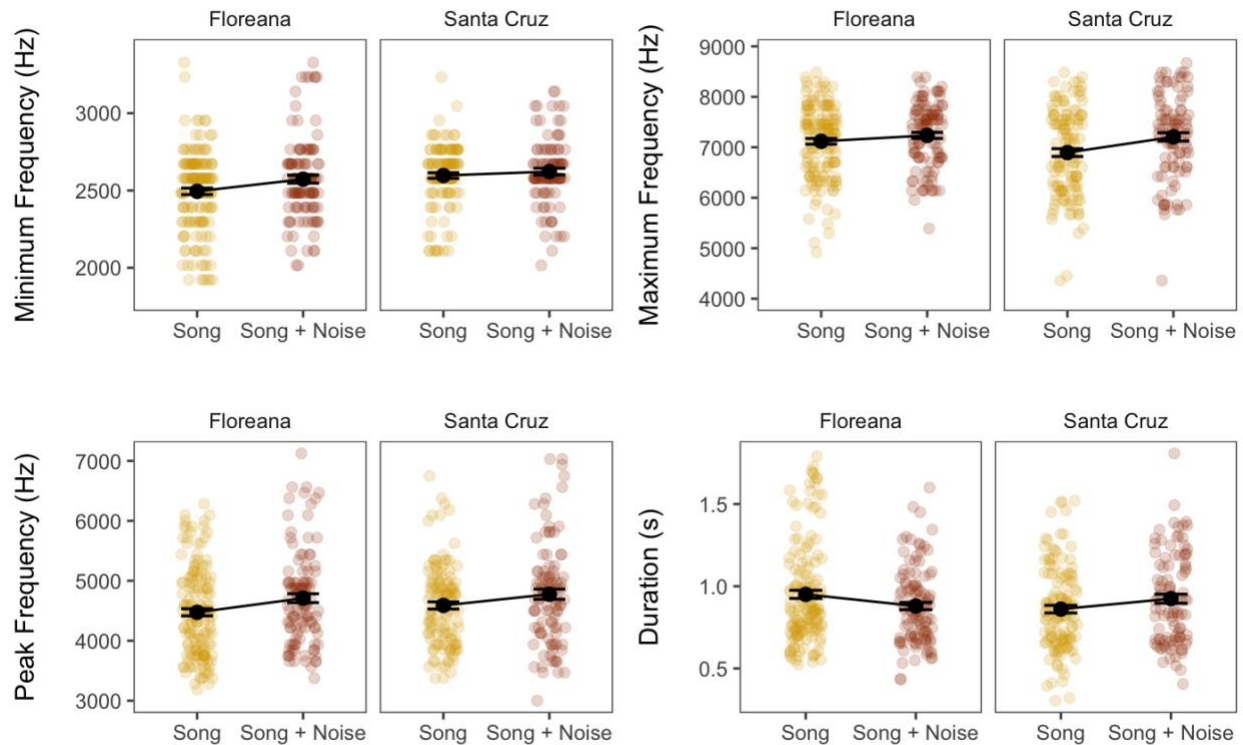


	<b>island [Santa Cruz]</b>	-246.46	-598.04 – 105.11	-1.42	0.163
	<b>treatment [Song + Noise] × island [Santa Cruz]</b>	193.92	-35.28 – 423.11	1.66	0.097
<b>Duration</b>	<b>(Intercept)</b>	0.95	0.87 – 1.03	23.75	<b>&lt;0.001</b>
	<b>treatment [Song + Noise]</b>	-0.09	-0.15 – -0.03	-3.05	<b>0.002</b>
	<b>island [Santa Cruz]</b>	-0.09	-0.21 – 0.04	-1.40	0.170
	<b>treatment [Song + Noise] × island [Santa Cruz]</b>	0.16	0.08 – 0.25	3.68	<b>&lt;0.001</b>
<b>Peak frequency</b>	<b>(Intercept)</b>	4493.20	4317.17 – 4669.23	50.16	<b>&lt;0.001</b>
	<b>treatment [Song + Noise]</b>	186.74	5.06 – 368.43	2.02	<b>0.044</b>
	<b>island [Santa Cruz]</b>	65.72	-212.16 – 343.60	0.48	0.634
	<b>treatment [Song + Noise] × island [Santa Cruz]</b>	6.44	-260.89 – 273.78	0.05	0.962

313 **Table 2.** Effects of island type and treatment on song structure parameters. Results are shown for

314 LMMs with island and treatment as fixed effects, and territory ID as a random effect.

315



316

317 **Figure 3.** Song structure variables depending on treatment and island a) Minimum frequency  
318 (Hz), b) Maximum Frequency (Hz), c) Peak Frequency, and d) Duration. Each coloured circle  
319 represents a song, black dots in the middle and intervals show group means  $\pm$  SE.

