Transgenerational effects impact the vulnerability of a host-parasitoid system to rising temperatures.

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Abstract

Transgenerational effects, where ancestral experience of environmental conditions influences the performance in subsequent generations, are hypothesised to have substantial consequences for responses to climate change. However, any apparent detriment or advantage these processes generate for a focal species may be counteracted by concurrent effects upon interacting species. Using an experimental *Drosophila*-parasitoid model system, we determined how the parental thermal environment impacts the performance of both hosts and parasitoids. We found substantial responses in both trophic-levels, with potential evidence for both condition-transfer effects and adaptive transgenerational plasticity. We used these results to parameterise discrete-time simulation models to explore how transgenerational effects of thermal conditions and temporal autocorrelation in temperature are expected to impact the time to extinction for this host-parasitoid system under climate change. The models predicted that transgenerational effects would significantly hasten the time to extinction, largely through a reduction in estimated average performance. Under a simple model, there was an additional hastening of extinction derived from the interaction between the coupled transgenerational effects, although no interactive effects were detectable in an alternative model with greater buffering capacity. Our research demonstrates a new mechanism underlying how tightly-coupled interacting species will respond to climate change and highlights the need to quantify and contextualise thermal transgenerational effects.

Introduction

Transgenerational effects, where the conditions experienced in earlier generations impact the performance in subsequent generations without evolutionary change (Galloway & Etterson, 2007), are increasingly recognised as an influential determinate of climate change risk (Donelson et al., 2018). However, although it is widely appreciated that direct effects of climate change on species will be modulated by both the wider community context (Carroll et al., 2024; Chen & Lewis, 2023; Davis et al., 1998; Gilman et al., 2010; Tylianakis et al., 2008; Urban et al., 2016), the impact of transgenerational effects in a multispecies context has been underexplored (Malinski et al., 2024). Focal species responses may be offset or exacerbated by corresponding responses of interacting species, and since temperature is an important abiotic factor for consumer-resource interactions (Amarasekare, 2015), the strength of the interactions with other species that may also be directly influenced by transgenerational responses to a fluctuating climate. Transgenerational effects have important implications for understanding the impact of autocorrelation in environmental fluctuations (Heino et al., 2000; Petchey et al., 1997). Recent work has identified that thermal performance plasticity may be stronger than thermal limit plasticity (Pawar et al., 2024; Weaving et al., 2022), heightening the potential role of cross-generational effects where range limits are determined by interactions with other species rather than absolute environmental limits (Chen & Lewis, 2024; Paquette & Hargreaves, 2021).

Transgenerational effects can include both condition-transfer and adaptive plasticity effects through anticipatory effects, which can be challenging to disentangle (Sánchez☐Tójar et al., 2020). Condition-transfer, the direct modulation of the amount of investment the parental generation can make in the next by the suitability of the conditions is likely the most widespread form of transgenerational effect (Bonduriansky & Crean, 2018). However, where environmental fluctuations are predictable (Burgess & Marshall, 2014), this information could be utilised to better anticipate the environment experienced by the next generation, reducing the lag between environmental signal and phenotypic response (Cavieres et al., 2020). Such priming could significantly alter the extinction risk posed by climate change (Salinas & Munch, 2012), although the supporting evidence for their prevalence is still relatively weak (Sánchez☐Tójar et al., 2020; Uller et al., 2013) and the magnitude and direction context-dependent (Baker et al., 2019; Yin et al., 2019). Furthermore, identifying a transgenerational effect is only the first step to demonstrating its impact – the effect must be contextualized into an environmental (Burgess & Marshall, 2014) and biotic context (Walther, 2010).

Here we investigate the impact of transgenerational effects in a model experimental insect host-parasitoid system. Parasitoids are ectotherms whose larvae develop by feeding in or on the bodies of other arthropods, representing ~10% of described insect species (Eggleton & Belshaw, 1997; Godfray, 1994). By strongly regulating arthropod host population dynamics and promoting coexistence

between different host species, parasitoids play important roles in maintaining biodiversity and influencing community structure (Godfray, 1994). Host-parasitoid interactions thus represent an ecologically important subset of biotic interactions, yet are relatively underrepresented in literature on biotic responses to climate change (Furlong & Zalucki, 2017; Hance et al., 2007; Jeffs & Lewis, 2013; Thierry et al., 2019). Host-parasitoid interactions offer an experimentally amenable window to start to parse ecosystem complexity. Since the reproduction of parasitoids is closely linked to their success finding and attacking hosts (Hassell, 2000), the dynamics are tightly coupled, with the high potential for emergent effects to arise in their response to climate change (Furlong & Zalucki, 2017).

