Daphnia-associated bacterial communities correlate with diet

quantity, environmental conditions and epidemic size across

natural outbreaks

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Running head

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Abstract

Zooplankton-associated microbiomes play an important role for host health and contribute to ecosystem processes such as nutrient cycling. Yet, few studies have assessed how environmental gradients and biotic interactions, including parasitism and diet, may shape the microbiome composition of wild zooplankton. Here, we analysed the microbiomes of water fleas from the *Daphnia longispina* species complex using 16S rRNA gene sequencing and a long-term field dataset spanning six sampling events over 13 years. Sampling coincided with outbreaks of the virulent eukaryotic gut parasite *Caullerya mesnili.* Additionally, we explored how microbiome structure varied in relation to water parameters, phytoplankton density (i.e. *Daphnia* diet), and zooplankton density and community structure. *Daphnia* microbiomes displayed strong temporal variation, and comparatively small differences based on host infection status. Microbiome beta diversity correlated with phytoplankton density but not with its community composition, including green algae, protists and cyanobacteria. Environmental conditions, including temperature, dissolved oxygen and cyanobacterial abundance - previously found to drive *Caullerya* epidemics - were also associated with distinct microbiome structures. Importantly, microbiome beta diversity co-varied with infection prevalence, suggesting a link between microbiome shifts, epidemic size, and environmental conditions driving large epidemics. Dominant bacterial taxa correlated with *Daphnia* density, whereas the phylogenetic composition of rare taxa was associated with total zooplankton density. These findings demonstrate the dynamic nature of *Daphnia* microbiomes and suggest potential mechanisms by which they may mediate disease dynamics, particularly through associations with diet quantity, temperature, and host population density.

Introduction

Zooplankton constitute an essential component of aquatic ecosystems, serving as both consumers and prey. Their interactions with predators, phytoplankton and parasites can shape aquatic ecosystem functioning (Bakhtiyar et al., 2020). Among these interactions, parasites represent an often overlooked but critical functional group, influencing ecosystem processes either directly through biomass generation for consumption, or indirectly by altering host population dynamics and traits (Preston et al., 2016). For instance, zooplankton-parasite associations may play a role in nutrient cycling within aquatic ecosystems (Cáceres et al., 2014; Frost et al., 2008).

Zooplankters also harbour communities of diverse, symbiotic microbes that vastly outnumber free-living microbes in the water column (Tang, 2005). For instance, copepods selectively enrich specific bacterial taxa from their environment, which then disperse back into the water, shaping bacterioplankton community composition (Shoemaker et al., 2019). Bacterial epibionts associated with water fleas (*Daphnia*) contribute to nutrient cycling and the transfer of dissolved organic carbon within the pelagic food web (Eckert & Pernthaler, 2014). Zooplankton-associated bacteria can also help their copepod hosts in digesting toxic cyanobacteria (Gorokhova et al., 2021), and serve as a mechanism of adaptation to toxic diets in other zooplankters such as *Daphnia* (Macke et al., 2017). These general symbiotic relationships between zooplankton and bacteria can exert cascading effects on ecosystem health and functioning, underscoring the importance of characterizing the abiotic and biotic factors such as diet and parasitism, that potentially shape the diversity and composition of bacterial communities.

Host-associated bacterial communities (or microbiomes) influence the health and performance of various aquatic organisms (reviewed in Infante-Villamil et al., 2021). In addition to host traits, parasite traits and environmental conditions, microbiomes are proposed as the fourth vertex of a disease pyramid (Bernardo-Cravo et al., 2020). However, unravelling the forces that potentially influence microbiomes in natural settings is challenging due to their temporal variability and the complex interplay of deterministic (e.g. abiotic environmental gradients, biotic interactions of host species) and stochastic (e.g. horizontal transmission) processes (Sze et al., 2020). Wild host microbiomes can fluctuate over various timescales, spanning from hours to days within a single host generation (Caughman et al., 2021; Minich et al., 2020) to

longer timescales of months (Krajacich et al., 2018; Pierce & Ward, 2019). The relatively long-term

temporal variation in host microbiomes may be linked to seasonal environmental changes (Sharp et al., 2017;

Zhu et al., 2022) or occur independently of seasonal patterns (Epstein et al., 2019). Environmental factors

such as temperature and dissolved oxygen influence the composition, diversity and function of aquatic host-

associated microbiomes (reviewed in Sehnal et al., 2021). While diet is a strong predictor of microbiome

composition in many fish species (Escalas et al., 2021), this relationship does not apply to other aquatic

organisms, such as fire salamanders (Wang et al., 2021) or marine zooplankton (Xu et al., 2024).

Additionally, horizontal transmission plays a major role in shaping host microbiomes in aquatic

environments (Russell, 2019), with host population density influencing microbial dispersal (Miller et al.,

2018).

