**A field experiment assessing the roles of drought, herbivory, and local climate on**

**cyanogenesis cline formation and local adaptation in *Trifolium repens***

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**Abstract**

Projecting how a population will adapt to environmental changes requires a mechanistic

understanding of the specific biotic and abiotic factors that impose selection on that population.

In white clover (*Trifolium repens*), clines in an antiherbivore defense mechanism, hydrogen

cyanide (HCN), form via variation in selection imposed by the environment. However, the

specific environmental factors that select for or against chemical defense phenotypes in white

clover remain unresolved. We performed a field experiment in high and low latitude study sites,

with a factorial manipulation of precipitation and herbivory at each site. These factors are

hypothesized to be important in driving HCN cline formation, so we investigated their effects on

fitness of a white clover F3 recombinant population, segregating for the alleles underlying the

HCN chemical defense phenotype. Surprisingly, we found precipitation and herbivory either did

not drive differential selection on HCN or its metabolic components, or did not impose selection

in a manner consistent with the maintenance of observed HCN clines. Instead, we find that the

production only one of the metabolic components of HCN, cyanogenic glycosides, resulted in a

fitness advantage, even when lacking the ability to produce HCN. This was most prominent at

the northern study site, which is again contrary to expectations. These results suggest additional

physiological roles that cyanogenic glycosides may play in adaptation and the evolutionary

ecology of white clover. This study highlights the importance of experimental manipulations of

environmental factors to understand how selection acts on genes underlying important

phenotypic traits, often in unexpected ways.

**Introduction**

Spatial and temporal variation in the environment frequently drives patterns of local adaptation

and the formation of clines (Endler, 1977; Kawecki and Ebert, 2004; Hereford, 2009). Both

biotic and abiotic factors can exhibit variable selection across heterogeneous environments, yet it

remains poorly understood how spatial variation in specific environmental factors drives local

adaptation and the evolution of clines (Nuismer et al., 2000; Siepielski et al., 2017; Briscoe

Runquist et al., 2020). This challenge can be partially resolved by conducting field experiments

that manipulate multiple environmental factors that are putative agents of selection acting across

spatially separated sites spanning an environmental gradient. These manipulative experiments

can serve to further our understanding of how populations adapt to multifaceted environments,

with the goal of determining the predictability of adaptation in current and future climates.

Selection imposed by the environment can vary spatially, resulting in genotype-by-

environment interactions ($G \times E$), whereby certain genotypes are advantageous under particular

biotic and abiotic conditions (Gillespie and Turelli, 1989). However, the relative impacts of

specific biotic and abiotic factors on selection are often difficult to discern, highly variable

among species, and/or highly dependent on environmental context (Macel et al., 2007; Siepielski

et al., 2017; Wright et al., 2018; Briscoe Runquist et al., 2020; Hargreaves et al., 2020). Biotic

and abiotic environments also often act concurrently to impose selection and influence

evolutionary patterns. A prominent hypothesis that integrates biotic and abiotic selection is the

latitudinal herbivory-defense hypothesis (LHDH; Johnson and Rasmann, 2011), which posits

that at lower latitudes, warmer and more stable climates support greater abundance and diversity

of herbivores, resulting in stronger selective pressure on plants due to herbivory (Coley, 1991;

Coley and Barone, 1996; Anstett et al., 2016; Moreira et al., 2018). In this case, a plant

functional trait (defense) is simultaneously influenced by multiple $G \times E$ interactions, where the

environment includes both biotic (herbivory) and abiotic factors (latitude, or by proxy,

temperature). However, evidence for and against the LHDH is mixed when examined at a large

scale (Moles et al., 2011; Anstett et al., 2016), and indeed the incidence of patterns in support of

this hypothesis might depend on the specific biotic and abiotic environmental context in question

(Anstett et al., 2014; Moreira et al., 2015; Kooyers et al., 2017; Anstett et al., 2018). These

complex results highlight the need for a continued focus on concurrent investigation of biotic and

abiotic factors as drivers of spatially variable selection on functional traits along environmental

gradients.

The cosmopolitan plant white clover (*Trifolium repens*) offers an ideal study system to

investigate environmental factors as drivers of selection on an ecologically important trait.

Cyanogenesis, or the production of the antiherbivore defense hydrogen cyanide (HCN), is a

chemical defense trait found in over 3000 plant species (Poulton, 1990). The production of HCN

in *T. repens* is typically controlled by two Mendelian inherited polymorphic loci (Corkill, 1940;

Atwood and Sullivan, 1943; Poulton, 1990). One locus governs the ability to produce cyanogenic

glycosides (CNglcs; namely, linamarin and lotaustralin), which is controlled by three linked

genes (*CYP79D15*, *CYP736A187*, and *UGT85K17*) in the metabolic pathway that produces

linamarin and lotaustralin (Corkill, 1940; Atwood and Sullivan, 1943; Olsen et al., 2008). We

hereafter refer to this locus as *Ac/ac* (Corkill, 1940; Atwood and Sullivan, 1943; Olsen et al.,

2008). The other locus (*Li/li*) governs the ability to produce a β-glycosidase called linamarase, a

hydrolyzing enzyme that cleaves the sugar moiety of CNglcs from the $-C\equiv N$ functional group to

release toxic HCN (Corkill, 1940; Atwood and Sullivan, 1943; Olsen et al., 2007). Deletion of

the functional *Ac* allele results in an inability to produce CNglcs while deletion of the functional

*Li* allele results in an inability to produce linamarase (Olsen et al., 2007; Olsen and Small, 2018).