Insects are likely to be amongst the most responsive taxa to climate change (Harvey et al., 2023) and the impact of temperature on parental generations on insect fitness has long been registered (Mousseau & Dingle, 1991). Thermal transgenerational effects have been demonstrated in *Drosophila* in a variety of contexts (Cavieres et al., 2020; Schiffer et al., 2013) and various responses to fluctuating temperature have been documented in parasitoid attack rates (Costaz et al., 2023; Ismaeil et al., 2013; Le Lann et al., 2021). Many of the processes intrinsic to parasitoids' life-history demonstrate temperature dependence; for instance, foraging traits, such as successfully locating and ovipositing in a host egg, and attack traits, such as evading the host's immune defences (Hance et al., 2007; Synodinos et al., 2021). A common prediction is that parasitism will increase with rising temperatures (Péré et al., 2013); however the evidence for this is mixed (Thierry et al., 2019, 2021). Due to their high trophic position and typical specialisation, parasitoid populations are also potentially particularly vulnerable, with their fate often tied to that of their hosts (Jeffs & Lewis, 2013).

To explore the role of transgenerational effects within a tightly-coupled trophic system responding to generation-scale variance in temperature, we first conducted highly replicated experimental trials to assess how the thermal performance curves of both host and parasitoid are influenced by the temperatures the previous generation is exposed to. Focussing on the impacts on the population dynamics under warming conditions, we then used the observed responses to parametrise simulation models to explore the impact of the observed transgenerational effects under climate change scenarios, specifically the potential for interactions between transgenerational effects across trophic levels. Any experimentally identified cross generational impacts can have at least three routes: 1) direct effects from a recorded shift in average performances of the focal species given variable parental conditions, 2) emergent dynamic impacts of the transgenerational effects and the autocorrelation regime, and 3) cross-trophic level effects from transgenerational effects on interacting species. We hypothesised that any identified transgenerational effects would have a mixture of direct, indirect and emergent impacts on the simulated population dynamics.

Methods

Study organisms

Drosophila (Diptera: Drosophilidae) provide a practical study system that have been widely used to study both thermal performance and community responses to climate change (Davis et al., 1998; Hoffmann & Sgrò, 2011; Rezende et al., 2020). In *D. melanogaster*, a significant interaction between parental and offspring thermal environment on the entire TPC has been previously identified (Cavieres et al., 2019; Gilchrist & Huey, 2001; Huey et al., 1995). Using *Drosophila*, rather than say algal models of phenotypic plasticity (Fey et al., 2021), allows for the discretisation of generations and a simplification of the modelling process.

We used *Drosophila sulfurigaster* (Duda, 1923) and a generalist hymenopteran parasitoid wasp *Asobara* (Braconidae: Alysiinae, strain: KHB, reference: USNMENT01557097, BOLD process ID:DROP043-21, species identifier: drop_Aso_sp8, Lue et al., 2021). *D. sulfurigaster* is a partially susceptible native host of *Asobara* parasitoids, allowing for some variation in the rate of parasitism across treatments. The parasitoid *Asobara*, attack the second-instar larval stage, with a single offspring emerging from each host pupa. Importantly *Asobara* (henceforth simply 'wasps') have approximately the same generation time as *D. sulfurigaster* (henceforth simply 'flies') simplifying both experiments and analysis. Prior to the experiment, the wasps had been maintained on highly susceptible *D. melanogaster* larvae.

Both host and parasitoid species were established from adults collected in the Wet Tropics bioregion of northeast Queensland, Australia. In their natural environment the Drosophila community suffers considerable parasitoid pressure with losses of between a third to a half of pupae (Jeffs et al., 2021) and laboratory studies have identified strong impacts on community dynamics (Terry et al., 2021). *Drosophila* and parasitoid cultures were originally exported under permit PWS2016-AU-002018 granted by the Department of Energy and Environment of the Australian Government. These cultures had been maintained at a constant 24°C at the Biology Centre, Czech Academy of Sciences, before being transferred to the Department of Biology, University of Oxford, where they were maintained at 25°C in 150ml bottles on standard cornflour–sugar-yeast–agar *Drosophila* food medium, in a 12:12h light: dark cycle.

Effect of temperature on fly fecundity

All experiments were carried out between September and December 2023 at the Department of Biology, University of Oxford. In the initial generation (G0), to generate a large number of flies at low density, approximately 20 adult flies were allowed to lay eggs for 24 hours in each of 48 bottles with 30ml of food medium. These bottles were randomly assigned to incubators set at one of three temperatures (19, 23, 27°C), under a 12:12h light:dark cycle (Figure 1). Temperature and humidity

loggers in each incubator showed that a controlled environment was maintained throughout the

experiments. Across all the experiments, once fly larvae emerged, pupation cards were added to

provide a vertical surface for pupation.