Despite growing interest in host-microbiome interactions, few studies have investigated microbiome changes

in wild host populations over time, particularly in relation to natural parasite epidemics. The microbiome of

the reef-building coral *Montastraea faveolata* undergoes strong abundance shifts in response to seasonal

variation and the onset of Yellow Band Disease (Kimes et al., 2013). In the fish host *Scomber japonicus,*

both the microbiome composition and abundance of a pathogen *Photobacterium damselae* correlated with

the summer conditions (Minich et al., 2020). Similarly, amphibian declines attributed to the disease

chytridiomycosis are linked to seasonal changes in skin microbiota (Schmeller et al., 2022), with the

abundance of specific bacterial taxa correlating with chytrid infection intensity in the frog *Philoria*

loveridgei (López et al., 2017). These examples illustrate how host microbiomes may both influence and

respond to disease dynamics in natural populations (Stencel, 2021).

The water flea *Daphnia* is a globally prevalent zooplankton genus with significant ecological roles, including

controlling phytoplankton populations (Hébert et al., 2016) and influencing bacterioplankton abundance (De

Corte et al., 2023) and community composition (Berga et al., 2015; Degans & De Meester, 2002).

Conversely, *Daphnia*-associated bacterial communities are also shaped by the surrounding bacterioplankton

(Eckert et al., 2021; Macke et al., 2020) and are crucial for host survival and reproduction (Sison-Mangus et

al., 2015). The presence of specific bacteria such as *Limnohabitans* sp. in the *Daphnia* microbiome affects

host abundance (Peerakietkhajorn et al., 2015). *Daphnia* microbiome compositions also fluctuate across

multiple timescales: influenced by diurnal feeding rates (Pfenning-Butterworth et al., 2022), host age in

laboratory conditions (Callens et al., 2018), and seasonal variations in wild populations (Hegg et al., 2021).

Moreover, *Daphnia* populations experience strong selection from various microparasites, which can regulate

their population densities (Ebert, 2005) and subsequently, their microbiome compositions. For instance,

Daphnia infected with the virulent eukaryotic gut parasite *Caullerya mesnili* (Class: Ichthyosporea; hereafter

Caullerya) experience reduced diversity and altered community composition of their associated bacteria

(Rajarajan et al., 2022).

Despite advancements in understanding host-microbiome and host-parasite interactions, few studies have

characterized their relative effects on wild zooplankton communities, or their association with long-term

disease dynamics. Here, we investigate *Daphnia* microbiome variation in Lake Greifensee across six

Caullerya epidemics spanning 13 years. *Caullerya* commonly parasitizes *Daphnia* guts and is widespread in

European lakes (Wolinska et al., 2007). This parasite reduces host fecundity and life span (Lohr et al., 2010),

thereby exerting strong selection pressure on host populations (González-Tortuero et al., 2016; Turko et al.,

2018). We analyse how *Daphnia* microbiome compositions correlate with *Caullerya* epidemic size, host

density and environmental conditions (temperature, dissolved oxygen and abundance of cyanobacteria in

lake water), which have been identified as key drivers of *Caullerya* epidemics in this lake (Tellenbach et al.,

2016). We also test if the community composition of *Daphnia*-associated bacteria correlates with the

composition and density of dominant pelagic phytoplankton (cyanobacteria, green algae, protists, golden

algae and diatoms), and zooplankton (copepods, cladocerans, rotifers). We hypothesize that (a) *Daphnia*

microbiome composition (i.e. dominant bacterial orders in *Daphnia*) and beta diversity differ between years

and by host infection status; (b) environmental gradients (temperature, dissolved oxygen and cyanobacterial

abundance), known to drive *Caullerya* epidemics, correlate with *Daphnia* microbiome structure and (c)

phytoplankton community composition predicts *Daphnia* microbiome structure, reflecting the role of diet in

shaping host-associated microbial communities.

Material & Methods

***Daphnia* sampling and processing:** *Daphnia* were collected from Lake Greifensee, Switzerland (N 47°20′41″, E 8°40′21″), using vertical towing with a 150 µm zooplankton net from 0-30 m depths at three sampling stations. Zooplankton samples from all stations were pooled into a single 10 L canister half-filled with surface lake water. Samples were collected during the peaks of natural *Caullerya* epidemics on six occasions over 13 years (10/10/2007, 11/08/2011, 29/08/2013, 04/09/2014, 24/08/2017, 10/11/2020), as part of a long-term *Daphnia* infection study (Turko et al., 2018). For clarity, these sampling events are hereafter referred to as "years", as they represent different calendar years around epidemic peaks (Fig. S1). Upon collection, zooplankton samples were sieved through a ~250 µm mesh. Adult specimens (>1 mm in size, independent of infection status) belonging to the *Daphnia longispina* species complex were gently lifted by the antennae and screened for *Caullerya* spores in the gut under 50× magnification (Lohr et al., 2010). Individuals with *Caullerya* spores were classified as 'infected', whereas those without visible spores were categorized as 'uninfected'. *Daphnia* with resting eggs (ephippia), or with other infections or rare co-infections involving *Caullerya* and other parasites, were excluded. Individual *Daphnia* were then transferred to wells containing 20 µL autoclaved MilliQ water and stored at -80 °C within eight hours of collection from the lake, until further processing.