## 320 Discussion

321 In the present study, we asked whether Galápagos yellow warblers show increased  
322 aggression and flexibility in their song when faced with a simulated intruder in experimentally  
323 increased noise. Unlike some of the previous studies showing an increase in aggression levels  
324 with increasing ambient noise (Akçay et al., 2020; Davies & Sewall, 2016; Phillips &  
325 Derryberry, 2018), or experimental noise (Grabarczyk EE & Gill, 2019; Önsal et al., 2022) we  
326 found no effect of experimental traffic noise on physical aggression. Instead, we found that birds  
327 on Santa Cruz sang songs with significantly higher minimum frequency than birds on Floreana

328 and, when experimentally exposed to traffic noise, birds in both islands increased the minimum  
329 frequency of their songs. The latter effect was more pronounced on Floreana. We also found an  
330 increase in song duration with experimental noise but only on Santa Cruz. The peak frequency of  
331 songs was also slightly increased with experimental noise compared to the song-only treatment.  
332 The other acoustic parameters (maximum frequency and bandwidth) did not differ between the  
333 islands or treatments. We discuss these results below.

334         The difference in the baseline (under no experimental noise) minimum frequencies  
335 between the islands may be due to long-term selection for higher frequency songs on Santa Cruz  
336 Island as traffic levels have increased over the past few decades. A similar explanation was  
337 proposed for the changes in frequency parameters in the white-crowned sparrow songs in the San  
338 Francisco Bay Area in the USA (Luther & Baptista, 2010; Luther & Derryberry, 2012). Whether  
339 such changes may have come about through preferential cultural learning of songs with higher  
340 frequencies is still debated (Moseley et al., 2018; Zollinger et al., 2017). Interestingly, the white-  
341 crowned sparrows in the San Francisco Bay returned to singing lower minimum frequency songs  
342 in response to the dramatic reduction in human activities and anthropogenic noise during  
343 COVID-19 shutdowns of Spring 2020 (Derryberry et al., 2020). This latter finding suggests that  
344 irrespective of whether the long-term changes in song frequencies in bird songs are due to  
345 developmental or other proximate mechanisms, the birds retain adaptive plasticity to adjust song  
346 features as adults.

347         An alternative hypothesis for why the birds on Santa Cruz may be singing songs with  
348 higher minimum frequencies may be due to differences in body size. Larger birds may produce  
349 songs with lower minimum frequency (Friis et al., 2022; Ryan & Brenowitz, 1985). Therefore, if  
350 there are differences in yellow warbler body size between the islands, such morphological

351 differences could explain differences in song frequency (Potvin, 2013). While we did not  
352 measure yellow warbler morphology across the islands, two previous studies did. Browne and  
353 colleagues (2010) found that there were some size differences between yellow warblers on  
354 different islands with Santa Cruz birds being smallest, although their sample sizes were relatively  
355 small (including for Floreana;  $n = 4$ ). Another study by Chaves and colleagues (2012), with a  
356 larger sample size did not find any morphological differences between yellow warblers from four  
357 Galápagos islands, but their dataset did not include birds from Floreana. Therefore, this eco-  
358 morphological hypothesis needs further data to test it fully.

359         Interestingly, birds on Floreana adjusted their minimum frequency more than the birds on  
360 Santa Cruz but birds on Santa Cruz increased the duration of their song more than the birds on  
361 Floreana. These findings point to a potentially complex effect of previous experience on adaptive  
362 adjustments on song. On one hand, the finding that birds in Santa Cruz increase the duration of  
363 their songs more than birds in Floreana is consistent with the hypothesis that long-term selection  
364 due to-- or previous individual experience with noise leads to the ability to rapidly and  
365 adaptively adjust songs features (Gentry et al., 2017; LaZerte et al., 2016). On the other hand, the  
366 minimum frequency patterns that show stronger adjustments on Floreana than in Santa Cruz  
367 appear to run against this hypothesis. A potential reason behind these seemingly contradictory  
368 findings may be that there is a limit to plasticity in minimum frequencies; perhaps birds on Santa  
369 Cruz are already singing at higher minimum frequencies (and/or higher amplitudes as we discuss  
370 below) and may not be able to further increase their minimum frequency. Instead, they may  
371 attempt to cope with the increased noise levels by increasing the duration of their songs.

