Title: Are super-predators also super-scary? A systematic review and meta-analysis of animal

responses to human interactions

Authors:

Shawn Dsouza[1*]

Email: shawn27.dsouza@pm.me

Kartik Shanker[1]

Email: kshanker@iisc.ac.in

Maria Thaker[1]

Email: mariathaker@iisc.ac.in

[1] Center for Ecological Sciences, Indian Institute of Science, Bengaluru, India.

*Corresponding Author:

Shawn Dsouza

Mob No.: +91 9821158693

Email: shawn27.dsouza@pm.me

Address: Center for Ecological Sciences, Indian Institute of Science, CV Raman Ave, Bangalore

– 560097

Abstract

Human induced rapid environmental change has been recognized as a global threat to natural

systems and the organisms that inhabit them. Human hunters and fishers interact with animals in

natural spaces as predators, and are more effective in capturing prey and reducing populations

than natural predators overall. On the other hand, seemingly benign interactions with humans,

such as tourism, may also be perceived as threatening and have negative impacts on animal

populations. Here, we provide a synthesis of the current knowledge of the combined effects of

lethal and non-lethal human interactions on the behavioural responses of animals in natural

spaces. We conducted a systematic review and meta-analysis of literature published over the past

three decades, and built a dataset to determine the relative effect of lethal and non-lethal human

interactions on foraging, vigilance, and movement behaviors of wild animals. Lethal human

interactions included hunting, fishing, and retaliatory killing in response to conflict, and non-

lethal human interactions were distinguished as active (snorkeling, scuba diving, pedestrians,

hiking, and tourists) or passive (roads, human settlements, sonar, and boat traffic). We also

considered how the trophic level of the species and body size could influence the effect of human

interactions on animal behavior. Our findings show that lethal human interactions had a

significant effect on the behaviors of animals, causing animals that are actively hunted by humans

to increase vigilance, reduce foraging, and alter movements. Both active and passive non-lethal

interactions had weaker and non-significant effects on altering these behaviours of animals.

Overall, our meta-analysis shows that the impact of non-lethal human activities in eliciting fear-

driven behavioural changes in prey does not seem to have broad empirical support. It also

suggests that the intensity of effect of the "human super-predator" may depend on the trophic

level of affected species, or the history of the human-animal interactions in that landscape, In

order to understand the nuances of these effects, systematic studies across more species and

geographic regions are needed.

Keywords: HIREC, risk − disturbance hypothesis, non − consumptive effects, predator − prey

interactions, fear

Contents

I. Introduction

Humans are a pervasive species that interact with and change their environment in a multitude of ways. Over the past two centuries, human development has led to the alteration of global ecosystems at a rate that has not been witnessed in the history of the planet. From climate change to the rampant extraction of biological resources, the effects of human-induced rapid environmental change (HIREC) are large in scale and magnitude (Estes et al. 2011; Sih 2013). Given their unique ecology, humans not only occupy every eco-region of the world, but have multiple functional roles that can shift trophic ecology in intensive and extensive ways. Animals therefore have no choice but to respond to human activities in the natural world.

Throughout most of their evolutionary history, humans have been hunter–gatherers (Treves and Naughton-Treves 1999). As predators, however, humans are extraordinarily deadly, as they out-compete most natural predators in the rate and efficacy of killing their prey (Darimont et al. 2009; 2015). Through social organisation and tool use, humans can target multiple trophic levels at high efficiency (Treves and Naughton-Treves 1999). Thus, humans have been referred to as generalist super-predators (Clinchy et al. 2016), taking a disproportionately large number of animals from ecosystems. In fact, the rate of mortality of prey species from human predators, such as hunters and fishers, are much greater in magnitude than the sum of kill by all natural predators in both terrestrial and marine ecosystems (Darimont et al. 2015). Thus, humans create truly risky conditions for both wild prey and predators in natural systems (Oriol-Cotterill et al. 2015).

In addition to hunting and fishing, humans also interact with wild animals through non–lethal activities (Boyle and Samson 1985). Non-lethal interactions of humans with animals can be

further distinguished as active or passive. Active non-lethal interactions include tourism, walking

in parks, hiking in nature reserves, mountaineering and off-road bicycling, and have been shown

to negatively influence animals even-though the humans in their habitats are not actively killing

them (e.g. Loehr et al. 2005; Newey 2007; Brown et al. 2012; Montero-Quintana et al. 2020).

Passive non-lethal interactions are even more pervasive, as they include alterations to

environments. Roads and settlements, in particular, have the greatest potential to influence the

behaviour of animals (e.g. Mehlhoop et al. 2002; Cappa et al. 2017, etc). Given the number of

bird collisions on glass (Loss, Will, and Marra 2015) and high occurrence of roadkill globally

(Benson, Mahoney, and Patterson 2015), animals are facing mortality from human interactions

that they may not perceive as risky.