Bacterial community profiling of *Daphnia*: DNA was extracted from whole *Daphnia* individuals using the HotSHOT method (Montero-Pau et al., 2008). Samples were incubated in an alkaline lysis solution at 95 °C for 30 min, cooled on ice for 3-4 min, and eluted in a neutralizing solution. DNA extraction was carried out in three blocks (Table S1). Three DNA extraction negative controls containing only extraction reagents were added in the third DNA extraction block. Extracted DNA was stored at -20 °C until further processing. Subsequent PCR steps were carried out simultaneously for all samples. Bacterial DNA (V4-V5 region of the 16S gene) was amplified using barcoded universal 16S primers 515F (5'-GTGCCAGCMGCCGCGGTAA-3') (Caporaso et al., 2011) and m806R (5'-GGACTACNVGGGTWTCTAAT-3') (Apprill et al., 2015) under the following cycling conditions: initial denaturation at 95 °C – 3 min; (98 °C – 20 s, 52 °C – 15 s, 72 °C – 15 s) 33 cycles; followed by a final extension at 72 °C for 5 min. This PCR step was done with three technical replicates for each sample, and 10 PCR no-template negative controls were included. Technical replicates were pooled before the next PCR step. For the second PCR, indexes from the Illumina Nextera XT

V2 Library Prep Kit were then added to the samples under the conditions: initial denaturation at 95 °C – 3

min; (95 °C – 30 s, 55 °C – 30 s, 72 °C – 30 s) 10 cycles; followed by a final extension at 72 °C for 5 min.

The location of DNA samples (including a total of 13 negative controls, i.e. 3 "blank" DNA extraction and

10 PCR no template samples) in the 384-well plate was randomized throughout the workflow, though

technical replicates in the first PCR step were in adjacent wells. Indexed PCR products were then purified

using magnetic Agencourt AMPure XP beads (Cat #A63881), quantified using a Qubit HS Assay,

normalized and pooled into a 1.2 nM library for amplicon sequencing. Paired-end sequencing was carried out

using MiSeq with 600 cycles (PE300) v3 run kit and 10% PhiX. Final sample sizes for analyses (Table S1)

were determined after bioinformatic processing (see below).

Pre-processing of reads: Sequencing generated 7.6M paired-end reads, after exclusion of one infected

sample from 2014 that had an unusually high number of reads (~720,000 as opposed to the average of 46,230

reads in biological samples). Initial pre-processing steps were performed on the Euler computing cluster at

ETH Zurich, as previously described (Rajarajan et al., 2022). Raw reads were 5'-end trimmed, merged, and

quality-filtered. Amplicons were denoised with UNOISE3 (Edgar & Flyvbjerg, 2015) and the resulting Zero-

radius Operational Taxonomic Units (ZOTUs) were assigned taxonomy using the non-Bayesian SINTAX

classifier v.138 (Edgar, 2016) against the Silva database v138 (Quast et al., 2013) with a confidence cut-off

of 0.75. Biological samples had 320 to 237,358 reads and negative controls had 412 to 2188 reads, with one

PCR no-template control unexpectedly showing 24,356 reads. Based on this distribution of reads in

biological samples versus negative controls, biological samples with <3000 reads were excluded from

downstream analyses. This resulted in a final sample size of 9-16 *Daphnia* per year and infection status,

totalling 71 uninfected and 90 infected individuals (Table S1).

Subsequent filtering steps were performed in R v4.0.2 (R Core Team, 2023) using the phyloseq (McMurdie

and Holmes 2013) and vegan (Dixon, 2003) packages. 62 ZOTUs (21 of unidentified phyla, 18 chloroplasts,

6 mitochondria, 16 singletons and 1 eukaryote) were filtered out. An additional 42 ZOTUs were removed

after rarefying samples to 3,017 reads using vegan::*rarefy_even_depth*, resulting in a final dataset of 423

ZOTUs. There was no consistent increase or decrease in ZOTU richness from older to newer samples,

indicating that long-term freezing did not impact microbiome composition (Fig. S2).