372         It is worth noting that the changes in frequency, while potentially adaptive, may also  
373 come about by increases in song amplitude. Nemeth and colleagues (2013) showed that

374 amplitude and frequency covary and that birds with greater song amplitude also have higher  
375 frequencies. In our study, we did not measure the amplitude of the songs, and therefore can't rule  
376 out an effect of the amplitude changes on song frequencies. Nemeth and Brumm, (2010) found  
377 that increasing amplitude is more effective in overcoming masking effects in noisy  
378 environments, which they measured experimentally in the great tit (*Parus major*) and blackbird  
379 (*Turdus merula*) populations. Specifically, higher amplitude songs travelled further whereas  
380 there was greater attenuation of higher pitched songs. In regard to the argument of efficacy, in  
381 urban habitats, the upward shift in song frequency increased the communication distance to a  
382 significantly lesser degree than amplitude (Nemeth & Brumm, 2010). Measuring amplitude in  
383 the field with freely moving animals is difficult but future studies should endeavour to  
384 disentangle these effects. Another fruitful avenue would be to measure the transmission of the  
385 songs from the two islands under different noise treatments to see if the frequency as well as  
386 duration adjustments do in fact lead to better transmission.

387         If song produced during territorial defence is an aggressive signal that is used to resolve  
388 agonistic interactions instead of physical aggression, then acoustic noise that interferes with  
389 effective signaling may result in increased aggression. Several studies across taxa have found  
390 increased levels of aggression in noisy environments (in humans: Alimohammadi et al., 2018;  
391 Moudon, 2009; in songbirds: Akçay et al., 2020; Phillips & Derryberry, 2018). In contrast, we  
392 found that despite the pronounced difference in traffic levels between islands, yellow warblers  
393 did not show island-level differences in their physical response behaviour to the simulated  
394 intruder, nor did they increase aggression when they were faced with intermittent traffic noise. It  
395 is possible that short-term and long-term adjustments in song frequency (and possibly amplitude)  
396 are sufficient to deal with interference from the traffic noise. It is also possible that yellow

397 warblers use signals in other modalities (such as visual modality, e.g. wing-out and tail-spread  
398 display) more effectively in increased noise (Akçay & Beecher, 2019; Önsal et al., 2022).

399         In addition to posing a problem for sound transmission (Parris & Schneider, 2009;  
400 Schmidt & Balakrishnan, 2015), traffic also poses a risk to survival (García-Carrasco et al.,  
401 2020; Grilo et al., 2021; Parris & Schneider, 2009). Globally, roadkill is becoming an increasing  
402 threat to the survival of animals (García-Carrasco et al. 2020; Grilo et al. 2021). In the  
403 Galápagos, yellow warblers make up by far the largest portion (up to >70% on the main road) of  
404 roadkill on Santa Cruz, despite having similar abundances to other songbirds (García-Carrasco et  
405 al. 2020). Why yellow warblers make such a high proportion of roadkill on Galápagos is unclear,  
406 although some ecological and morphological traits are likely to be important. A study in Portugal  
407 found that foliage-gleaning birds as well as to a lesser extent birds with high flight  
408 manoeuvrability had higher mortality risk due to vehicle collision (Santos et al. 2016). The  
409 combination of these risk factors in the Galapagos yellow warblers may be leading  
410 disproportionate mortality of this species on roads but this hypothesis needs further testing. A  
411 way that humans can help reduce roadkill and acoustic competition is to regulate speed limits –  
412 given vehicle velocity is positively correlated with noise (Cowan, 1993) and areas with lower  
413 speed limits have less roadkill (Farmer & Brooks, 2012; Hobday, 2010).

414         In summary, we found that traffic noise affects the song structure of the Galápagos  
415 yellow warblers, with some differences between islands with different traffic levels. On both  
416 islands, when exposed to traffic noise, birds increased the minimum frequency of their songs.  
417 And on the island with high traffic, resident birds also sang longer songs when exposed to a  
418 simulated intruder song with traffic noise. While the Galápagos Islands are a natural living  
419 laboratory that inspired Charles Darwin's formulation of the theory of evolution by natural

420 selection (Darwin, 1859; P. R. Grant & Grant, 2014), our study adds to the growing body of  
421 knowledge that the archipelago is affected by anthropogenic impacts that are expanding globally.  
422 The Galápagos yellow warbler, common in all the habitats of vegetated islands, is a good model  
423 system for studying the human impacts on this unique archipelago.