Other non-lethal activities such as air and sea travel may also affect animal behavior (e.g. Van der

Kolk et al. 2024; see Harris et al. 2019 for review). Although compelling evidence suggests that

non-lethal human activities can be extremely strong in changing animal behaviours (Boyle and

Samson 1985; selected examples: (Davis et al. 1997; Christopher M. Papouchis et al. 2001;

Nevin and Gilbert 2005; Westekemper et al. 2018; Ladle et al. 2019), many other studies show

that animals habituate quickly to human encounters (see Smith et al. 2024). Thus, it remains

unclear whether the non-lethal interactions of humans on animals are as strong as the lethal

interactions in eliciting fear and changing behaviours.

Given that humans have both lethal and non-lethal interactions with animals in the wild, the

individual and combined effects of these on animal behaviors can potentially be very high. In a

recent narrative review, Smith et al. (2024) outlined a framework describing the pathways by

which lethal and non-lethal human interactions can change phenotypic (behavioral and

physiological) responses of animals, with consequence for population growth and abundance.

They find that the perceived risk of humans can induce phenotypic changes in wild animals, but

as expected, the evidence is mixed (Smith et al 2024). What is still missing is a quantitative

assessment of the relative magnitude of effect for different types of human activities, which

would illustrate when and why animals respond to human-induced risk.

In this paper, we use a meta-analysis to synthesize the current evidence for human-induced

behavioural changes in wild animals. We separated studies based on whether the interactions

were lethal (hunting, fishing), active and nonlethal (ecotourism, hiking) and passive and

nonlethal (roads and human settlements). One of the clearest and fastest responses that prey use

to minimize risk is by shifting behaviour. Decades of research has shown that natural predators

change their foraging and vigilance patterns, as well as the movement and habitat use patterns of

prey (Brown and Kotler 2004; Sih 1984). Thus, we focused on changes in foraging behaviours

(bite rates and foraging time), vigilance (vigilance time and proportion) and movement

(displacement, home range size and rate) of animals due to human interactions. We predict that in

areas where humans hunt, animals would strongly invest in anti-predator behavioural strategies,

spending more time being vigilant, less time in foraging, and increasing their movement rate or

displacement (Figure 1). We also expect animals to perceive non-lethal activities of humans as

threatening but to a lesser degree; thus, we expect a lower intensity of anti predator responses to

active and passive non-lethal interactions of humans (Figure 1). We also tested the potential

influence of body size on the magnitude of behavioural effects based on the hypothesis that larger

animals would react to a greater extent to human disturbance as they are disproportionately

targeted by the lethal actions of humans.

From a systematic review of published studies, we derived effect sizes that describe the

magnitude of behavioural change across these categories of human interactions. We synthesise

our meta-analysis to determine whether non-lethal effects of humans that are active or passive

influence wild animal behaviours to the same extent as lethal human effects. Finally, we highlight

where gaps in the literature have yielded skewed conclusions about human impacts on the

behaviour of animals in natural spaces and suggestavenues for future research in this field.

II. Methods

To determine the strength of behavioural change (i.e. effect sizes) caused by human activities, we

conducted a systematic literature survey that included studies where foraging, vigilance or

movement behaviours were measured in animals in response to lethal human interactions

(hunting, including poaching and trapping, and fishing), nonlethal but active interactions

(walking in parks, hiking in nature reserves, off-road bicycling and animal watching), and

nonlethal but passive interactions (roads and human settlements).

1. Search and literature database

We conducted a scoping search using the web of science database and their advanced search

function. Our initial search term was "(risk OR fear) AND (human OR anthropogenic) AND

(behavio*) AND ALL=(predate*)". The search terms were applied across all fields. The initial

search returned 2319 results on 12th January 2023. We downloaded the data and conducted a

literature mapping analysis to identify additional search terms and updated our search string to

"(human OR anthropo*) AND (risk OR fear OR NCE OR trait mediated effect OR

nonconsumptive* effect) AND (behavio*) AND (predat*)". Using our updated search string, we

conducted searches on the SCOPUS and WOS database as well as grey literature databases such

as OpenThesis on 14th February 2023. We then used Rayyan, a literature management and

screening software to deduplicate our database (Ouzzani et al. 2016; Foo et al. 2021). After

deduplication we were left with 7562 abstracts. This review was not registered.

2. Initial screening

We piloted our initial screening protocol on 100 randomly selected papers using Rayyan

(Ouzzani et al., 2016). We decided inclusion and exclusion criteria by evaluating titles, abstracts,

and key words to determine the population studied, the exposure or intervention to the

population, the control if any and the outcome (see Appendix A.1 for detailed inclusion criteria).

The initial screening was conducted by three independent researchers. The three researchers

discussed conflicting decisions on inclusion or exclusion of abstracts until a consensus was

reached. We applied the finalised screening protocol to all abstracts in our database and were left

with 436 papers to further evaluate.