Subsequent steps were timed relative to developmental milestones, as time to eclosion is dependent on

the temperature treatment (nine days at 27°C, 11 days at 23°C and 15 days at 19°C). Since *D.*

sulfurigaster can take four days post-eclosion to mature enough for egg laying (J. Chen, *pers. obs.*), a

maturation time of five-days post-ecolosion was allowed before two male and two female G0 flies

were counted into acclimatisation vials containing food medium under light CO_2 anaesthesia. The

acclimatisation vials minimised the impact of the CO_2 anaesthesia, allowed acclimation to the new

temperature and better standardised egg-laying time in the subsequent test vial. Forty such

acclimatisation vials from each G0 temperature were randomly assigned to incubators set at one of

five temperatures (G1 temperatures: 19, 21, 23, 25, 27°C) for 24 hours (600 vials total). The

experimental temperatures were chosen to be representative of the study site the cultures of both

species originate from; where mean temperatures range from 21°C at the highest elevation sites, to

26°C at the lowest (Jeffs et al., 2021).

For the fecundity trials, the flies in each acclimatisation vial were transferred into standardised test

vials containing 5ml of food medium. Any deaths at this stage were noted, but otherwise considered

part of the performance assessment. These vials remained at their G1 test temperature. After 24 hours

of egg laying, the G0 adult flies were removed. Vials were checked regularly, and as G1 flies emerged,

they were removed, stored, then counted. In total, 15,755 flies were counted across the 600 trials.

Figure 1: Summary of experimental design for the fly temperature treatment experiment. G0 = Generation 0. G1 = Generation 1. Bottles containing approximately 20 flies were randomly assigned to incubators set at one of three G0 temperatures (19, 23, 27°C). After 24 hours to lay eggs, the adult flies were removed. Once G0 flies emerged and matured, two males and two females were counted into each acclimatisation vial. Acclimatisation vials were randomly assigned to incubators set at one of five temperatures (19, 21, 23, 25, 27°C) in a fully factorial design (n=40 per combination). After 24 hours of acclimatisation, flies were transferred into standardised test vials that remained in their assigned G1 test temperature to lay eggs for 24 hours. G1 emergences were regularly removed and counted.

Effect of temperature on wasp reproductive output

The wasp temperature treatment experiment followed a similar design (Figure 2). Standardised host vials were founded with five adult female flies, sufficient males, 5ml of standard food medium and a sprinkle of yeast granules to stimulate egg laying and reduce variance in egg laying rate. Vials were maintained at 25°C and adult flies were allowed to lay eggs for 48 hours before being removed. Wasps were added shortly after to attack the early instar larvae. Two male and two female wasps were added by aspiration to 48 standardised host wasp generation G0 host vials, and randomly assigned to one of the three G0 temperatures (19, 23, 27°C, 16 per temperature).

Due to differences in development time, the timing of G1 was staggered. Upon the first observed adult wasp emergence from each G0 treatment, G1 standardised host vials were prepared, to which two male and two female G0 wasps were added. These vials were randomly assigned to one of the five G1 test temperatures (19, 21, 23, 25, 27°C), following the fully factorial design. There were sufficient G0 wasps for 16-20 replicates for each treatment combination (total n=277). Test vials were checked regularly and both flies and wasps removed, stored and counted. In total, 6564 flies and 18,453 wasps were counted.

Figure 2: Summary of experimental design for wasp temperature treatment experiment. G0 = Generation 0. G1 = Generation 1. To initiate G0, two male and two female wasps were added to standardised vials of hosts. The vials were stored in incubators at either 19, 23, 27°C. Two male and two female wasps recently emerged were transferred to G1 host vials randomly assigned to G1 temperatures (19, 21, 23, 25, 27°C), in a full factorial design. Each treatment combination had between 16 and 20 replicates. As G1 wasps and flies emerged they were removed, stored and then counted.

Statistical analysis

All statistical analyses were performed in R version 4.4.0 (R Core Team, 2024). Raw data, code, fitted

model objects and markdown documents covering all steps are publicly archived on

Zenodo https://doi.org/10.5281/zenodo.13327764 and available through GitHub at

https://github.com/jcdterry/MultiGenHeatEffects_public.