#### 424 **Data availability**

425 Data and R-code to reproduce analyses can be found in the supplementary materials.

426

#### 427 **References**

- 428 Akçay, Ç., & Beecher, M. D. (2019). Multi-modal communication: Song sparrows increase  
429 signal redundancy in noise. *Biology Letters*, *15*(10), 20190513.  
430 <https://doi.org/10.1098/rsbl.2019.0513>
- 431 Akçay, Ç., Porsuk, Y. K., Avşar, A., Çabuk, D., & Bilgin, C. C. (2020). Song overlapping, noise,  
432 and territorial aggression in great tits. *Behavioral Ecology*, *31*(3), 807–814.  
433 <https://doi.org/10.1093/beheco/araa030>
- 434 Alimohammadi, I., Kanrash, F. A., Abolaghasemi, J., Afrazandeh, H., & Rahmani, K. (2018).  
435 *Effect of Chronic Noise Exposure on Aggressive Behavior of Automotive Industry*  
436 *Workers*. *9*(4).
- 437 Araya-Salas, M., & Smith-Vidaurre, G. (2017). warbleR: an R package to streamline analysis of  
438 animal acoustic signals. *Methods in Ecology and Evolution*, *8*(2), 184-191.
- 439 Bermúdez-Cuamatzin, E., Ríos-Chelén, A. A., Gil, D., & Garcia, C. M. (2011). Experimental  
440 evidence for real-time song frequency shift in response to urban noise in a passerine bird.  
441 *Biology Letters*, *7*(1), 36–38. <https://doi.org/10.1098/rsbl.2010.0437>

- 442 Bowen, A. E., Gurule-Small, G. A., & Tinghitella, R. M. (2020). Anthropogenic noise reduces  
443 male reproductive investment in an acoustically signaling insect. *Behavioral Ecology and*  
444 *Sociobiology*, 74(8), 103. <https://doi.org/10.1007/s00265-020-02868-3>
- 445 Bradbury, J. W., & Vehrencamp, S. L. (2011). *Principles of animal communication* (2nd ed).  
446 Sinauer Associates.
- 447 Browne, R. A., Collins, E. I., & Anderson, D. J. (2010). Morphological variation of Galapagos  
448 island populations of the yellow warbler *Dendroica petechia aureola*. *Galapagos*  
449 *Reports*. 67. 9-13.
- 450 Brumm, H. (Ed.). (2013). *Animal Communication and Noise* (Vol. 2). Springer.  
451 <https://doi.org/10.1007/978-3-642-41494-7>
- 452 Brumm, H., & Slater, P. J. B. (2006). Ambient noise, motor fatigue, and serial redundancy in  
453 chaffinch song. *Behavioral Ecology and Sociobiology*, 60(4), 475–481.  
454 <https://doi.org/10.1007/s00265-006-0188-y>
- 455 Brumm, H., & Zollinger, S. A. (2011). The evolution of the Lombard effect: 100 years of  
456 psychoacoustic research. *Behaviour*, 148(11–13), 1173–1198.  
457 <https://doi.org/10.1163/000579511X605759>
- 458 Brumm, H., & Zollinger, S. A. (2013). Avian Vocal Production in Noise. In H. Brumm (Ed.),  
459 *Animal Communication and Noise* (Vol. 2, pp. 187–227). Springer Berlin Heidelberg.  
460 [https://doi.org/10.1007/978-3-642-41494-7\\_7](https://doi.org/10.1007/978-3-642-41494-7_7)
- 461 Brumm, H., Zollinger, S. A., Niemelä, P. T., & Sprau, P. (2017). Measurement artefacts lead to  
462 false positives in the study of birdsong in noise. *Methods in Ecology and Evolution*,  
463 8(11), 1617–1625. <https://doi.org/10.1111/2041-210X.12766>
- 464 Catchpole, C. K., & Slater, P. J. B. (2008). *BIRD SONG BIOLOGICAL THEMES AND*



465            *VARIATIONS.*

466   Chaves, J. A., Parker, P. G., & Smith, T. B. (2012). Origin and population history of a recent  
467            colonizer, the yellow warbler in Galápagos and Cocos Islands: Phylogeography of yellow  
468            warblers in Galápagos. *Journal of Evolutionary Biology*, 25(3), 509–521.

469            <https://doi.org/10.1111/j.1420-9101.2011.02447.x>

470   Colombelli-Négrel, D., Akçay, Ç., & Kleindorfer, S. (2023). Darwin’s finches in human-altered  
471            environments sing common song types and are more aggressive. *Frontiers in Ecology*  
472            *and Evolution*, 11, 1034941. <https://doi.org/10.3389/fevo.2023.1034941>

473   Common, L. K., Sumasgutner, P., Sumasgutner, S. C., Colombelli-Négrel, D., Dudaniec, R. Y.,  
474            & Kleindorfer, S. (2022). Temporal and spatial variation in sex-specific abundance of the  
475            avian vampire fly (*Philornis downsi*). *Parasitology Research*, 121(1), 63–74.