Therefore, possession of at least one dominant allele (*Ac* and *Li*) at each locus results in the

presence of CNglcs and linamarase, respectively, and the ability to produce HCN. Plants that are

homozygous recessive *ac* lack CNglcs, and plants that are homozygous recessive *li* lack

linamarase (Corkill, 1940; Atwood and Sullivan, 1943). This genetic architecture results in four

cyanogenesis phenotypes (referred to as "cyanotypes"): one cyanotype that can produce both

CNglcs and linamarase (HCN$^+$/cyanogenic; AcLi); one that can produce only CNglcs

(HCN$^-$/acyanogenic; Acli); one that can produce only linamarase (HCN$^-$; acLi); and one that can

produce neither CNglcs nor linamarase (HCN$^-$; acli).

Clines in cyanogenesis in *T. repens* have been demonstrated across many types of

environmental gradients (Daday, 1954a; Daday, 1954b, Daday, 1958; Ganders, 1990; Thompson

et al., 2016). In general, the formation of clines in cyanogenesis relies on shifts in the ratio of

benefits to costs of HCN production along environmental gradients, altering selection for or

against cyanogenesis. Many hypotheses have been proposed to describe the agents of selection

driving cyanogenesis cline formation based on the costs and benefits of HCN production. For

example, herbivory and soil moisture have been hypothesized as potential environmental factors

driving latitudinal cyanogenesis cline formation based on the benefits of HCN or its metabolic

components across environments. Specifically, cyanogenesis may be favoured at lower latitudes

due to greater herbivore pressure (Daday, 1965; Angseesing, 1974; Dirzo and Harper, 1982;

Kakes, 1989; Santangelo et al., 2019), while CNglcs (and possibly linamarase) could provide

additional benefits of nitrogen recycling under low soil moisture conditions, which could vary

predictably across the range of white clover (Foulds and Grime, 1972a,b; Kooyers et al., 2014;

Albano and Johnson, 2023; Hendrickson et al., 2025). Freezing temperatures have been similarly

hypothesized as drivers of latitudinal cyanogenesis cline formation based on exacerbated costs of

HCN or its metabolic components in colder environments (Daday, 1965; Olsen and Ungerer,

2008; Kooyers et al., 2018; Fadoul et al., 2023). However, conflicting evidence is present in the

literature for each of these hypothesized environmental factors, with common garden studies

across latitudes also not definitively uncovering causative selection pressures (Wright et al.,

2018; Albano et al., 2024). The importance of multiple biotic and abiotic factors in imposing

selection has not yet been directly tested using manipulative experiments in the field. Therefore,

to improve our understanding of the mechanisms driving local adaptation and the evolution of

observed clines, it is important to perform field experiments across environments that manipulate

the effects of multiple environmental factors on a plant species within known genetic control of

an ecologically important trait.

To try and resolve uncertainty in the forces driving the evolution of HCN clines in *T.*

*repens*, we first sought to test whether there is a latitudinal cline in herbivory and HCN

frequency across an extended latitudinal gradient. We then conducted a manipulative field

experiment between disparate latitudes, creating two study sites in the northern and southern

distributions of the introduced range of *T. repens* in North America and performing a factorial

manipulation of herbivory and precipitation within each site. This experiment tests the

overarching hypothesis that biotic and abiotic factors concurrently drive the evolution of

adaptive clines in cyanogenesis along latitudinal gradients. Our specific research questions are:

(Q1) Is there a cline in herbivory and HCN frequency across a latitudinal gradient in the

introduced North American range of *T. repens*? (Q2) Is selection for or against cyanogenesis or

its metabolic components influenced by the manipulation of precipitation or herbivory? (Q3) Is

selection on cyanogenesis or its metabolic components by precipitation or herbivory also

dependent on the latitude of the study site? (Q4) Is the outcome of any observed selection on

cyanogenesis or its metabolic components consistent with the maintenance of the cyanogenesis

polymorphism within populations or the formation of cyanogenesis clines across latitudinal

gradients? We expect to observe strong evidence of latitudinal clines in herbivory and HCN

frequency, and selection due to precipitation and herbivory that is consistent with the

maintenance of the cyanogenesis polymorphism and the formation of cyanogenesis clines across

latitude in North America, based on three specific predictions (**Figure 1**). First, we predict

exacerbated costs of producing HCN, CNglcs, and/or linamarase at the higher latitude site due to

seasonal freezing temperatures. Second, when herbivores are reduced, we predict a larger

beneficial effect on fitness in *T. repens* unable to invest in HCN, CNglcs, and/or linamarase in

the lower latitude site, due to preferential investment into growth and reproduction as opposed to

defense in plants that are already experimentally defended. Third, we predict *Ac* plants to

experience a fitness advantage under drought, and that this advantage could vary between the

high and low latitude sites based on differences in the level of precipitation and

evapotranspiration occurring throughout the duration of the experiment, with both expected to be

higher at the low latitude site, in general. Our results provide insight into the potential interplay

between biotic and abiotic environments in the formation of clines in an ecologically important

trait within a widespread plant species.