Environmental data for Lake Greifensee: Sampling dates for physicochemical, zooplankton, and

phytoplankton data are listed in Table S1. Depth-resolved physicochemical data for Lake Greifensee were

obtained from the Office of Waste, Water, Energy and Air of the Canton of Zurich (AWEL,

http://www.awel.zh.ch/). Temperature and dissolved oxygen measurements were averaged for the top 15 m

of the lake. Phytoplankton were sampled from 0-20 m depth using a Schröder bottle. Phytoplankton

(including cyanobacterial) cell counts in lake water were measured with the Utermöhl method (Utermöhl,

1958) using sedimentation chambers. Dominant phytoplankton and microbial groups - diatoms (Fig S3),

golden algae (Fig. S4), green algae (Fig. S5), protists (Fig. S6) and cyanobacteria (Fig. S7) were identified to

the lowest possible taxonomic level based on morphology, using inverted light microscopy. Zooplankton

samples were collected with three vertical tows of a 95 µm plankton mesh at the same locations where

Daphnia samples were taken, and were preserved in 100% ethanol. To assess *Caullerya* infection rates, 200

adult *Daphnia* were visually screened for *Caullerya* spores in their gut (see above for details).

Statistical analyses: Statistical analyses were performed in R v4.0.2 (R Core Team, 2023) using the

phyloseq (McMurdie & Holmes, 2013) and vegan (Dixon, 2003) packages. Variation in beta diversity

(Unifrac distance and Bray-Curtis Dissimilarity) by year and infection status was tested with two-way

PERMANOVAs (9999 perm) using vegan::*adonis.* The proportion of variation explained by year and

infection status was quantified with a db-RDA using the function vegan::*capscale.* Given the substantial

interannual variation in bacterial communities (see Results), subsequent analyses of infection status-effects

were carried out separately for each year. Dominant bacterial orders were compared by infection status,

where dominant orders were defined as those comprising more than 1% of the dataset or present in more

than 50% of all samples. Orders not meeting these criteria were lumped into the category "Other".

Differential abundance of bacterial orders by infection status was tested using the DESeq2 package, with *p*-

values adjusted for multiple comparisons using the "fdr" method (Love et al., 2014). Beta diversity in

Daphnia bacterial communities was correlated with epidemic size (i.e. population infection rates),

temperature, dissolved oxygen, and cyanobacterial cell counts using vegan::*envfit*, with *p*-values adjusted for

multiple comparisons using the Bonferroni method. To test for potential associations between *Daphnia*

bacterial communities and plankton composition in the lake, Jaccard dissimilarity matrices of *Daphnia*

bacterial communities were correlated with those of (a) phytoplankton communities and (b) zooplankton

communities, using a Mantel test.

Results

<u>Bacterial community beta diversity differs between years and marginally by *Caullerya*-infection status</u>

Analyses of beta diversity (Bray-Curtis Dissimilarity and Unifrac distance) revealed that the primary driver

of variation in bacterial communities was the sampling year, accounting for about 30% of the observed

variation (Table 1). In contrast, infection status accounted for only 1% of the variation. The beta diversity of

dominant taxa in *Daphnia*-associated bacterial communities, assessed by Bray-Curtis Dissimilarity (which

weighs bacterial taxa by their relative abundances), differed significantly between years and marginally

between infected and uninfected *Daphnia*. When phylogenetic relatedness between ZOTUs was considered

using Unifrac distance, both year and infection status had significant effects on beta diversity (Fig. 1, Table

1). However, there was no significant interaction between year and infection status in shaping overall

microbiome community structure as assessed with beta diversity.

Table 1. Two-way PERMANOVAs using the beta diversity indices Bray-Curtis dissimilarity (top section)

and Unifrac distance (lower section) between years and by and infection status (9999 permutations).

Significant results (*$p < 0.05$) are indicated with an asterisk. The "% Variation" column indicates the

percentage of variation explained by each factor, as determined by db-RDA. Significant variation in beta

diversity by year (but not infection status) may be due differential dispersion of bacterial communities

among years, as indicated by vegan::betadisper.

	Bray Curtis Dissimilarity						**% Variation**
	Df	Sums Sq	Mean Sq	F value	R-squared	*p* value	
year	5	15	2.99	15.36	0.327	0.0001*	33
infection	1	0.4	0.408	2.09	0.009	0.0581	0.92
Residuals	156	30.4	0.195		0.664		
Total	162	45.7			1		
	Unifrac Distance						**% Variation**
	Df	Sums Sq	Mean Sq	F value	R-squared	*p* value	
year	5	11.1	2.219	13.7	0.303	0.0001*	30.28
infection	1	0.3	0.292	1.8	0.008	0.0337*	0.79
Residuals	156	25.3	0.162		0.689		
Total	162	36.7			1		

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Figure 1. Nonmetric Multi-Dimensional Scaling (NMDS) plots of beta diversity (A) Bray-Curtis

Dissimilarity and (B) Unifrac distance in individual *Daphnia*, grouped by sampling year and *Caullerya*

infection status. Data point shapes indicate infection status, while colors indicate the year of sampling.

Arrows show abiotic (temperature, dissolved oxygen) and biotic (*Caullerya* epidemic size, zooplankton, and

phytoplankton densities) variables that significantly correlate with beta diversity. The length of arrows

reflects the strength of correlation, and are scaled to the respective NMDS1 and NMDS2 axes, which differ

for the two diversity metrics. Ellipses show 95% confidence intervals for each year.