476            <https://doi.org/10.1007/s00436-021-07350-1>

477   Courter, J. R., Perruci, R. J., McGinnis, K. J., & Rainieri, J. K. (2020). Black-capped chickadees  
478            (*Poecile atricapillus*) alter alarm call duration and peak frequency in response to traffic  
479            noise. *PLOS ONE*, 15(10), e0241035. <https://doi.org/10.1371/journal.pone.0241035>

480   Cowan, J.P., (1993). Handbook of environmental acoustics. John Wiley & Sons, pp.150-151.

481   Darwin, C. (1859) *On the Origin of Species by Means of Natural Selection*, Vol 167. John  
482            Murray, London

483   Davies, S., & Sewall, K. B. (2016). Agonistic urban birds: Elevated territorial aggression of  
484            urban song sparrows is individually consistent within a breeding period. *Biology Letters*,  
485            12(6), 20160315. <https://doi.org/10.1098/rsbl.2016.0315>

486   De Jong, K., Amorim, M. C. P., Fonseca, P. J., Fox, C. J., & Heubel, K. U. (2018). Noise can  
487            affect acoustic communication and subsequent spawning success in fish. *Environmental*

- 488            *Pollution*, 237, 814–823. <https://doi.org/10.1016/j.envpol.2017.11.003>
- 489 De León, L. F., Sharpe, D. M. T., Gotanda, K. M., Raeymaekers, J. A. M., Chaves, J. A.,  
490 Hendry, A. P., & Podos, J. (2019). Urbanization erodes niche segregation in Darwin’s  
491 finches. *Evolutionary Applications*, 12(7), 1329–1343. <https://doi.org/10.1111/eva.12721>
- 492 Derryberry, E. P., Phillips, J. N., Derryberry, G. E., Blum, M. J., & Luther, D. (2020). Singing in  
493 a silent spring: Birds respond to a half-century soundscape reversion during the COVID-  
494 19 shutdown. *Science*, 370(6516), 575–579. <https://doi.org/10.1126/science.abd5777>
- 495 Duquette, C. A., Loss, S. R., & Hovick, T. J. (2021). A meta-analysis of the influence of  
496 anthropogenic noise on terrestrial wildlife communication strategies. *Journal of Applied*  
497 *Ecology*, 58(6), 1112–1121. <https://doi.org/10.1111/1365-2664.13880>
- 498 Falls, J. B., & Brooks, R. J. (1975). Individual recognition by song in white-throated sparrows.  
499 II. Effects of location. *Canadian Journal of Zoology*, 53(10), 1412–1420.  
500 <https://doi.org/10.1139/z75-170>
- 501 Farmer, R. G., & Brooks, R. J. (2012). Integrated risk factors for vertebrate roadkill in southern  
502 Ontario. *The Journal of Wildlife Management*, 76(6), 1215–1224.  
503 <https://doi.org/10.1002/jwmg.358>
- 504 Foltz, S. L., Ross, A. E., Laing, B. T., Rock, R. P., Battle, K. E., & Moore, I. T. (2015). Get off  
505 my lawn: Increased aggression in urban song sparrows is related to resource availability.  
506 *Behavioral Ecology*, 26(6), 1548–1557. <https://doi.org/10.1093/beheco/arv111>
- 507 Grabarczyk EE, & Gill SA (2019) Anthropogenic noise affects male house wren response to but  
508 not detection of territorial intruders. *PLoS One* 14:e0220576
- 509 Galápagos Conservancy. (2023). Floreana Island. [Online]. Galápagos Conservancy. Available  
510 at: [https://www.galapagos.org/about\\_galapagos/the-islands/floreana-island/](https://www.galapagos.org/about_galapagos/the-islands/floreana-island/) [Accessed 24

- 511 November 2023].
- 512 Gallego-Abenza, M., Mathevon, N., & Wheatcroft, D. (2020). Experience modulates an insect's  
513 response to anthropogenic noise. *Behavioral Ecology*, *31*(1), 90–96.  
514 <https://doi.org/10.1093/beheco/arz159>
- 515 García-Carrasco, J.-M., Tapia, W., & Muñoz, A.-R. (2020). Roadkill of birds in Galapagos  
516 Islands: A growing need for solutions. *Avian Conservation and Ecology*, *15*(1), art19.  
517 <https://doi.org/10.5751/ACE-01596-150119>
- 518 Gentry, K. E., Derryberry, E. P., Danner, R. M., Danner, J. E., & Luther, D. A. (2017).  
519 Immediate signaling flexibility in response to experimental noise in urban, but not rural,  
520 white-crowned sparrows. *Ecosphere*, *8*(8). <https://doi.org/10.1002/ecs2.1916>
- 521 Gotanda, K. M. (2020). Human influences on antipredator behaviour in Darwin's finches.  
522 *Journal of Animal Ecology*, *89*(2), 614–622. <https://doi.org/10.1111/1365-2656.13127>
- 523 Grabarczyk, E. E., & Gill, S. A. (2019). Anthropogenic noise affects male house wren response  
524 to but not detection of territorial intruders. *PLoS ONE*, *14*(7), 1–17.  
525 <https://doi.org/10.1371/journal.pone.0220576>
- 526 Grade, A. M., & Sieving, K. E. (2016). When the birds go unheard: Highway noise disrupts  
527 information transfer between bird species. *Biology Letters*, *12*(4), 20160113.  
528 <https://doi.org/10.1098/rsbl.2016.0113>
- 529 Grant, B. R., & Grant, P. R. (1979). Darwin's finches: Population variation and sympatric  
530 speciation. *Proceedings of the National Academy of Sciences*, *76*(5), 2359–2363.  
531 <https://doi.org/10.1073/pnas.76.5.2359>
- 532 Grant, P. R., & Grant, B. R. (2014). *40 years of evolution: Darwin's finches on Daphne Major*  
533 *Island*. Princeton Univ. Press.