Flies

To test the statistical support for parental generation (G0) temperature influencing fly fecundity in the

subsequent generation (G1), multiple models were fit and compared by AIC, with and without the

inclusion of G0 temperature. A thermal performance curve likelihood function (Equation 1) based on

the deutsch_2008 function (Deutsch et al., 2008) in the rTPC package (Padfield, 2023) was

customised to fit to discrete count data, and fit using the bbmle package (Bolker et al., 2023).

$$N \sim R_{max}\left(1 - \left(\frac{T - T_{opt}}{T_{opt} - CT_{max}}\right)^2\right) \text{for } T > T_{opt}$$

(Equation 1)

where N is the number of flies expected and T is the G1 temperature (°C). The three fitted parameters

are: the optimum temperature T_{opt}, the maximum reproduction at optimum temperature R_{max}, a curve

width parameter a. The critical thermal maximum CT_{max} fell outside the range of the experimental

trials and was fixed at 30°C based on previous studies in this species (Chen & Lewis, 2024). To

account for the overdispersion in the observations a negative binomial error function was used.

The model variants were tested where the three fitted parameters (T_{opt}, R_{max}, a) were fixed regardless

of the G0 temperature (Model 1b), with the parameters having a linear dependence on G0 temperature

(Model 1c), and with the parameters having separate values for each of the three G0 temperature

levels (Model 1d).

Wasps

To test the statistical support for G0 temperature influencing wasp reproductive output, a similar

approach was followed. To account for variation in the numbers of larvae in the standardised host

vials, the proportion of wasps of the total emerged insects (wasp count/ (wasp count + fly count)) was

used, assuming a binomial distribution. This approach assumes that the combined count was

independent of G1 temperature, which was supported (linear model of Total Count ~ G1, $F_{1,275} =$

0.373, p = 0.541) .

Unlike for fly fecundity, there is no established thermal performance model to use for parasitoid

wasps. Based on inspection of the data, a monotonic relationship between temperature and fecundity

appeared justified within the trial range and binomial generalised linear models were fit. Multiple

models were fit and compared by AIC: without inclusion of G0 (Model 2a), with G0 temperature as a

linear predictor (Model 2b), with an interaction between G0 and G1 temperature (Model 2c). To

examine if the effect of G0 was non-linear, separate models for each of the three G0 temperatures

were fit and their likelihoods combined (Model 2d).

Simulation Modelling

To explore how transgenerational effects of thermal conditions and autocorrelation in temperature

may impact the time to extinction of this host-parasitoid system under a high-emissions climate

change scenario, we built simulation models incorporating data from the experiments. Both host and

parasitoid species are assumed to have discrete and synchronised generations in a Nicholson-Bailey

framework (Hassell, 2000).

Separate simulation models were built parameterised either with the model fits with or without

including transgenerational effects in either species performance. Two modelling frameworks were

used to explore the potential impact on dynamics, one very simple using Type II functional response,

the second using a somewhat more realistic representation of spatial aggregation.

Model A – Type II Functional Response

The potential fly population in the next time step N_{t+1}^*, given the current fly population density N_t is

based on a logistic growth model:

$$N_{t+1}^* \ = \ fr(T)N_t \left(1 - \frac{N_t}{K_H}\right)$$

231   (Equation 2)

232   where potential fly population growth rate in a particular environment $r(T)$ is scaled by a failure rate

233   f, and K_H is a host carrying capacity term.

234   The number of flies lost to parasitism followed a Type II functional response:

$$N_{t+1} \ = \ N_{t+1}^* \ - \ P_t \frac{\alpha(T)N_{t+1}^*}{1 + \alpha(T)hN_{t+1}^*}$$

235   (Equation 3)

236   where h is a handling time, and $\alpha(T)$ is the temperature dependent attack rate.

237   The wasp population in the next generation was calculated similarly, with the addition of a small

238   immigration component (d) to maintain some wasp threat throughout:

$$P_{t+1} = P_t \frac{\alpha(T)N_{t+1}^*}{1 + \alpha(T)hN_{t+1}^*} + d.$$

239   (Equation 4)

240   **Model B – Proportion surviving under Spatial Aggregation**

241   An alternative model follows the overall structure of Model A, but modelled the conversion of

242   potential hosts as a proportion that survive assuming aggregated parasitoid attacks determined by the

243   parameter k, following Hassell (2000) and Meisner et al. (2014):

$$N_{t+1} \ = \ N_{t+1}^* \left[1 + \frac{\alpha(T)P_t}{k(1 + \alpha(T)hN_{t+1}^*)}\right]^{-k}.$$

244   (Equation 5)

245 Taking the wasp population as those that do not survive, the wasp population size in the next
246 generation is then:

$$P_{t+1} = N^*_{t+1} \left(1 - \left[1 + \frac{\alpha(T)P_t}{k(1 + \alpha(T)hN^*_{t+1})} \right]^{-k} \right) + d.$$

247   (Equation 6)

Experimentally Derived Temperature Dependence

The fly population growth rate $r(T)$ followed the environmental performance curve fit to the empirical
data (Equation 1). Without the inclusion of transgenerational effects, the values of the thermal
performance curve parameters were fixed to those identified with Model 1b (Table S3). When
transgenerational effects are included, linear interpolation is used between and beyond the curve
parameter values identified by the best fit model (Model 1d, Figure S1). To prevent negative R_{max}
values, the expression was lower-bounded at 0.1.