3. Full text screening and data collection

After conducting initial screening, the screened database was downloaded from Rayyan (Ouzzani

et al., 2016). We then cleaned the database and extracted PICO data in a workable format using

R. Papers chosen for inclusion during our initial screening were full-text screened. We

determined our protocol based on variables in our initial screening, the methods used, study

design and reporting (Appendix A.2) (Foo et al., 2021). We collected data on the study species

identity and trophic level, location of the study, treatment, controls, responses, mean effects,

variance measures and sample size. We piloted the protocol on 39 papers. Where results were not

clearly reported we either used published data to calculate mean effects and variance measures or

used a plot digitiser to extract data from figures. Studies in which data extraction was not

possible were excluded from our meta-analysis. We included 71 studies at this stage.

4. Backward and forward search

We identified key studies and reviews during our screening for backward and forward search.

Backward search identified papers cited by these studies and reviews, and forward search

identified papers that cite these key papers. We added the results to our existing database and

deduplicated it to create the final paper database for the review (Foo et al., 2021). After screening

the additional abstracts, we added an additional 40 studies to our final selection. Of these, 14

were added after full-text screening. We have included a PRISMA (EcoEvo) checklist and

flowchart in Appendix A and E (Page et al. 2021; O'Dea et al. 2021).

5. Types of studies

There were several types of study approaches that measured animal responses to human

activities. The majority of studies were observational or natural experiments. This included

spatial or temporal contrast. Spatial contrasts included designs in which behaviors were compared

across protected and unprotected areas for lethal interactions (n = 8), areas with and without

tourist for active interactions (n = 6) and areas close and far away from roads or settlements for

passive interactions (n = 4). Studies also employed auditory stimuli (n = 5) for example road

noise or human speech in areas with hunting pressure or high traffic to test for the effect of lethal

interactions and passive interactions respectively. Other experimental studies employed direct

pedestrian disturbance to simulate active interactions (n = 3) and lethal interactions in areas with

hunting (n = 3). Four studies used temporal contrast in human activities such as hunting closures

and tourist seasons to determine the effects of active (n = 1) and lethal (n = 3) interactions.

Finally, ten studies did not implement or utilize a contrast for their comparison and instead

determined correlations between human activities for active interactions and distance from road

for passive interaction to study their effects on behavior.

We have focused on studies that deal with foraging, vigilance and movement behaviors. Studies

dealing with foraging and vigilance behaviors used focal individual sampling and scan sampling.

The measured behaviors were time spent vigilant, time spent foraging, frequency of vigilance

behavior in a group, rate of individual vigilance behavior, and foraging (bite) rates. We included

movement studies that measured movement rates (speed), displacement (distance traveled), and

home range size using GPS tags, collars and transmitters. Home range size was included as it is

function of displacement and speed in a given sampling period.

6. Meta-analysis

We used studies that reported all necessary metrics for comparison including treatment means,

measures of uncertainty, and sample sizes for further analysis (Mikolajewicz and Komarova

2019, Appendix A.2). We did not include papers that only reported test statistics in our meta-

analysis. We gathered secondary data including trophic level, dietary guild and body mass, on

the species from each study from mammalbase (Lintulaakso 2021), animaltraits database

(Herberstein et al. 2022), handbook of the birds of the world (Del Hoyo et al. 1992) and fishbase

(Froese and Pauly 2010).

We divided the studies into categories based on their study designs as experimental or

observational. We converted proportional and percentage data (means and measures of

uncertainty) to absolute values (number of individuals or time in seconds), and we converted all

measures of uncertainty to variance before analysis. We calculated standardised mean difference

for studies that reported an explicit treatment and control (negative or positive), as

$d = \dfrac{x_{control} - x_{treatment}}{s_{pooled}}$, where x_i is the mean of treatment and control outcomes respectively and

$s_{pooled} = \sqrt{\dfrac{s^2_{control}(n_{control} - 1) + s^2_{treatmen}(n_{treatment} - 1)}{n_{treatment} + n_{control} - 2}}$, here s^2_i is the sample variance and n_i is the

sample size. All the correlational studies we included reported unstandarized regression

coefficients (b), we thus calculated the standardised mean differences from b as $d = \dfrac{b}{s_{pooled}}$, where

$s_{pooled} = \sqrt{\dfrac{s^2_y(N - 1) - b^2\left(\dfrac{n_{treatment}\, n_{control}}{n_{treatmen} + n_{control}}\right)}{N - 2}}$ here N is the total sample size (Lipsey and Wilson

2001). The model coefficients of studies which used a treatment:control contrast were multiplied

by -1 to maintain consistency across studies. We assigned each study a unique identifier (study

ID); when we extracted multiple data points from studies, we assigned a unique identifier to each

datapoint (data ID).