**Methods**

*Study system*

*Trifolium repens* is a cosmopolitan perennial legume and obligate outcrosser that can reproduce

sexually via flowers and seeds and clonally via stolons, often forming large clonal mats in lawns,

fields, and other human-disturbed areas (Burdon, 1983). These multiple modes of reproduction

have contributed to the global range expansion of *T. repens* from its native Eurasia to multiple

introduced ranges on many other continents, including North America (Daday, 1958; Kjærgaard,

2003). We tested for the presence of latitudinal clines in HCN, the prominent chemical defense

mechanism in *T. repens* described above, through field surveys of populations across a large

latitudinal transect of the introduced range of North America (**Supplementary Material**)

*Experimental design*

All plant material for the experiment is the product of an F3 generation of *T. repens* seeds

(methods for experimental plant generation found in **Supplementary Material**; **Figure S1**). We

addressed our research questions using experimental manipulations of precipitation and

herbivory at two study sites, located in King Township, Ontario, Canada (44°1'40" N, 79°39'29"

W) and Lafayette, Louisiana, USA (30°18'24" N, 92°00'33" W). Historical and current climatic

characteristics of each study site can be found in **Table S1**. F3 seedlings were germinated in peat

pots (8 cm diameter, 591 mL volume) filled with Promix LP15 (Premier Tech Horticulture,

Rivière-du-Loup, Québec, Canada) and established for 8 weeks in a greenhouse near each study

site. At the northern study site, plants were established in a climate-controlled greenhouse at

25°C day, 15°C night, and a 14 h:10 h light:dark cycle. At the southern study site, plants were

established in a non-climate-controlled greenhouse at ambient conditions. Once cyanotypes were

confirmed for each seedling, individuals of each of the four cyanotypes were randomly selected

for planting into each of 96 plots, for a total of 384 plants per study site (**Figure S2A**). Each plot

was then overlaid with a 1.52 m × 1.82 m sheet of black landscaping fabric, into which four

holes were cut to allow for the peat pots containing each individual plant to be embedded into the

natural soil at each study site (**Figure S2A**). Individual plants were then randomly assigned and

planted into one of the four holes in each plot in spring 2021, with a target of one individual of

each cyanotype within each subplot. This target was achieved in Ontario, but in Louisiana, a lack

of germination success for some cyanotypes prevented each plot from having exactly one

individual of each cyanotype, and seedlings from cyanotypes with an excess number of available

individuals were used to ensure four plants were present in each subplot. In total, 50 subplots

contained one of each cyanotype, 42 subplots contained no more than two individuals of the

same cyanotype, and 4 subplots contained no more than three individuals of the same cyanotype.

In Ontario, our study used a split-plot design, with herbivory manipulated at the whole

plot level and precipitation manipulated at the subplot level. The 48 whole plots at each study

site were arranged as 96 paired subplots, in which one subplot was assigned as receiving the

precipitation reduction (drought) treatment, while the other subplot remained as an ambient

precipitation control. A herbivore reduction treatment was then randomly assigned to 24 of the

48 whole plots, resulting in each of the 96 subplots being assigned one of four treatment

combinations (**Figure S2A**). In Louisiana, design was similar, but the combination of

precipitation reduction and herbivore reduction treatments (four treatment combinations between

the two 2-level manipulations) was applied at the subplot level. To manipulate precipitation,

rainout shelters were constructed following the guidelines of Kundel et al. (2018), with some

modifications based on available materials and experiment requirements. Rainout shelters were

constructed with a wooden frame to dimensions of 1.52 m × 1.82 m, to match the dimensions of

the landscaping fabric covering each subplot, and were raised in the centre to allow precipitation

to run off the subplot (**Figure S2B,C**). Clear, corrugated polycarbonate strips were attached to

the frame to cover a defined percentage of the area of the subplot and thereby either act as a

precipitation reduction treatment or a control, depending on the orientation in which the

polycarbonate strips were attached to the frame (i.e., concave down to allow precipitation onto

the plants versus concave up to capture precipitation and carry away from the plants; **Figure**

**S2D**). For the 2021 growing season, an equal number of transparent, corrugated polycarbonate

strips (10) were attached to both the precipitation reduction shelters and the control shelters and

covered 50% of the subplot area with the intention of excluding ~50% of precipitation. Soil

moisture (percent volume) was measured periodically across each growing season using a soil

moisture meter (Delta-T SM150 Soil Moisture Kit and Moisture Sensor; Delta-T Devices,

Burwell, UK), with three replicate measurements taken in each subplot in Ontario and four

replicate measurements taken in each subplot in Louisiana (**Figure S3**). In year 1 (2021), the

rainout shelters did not adequately reduce soil moisture in either Ontario (precipitation reduction

mean: $29.2 \pm 0.3\%$ volume, control mean: $29.5 \pm 0.3\%$ volume; $P = 0.504$) or Louisiana

(precipitation reduction mean: $37.7 \pm 0.8\%$ volume, control mean: $36.7 \pm 0.8\%$ volume; $P =$

0.423; **Table S2**). Therefore, in year 2 (2022), shelters were modified to contain 5 transparent,

corrugated polycarbonate strips on the control shelters and 15 strips on the precipitation

reduction shelters to cover ~75% of the area of treated subplots and allowing for a target of

excluding ~75% of precipitation. Additionally, at the Ontario study site only, a drainage pipe

was attached to the base of the rainout shelters on the precipitation reduction side of whole plots

to catch precipitation and allow it to evaporate. In year 2, soil volumetric water content in

precipitation reduction subplots was 18% lower than in control subplots in Ontario (precipitation

reduction mean: $29.6 \pm 0.2\%$, control mean: $36.1 \pm 0.2\%$; $P < 0.001$), while rainout shelters still

did not adequately reduce soil volumetric water content in Louisiana (precipitation reduction

mean: $32.0 \pm 1.1\%$, control mean: $32.4 \pm 1.2\%$; $P = 0.841$; **Table S2**). The years of the

experiment were particularly wet in Louisiana, subjecting this study site to flooding during any

moderate rain, likely causing the limited effectiveness of the precipitation reduction treatment

throughout the experiment (**Table S1**).