<u>Relative abundances of dominant bacterial orders differ between *Caullerya*-infected and uninfected</u>

<u>*Daphnia*</u>

Bacterial communities of individual *Daphnia* comprised of 16 dominant orders across all sampling years,

with substantial variation among individuals sampled within the same year (Fig. 2). The most abundant

bacterial order across all *Daphnia* individuals was Burkholderiales ($51.29 \pm 24.94\%$, mean $\pm$ SD), followed

by Rickettsiales ($15.1 \pm 26.22\%$) and Flavobacteriales ($14.58 \pm 12.73\%$). Differences in the relative

abundances of dominant bacterial orders between *Caullerya*-infected and uninfected *Daphnia* occurred

during the epidemic years 2007, 2013 and 2017. In 2007, Caulobacterales was more abundant in uninfected

Daphnia ($2.94 \pm 2.36\%$) compared to infected ones ($0.96 \pm 0.7\%$). Conversely, Chitinophagales was more

abundant in infected *Daphnia* in both 2007 ($1.34 \pm 1.22\%$) and 2013 ($2.51 \pm 3.97\%$), compared to uninfected

ones ($0.45 \pm 0.46\%$ and $1.5 \pm 2.9\%$, respectively). In 2013, Rickettsiales was more abundant in infected

Daphnia ($5.18 \pm 10.28\%$) than in uninfected ones ($2.12 \pm 4.7\%$). In 2017, Aeromonadales was more

abundant in uninfected *Daphnia* ($4.64 \pm 13.25\%$) compared to infected individuals ($1.66 \pm 3.22\%$), while

Pirellulales showed higher abundance in infected *Daphnia* ($0.38 \pm 0.59\%$) compared to uninfected ones (0.07

$\pm 0.12\%$, Table 2).

Figure 2. Relative abundances of dominant bacterial orders in individual *Daphnia* across years, with *Caullerya*-infected *Daphnia* in the top row and uninfected *Daphnia* in the bottom row. Each vertical bar corresponds to the bacterial community composition of a single *Daphnia* individual. The category "Other" includes bacterial orders comprising <1% of the dataset and present in <50% of samples. Missing bars indicate samples excluded from analyses (see Methods).

249 **Table 2**. Wald test results comparing the relative abundances of dominant bacterial orders between *Caullerya*-infected and uninfected *Daphnia*, analyzed separately

250 for each year. *p*-values were adjusted for multiple comparisons using the "fdr" method, with significant values (*p*<0.05) shown in bold. The "log2Change" column

251 indicates the logarithmic fold-change in taxon abundance, where negative numbers indicate higher abundance in infected *Daphnia*, and positive numbers indicate

252 higher abundance in uninfected *Daphnia*. The absolute "log2Change" values reflect the magnitude of the difference.

Bacterial Orders	2007 log2Change	*p* value	2011 log2Change	*p* value	2013 log2Change	*p* value	2014 log2Change	*p* value	2017 log2Change	*p* value	2020 log2Change	*p* value
Absconditabacteriales	-1.96	0.12	-0.96	0.48	1.48	0.97	0.28	0.80	-1.55	0.17	-0.94	0.71
Aeromonadales	-1.11	0.38	-1.31	0.37	-0.53	0.97	1.84	0.25	**3.54**	**0.02**	-1.20	0.75
Bacillales	0.03	0.99	0.38	0.84	-0.01	1.00	0.88	0.25	0.93	0.10	0.32	0.75
Burkholderiales	-1.02	0.08	0.76	0.37	0.00	1.00	-0.57	0.25	0.06	0.92	-0.08	0.94
Caulobacterales	**1.03**	**0.04**	0.52	0.96	0.03	1.00	0.06	0.98	0.27	0.68	-0.48	0.71
Chitinophagales	**-2.04**	**0.04**	-0.09	0.97	**-3.09**	**0.00**	-0.42	0.44	-0.20	0.68	0.64	0.57
Enterobacterales	0.08	0.94	-1.67	0.37	0.47	0.97	1.45	0.25	0.97	0.17	1.04	0.57
Flavobacteriales	-1.07	0.38	0.30	0.84	-0.14	1.00	-0.40	0.33	-0.44	0.62	-0.61	0.57
Pirellulales	-0.02	0.99	-0.54	0.81	1.74	0.63	0.64	0.75	**-2.91**	**0.02**	-1.95	0.57
Pseudomonadales	0.29	0.38	-0.52	0.51	-0.02	1.00	0.89	0.19	1.20	0.09	-0.60	0.57
Rhizobiales	0.32	0.55	0.13	0.97	0.07	0.98	0.96	0.75	0.46	0.62	-0.05	0.96
Rhodobacterales	0.48	0.61	0.18	0.97	-0.25	0.98	2.05	0.49	-1.50	0.34	-0.68	0.71
Rickettsiales	1.98	0.12	0.38	0.93	**-1.88**	**0.05**	1.64	0.19	-0.25	0.92	1.03	0.57
Sphingobacteriales	-2.11	0.08	1.08	0.37	-0.60	0.97	-2.64	0.19	-0.41	0.68	0.65	0.71
Sphingomonadales	1.30	0.12	0.11	0.97	0.11	0.98	-0.29	0.89	-0.04	0.92	0.04	0.96
Verrucomicrobiales	-0.82	0.55	-0.89	0.41	-0.68	0.98	0.92	0.44	0.88	0.34	1.22	0.57
Other	0.28	0.51	-0.60	0.37	0.11	0.97	0.52	0.42	-0.47	0.55	-0.49	0.57