We first fit an intercept only multilevel meta-analytic (MLMA) model with species, study and

data id as random effects to determine the degree of heterogeneity across studies after accounting

for known sources of variation as follows:

$$y_{ijk} = \mu + u_{ijk} + s_i + d_j + \eta_{ij}$$

where, y_{ij} is the effect from j^{th} data point of the ith study, μ is the overall mean across studies, u_{ij}

of the ith estimate withing the jth study, s_i is the random effect due to the ith study, d_j is the random

effect due to the jth data point and η_{ij} is the error of the jth estimate from the ith study. The random

effect framework assumes that each study has its own true effect size which is derived from a

population of true effect sizes. Since this meta-analysis involves a comparison across taxa in

different habitats and geographies, the random effect framework is appropriate (Nakagawa et al.

2022; Noble et al. 2022). We also fit a phylogenetic model with relatedness as random effect.

$$y_{ijk} = \mu + u_{ij} + s_i + d_j + p_k + \eta_{ij}$$

where, p_k is random effect due to the kth species. Phylogeny did not capture significant variation

across all outcomes (Chamberlain et al. 2012; Cinar, et al. 2022). We thus used a model with only

study identity, and data identity as random effects. We used the Q statistic to determine left over

variation. We used both I^2 and H^2 to determine heterogeneity across studies (Harrer et al. 2021).

We used the final MLMA to determine whether there was an overall significant effect across

studies. Finally, we used the MLMA to perform a multiple regression to determine the effect of

type of human interactions and animal size on the observed effects.

7. Sensitivity analysis

Our analysis draws from a global cross taxonomic dataset and, thus, there are multiple possible

sources of non-independence. Bias in terms of the taxa studies and geographic location of study

meant that certain populations were represented multiple times in our data set. Phylogenetic

relatedness may also lead to non-independence across studies (Noble et al. 2017). We account for

potential non-independence by implementing the MLMA with a random effect. We evaluated

publication bias visually in our dataset using a funnel plot and used regression test for funnel plot

asymmetry (Nakagawa et al. 2022, Appendix C). We also used a multilevel meta-regression on

sample variance with study identity and data identity as random effect to determine publication

bias. As significant results have a tendency to be published first, we plotted the effect size against

the year of publication to determine any time-lag effects. To determine the robustness of our

analysis, we used a leave-one-out method where we dropped each study from our data set in

sequence and ran the MLMA to determine if any single study had a disproportionate effect on the

summary effect size (Nakagawa et al. 2022; Noble et al. 2017) (Appendix D). All data

preparation and analysis were conducted in R – 4.0.1 (R Core Team 2024). Data and code are

available at https://www.github.com/cheesesnakes/superpredator-or-superscary.

III. Results

After screening, we included 85 studies in our systematic review, the majority of which were

published in the past two decades (Appendix A5). Of these, 44 studies reported enough

information to be included in the meta-analysis, covering 38 species and families. These studies

were widely distributed across geographies and habitats ranging from 78ºN to 43ºS and 169ºE to

123ºW (Figure 2). Across outcomes, 24 studies reported measures of foraging behaviours, which

included bite rate and time spent foraging, 29 studies focused on vigilance behaviours, which

included time spent vigilant and vigilance rate, and 9 studies dealt with movement behaviour,

which involved home range size, displacement and rate of movement (Figure 3). The majority of

these papers (n = 36) studied primary consumers, such as *Cervus elaphus* and *Capreolus*

capreolus (Figure 4, Appendix A.6). Only 7 studies included secondary and tertiary consumers

(Figure 4), thus statistical comparisons across trophic levels and functional guilds were omitted.

The body size of the animal did not significantly influence the effect of human activities on

foraging (β = - 4.139 x 10^{-6} ± 4.352 x 10^{-5}, t_{34} = -0.09, p = 0.924), movement (β = 0.006 ± 0.006,

T_5 = 0.948, p = 0.288), or vigilance behaviour (β = -4.139 x 10^{-6} ± 4.347 x 10^{-5}, t_{46} = -0.0517, p =

0.961). Overall, animals responded similarly to human disturbance regardless of their size.

1. Changes in behaviours across types of human interactions

The overall effect of human interactions across studies on foraging behaviours was significantly

negative (SMD = -0.411, 95% CI = -0.893 – 0.071, T_{22} = -1.73, p = 0.092) and on vigilance was

significantly positive (SMD = 1.31, 95% CI = 0.376 – 2.237, T_{27} = 2.83 p = 0.007). However,

across studies, the effect of human interactions on movement behaviour across studies was not

significant (SMD = 0.012, 95% CI = -0.694– 0.718, T_6 = -0.04, p = 0.971, Appendix C.2 – C.4,

Figure 4). Upon fitting the MLMA model, there was still significant heterogeneity among studies

that was not accounted for by random effects (Appendix C).

Foraging behaviour of animals was negatively affected by lethal human interactions (SMD = -

1.65, 95% CI = -3.56 - 0.27), but not active non-lethal interactions (SMD = 0.071, 95% CI = -

0.38 – 0.53) or passive non-lethal interactions (SMD = -0.688, 95% CI = -1.83 – 0.46). Five data

points showed a significant negative effect of active interactions on foraging behaviour and 6

data points showed a significant positive effect. Similarly, 5 data points showed a significant

negative effect of passive non-lethal interactions on foraging and 2 data points showed a

significant positive effect of passive interactions on foraging.