To manipulate herbivory, a combination of insecticide and molluscicide was applied

biweekly throughout the 2021 and 2022 growing seasons to deter feeding by all common *T.*

*repens* herbivores, following the protocol of Santangelo et al. (2019). Briefly, a 0.0092%

solution of Coragen® insecticide (Dupont Chemical, Mississauga, Ontario, Canada) was applied

to all aboveground tissue of treated plants until beading of the liquid occurred on leaves, using a

pressurized pesticide sprayer, with an equal volume of water being applied to control plants.

Additionally, ~800 mg of Safer's® Slug and Snail Killer molluscicide pellets (Safer Brand,

Lancaster, Pennsylvania, USA) were scattered across the surface of each treated plant. These

treatments were shown to be effective at reducing herbivory at both study sites (see **Results**) and

therefore were not modified between 2021 and 2022.

All plants were maintained for two full growing seasons at each study site, from spring

2021 until autumn 2022, with scheduled measurements of plant fitness being conducted across

each growing season. Whether a plant survived and/or flowered was assessed weekly, with ripe

fruits also being counted and collected weekly. Seeds were then separated using a thresher

(Precision Machine Co., Lincoln, NE, USA), and fecundity was measured as total seed set mass

per plant. Photos were taken of each plant monthly and plant area was determined using Easy

Leaf Area (Version 2.0; Easlon and Bloom, 2014). The maximum plant area across each growing

season was determined and growth rate was calculated as $(\ln[A_{max}] - \ln[A_1])/(t_{max} - t_1)$, where

$A_{max}$ represents the maximum plant area, $A_1$ represents the initial plant area (i.e., after one month

of growth in the field), and $(t_{max} - t_1)$ represents the number of days between the photos taken for

maximum plant area and initial plant area. Herbivory was assessed visually to the nearest 1%

from 5 fully-expanded trifoliate leaves per plant (or fewer if a plant had less than 5 fully-

expanded leaves) each month following the methods of Johnson et al. (2016). In Louisiana,

aboveground plant tissue outside of an 8 cm diameter area around the planting location of each

individual was manually removed in January 2022 to prevent plants from growing too large and

competing with neighbouring individuals during the second growing season. In Ontario, winter

freezing temperatures result in a natural dieback of vegetation after each growing season, so

manual tissue removal was not necessary.

*Statistical analysis*

All statistical analyses were conducted in R Version 4.4.1 (2024) and R Studio Version

2024.09.1 (R Development Core Team, 2024). All response variables were then analyzed

through general linear mixed effects models using the function *glmmTMB* in the *glmmTMB*

package (Version 1.1.9; Brooks et al., 2017), with $\chi^2$ and *p* values calculated using the *Anova*

function in the *car* package (Version 3.1-2; Fox and Weisberg, 2019) with Type II sums-of-

squares. For significant main effects of cyanotype and any significant interactions of cyanotype

with one or more of the manipulated variables, custom contrasts were created using the *emmeans*

and *contrast* functions in the *emmeans* package (Version 1.10.4; Length, 2024). For models

using quantitative measures of fitness or herbivory, all transformations necessary to meet

assumptions of normality and homogeneity of variance can be found in **Table S3**.

Each of the response variables (survival, whether plants produced flowers, whether plants

produced seeds, number of flower heads, seed set mass, maximum plant area, growth rate, and

herbivory) were assessed to determine the effects of cyanotype and each of our experimental

manipulations on plant fitness. The number of flower heads and seed set mass produced by each

plant were both zero-inflated, so prior to analyzing these variables, it was necessary to first

analyze whether plants flowered or produced seeds, and then subset the number of flower heads

and seed set mass variables to only include values greater than zero, as has been done previously

(Anderson et al., 2015; Ferris and Willis, 2018; Albano and Johnson, 2023; Scharnagl et al.,

2023). In Ontario, all models consisted of the following structure:

$Response \sim Cyanotype\ (C) + Precipitation\ (P) + Herbivores\ (H) + C{:}P + C{:}H + P{:}H$

$\qquad\qquad + (1|Herbivores/Whole\ Plot\ ID)$

In Louisiana, all models consisted of the following structure:

$Response \sim Cyanotype\ (C) + Precipitation\ (P) + Herbivores\ (H) + C{:}P + C{:}H + P{:}H$

$\qquad\qquad + (1|Herbivores{:}Precipitation/Whole\ Plot\ ID)$

The split-plot experimental design necessitated the use of nested random effects to account for

non-independence of replicates within each whole plot in Ontario, while the different treatment

application in Louisiana necessitated an interaction term between the herbivore and precipitation

reduction treatments as part of the structure of random effects. Models including additional

nested random effects to account for non-independence of replicates within each subplot either

experienced convergence issues or did not alter our conclusions and were therefore not used.

Cyanotype (AcLi, Acli, acLi, or acli), precipitation manipulation (reduced versus control) and

herbivore manipulation (reduced versus control) were included as fixed effects along with all

two-way interactions. The three-way interaction was also originally included in all models but

was not statistically significant at $p < 0.05$ in any cases and therefore was removed from all

models to increase statistical power in testing the remaining fixed effects. Data from each of the

two growing seasons were analyzed separately due to the difference in how the precipitation

manipulation was conducted between the 2021 and 2022 growing seasons and the manual

removal of aboveground tissue in Louisiana after the 2021 growing season. Data from each of

the two study sites were analyzed separately due to some subplots not containing exactly one

individual of each cyanotype in Louisiana, along with the differences in treatment application.