To represent temperature dependence in the attack rate $\alpha(T)$ of the wasps, the proportion models fit to
the experimental data ($\eta(T)$) were scaled by a maximum wasp attack coefficient t_p:

$$\alpha(T) = \eta(T)t_p$$

257   (Equation 7)

258 Model 2a (which did not include G0 temperature as a predictor) was used for $\eta(T)$ without the
259 inclusion of transgenerational effects. To include transgenerational effects, a model equivalent to
260 Model 2d was used which interpolated between the fitted values of the parameters at the three G0
261 with a quadratic function.

Simulation model parameter calibration
For both models, remaining parameters were selected based on a combination of experimentally
derived values, previous literature and manual tuning to generate plausible dynamics prior to the
introduction of simulated climate change (Table S3). Fly carrying capacity K_n was set sufficiently
large to overcome demographic stochasticity, whilst fly failure rate f was set on the same order of
magnitude as R_{max} to stabilise the overall growth rate and prevent overly-explosive exponential
growth. Wasp handling time h was set at 0.01 as, on average, the wasps can parasitise a maximum of
approximately 100 flies – due to egg limitation (maximum attack rate is $1/h$) (Godfray, 1994; Hassell,
2000). The wasp attack coefficient t_p is a scaling factor that defines wasp attack rate and was tuned to
generate an initial rate of parasitism of approximately 50% in line with field studies of this system
(Jeffs et al., 2021). The wasp spatial aggregation term k was set last, with simulations run at a range of

values around that in Meisner et al. (2014) to generate plausible dynamics with the populations

varying moderately around their carrying capacities.

Generating simulation temperature regime

Simulated temperature trajectories were built on a 500 generation burn-in period with mean

temperature 24°C (to settle any initial transient dynamics), followed by a linear rise of 5°C over 2000

generations (approximately 80 years). On top of this, Gaussian noise ($\sigma = 1$) was superimposed, with

an autocorrelation of either 0 or +0.8 (Figure 5). To prevent numerical errors outside the bounds of the

extrapolation, temperatures were capped at 30°C, the flies thermal maximum.

These values were selected to align with the ranges and coefficient of variation observed in fortnightly

averages from the summer months at a long-term meteorological station near the collection site of the

study species, Koombooloomba Dam (Australian Government Bureau of Meteorology, 2023) and a

high emissions warming scenario (IPCC, 2023).

Simulation trials

The simulation models were used to investigate the time to extinction of the hosts under a suite of

model assumptions: the inclusion of transgenerational effects in each of the flies and wasps and the

inclusion of positive autocorrelation in temperature between generations. Each model scenario was

run 100 times with consistent sets of random seeds to generate the temperature trajectories used across

scenarios. Extinction was defined as the first generation in which the fly population fell below one.

To test if there was an interaction between the impact of transgenerational effects (in the flies and

wasps) and autocorrelation in temperature on (logged) time to extinction, a linear mixed effects model

was fit using the *lme4* package (Bates et al., 2015). By taking the logarithm, the effect of the

explanatory variables on the proportional change in time to extinction could be modelled, thus

assuming a multiplicative rather than additive relationship between variables. The full model included

as fixed effects autocorrelation (0 or 0.8) and the inclusion of transgenerational effects for each

species (ON/OFF), along with all interaction terms. Temperature trajectory seed ID was included as a

random effect. The significance of the interaction terms was tested with a Type III ANOVA with

Satterthwaite's method via the *lmerTest* package (Kuznetsova et al., 2017).

Results

Experimental Evidence

For the fly temperature treatment experiment, the best model fit included G0 temperature as an

explanatory variable with all parameters able to vary with the three levels of G0 (Table 1, Model 1d),

indicating a significant impact of transgenerational effects of thermal conditions on fly performance

(Figure 3). As G1 temperature increased, fecundity increased for all three fly TPCs but at distinct rates.

At a G1 temperature of 19°C, the greatest predicted fecundity was seen when G0=19°C. At all

temperatures above 19°C, the greatest predicted fecundity was seen when G0=23°C. When G0=27°C,

performance was low across the temperature range, suggesting the presence of a non-adaptive

condition-transfer effect.