Across studies, lethal interactions had a significant positive effect on vigilance behaviours of

study animals (SMD = 3.192, 95% CI = 0.85 – 5.53) but active non-lethal interactions did not

have a significant effect (SMD = 0.57, 95% CI = -0.46 – 1.61) and nor did passive non-lethal

interactions (SMD = 0.44, 95% CI= -1.45 – 2.34, Figure 4). There was a significant positive

effect of active interactions on vigilance in 13 data points and significant positive effect of

passive interactions in 5 data points (Figure 4). There was no significant effect of human

interactions on movement.

Overall, when we consider all studies in this analysis, foraging behaviours were significantly

more suppressed by lethal human activities than by active non-lethal interactions (β = -1.141 $\pm$

0.529, t_{34} = -2.155, p = 0.038, figure 5). The magnitude of behavioural effects on vigilance

behaviour due to lethal human interactions was significantly higher that of active non-lethal

interactions (β = 2.359 $\pm$ 0.983, t_{46} = 2.398, p = 0.02, figure 5). There was no significant

difference in the magnitude of interactions between active and passive non – lethal human

interactions. There were no significant differences in the effect of human interactions on

movement behaviours across the three types of interactions (Appendix D, figure 5).

IV. Discussion

Humans occupy multiple niches across almost every ecosystem. As predators, humans are more deadly than most other predators in both marine and terrestrial systems because they can target multiple trophic levels simultaneously and can extract animals at a faster rate than natural predators (Darimont et al. 2015). Humans also occupy natural spaces for seemingly benign activities such as resource extractions, tourism, hiking, and so on. Thus, humans can be predators in the classic sense, but their non-lethal activities, such as their presence in the environment or modification of the environment, can also be perceived as threatening. The mere presence of humans has been shown to alter animal behaviour, a phenomenon considered under the risk disturbance hypothesis (Frid and Dill, 2002). Understanding how animals perceive and respond to humans in their environments is vital for successful management of animal populations as well as the management of ecosystems they inhabit. The review and meta-analysis presented here covers a wide range of studies both geographically and taxonomically over the past three decades (Figure 2, Appendix A).We find that the magnitude of effects on movement, foraging and vigilance behaviours of wild animals depends significantly on the type of human interaction with lethal human-wildlife interactions creating the greatest and only overall significant effects on behaviours compared to both active and passive non-lethal interactions (Figure 5).

1. Lethal interactions: hunting, fishing, etc.

Humans have hunted other for a majority of their evolutionary history (Treves and Naughton-Treves 1999). In this context, hunting includes fishing, and trapping that result in removal of prey individuals through direct human intervention. Darimont et al.'s (2015) analysis revealed that humans have killed a larger biomass of prey species than most other predators in their respective

ecosystems. The efficiency and scale at which humans can kill prey adds to their perceived

fierceness, a predator trait (Wirsing et al. 2021). Thus, prey should respond to human hunters

very strongly (Clinchy et al. 2016; Crawford et al. 2022). Clinchy et al. (2016) showed that in

areas where badgers are hunted, badgers respond to playbacks of human voices by significantly

reducing their foraging activity. Our meta-analysis showed that, as expected, lethal interactions

had a significant and adverse effect on both foraging and vigilance behavior of animals across

taxa and geographical locations. In line with the foraging – vigilance trade off (Brown and Kotler

2004), our metanalysis shows that animals significantly reduced foraging and significantly

increased vigilance in response to human hunters. We find an overall larger magnitude of effect

on vigilance rather than foraging (Figure 5), but this is most likely because fewer studies have

reported the effects of lethal interactions on foraging behavior (Figure 3).

The behavioral response to lethal interactions was largest for large game species such as deer

(*Dama dama*), elk (*Cervus* elaphus) and moose (*Alces alces*), whereas the response of terrestrial

predators, such as puma (*Puma concolor*), hyena (*Crocuta crocuta*) and badgers (*Meles meles*),

was limited (Suraci et al. 2019; Clinchy et al. 2016; Pangle and Holekamp 2010). Across prey

species, the type of response was varied. Pecorella et al. (2016) showed that fallow deer (*Dama*

dama) were more vigilant and foraged less in areas where culling was allowed than where culling

was not allowed in Maremma Regional Park (Italy). Large ungulates such as Sable antelope

(*Hippotragus niger*) and greater Kudu (*Tragelaphus stepsiceros*) also showed large increases in

vigilance in the areas around Hwange National Park (Crosmary et al. 2012). Fallow deer showed

a proportional decrease in foraging as vigilance increased (Pecorella et al. 2016), however, this

trade-off is not always proportional. Red deer (*Cervus elaphus*) in the Scottish Highlands showed

a disproportionate increase in vigilance behaviors relative to the decrease in foraging behavior in

response to lethal interactions (Jayakody et al. 2008). Conversely, Moose (*Alces alces*) showed a

larger decrease in foraging behavior than increase in vigilance (Bhardwaj et al. 2022). Notably,

roe deer (*Capreolus capreolus*) showed large changes in movement behavior as opposed to

foraging and vigilance (Benhaiem et al. 2008; Picardi et al. 2019).