**Results**

*Field surveys*

Our field surveys of *T. repens* populations identified the presence of clines in both herbivory and

plant defense across a north-south temperature gradient in North America. Mean annual

temperature was strongly negatively correlated with latitude ($r^2 = 0.995$; $P < 0.001$), with higher

latitudes being associated with lower mean annual temperatures (**Figure 2A**). The proportion of

cyanogenic *T. repens* plants was strongly positively correlated with mean annual temperature ($\chi^2$

$= 21.624$; $P < 0.001$), with warmer climates (i.e., lower latitudes) being associated with greater

incidence of cyanogenesis (**Figure 2B**). The percent of leaf area consumed was weakly

positively correlated with mean annual temperature ($r^2 = 0.065$; $P = 0.075$; **Figure 2C**), with

warmer climates (i.e., lower latitudes) being associated with marginally higher levels of

herbivory.

*Plant sexual fitness in Year 1 (2021)*

In Ontario, cyanotype affected plant fitness, but experimental manipulations (precipitation and

herbivores) did not affect selection on *T. repens* cyanotypes (**Table 1**). Specifically, cyanogenic

plants had 54% larger seed set mass than acyanogenic plants and *Ac* (CNglc-producing) plants

had 56% larger seed set mass than *ac* plants, while *Li/li* did not affect seed set mass, and

cyanotype did not affect whether plants produced seeds (**Figure 3**). We found limited evidence

of an effect of the precipitation manipulation on the sexual fitness of *T. repens*, potentially due to

the lack of effectiveness of this treatment in 2021. Similarly, the herbivore manipulation did not

affect the sexual fitness of *T. repens* in Ontario in 2021, even though it reduced herbivory by

54%.

In Louisiana, cyanotype affected whether plants produced seeds but not their seed set

mass, while the herbivore manipulation affected plant fitness and selection on *T. repens*

cyanotypes (**Table 1**; **Figure 3**). Cyanogenic (HCN-producing) plants were 53% more likely to

produce seeds than acyanogenic plants and *Li* (linamarase-producing) plants were 44% more

likely to produce seeds than *li* plants, while *Ac/ac* did not affect whether plants produced seeds

(**Figure 3**). We found strong evidence of an effect of the herbivore manipulation on the sexual

fitness of *T. repens*, with herbivore reduction resulting in 2.5× more individuals producing seeds

and 3.2× greater seed set mass. Cyanotype and herbivore manipulation also interacted to affect

the sexual fitness of *T. repens*, with AcLi, Acli, and acli plants producing 4.5×, 11.2×, and 5.8×

larger seed set masses, respectively, when herbivores were reduced versus the control, while acLi

plants exhibited no sexual fitness advantage when herbivores were reduced (**Figure 4**). The

precipitation manipulation had no effect on fitness or selection, again likely due to the lack of

effectiveness of this experimental treatment in 2021.

*Plant sexual fitness in Year 2 (2022)*

In Ontario, cyanotype affected plant fitness, but we only found limited evidence of the

precipitation manipulation affecting selection on *T. repens* cyanotypes and no evidence of an

effect of the herbivore manipulation on selection (**Table 1**). Specifically, cyanogenic plants were

15% more likely to produce seeds than acyanogenic plants and *Ac* plants were 27% more likely

to produce seeds than *ac* plants (**Figure 3**). Furthermore, cyanogenic plants produced 27%

greater seed set mass than acyanogenic plants and *Ac* plants produced 41% greater seed set mass

than *ac* plants. *Li/li* did not affect seed set mass or whether plants produced seeds (**Figure 3**).

Precipitation reduction resulted in 36% smaller seed set mass, while cyanotype and precipitation

manipulation also interacted to affect whether plants produced seeds (**Table 1**), with acli plants

54% more likely to produce seeds under reduced precipitation, while AcLi, Acli, and acLi plants

were not affected by precipitation reduction (**Figure 5**). Herbivore reduction resulted in a 21%

higher proportion of individuals producing seeds, but it did not affect seed set mass (**Table 1**).

In Louisiana, cyanotype affected plant fitness, but experimental manipulations

(precipitation and herbivores) did not affect selection on *T. repens* cyanotypes (**Table 1**).

Specifically, *Ac* (CNglc-producing) plants had 32% larger seed set mass than *ac* plants, while

cyanogenesis and *Li/li* did not affect seed set mass, and cyanotype did not affect whether plants

produced seeds (**Figure 3**). We again found limited evidence of an effect of the precipitation

manipulation on the sexual fitness of *T. repens*, despite the rainout shelter modifications made

prior to the 2022 growing season. Similarly, the herbivore manipulation did not affect the sexual

fitness of *T. repens* in Louisiana in 2022, even though it reduced herbivory by 17%.

*Additional reproductive and growth traits in Years 1 and 2*

In general, results for whether plants flowered and the number of flower heads produced were

qualitatively similar to whether plants produced seeds and seed set mass across both growing

seasons at both the Ontario (**Tables S4**, **S6**) and the Louisiana (**Tables S5**, **S7**) study sites. We

also did not find evidence of selection for or against *T. repens* cyanotypes due to either

precipitation or herbivore manipulation based on measures of survival or vegetative fitness

(maximum plant area or growth rate; **Tables S4-S7**).