Table 1: **Comparison of models fitted to the fly data.** Model 1c uses a linear dependence of T_{opt}, R_{max} and a

on G0 temperature. Model 1d fits different values of T_{opt}, R_{max} and a for each of the three G0 temperatures.

Models 1e, 1f and 1g are the same as 1d but each prevent one of the parameters from varying with G0

temperature to test whether the inclusion of temperature dependence is necessary for all curve parameters.

Model	Error Distribution	Temperature dependence of parameters (T_{opt}, R_{max}, a)	N parameters	Log-Likelihood	ΔAIC
1a	Poisson	None	3	-7737.23	10700.77
1b	Negative Binomial	None	4	-2502.36	233.03
1c	Negative Binomial	$\theta = \theta_{i,19} + (T-19)\theta_{i,T}$	7	-2429.22	92.76
1d	Negative Binomial	$\theta_i = \begin{cases} If\ T = 19, \theta_{i,19} \\ If\ T = 23, \theta_{i,23} \\ If\ T = 27, \theta_{i,27} \end{cases}$	10	-2379.85	-
1e	Negative Binomial	Like 1d for R_{max} and a, and 1b for T_{opt}	8	-2427.30	90.91
1f	Negative Binomial	Like 1d for T_{opt} and a, and 1b for R_{max}	8	-3965.30	3166.91
1g	Negative Binomial	Like 1d for T_{opt} and R_{max}, and 1b for a	8	-3756.14	2748.59

315
Figure 3: Fly Thermal Performance Data and Fitted Curves. Shows the effect of temperature on the

fecundity of flies in G1. Separate panel for each G0 temperature. Blue lines represent predictions based on best

fit thermal performance curve model. Parameter estimates and 95% confidence intervals are in the

supplementary information (Table S1)).

For the wasp temperature treatment experiment, the best supported model included G0 temperature as

a three-level factor and allowed it to interact with the effect of G1 temperature (Table 2, Model 2d),

indicating a significant transgenerational impact on wasp performance (Figure 4). In comparison to

the flies, the association between G1 temperature and wasp performance was more variable depending

on G0 temperature. When G0=19°C, as G1 temperature increased, the predicted proportion of wasps

decreased. In contrast, when G0=27°C, as G1 temperature increased, the predicted proportion of

wasps increased. This could be interpreted as an adaptive transgenerational effect.

Table 2: Description and comparison of models fitted to the wasp thermal performance data. Model 2d,

including distinct responses to each G0 temperature, was strongly supported. G0 = Generation 0. G1 =

Generation 1.

Model	G0	Interaction	Explanatory variable(s)	n	Likelihood	ΔAIC
2a	Absent	Absent	G1	2	-4913.71	1364.77
2b	Linear	Absent	G1+G0	3	-4552.96	645.28
2c	Linear	Present	G1*G0	4	-4397.26	335.87
2d	Factor	Present	G1*G0	6	-4227.32	-

Figure 4: Wasp Thermal Performance Curves. Shows the effect of temperature on the proportion of wasps
(wasp count/total count, where total count includes the flies and G1 wasps). Black points represent raw data.
Blue lines represent predictions based on best fit model. Parameter estimates and 95% confidence intervals are
in the supplementary information (Table S2).

Simulation model results

Both autocorrelation and the full inclusion of transgenerational effects markedly hastened the
extinction of the fly host in the simulations under both models (Figure 5, Table S4).
Converting from generations to years, accounting for transgenerational effects hastened the
time-to-extinction by approximately 30 years.

In the simpler Model A, autocorrelation and each of the transgenerational effects (i.e. the fly
and the wasp) had a significant interactive effect (three-way interaction term, $F_{1,693} = 10.86$, p
= 0.001, Table S5). Taken separately, transgenerational effects on the fly had a markedly
greater impact. Transgenerational effects on the wasp alone had minimal impact, but in
combination with transgenerational effects on the fly and autocorrelation, further hastened
extinction (Figure 5, top left). However, under Model B there were no detectable interactive
effects between either of the transgenerational effects and autocorrelation on the log-time-to-
extinction ($F_{1,693} = 0.0083$, p = 0.927, Table S5).

Figure 5: Density histograms of the observed distribution of time to extinction (in generations), over 100 simulations the host-parasitoid system under different scenarios and models. There is a near perfect overlap between scenarios with transgenerational effect modelled for 'wasp-only' and 'none' (blue and green, combining to dark turquoise), as well as 'both' and 'fly-only' (yellow and, red combining to orange), with the exception of the top left panel. Geometric means are given in Table S4. Note the logarithmic x-scale.