- 534 Grilo, C., Borda-de-Água, L., Beja, P., Goolsby, E., Soanes, K., Le Roux, A., Koroleva, E.,  
535 Ferreira, F. Z., Gagné, S. A., Wang, Y., & González-Suárez, M. (2021). Conservation  
536 threats from roadkill in the global road network. *Global Ecology and Biogeography*,  
537 *30*(11), 2200–2210. <https://doi.org/10.1111/geb.13375>
- 538 Hage, S. R., Jiang, T., Berquist, S. W., Feng, J., & Metzner, W. (2013). Ambient noise induces  
539 independent shifts in call frequency and amplitude within the Lombard effect in  
540 echolocating bats. *Proceedings of the National Academy of Sciences*, *110*(10), 4063–  
541 4068. <https://doi.org/10.1073/pnas.1211533110>
- 542 Halfwerk, W., & Slabbekoorn, H. (2015). Pollution going multimodal: The complex impact of  
543 the human-altered sensory environment on animal perception and performance. *Biology*  
544 *Letters*, *11*(4), 20141051. <https://doi.org/10.1098/rsbl.2014.1051>
- 545 Harvey, J. A., Chernicky, K., Simons, S. R., Verrett, T. B., Chaves, J. A., & Knutie, S. A.  
546 (2021). Urban living influences the nesting success of Darwin’s finches in the Galápagos  
547 Islands. *Ecology and Evolution*, *11*(10), 5038–5048. <https://doi.org/10.1002/ece3.7360>
- 548 Hobday, A. J. (2010). Nighttime driver detection distances for Tasmanian fauna: Informing  
549 speed limits to reduce roadkill. *Wildlife Research*, *37*(4), 265.  
550 <https://doi.org/10.1071/WR09180>
- 551 Hobson, K. A., & Sealy, S. G. (1989). Mate Guarding in the Yellow Warbler *Dendroica*  
552 *petechia*. *Ornis Scandinavica*, *20*(4), 241. <https://doi.org/10.2307/3676487>
- 553 Instituto Nacional de Estadística y Censos (INEC). 2016. Censo de Población y Vivienda  
554 Galápagos, 2015. INEC, Quito, Ecuador. [online] URL:  
555 <https://www.ecuadorencifras.gob.ec/censo-de-poblacion-y-vivienda-galapagos/>
- 556 Kleindorfer, S., Common, L. K., & Sumasgutner, P. (2021). Nesting Success and Nesting Height

- 557 in the Critically Endangered Medium Tree Finch (*Camarhynchus pauper*). *Birds*, 2(4),  
558 427–444. <https://doi.org/10.3390/birds2040032>
- 559 Kleist, N. J., Guralnick, R. P., Cruz, A., & Francis, C. D. (2016). Anthropogenic noise weakens  
560 territorial response to intruder’s songs. *Ecosphere*, 7(3).  
561 <https://doi.org/10.1002/ecs2.1259>
- 562 K. Lisa Yang Center for Conservation Bioacoustics at the Cornell Lab of Ornithology. (2022).  
563 Raven Pro: Interactive Sound Analysis Software (Version 1.6.4) [Computer software].  
564 Ithaca, NY: The Cornell Lab of Ornithology. Available from  
565 <https://ravensoundsoftware.com/>.
- 566 Kricher, J. C., & Loughlin, K. C. (2022). *Galápagos: A natural history* (Second edition).  
567 Princeton University Press.
- 568 Kunc, H. P., Morrison, K., & Schmidt, R. (2022). A meta-analysis on the evolution of the  
569 Lombard effect reveals that amplitude adjustments are a widespread vertebrate  
570 mechanism. *Proceedings of the National Academy of Sciences*, 119(30), e2117809119.  
571 <https://doi.org/10.1073/pnas.2117809119>
- 572 Kunc, H. P., & Schmidt, R. (2021). Species sensitivities to a global pollutant: A meta-analysis on  
573 acoustic signals in response to anthropogenic noise. *Global Change Biology*, 27(3), 675–  
574 688. <https://doi.org/10.1111/gcb.15428>
- 575 Lawson, S. L., Enos, J. K., Fernandez-Duque, F., Kleindorfer, S., Ward, M. P., Gill, S. A., &  
576 Hauber, M. E. (2023). Absence of Referential Alarm Calls in Long-term Allopatry from  
577 the Referent: A Case Study with Galapagos Yellow Warblers. *Behavioral Ecology and*  
578 *Sociobiology*, 77(8), 99. <https://doi.org/10.1007/s00265-023-03372-0>
- 579 LaZerte, S. E., Slabbekoorn, H., & Otter, K. A. (2016). Learning to cope: Vocal adjustment to