2. Active non-lethal disturbance

According to the risk disturbance hypothesis, animals are expected to perceive non-lethal human

disturbance similarly to the threat of predation (Frid and Dill 2002). For instance, bird species

have been found to reduce foraging and increase vigilance in response to pedestrians in urban

parks of Madrid (Fernández-Juricic et al. 2001). These behavioural changes are akin to how birds

would respond to a natural predator, suggesting that human presence is perceived as a threat.

Thus, animals may respond to human disturbances in the same way that they do to mitigate

predation risk from natural predators. Multiple studies have sought to test the risk disturbance

hypothesis across contexts. Price (2008), in a review, showed that human disturbance in the form

of pedestrians alters the flight, foraging and vigilance behaviours of multiple bird species.

Stansell et al. (2022) showed that dark-eyed juncos (*Junco hyemalis*) are more shy in urban areas

than in their wild habitat.

However, the relative magnitude of perceived risk from lethal humans, and non-lethal humans

may greatly differ (Mols et al. 2022). Our analysis shows variation across studies in terms of the

magnitude and direction of observed effects for both types of non-lethal interaction with humans,

but the effects of active interactions were small for most studies. In addition, lethal interactions

with humans elicited the greatest magnitude of effects across outcomes. We hypothesize that this

could be attributed to the divergent ways in which animals perceive benign humans in their

environment (Figure 1, Smith et al. 2021).

In some cases, the presence of humans may benefit some species by altering interactions between

them; female grizzly bears with cubs have been observed using areas with recreational human

activity to avoid depredation by aggressive males (Ladle et al. 2019). In particular, they may

influence the interactions between predators and prey (Moller 2008; Shannon et al. 2014;

Proudman et al. 2021). For example, humans may deter predators by their mere presence and

create safe havens for prey who are ambivalent to human presence to forage. This is known as the

human shield hypothesis (sensu Berger 2007). However, lack of data across trophic levels makes

it difficult to test this hypothesis (Reimers, Lund, and Ergon 2011; Ward et al. 1997; Uchida and

Blumstein 2021).

In addition, a large number of studies in our analysis showed no significant effect of active non-

lethal interactions on foraging (n = 10) and vigilance (n = 9) behavior, i.e., the effect sizes were

close to 0. We hypothesise that this subdued response to humans may be attributed to habituation

as a function of frequency of interaction. Many studies have pointed to substantial habituation to

tourist interactions, such as with mammals in the African Savannah (Knight 2009), killer whale

and Hector's dolphin interactions with whale watchers (Bejder et al. 1999, Williams et al. 2002)

and penguin interactions with researchers (Shelton et al. 2004). More specifically, Uchida and

Blumstein (2021) showed that yellow bellied marmots (*Marmota flaviventer*) in highly disturbed

forests grew habituated to humans after repeated exposure and reduced their flight initiation

distances; however, marmots in groups that were habituated gained less body mass over the

duration of the study. On the other hand, mammal population densities are on the whole higher

near human settlements, indicating that the benefits of associated resources may outweigh the

negative effects of exposure (Tucker et al, 2021). Overall, there is a need to examine both the

immediate effects of non-lethal human disturbance as well as the long-term positive or negative

fitness consequences of such exposure.

3. Passive non-lethal interactions

Humans are the most elaborate ecosystem engineer on the planet, completely altering

environments to make it more suitable for themselves. Predation risk is highly dependent on

whether animals perceive these environments as safe or threatening (Wirsing et al. 2021). Van der

Kolk et al. (2024) found that various shore birds altered their flight responses based on the

frequency of airplanes over their nesting areas, with intermittent disturbance having the greatest

negative effect. Similarly, multiple studies have highlighted the negative effects of ship sonar on

cetacean behavior (see Harris et al. 2018 for a meta-analysis and review).

By altering habitats, human development alters the distribution of risk in the landscape. For

example, roads can clear away thickets or other cover that ambush predators use thus making it

"safer" for some prey species. Thus, one would expect prey species to congregate around roads

and invest more in foraging. We see evidence for this in guancos (*Lama guanicoe*) and mice

(*Peromyscus leucopus*) (Cappa et al. 2017; Grignolio et al. 2011). Mehlhoop et al. (2002) showed

that moose (*Rangifer tarandus*) in Norwegian forests avoid human infrastructure in areas where

hunting is permitted. On the other hand, Cappa, Giannoni, and Borghi (2017) showed that

guancos (*Lama guanicoe*) in large groups were less vigilant and foraged closer to roads in an

Argentine wildlife reserve and thus may perceive roads as beneficial. Widespread habitat

alteration has also led to temporal shifts in activity of multiple taxa globally (Gaynor et al. 2018).