Daphnia bacterial community composition correlates with environmental conditions and plankton

density

Both Bray-Curtis dissimilarity and Unifrac distance metrics of _Daphnia_ bacterial communities showed

significant correlations with temperature, dissolved oxygen content, and _Caullerya_ epidemic size. Moreover,

these metrics were significantly associated with _Daphnia_ density and the densities of several phytoplankton

and microbial groups in the lake water, such as cyanobacteria, green algae and protists (Fig 1, Table 3). In

contrast, only Unifrac distance of _Daphnia_ bacterial communities was significantly correlated with total

zooplankton density and the density of golden algae in the lake water.

Table 3 Correlation of *Daphnia* bacterial community beta diversity indices (Bray-Curtis dissimilarity and Unifrac distance) with abiotic (temperature, dissolved oxygen) and biotic factors (*Caullerya* epidemic size, and densities of phytoplankton and zooplankton groups in lake water), analyzed using vegan::*envfit* (9999 permutations). *p*-values were adjusted for multiple comparisons using the Bonferroni method. Significant results (*$p < 0.05$) are indicated with an asterisk.

Bray-Curtis dissimilarity		
Variable	**r-squared**	***p*-adjusted**
Dissolved oxygen (mg/L)	0.2069	0.001*
Green algal density	0.1649	0.001*
Temperature (°C)	0.1325	0.001*
Protist density	0.0756	0.029*
Golden algal density	0.061	0.376
Diatom density	0.0566	0.696
Epidemic size	0.0456	0.011*
Daphnia density	0.0436	0.001*
Cyanobacterial density	0.0378	0.001*
Total zooplankton density	0.0089	1
Unifrac distance		
Variable	**r-squared**	***p*-adjusted**
Temperature (°C)	0.2776	0.001*
Green algal density	0.2771	0.001*
Cyanobacterial density	0.2209	0.001*
Protist density	0.1718	0.001*
Dissolved oxygen (mg/L)	0.1376	0.001*
Daphnia density	0.1102	0.001*
Epidemic size	0.0705	0.001*
Diatom density	0.023	1
Total zooplankton density	0.0208	0.001*
Golden algal density	0.0087	0.001*

Daphnia bacterial community composition does not correlate with diet (phytoplankton) or zooplankton community composition

Jaccard dissimilarity matrices of *Daphnia* bacterial communities did not exhibit significant correlations with those of phytoplankton, including green algae, golden algae, protists, cyanobacteria, or diatoms (*$p > 0.05$*, Mantel test). Similarly, *Daphnia* bacterial community composition showed no significant correlation with the composition of pelagic zooplankton, including copepods, rotifers, and *Daphnia* (Table 4).

276   **Table 4.** Mantel tests assessing the correlation between presence-absence beta diversity (Jaccard

277   dissimilarity) of *Daphnia*-associated bacteria and phytoplankton taxa (representing *Daphnia* diet) in lake

278   water, across years. The ZOTUs of individual *Daphnia* bacterial communities were pooled for each year.

279   Bacterial community composition of dominant taxa in lake water was determined using microscopy.

Jaccard dissimilarity		
Taxonomic group	**r**	***p*-value**
Cyanobacteria	0.31	0.22
Green algae	0.40	0.09
Golden algae	0	0.37
Protists	0.26	0.16
Diatoms	-0.26	0.78
Zooplankton	-0.19	0.69

Discussion

We compared the bacterial community (microbiome) composition of wild *Caullerya*-infected and uninfected

Daphnia collected across six natural epidemics spanning a remarkable 13-year period. This study is the first

to investigate potential biotic (parasitism, diet) and abiotic drivers of wild aquatic animal microbiomes over

such an extensive timeframe. As with all observational field investigations, the relationships between

Daphnia-associated bacterial diversity and its potential environmental drivers remain correlational.

Nevertheless, many of the investigated factors correlated significantly with microbiome composition,

highlighting the dynamic nature of *Daphnia* microbiomes in response to environmental conditions.