- 580 urban noise is correlated with prior experience in black-capped chickadees. *Proceedings*  
581 *of the Royal Society B: Biological Sciences*, 283(1833), 20161058.  
582 <https://doi.org/10.1098/rspb.2016.1058>
- 583 Liu, Y., Zollinger, S. A., & Brumm, H. (2021). Chronic exposure to urban noise during the vocal  
584 learning period does not lead to increased song frequencies in zebra finches. *Behavioral*  
585 *Ecology and Sociobiology*, 75(1), 3. <https://doi.org/10.1007/s00265-020-02942-w>
- 586 Logue, D. M., Abiola, I. O., Rains, D., Bailey, N. W., Zuk, M., & Cade, W. H. (2010). Does  
587 signalling mitigate the cost of agonistic interactions? A test in a cricket that has lost its  
588 song. *Proceedings of the Royal Society B: Biological Sciences*, 277(1693), 2571–2575.  
589 <https://doi.org/10.1098/rspb.2010.0421>
- 590 Luther, D., & Baptista, L. (2010). Urban noise and the cultural evolution of bird songs.  
591 *Proceedings of the Royal Society B: Biological Sciences*, 277(1680), 469–473.  
592 <https://doi.org/10.1098/rspb.2009.1571>
- 593 Luther, D., & Derryberry, E. P. (2012). Birdsongs keep pace with city life: Changes in song over  
594 time in an urban songbird affects communication. *Animal Behaviour*.
- 595 Márquez, L. (2000). Mortalidad de las aves vía Puerto Ayora - Canal de Itabaca 1999. Colegio  
596 Loma Linda. Colegio Loma Linda, Puerto Ayora, Santa Cruz, Galápagos, Ecuador.
- 597 Maynard Smith, J., & Harper, D. (1988). The evolution of aggression: Can selection generate  
598 variability? *Philosophical Transactions of the Royal Society of London. B, Biological*  
599 *Sciences*, 319(1196), 557–570. <https://doi.org/10.1098/rstb.1988.0065>
- 600 Mitra, S. S. (1999). Ecology and Behavior of Yellow Warblers Breeding in Rhode Island's Great  
601 Swamp. *Northeastern Naturalist*, 6(3), 249. <https://doi.org/10.2307/3858601>
- 602 Moseley, D. L., Derryberry, G. E., Phillips, J. N., Danner, J. E., Danner, R. M., Luther, D. A., &

- 603 Derryberry, E. P. (2018). Acoustic adaptation to city noise through vocal learning by a  
604 songbird. *Proceedings of the Royal Society B: Biological Sciences*, 285(1888), 20181356.  
605 <https://doi.org/10.1098/rspb.2018.1356>
- 606 Moudon, A. V. (2009). Real Noise from the Urban Environment. *American Journal of*  
607 *Preventive Medicine*, 37(2), 167–171. <https://doi.org/10.1016/j.amepre.2009.03.019>
- 608 Nemeth, E., & Brumm, H. (2010). Birds and Anthropogenic Noise: Are Urban Songs Adaptive?  
609 *The American Naturalist*, 176(4), 465–475. <https://doi.org/10.1086/656275>
- 610 Nemeth, E., Pieretti, N., Zollinger, S. A., Geberzahn, N., Partecke, J., Miranda, A. C., & Brumm,  
611 H. (2013). Bird song and anthropogenic noise: Vocal constraints may explain why birds  
612 sing higher-frequency songs in cities. *Proceedings of the Royal Society B: Biological*  
613 *Sciences*, 280(1754), 20122798. <https://doi.org/10.1098/rspb.2012.2798>
- 614 Oviedo, M., J. Agama, E. Buitrón, and F. Zavala. 2009. The first complete motorized vehicle  
615 census in Galapagos. *Galapagos Report 2009-2010*:48-53.
- 616 Önsal, Ç., Yelimlieş, A., & Akçay, Ç. (2022). Aggression and multi-modal signaling in noise in  
617 a common urban songbird. *Behavioral Ecology and Sociobiology*, 76(7), 102.  
618 <https://doi.org/10.1007/s00265-022-03207-4>
- 619 Parris, K. M., & Schneider, A. (2009). Impacts of Traffic Noise and Traffic Volume on Birds of  
620 Roadside Habitats. *Ecology and Society*.
- 621 Phillips, J. N., & Derryberry, E. P. (2018). Urban sparrows respond to a sexually selected trait  
622 with increased aggression in noise. *Scientific Reports*, 8(1), 7505.  
623 <https://doi.org/10.1038/s41598-018-25834-6>
- 624 Potvin, D. A., & Mulder, R. A. (2013). Immediate, independent adjustment of call pitch and  
625 amplitude in response to varying background noise by silvereyes (*Zosterops lateralis*).