Conflict with predator species can also make human settlements safe spaces for consumers to

forage (eg. wild boar, Podgorski et al. 2013). Conversely, development may also open up habitats

making it more difficult for prey individuals to escape predation (Smith, et al. 2016). Human

development may also pose a direct threat to animals, especially in the case of conflict with

livestock or roadkill (Loss et al. 2015, van der Kolk et al. 2024, Harris et al. 2018). Across 468

mammal species, Tucker et al (2021) find an increase in population densities near human

settlements, and speculate that more resources and lower competition/predation may lead to this

pattern. However, they do caution that while some species may get benefits, many others be

negatively affected and decline. Similarly, our meta-analysis found considerable variation across

studies on the effect of passive interactions on foraging, vigilance and movement behaviors

(Figure 4). Thus, the response of animals to passive interactions with roads and human

settlements may be highly context and species dependent.

4. Context dependence of the behavioural effects human interactions on animals

Animals can make mistakes in how they respond behaviourally to human interactions (Smith et

al. 2021), by overreacting or underreacting. Animals can overreact to non-lethal interactions

leading to high energy investment and even chronic stress responses (e.g. (Blumstein 2003;

Nunes et al. 2019; Westekemper et al. 2018; review: Larson et al. 2016). On the other hand,

misunderstanding lethal interactions can lead to the death of the individual. Thus, there are a

number of contingencies that may apply to the expression of risk avoidance and mitigation

behaviours. For example, the openness of the habitat, predator traits such as habitat domain,

hunting mode and fierceness, and prey body state are some important variables to consider

(Schmitz et al. 2001, Wirsing et al. 2021). While all of these also apply to the human predator,

there may be some additional considerations for a predictive framework (Smith et al. 2024).

i. Trophic level

Human hunters can augment existing predation risk from natural predators, create new predation

risk or both depending on which animals they target (Supplementary Table B2). Hunters tend to

disproportionately target predatory and larger species for their higher social and economic value.

For example, marine apex predators such as sharks, barracuda and tuna are highly valuable in fish

markets compared to consumer species such as sardines (Pauly et al. 1998; Jackson et al. 2001).

Similarly, in many developed countries large herbivores such as deer are hunted for sport or

culling (Pecorella et al. 2016). Thus, interactions with humans may be perceived differently

across trophic levels.

When humans target top predators, they create new risk that these animals have not previously

experienced. Predators may or may not be equipped to respond to new threats leading to many

extirpations (e.g. Wolves in Yellowstone (Laundré, Hernández, and Altendorfet al. 2001), bears in

Alaska (Merkle et al. 2013), puma in California and Florida (Nickel et al. 2021)). Thus, predators

must respond as naive prey when targeted by humans. The severity and direction of the response

may be proportional to the extent and duration of persecution by humans. Predators may be

overly cautious around human hunters leading to greater missed opportunities. Alternatively,

predators may not be cautious enough around humans leading to greater direct mortality (Smith

et al. 2021). The direct removal of predator individuals and their behavioural responses to the risk

of the human super-predator may have beneficial effects for consumers effectively subtracting

predation risk for these animals (e.g. mesopredator release and population explosions of large

herbivores, see Terbourg et al. 2015).

The majority of studies we found in our search focused on lower trophic levels and not predators

(Appendix A). This may be because it is particularly difficult to observe foraging and vigilance

behaviours in predators (but see Smith et al. (2017) for examples of human disturbance effects on

Puma concolor foraging behaviour). We were unable to test our hypotheses regarding predator

behavioural responses to humans in this meta – analysis, however, we highlight this as an

important topic for future research. Incorporating other behaviours such as movement and habitat

selection may allow us to gain further insights into how predators themselves respond to the

threat of predation.

Humans may also compete with top predators by targeting consumers (eg. deer hunters in

California). This may have a multiplicative effect on predation risk perceived by consumer

individuals. Consumers may have fewer safe habitats to forage and thus may experience a higher

energy cost of anti-predator strategies. There may also be mutualisms between predators and

humans creating areas of higher predation success (corralling) or even subsidizing predator

energy costs (cattle depredation). In a majority of studies, humans were not the only predator

present. For example, humans compete with puma in the Santa Cruz Mountains of California,

with wolves in parts of north America, and with bears in the arctic. Thus, hypotheses about

multiplicative effects of human predation risk and competition with top predators may be testable

in the field.