We found that *Daphnia* microbiomes differ significantly between years, both in the relative abundances of

dominant taxa and their phylogenetic composition. Temporal differences in microbiome beta diversity were

more pronounced than differences between infected and uninfected *Daphnia*. While infection status

significantly influenced the phylogenetic composition of rare bacterial taxa (Unifrac distance), its effect on

dominant taxa (Bray-Curtis dissimilarity) was marginal. Certain bacterial orders such as Chitinophagales

exhibited infection-related changes in relative abundance, but this pattern was inconsistent across years.

Specifically, infection-associated shifts in dominant bacterial orders were evident in only three of the six

surveyed epidemics (2007, 2013 and 2017), with the specific bacterial orders affected varying annually.

However, Burkholderiales and Flavobacteriales taxa dominated *Daphnia* microbiomes in all surveyed years.

Environmental conditions previously shown to drive *Caullerya* epidemics - including cyanobacterial

abundance and water temperature (Tellenbach et al., 2016), as well as *Caullerya* epidemic size itself,

correlated significantly with *Daphnia* microbiome compositions. These findings provide the first evidence of

a link between aquatic host microbiomes and epidemic size over a long-term, natural study. Our results

generally align with studies in other aquatic animals, such as the frog *Philoria loveridgei* (López et al., 2017)

and the sponge *Porites astreoides* (Meyer et al., 2014), where temporal variation in host microbiomes was

observed, with infection-associated taxa differing by year. Lastly, contrary to our hypothesis, the

composition of the *Daphnia* diet (phytoplankton community) did not associate with *Daphnia* microbiome

composition, although the abundance of certain phytoplankton groups (green algae and cyanobacteria, but

not diatoms) correlated with microbiome structure (beta diversity).

Caullerya epidemics are associated with temporal shifts in *Daphnia* population-genetic structure (Schoebel

et al., 2011; Turko et al., 2018). Given that *Daphnia* genotypes can harbour distinct microbiomes (Frankel-

Bricker et al., 2020; Rajarajan et al., 2023), our findings suggest complex interactions between host genetics,

infection status and microbiome compositions. For instance, the order Rickettsiales was enriched in infected

Daphnia in 2013 but not in other years, including 2020 in the present study. In contrast, we previously noted

a putative increase in the relative abundance of Rickettsiales in uninfected compared to *Caullerya*-infected

Daphnia, among animals collected at a different timepoint during the *Caullerya* epidemic in 2020 (Rajarajan

et al., 2022). Similarly, we found no differences in the relative abundance of Enterobacterales between

infected and uninfected animals in the present study, including November 2020 (see Fig. S1), whereas

Enterobacterales was enriched in infected *Daphnia* from December 2020 (Rajarajan et al., 2022). These

patterns suggest that *Daphnia* microbiomes may shift over the course of a single epidemic, consistent with

the strong temporal variation detected in our study. Such shifts might be potentially be related to host clonal

turnover during epidemics (Turko et al., 2018).

Our results suggest a link between environmental conditions driving *Caullerya* epidemics (temperature,

cyanobacterial density) and *Daphnia* microbiomes. Temperature influences *Caullerya* infection prevalence

under experimental settings (Schoebel et al., 2011), and natural epidemics are associated with both

temperature and high cyanobacterial abundance in lake water (Tellenbach et al., 2016). Moreover, *Daphnia*

are more susceptible to *Caullerya* infection when fed a diet consisting of cyanobacteria (regardless of

toxicity), as confirmed with a laboratory experiment (Tellenbach et al., 2016). These factors (temperature,

cyanobacterial diet) influence not only *Caullerya* infection, but also the composition of *Daphnia*

microbiomes (Akbar et al., 2021; Frankel-Bricker et al., 2020; Sullam et al., 2018). In the present study,

temperature and cyanobacterial density in the lake water correlated significantly with microbiome

composition during *Caullerya* epidemics, with the additional correlation with epidemic size itself suggesting

that microbiomes generally differed between *Daphnia* from small versus large epidemics. *Daphnia*

microbiomes can mediate tolerance to toxic cyanobacteria (Macke et al., 2017), potentially promoting grazer

fitness. This raises the possibility of a bottom-up effect whereby *Daphnia* microbiomes shape host traits that

in turn influence phytoplankton abundance. However, this hypothesis requires empirical validation

(Decaestecker et al., 2024). The observed correlation between microbiome structure, temperature, and

cyanobacterial abundance thus supports the idea that host microbiomes were associated with the specific

environmental conditions (temperature and cyanobacterial abundance) that additionally drive disease

epidemic size.