- 626 *Behavioral Ecology*, 24(6), 1363–1368. <https://doi.org/10.1093/beheco/art075>
- 627 Pinheiro J, Bates D, R Core Team (2022). nlme: Linear and Nonlinear Mixed Effects Models.  
628 R package version 3.1-158, <https://CRAN.R-project.org/package=nlme>.
- 629 R Core Team (2023). R: A language and environment for statistical computing. R Foundation for  
630 Statistical Computing, Vienna, Austria. URL <https://www.R-project.org/>.
- 631 Revelle, W. R. (2017). psych: Procedures for personality and psychological research.
- 632 Ríos-Chelén, A. A., Quirós-Guerrero, E., Gil, D., & Macías Garcia, C. (2013). Dealing with  
633 urban noise: Vermilion flycatchers sing longer songs in noisier territories. *Behavioral*  
634 *Ecology and Sociobiology*, 67(1), 145–152. <https://doi.org/10.1007/s00265-012-1434-0>
- 635 Rivkin, L. R., Johnson, R. A., Chaves, J. A., & Johnson, M. T. J. (2021). Urbanization alters  
636 interactions between Darwin’s finches and *Tribulus cistoides* on the Galápagos Islands.  
637 *Ecology and Evolution*, 11(22), 15754–15765. <https://doi.org/10.1002/ece3.8236>
- 638 Roca, I. T., Desrochers, L., Giacomazzo, M., Bertolo, A., Bolduc, P., Deschesnes, R., Martin, C.  
639 A., Rainville, V., Rheault, G., & Proulx, R. (2016). Shifting song frequencies in response  
640 to anthropogenic noise: A meta-analysis on birds and anurans. *Behavioral Ecology*,  
641 27(5), 1269–1274. <https://doi.org/10.1093/beheco/arw060>
- 642 Salgado-Ortiz, J., Marra, P. P., Sillett, T. S., & Robertson, R. J. (2008). Breeding ecology of the  
643 mangrove warbler (*Dendroica petechia bryanti*) and comparative life history of the  
644 yellow warbler subspecies complex. *The Auk*, 125(2), 402–410.  
645 <https://doi.org/10.1525/auk.2008.07012>
- 646 Santos, S. M., Mira, A., Salgueiro, P. A., Costa, P., Medinas, D., & Beja, P. (2016). Avian trait-  
647 mediated vulnerability to road traffic collisions. *Biological Conservation*, 200, 122-130.  
648 doi: <https://doi.org/10.1016/j.biocon.2016.06.004>



- 649 Schmidt, A. K. D., & Balakrishnan, R. (2015). Ecology of acoustic signalling and the problem of  
650 masking interference in insects. *Journal of Comparative Physiology A*, 201(1), 133–142.  
651 <https://doi.org/10.1007/s00359-014-0955-6>
- 652 Searcy, W. A., Anderson, R. C., & Nowicki, S. (2006). Bird song as a signal of aggressive intent.  
653 *Behavioral Ecology and Sociobiology*, 60(2), 234–241. [https://doi.org/10.1007/s00265-](https://doi.org/10.1007/s00265-006-0161-9)  
654 [006-0161-9](https://doi.org/10.1007/s00265-006-0161-9)
- 655 Slabbekoorn, H., & Peet, M. (2003). Birds sing at a higher pitch in urban noise. *Nature*,  
656 424(6946), 267–267. <https://doi.org/10.1038/424267a>
- 657 Snow, D. W. (1966). Annual Cycle of the Yellow Warbler in the Galapagos. *Bird-Banding*,  
658 37(1), 44. <https://doi.org/10.2307/4511232>
- 659 Verzijden, M. N., Ripmeester, E. A. P., Ohms, V. R., Snelderwaard, P., & Slabbekoorn, H.  
660 (2010). Immediate spectral flexibility in singing chiffchaffs during experimental exposure  
661 to highway noise. *Journal of Experimental Biology*, 213(15), 2575–2581.  
662 <https://doi.org/10.1242/jeb.038299>
- 663 Zollinger, S. A., Slater, P. J. B., Nemeth, E., & Brumm, H. (2017). Higher songs of city birds  
664 may not be an individual response to noise. *Proceedings of the Royal Society B:*  
665 *Biological Sciences*, 284(1860), 20170602. <https://doi.org/10.1098/rspb.2017.0602>