In some cases, humans may target multiple trophic levels simultaneously. Trawl fishing, for

example, is particularly non-selective. It is difficult to determine the effects of such ubiquitous

predation risk. Suraci et al. (2019) tested effects of lethal human interactions across multiple

trophic levels in the Santa Cruz Mountains and found evidence for a behavioural cascade. To our

knowledge, no other study has simultaneously tested behaviours across multiple trophic levels.

ii. Body mass

As with predators, larger animals may also be disproportionately targeted for economic reasons

and as trophies. Larger and healthy animals are rarely targeted by predators and thus have little

reason to invest in anti-predator behaviour. Thus, we expected the magnitude of effects to

increase with the body mass of the study animal. This, however, was not the case in our analysis

as the change in effect size was negligible across outcomes.

iii. History

The disproportionate targeting of apex predators has led to the local extirpation of many species

in the global north. Though conservation efforts have led to the creation of no-take and no-

hunting zones in the past few decades, a history of hunting may still affect how animals perceive

humans in their environment (Clinchy et al. 2016). For example, pPlayback experiments of

extirpated predators indicates that fear may persist across generations (Clinchy et al. 2016).

However, the degree of response by animals may depend on a number of factors, starting from

their own responses (whether innate or learned), nature of hunting, duration of hunting cessation,

which species were historically targeted and the nature of current human interactions. Repeated

exposure and learned tolerance may also play a role in response attenuation (Uchida and

Blumstein, 2021). Thus, it may be worthwhile to investigate whether the animal response over

time to the cessation of human activity, both lethal and non-lethal.

5. Using fear of humans for management

The fear of human hunters can also be valuable tool for conservation and management. Miller

and Schmitz (2019) proposed a framework for reducing the conflict between large carnivores

(*Panthera tigris*) and cattle herders in central India. They showed that by utilizing tigers' intrinsic

fear of humans in the central Indian landscape and common human signs such as traps, fences

and scents, conflict can be significantly reduced. Blackwell et al. (2016) proposed using fear

induced by periodic playbacks of human noises to deter animals from roads and reduce the toll of

road kill. Gaynor et al. (2021) proposed an applied ecology of fear framework for using non-

consumptive behavioural effects of the human super predator for conservation and management.

If indeed non-lethal cues work best when reinforced by actual lethal actions, a combination of

human cues and periodic and regulated hunting may offer a viable model for the management of

wild populations and the reduction of conflict.

V. Conclusions

The effects of natural animal predators on prey have been studied and debated for many decades

(Lotka 1920; Sih 1984; Dill, Heithaus, and Walters 2003; Steneck 2012; S.L. Lima et al. 2021). It

is clear that through their consumption of prey, natural predators create a dangerous environment

to which prey respond (Lima and Dill 1990; Laundré 2010). Thus, a wide range of anti-predator

responses have evolved and are induced in animals to minimize the probability of being killed

(Lima and Dill 1990). Recent evidence suggests that the fear of predators itself even without

direct threat of being killed changes prey responses (i.e. non-consumptive effects of predators),

with potential consequences to fitness (Schmitz, Beckerman, and O'Brien 1997; Heithaus et al.

2008; Hawlena and Schmitz 2010; Clinchy, Sheriff, and Zanette 2013; Heithaus et al. 2008;

Sheriff et al. 2020; Michael Clinchy, Sheriff, and Zanette 2013).

We have presented an extensive synthesis of the available literature on the effect of human

interactions on animal behaviour. Our general conclusions and suggestions for future research are

listed below:

1. While lethal interactions with humans have a profound effect on animal behaviour, there

is mixed evidence of such effects from non-lethal interactions.

2. There is considerable variation in the response of animals to non-lethal interactions that

has yet to be disentangled. Factors such as frequency of interactions, history of hunting,

and trophic level may attenuate the response of animals to human disturbance.

3. The effects of human interactions on animal fitness and movement would be an important

consideration for the management of wild populations.

4. Future research may benefit from contrasting humans as lethal predators with natural

predators in the same space and across trophic levels, given the potential for

compounding effects on animals and ecosystems.

Acknowledgments

Funding for this research was provided by the Prime Ministers Research Fellowship () to SD with

some support from DBT-Wellcome Trust India Alliance grant (IA/I/19/2/504639) to MT. We

would like to thank Dr. Nitya Mohanty and Dr. Daniel Noble for their invaluable inputs during

the drafting of this manuscript.

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 Figures and tables

Figure 1: The different scenarios in which humans interact with animals in wild spaces and the
hypothesised consequences on animal behavior. We expect a gradient of response with the
strongest behavioral changes resulting from lethal interactions with humans, followed by active
interactions and passive interactions. The scenarios are: a) A hunter that targets top predators
(top) b) A group of ecotourists watching animals in their natural habitat (middle) c) A road cutting
across a natural habitat (bottom).

945 **Figure 2:** Geographical distribution of studies included in the systematic review.

Figure 3: Distribution of included studies across type of interaction, treatment, behavior and

948   measured outcome (in order).

Figure 4: The effect of human disturbance (lethal, active disturbance, passive disturbance) across taxa and trophic levels. Negative values on the x-axis indicate a reduction in that behaviour, whereas positive values indicate an increase in that behaviour, relative to control conditions where no human activity is present.

Figure 5: Summary effect of human interactions on foraging, movement, and vigilance behaviour across studies included in this meta-analysis. For movement, active disturbance was excluded due to low sample size.