*Herbivory in Years 1 and 2*

In Ontario in 2021, herbivory (measured as the percentage of leaf area consumed) was, as

expected, strongly affected by the reduction of herbivores, which resulted in 57% less leaf tissue

consumed (herbivore reduction mean: 6.0 ± 0.5% leaf area consumed, control mean: 14.0 ± 0.8%

leaf area consumed; **Table S4**). Cyanogenic plants experienced 33% less herbivory than

acyanogenic plants (**Figure 6**). Surprisingly, acyanogenic *Ac* (Acli) plants experienced 46% less

herbivory than *ac* plants, which is contrary to expectations because solely producing CNglcs

does not result in *T. repens* plants being able to produce HCN as a defense mechanism (**Figure**

**6**). Similarly, in Ontario in 2022, herbivory was strongly affected by the reduction of herbivores,

which resulted in 57% less leaf tissue consumed (herbivore reduction mean: 10.0 ± 0.7% leaf

area consumed, control mean: 17.8 ± 1.0% leaf area consumed; **Table S6**). Cyanogenic plants

experienced 25% less herbivory than acyanogenic plants (**Figure 6**). Acyanogenic *Ac* (Acli)

plants experienced 27% less herbivory than *ac* plants, which again, is contrary to expectations

(**Figure 6**).

In Louisiana in 2021, herbivory was strongly affected by the reduction of herbivores,

which resulted in 87% less leaf area consumed (herbivore reduction mean: 2.2 ± 0.3% leaf area

consumed, control mean: 16.6 ± 1.3% leaf area consumed; **Table S5**). However, cyanotype did

not affect the amount of herbivory on *T. repens* plants, indicating a lack of effectiveness of HCN

in defending plant tissue at this site (**Figure 6**). In Louisiana in 2022, neither the herbivore

manipulation nor cyanotype alone affected the amount of herbivory on *T. repens* plants (**Table**

**S7**; **Figure 6**). However, cyanotype and the herbivore manipulation  did interact to affect

herbivory, with AcLi, Acli, and acLi plants experiencing 47%, 29%, and 23% reductions in leaf

area consumed, respectively, when herbivores were reduced versus the control, while acli plants

exhibited 15% greater herbivory when herbivores were reduced (**Table S7**).

**Discussion**

Our findings indicate the presence of a latitudinal cline in plant defense, with a reduced

frequency of cyanogenic *T. repens* plants in colder North American climates, in accordance with

the LHDH. Additionally, we provide the first direct evidence of a positive relationship between

mean annual temperature and herbivory in *T. repens*, indicating greater herbivore pressure at

lower latitudes, despite greater incidence of cyanogenesis, also in accordance with the LHDH

(Q1). Our manipulative experiment clearly shows that the metabolic components of HCN are

under natural selection, but our results do not support the long-standing hypotheses that

precipitation and herbivory impose selection on the presence/absence of HCN or its metabolic

components in *T. repens* and drive the formation of cyanogenesis clines. Despite clear evidence

that reduced precipitation and herbivory both decrease *T. repens* fitness, we found only limited

evidence of precipitation and herbivory imposing differential selection on *T. repens* cyanotypes

(Q2). We did find some evidence of differential effects of cyanotype that depended on the study

site location, with greater beneficial effects of cyanogenesis and the ability to produce CNglcs on

plant fitness at the higher latitude study site, which is contrary to expectations (Q3). In the

limited instances when herbivory or precipitation did interact to affect *T. repens* fitness, the

outcomes were also contrary to expectations, with demonstrated fitness advantages for

unexpected cyanotypes based on the observed latitudinal cyanogenesis cline (Q4). Here we

discuss our results and their implications in the context of plant-herbivore interactions and

latitudinal cline formation for the presence of HCN and its metabolic components in *T. repens*.

*Limited evidence of precipitation as a driver of cyanogenesis clines*

The ability of plants to invest in essential functions like vegetative growth, reproduction, and

defense is dependent on the costs and benefits of those investments. Soil moisture availability is

a key environmental factor determining the balance of costs and benefits of these functions by

impacting resource availability and therefore, the ability of plants to invest into both primary and

secondary metabolic processes (Coley et al., 1985; Mundim and Pringle, 2018). Unexpectedly,

precipitation had inconsistent effects on *T. repens* fitness. This may be at least partially due to

the design of our rainout shelters, which only reduced precipitation in Ontario in 2022, and/or

due to consistent flooding that occurred throughout the entire experiment in Louisiana. We also

recognize the importance of extreme events in driving natural selection (Grant et al., 2017), with

variation in climatic and experimental conditions between years in our experiment possibly

explaining observed inconsistencies (**Table S1**, **Table S2**). Nonetheless, we can conclude that

even in instances where precipitation did exhibit strong effects on quantitative measures of plant

fitness (e.g., precipitation reduction resulted in decreased seed set mass in Ontario in 2022), there

were no differential effects on fitness of *T. repens* cyanotypes due to the precipitation

manipulation. Thus, plants with different capacities to invest in plant defense did not differ in

their ability to invest in growth or reproduction based on the impact of precipitation reduction on

resource acquisition, which is contrary to hypothesis surrounding the involvement of water

availability in cyanogenesis cline formation. More specifically, our findings are contrary to

recent evidence that *Ac* plants exhibit a fitness advantage when experiencing moderate and

consistent drought conditions (Kooyers et al., 2014; Albano and Johnson, 2023), potentially due

to the ability of plants to recycle N-rich compounds such as CNglcs during periods of low soil

moisture (see further discussion below; Møller, 2010; Machingura et al., 2016). Our results are

partially supported, however, by classic work on the effects of soil moisture on *T. repens*

cyanotypes, with *ac* plants exhibiting greater survival and similar dry biomass compared to *Ac*

plants experiencing drought, possibly due to linkage of *Ac/ac* with other loci involved in

environmental responses (Foulds and Grime, 1972a). As a whole, we are unable to conclude that

additional benefits aside from defense, such as greater facilitation of N recycling under drought

by *Ac* plants, are involved in cyanogenesis cline formation in *T. repens*.