Beyond cyanobacteria, the abundance of certain phytoplankton groups (green algae and protists), which are

grazed on by *Daphnia,* associated with distinct *Daphnia* microbiome compositions. In natural settings,

Daphnia-associated bacterial communities vary independently of the composition of bacterioplankton

communities (Hegg et al., 2021), whereas laboratory investigations have shown how diet quality shapes their

taxonomic composition (Akbar et al., 2021) and functional activity (Li et al., 2021). However, our data did

not support the idea that similarity in diet composition (algal and protist communities in lake water) predicts

host microbiome composition. This result was contrary to our expectations, as well as reports in marine

zooplankton (Xu et al., 2024). Nevertheless, high densities of green algae and protists were associated with

distinct *Daphnia* microbiome structures in 2007 and 2013 (see Fig. 1 & Fig. S2). In contrast, the density of

diatoms (e.g. *Asterionella* sp.*, Fragillaria* sp. and *Melosira* sp., see Fig. S3) and golden algae (e.g. *Erkenia*

sp. (=*Chrysochromulina* sp.), *Uroglena* sp. and *Dinobryon* sp., see Fig. S4) showed no relationship with

Daphnia microbiome structure, possibly due to their large size, colonial form or motility, making them less

ingestible for smaller *Daphnia* species (Cáceres, 1998; Talling, 2003; Wagner, 2008; Yoshida et al., 2001)

(albeit golden algal density weakly correlated with the phylogenetic composition of rare bacterial taxa within

the *Daphnia* microbiome). High densities of green algae trigger increased *Daphnia* ingestion rates (Porter et

al., 1982) which in turn may drive changes in microbiome compositions (Pfenning-Butterworth et al., 2022).

Our results posit that food abundance, rather than its specific dietary composition, may play a key role in

shaping microbiome structures in wild *Daphnia* populations, consistent with findings that low food

availability reduces microbial richness in *Daphnia* (Freese & Schink, 2011; Ichige & Urabe, 2023).

Daphnia population density also correlated with the microbiome beta diversity, with microbiomes

consistently dominated by Burkholderiales. Burkholderiales taxa are widespread in aquatic hosts including

sponges (Paix et al., 2024), corals (Ziegler et al., 2019), and fish (Pérez-Pascual et al., 2021), and are

facilitated by 'symbiotic islands' or genomic regions promoting colonization and survival within hosts

(Stillson et al., 2022). Beneficial Burkholderiales strains may enhance *Daphnia* fitness, potentially driving

high host abundances in natural environments (Peerakietkhajorn et al., 2015). Alternatively, high *Daphnia*

population densities may facilitate horizontal transmission of microbiota (Miller et al., 2018). Both

freshwater (Eckert et al., 2021; Rajarajan et al., 2023) and marine (Boscaro et al., 2022) zooplankton

microbiomes are strongly shaped by their environment, with no evidence of phylosymbiosis - i.e.

microbiome compositions are not determined by host species identity or relatedness. Horizontal transmission

between individuals may therefore be a key factor shaping the microbiomes of wild zooplankton (Russell,

2019) and is known to shape *Daphnia* microbiomes in laboratory conditions (Cooper et al., 2022). Thus,

high host population densities may facilitate horizontal transmission of symbionts, increasing their

prevalence in hosts. In addition to *Daphnia* population density, we found that the phylogenetic composition

of rare, but not dominant, bacteria in wild *Daphnia* microbiomes was correlated with the total density of

other zooplankton species in the lake water. This suggests that the presence and abundance of other species

in a host community may also influence the structure of wild host microbiomes, although to a lesser extent.

In conclusion, our study highlights the strong temporal variation in *Daphnia* bacterial community

composition. Diet quantity, rather than composition, and host population density emerge as key factors

potentially shaping these communities. Additionally, epidemic size and environmental conditions that drive

disease outbreaks, such as temperature and cyanobacterial abundance, were also associated with microbiome

structure. Thus, wild *Daphnia* microbiomes may be linked to host health and disease dynamics through

complex interactions involving temperature, diet quantity, and population density.

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Author Contribution

Amruta Rajarajan: Conceptualization (equal); Data curation: (lead); Formal analysis (lead); Investigation (lead); Writing – original draft preparation (lead); Writing – review & editing (equal). **Justyna Wolinska**: Conceptualization (equal); Formal analysis (supporting); Writing – original draft preparation (supporting); Writing – review & editing (equal); Funding acquisition (equal); Supervision (equal). **Jean-Claude Walser**: Data curation (supporting); Formal analysis (supporting); Methodology (equal); Writing – review & editing (supporting). **Nadine Tardent**: Investigation (supporting); Writing – review and editing (supporting). **Silvana Käser**: Investigation (supporting); Methodology (equal); Writing – review and editing (supporting). **Esther Keller**: Investigation (supporting); Methodology (equal); Writing – review and editing (supporting). **Piet Spaak**: Conceptualization (equal); Formal analysis (supporting); Writing – original draft preparation (supporting); Writing – review and editing (supporting); Funding acquisition (equal); Supervision (equal).

Data availability statement

Raw 16S rRNA gene sequences of *Daphnia* samples included in the final analysis are available in the Sequence Read Archive (SRA) under BioProject ID PRJNA1252098. Long-term phytoplankton and zooplankton community data and processed 16S rRNA gene sequence data analysed in this manuscript, as well as R code are available at https://doi.org/10.25678/0008GD.