Discussion

These results demonstrate how transgenerational effects can have substantial impacts on the consequences of climate change through a variety of pathways. Despite hope and interest in the potentially resilience-enhancing nature of adaptive responses, our results highlight the negative aspects of transgenerational effects. Modelling is crucial to contextualise the various processes (Terblanche & Hoffmann, 2020), but the dependence on model form emphasises the need for caution and further theoretical developments.

Two key caveats need to be reiterated when interpreting these results: firstly, our focus was on identifying potential impacts of cross-generational transfer of population dynamics rather than isolating physiological mechanisms. As such, although the nature of these responses suggests contributions from both condition-transfer effects and potential adaptive transgenerational plasticity, these cannot be cleanly differentiated within our design. This is a fundamental challenge – while the observed effects are 'transgenerational' from the perspective of population models tracking adults, it

is plausible that the most important underlying physiological processes are can be considered within-
generational (Donelson et al., 2018) from other perspectives. Secondly, these models are not intended
to be specific representations of particular real-world scenarios or locations. The experiments focus on
just two species from a complex community (Chen & Lewis, 2023, 2024; Jeffs et al., 2021) and do
not account for many features, from microclimates (Pincebourde & Woods, 2020), the impact of
seasonality (Tougeron et al., 2020), to the potential for evolution (Rudman et al., 2022).

Experimental observations

We found clear experimental evidence that the parental thermal environment had a significant and
substantial impact on subsequent reproductive output in both the hosts and parasitoids. When the
parental fly generation was maintained in a thermally stressful environment (G0=27°C), subsequent
reproductive output was lower across the temperature range (Figure 3), in line with previous results
using *D. melanogaster* that identified an elevated cost to living in an extreme thermal environment,
with high and constant temperatures reducing the thermal performance of the offspring generation
(Cavieres et al., 2019). The most direct explanation is that G0 flies that developed in stressful
temperatures were not physiologically able to take advantage of more clement conditions in the G1
generation. On average, across all offspring generation temperatures, the best fly performance was
when the parental generation was maintained at an intermediate temperature (G0=23°C, Figure 3).
This indicates a condition-transfer effect and aligns with the optimum parental temperature hypothesis
that parents kept at intermediate temperatures produce fitter offspring compared to parents at
temperature extremes (Cohet & David, 1978; Gilchrist & Huey, 2001). Although it appears that the
thermal optimum was beyond the temperature range of the experiment, it seems the majority of the
impact was from changes in performance, with relatively little change in thermal optimum, in line
with results from other species in this community (MacLean et al., 2019). Although our experimental
design cannot distinguish developmental success rates from the number of eggs laid, it does reiterate
the importance of accounting for potential carry-over effects when assessing thermal tolerances
(Jenkins & Hoffmann, 1994; Schiffer et al., 2013; Sgrò and & Hoffmann, 1998).

There was also some indication of adaptive cross-generational effects where offspring facing the same
environmental conditions as their parents have a higher fitness than offspring facing different
environmental conditions (Gilchrist & Huey, 2001) in anticipation of what they are likely to
experience (Cavieres et al., 2020; Tariel et al., 2020). For instance, when the parental generation was
maintained in a thermally stressful environment (G0=27°C), the proportional decrease in fecundity
was lower when the offspring generation was also in a thermally stressful environment (Figure 3).
This could indicate the possibility of some anticipatory effect, although it is overwhelmed by the non-
adaptive transfer of poor condition. Furthermore, the greatest thermal performance at 19°C was when
the parental generation was also maintained at 19°C (Figure 3). This highlights how while it is
important to differentiate between anticipatory and condition-transfer effects when understanding the

evolutionary causes and consequences of transgenerational effects (Sánchez Tójar et al., 2020), both
can operate simultaneously.

The assay we used (with a discrete window of egg laying time) is especially sensitive to the rate of
egg laying but does not capture longevity of the G0 flies. While fly fecundity was lower when the
parental generation was maintained in a thermally stressful environment, demographic trade-offs may
have meant other traits, such as thermal tolerance (Bozinovic et al., 2011) or longevity (Ismaeil et al.,
2013), were not. In a previous study, *D. melanogaster* subjected to a variable and stressful thermal
environment exhibited reduced fecundity, yet showed increased thermal tolerance compared to flies
kept at a constant temperature (Cavieres et al., 2020).

The wasp thermal performance curves show clear evidence for improvements in thermal performance
where there was a close temperature match between the generations. When the parental generation
was maintained at a low temperature (G0=19°C), the greatest wasp proportion was when the G1
temperature was also low (Figure 4). In comparison, when the parental generation was maintained at a
high temperature (G0=27°C), the greatest wasp proportion was when the temperature was also high.
This acclimation would be expected to reduce their vulnerability to climate change despite many of
their life-history traits, such as a high trophic position and typical trophic specialisation, increasing
their susceptibility (Jeffs & Lewis, 2013). Furthermore, this result has potential implications for the
efficiency and maintenance of biological control programmes utilising parasitoids under climate
change, with parasitoids the most widely-used agents for the biocontrol of insect pests (Hochberg &
Ives, 2000; Ismaeil et al., 2013).