In our study, a sole instance of differential effects of precipitation reduction on *T. repens*

cyanotypes (at the Ontario study site in 2022) resulted in increased likelihood of flowering and

producing seeds in acli plants, but not in any other cyanotype. However, when precipitation was

reduced, acli plants were still not more likely than Acli or AcLi plants to flower or produce

seeds, which would be expected if acyanogenic plants are selected for in colder climates.

Furthermore, our results provide little evidence of selection on HCN, CNglcs, or linamarase due

to soil moisture in terms of quantitative measures of fitness (i.e., number of flower heads or seed

set mass). The literature remains conflicted on how soil moisture is likely to influence selection

on cyanogenesis or its metabolic components in *T. repens*. Foulds and Grime (1972a,b) found a

lack of flowering success for cyanogenic plants under drought and a lower frequency of *Ac* in

droughted environments, while Kooyers et al. (2014) found a higher frequency of *Ac* in

droughted environments. None of these previous studies found an effect of soil moisture

availability on the frequency of *Li* or HCN in *T. repens* populations. Overall, our results are

inconsistent with the maintenance of cyanogenesis clines across latitudinal gradients and we are

unable to provide more definitive insight into the conflicting evidence from the literature

concerning the ability of soil moisture/precipitation to drive the formation of *T. repens*

cyanogenesis clines. Our methods have improved on previous experiments by manipulating *Ac*

and *Li* to be on a randomized genetic background to minimize the influence of other genes

involved in environmental response that may have affected previous results. However, because

our F3 generation is the product of an original cross of only two parent plants and only a single

generation of random crossing, it is possible that loci that are very tightly linked to *Ac*/*ac* or *Li*/*li*

are influencing our results as well (Kuo et al., 2024a). We are also unable to account for copy

number variation at either locus in the parent plants or among the F3 generation individuals used

in the experiment, which could affect rapid adaptation across environmental gradients (Kuo et

al., 2024b). A genome-wide association analysis could help elucidate if nearby genomic regions

or copy number variation at cyanogenesis loci influence drought response in *T. repens*.

*Costs and benefits of cyanogenesis and its metabolic components as a chemical defense*

*mechanism*

The benefits associated with plant chemical defense production are likely to be dependent on

environmental context. For example, under the LHDH, the benefits of antiherbivore defense are

predicted to be more prominent at lower latitudes due to greater herbivore pressure imposing

stronger selection in these environments (Coley, 1991; Coley and Barone, 1996; Johnson and

Rasmann, 2011). In line with the LHDH, we demonstrated the presence of a latitudinal cline in

herbivory that coincides with clines in the frequency of cyanogenesis in *T. repens* populations,

with greater herbivory and cyanogenesis frequency being associated with warmer temperatures

(i.e., lower latitudes). If the LHDH is further supported by the results of our manipulative

experiment, in Louisiana (where experimental release of herbivore pressure is hypothesized to be

most substantial), we would expect interactions between cyanotype and herbivore manipulation,

with the reduction of herbivores resulting in the greatest benefit for plants that invest the least

amount of resources into defense, and vice versa. Instead, any demonstrated interactions between

cyanotype and herbivore manipulation in Louisiana were not concordant with expectations,

because acli plants did not benefit the most from the experimental exclusion of herbivores, nor

did AcLi plants benefit the least (**Figure 4**).

Our results provide stronger evidence of the presence of CNglcs providing a fitness

advantage to *T. repens* than the presence of HCN itself. This is an unexpected result because

without the ability to produce linamarase, *Ac* plants are unable to endogenously produce HCN at

the moment that cell lysis occurs when herbivores damage plant tissue, which is the typical

mechanism by which cyanogenic (AcLi) plants exhibit feeding deterrence. However, Acli plants

experienced equal deterrence of herbivory to AcLi plants in Ontario, despite a lack of HCN

production. Previous work does indicate the potential for CNglc-producing plants to be able to

elicit a toxic effect on herbivores later in the feeding process that could result in deterrence of

feeding (Dirzo and Harper, 1982; Kakes, 1989; Desroches et al., 1997; Tattersall et al., 2001;

Pankoke et al., 2012). This mechanism is a result of β-glycosidases produced by herbivores,

some of which may unintentionally hydrolyze CNglcs to produce HCN in the herbivore's gut. In

fact, previous studies involving feeding assays on *T. repens* tissue indicate the potential for

feeding deterrence by CNglc-producing plants (Dirzo and Harper, 1982; Kakes 1989). Dirzo and

Harper (1982) also determined that excised guts of slug species frequently produce HCN when

exposed to chemical CNglcs. There is evidence for this downstream antiherbivore effect in

numerous plant study systems that are able to produce various CNglcs as well, including vicine

in broad beans (Desroches et al., 1997), iridoid glycosides in *Plantago lanceolata* (Pankoke et

al., 2012), and dhurrin in *Arabidopsis thaliana* and other plant species (Tattersall et al., 2001;

Yadav et al., 2023). However, previous feeding trials using *T. repens* also provide evidence of a

lack of feeding deterrence in CNglc-producing plants (Dirzo and Harper, 1982; Fadoul et al.,

2023), indicating a level of heterogeneity either within or among herbivore species in their

production of β-glycosidases that could cause HCN production in the gut post-feeding. A future

study conducting a direct test of this mechanism on multiple herbivore communities in natural

settings is critical to assessing herbivory as a potential hypothesis to describe the consistent

presence of latitudinal clines in HCN (and in the frequency of *Ac*) in *T. repens* populations.