Time to extinction: Role of transgenerational effects and autocorrelation in temperature

The simulation models predict that including transgenerational effects of thermal conditions
significantly hastens the time to extinction of this host-parasitoid system. The substantial negative
effect of including the transgenerational effects in the simulation models highlights the predominance
of negative condition-transfer over any potential adaptive transgenerational plasticity through
anticipatory effects. Autocorrelation, even without transgenerational effects, magnifies the threat
posed by environmental stochasticity near a limit (Lande et al., 2003). By prolonging periods of
adverse conditions and reducing the presence of temporal refugia, positive temporal autocorrelation
can amplify environmental stress and exacerbate the risk of extinction (Postuma et al., 2020).

Without autocorrelation, there *should* be no special dynamic consequences of transgenerational effects.
The observed strong impact seen in the scenarios without autocorrelation show that effectively
averaging across assays of the thermal performance of flies raised in more suitable environment is
inflating the estimated performance at higher temperatures. The additional effects from the
combination of autocorrelation and transgenerational effects are the direct dynamic impact from the
transgenerational effects. The surprising absence of any detectable emergent interaction within Model

B highlights the context dependency of whether experimentally identified effects will have

population-level consequences, or whether emergent effects will be absorbed by other buffering

mechanisms (such as the modelled spatial clumping through the negative-binomial term). By contrast,

in the simpler Model A there was a clearly identifiable additional interaction effect, although this

dynamic component was considerably smaller than the independent impacts of both autocorrelation

and the direct effects of transgenerational effects. This effect was negative, again highlighting how

any benefits of acclimation were overwhelmed by condition-transfer effects. The distinct responses

under the two models, both of which are relatively standard representations of host-parasitoid

interactions, highlights the caution needed in scaling from experimental results to real populations.

The modelling transgenerational effects observed in the wasps had essentially no impact on the time

of extinction on their own, but under Model A did have an impact in combination with the

transgenerational effects observed in the flies. The double-effect of sequential hot generations of

reduced host and improved wasp performance brought forward the point at which the host population

was pushed below the extinction threshold.

Linking experiments, physiological mechanisms and dynamic consequences

Precisely distinguishing the multiple possible contributing factors that can underly acclimatory and

transgenerational effects is a significant challenge (Terblanche & Hoffmann, 2020), further

confounded by difficulties precisely delimitating the boundaries between generations (Burggren, 2015;

Lee et al., 2020) and the potential for influences to span multiple generations (Tariel et al., 2020). For

example, environmental impacts on primordial germ cells that appear as examples of

transgenerational plasticity could be due to within-generation plasticity in offspring during early

stages of development (Donelson et al., 2018).

Our broad-scale approach allows us to deploy the high replication necessary to confidently identify

effects but cannot differentiate physiological mechanisms, and the experimental requirements to fully

partition all potential different processes can quicky become unfeasible at scale. While further

experiments could address some of the gaps, for example transferring a controlled number of eggs to

differentiate the impact of adult condition influencing egg laying rate from effects related to offspring

development at different temperatures, it is hard to assess how much difference such extra information

would bring. As such, our results should be taken as experimentally-informed potential scenarios,

rather than precise forecasts. In scenarios with stochastic environments and multiple interacting

species, model predicted outcomes are readily qualitatively changed under different parameter

combinations (Terry et al., 2022). Closing the complexity gap and determining the appropriate level of

abstraction to tackle these processes at scale will be a key challenge for future research. Our

experiments were designed to align with models of host-parasitoid dynamics that conventionally

operate on an adult-adult basis (Nicholson & Bailey, 1935), but demarking the generational boundary

at a different stage (e.g. eggs – eggs) could well give different results. Insect life stages are known to

be impacted differently by temperature (Kingsolver et al., 2011; Pawar et al., 2024) and so our
generational scale thermal treatment could be missing important consequences by not accounting for
demographic structuring (Amarasekare & Coutinho, 2013).

Conclusion

479 Our results highlight the substantial, but context-dependent, impact multi-trophic transgenerational
480 effects can have on population dynamics under climate change. These findings underscore the need to
481 quantify and contextualise transgenerational effects when understanding species' responses to climate
482 change and show the need to both move beyond demonstrations that such effects exist to improving
483 our theoretical understanding of the scenarios in which they are influential.

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