While CNglcs may be providing a fitness advantage to *T. repens* plants through a

reduction of herbivory despite a lack of HCN production, it is unlikely that CNglcs acting as a

chemical defense mechanism is their only benefit to plant performance. In the sole instance of

herbivore manipulation imposing differential effects on the fitness of *T. repens* cyanotypes (at

the Louisiana study site in 2021), the reduction of herbivores resulted in the greatest benefit for

Acli plants, indicating a distinct selective advantage of CNglc production that outweighs its cost,

even when herbivory is experimentally reduced by over 50%. It is possible that the recycling of

CNglcs into usable N through β-cyanoalanine synthase serves as this additional benefit to CNglc

production (Machingura et al., 2016). The ability to produce CNglcs has been shown to be

particularly advantageous under low soil moisture in a greenhouse (Albano and Johnson, 2023),

which was not the case in the current study conducted in natural conditions. However, it remains

a possibility that this additional source of N is universally beneficial to a plant's ability to invest

in growth and/or reproduction, regardless of soil moisture conditions. By contrast, we did not

find consistent evidence of the ability to produce linamarase providing a fitness benefit to *T.*

*repens* plants aside from the contribution of linamarase to HCN production in AcLi plants. Thus,

linamarase may incur an allocation cost to plants in the absence of HCN and herbivores. Further

work directly addressing all potential costs and benefits of CNglcs and linamarase is needed to

more definitively test hypotheses surrounding the formation of cyanogenesis clines in *T. repens*.

*Maintenance of cyanogenesis polymorphisms and clines*

We find evidence of epistatic interactions between the *Ac*/*ac* and *Li*/*li* loci but not in a manner

that explains the maintenance of the cyanogenesis polymorphism within populations. One such

epistatic interaction is the increase in flowering success of acli plants under drought that is not

present in any other cyanotype, indicating an absence of both metabolic components of

cyanogenesis is required for drought to affect flowering in *T. repens*. An additional epistatic

interaction is evidenced by fitness benefits (and greater relative deterrence of herbivory) for

cyanogenic plants in Ontario but not in Louisiana, indicating the ability to produce both

metabolic components of cyanogenesis provides fitness advantages that are variable among *T.*

*repens* populations. However, these results conflict with expected trends that would explain the

maintenance of cyanogenesis polymorphisms within populations, which would require AcLi and

acli cyanotypes to be favoured over Acli and acLi cyanotypes (Kooyers and Olsen, 2012).

Our results also do not provide evidence that either directional or epistatic selection

would create or maintain cyanogenesis clines across a latitudinal gradient. In fact, our field

sampling efforts clearly demonstrate fixation of both *Ac* and *Li* alleles in Louisiana, as expected,

while surprisingly, our experimental results suggest that *Ac* alleles should be fixed at both

northern and southern study sites. These results are contrary to the LHDH and to previous

evidence of exacerbated ecological or allocation costs of HCN production in colder climates as

drivers of cyanogenesis cline formation (Daday, 1965; Kooyers et al., 2018). Results of previous

studies on deviations from the LHDH do provide some support for our findings, with greater

expression of plant defenses in colder climates or plant defenses exhibiting selection patterns

based on other environmental factors and/or other plant processes, although these results are

often highly variable (Moles et al., 2011; Kooyers et al., 2017; Anstett et al., 2018). It is also

possible that rare and stressful extreme climatic events that did not occur during the time period

of our experiment select for *ac* and *li* alleles, which could be at least partially responsible for the

maintenance of the cyanogenesis polymorphism within most natural *T. repens* populations and

cyanogenesis cline formation. Taken together, these findings indicate that understanding the

formation of clines in plant defense across complex environmental gradients requires a

multifaceted approach to understanding the costs, benefits, and trade-offs associated with plant

defenses among populations.

*Limitations*

The clearest limitation of our study is the observed lack of effectiveness of the precipitation

manipulation in significantly reducing soil moisture availability compared with control subplots

at both study sites in 2021. This restricts our ability to detect effects of precipitation

manipulation on *T. repens* fitness and to assess precipitation as an agent of selection on

cyanogenesis and its metabolic components in *T. repens*. Rainout shelters were modified after

the 2021 growing season to reduce precipitation more substantially in the treated subplots and

limit the amount of unintentional precipitation reduction in the control subplots. These

modifications resulted in significant differences in soil moisture between precipitation reduction

subplots and control subplots in Ontario in 2022, but not in Louisiana, likely because of

uncharacteristically high rainfall and soil moisture across the growing season at this study site

(**Table S1**). However, despite the ineffectiveness of the treatment, in instances where

precipitation reduction did have significant effects on *T. repens* fitness in Ontario in 2022, we

did not find evidence of the precipitation manipulation imposing differential selection on *T.*

*repens* cyanotypes. This result increases confidence in our evidence against precipitation as a

driver of the formation of cyanogenesis clines in *T. repens*, despite mixed effectiveness of the

precipitation manipulation.

An additional potential limitation of our study is that individuals were not transplanted

into the field until after eight weeks of growth in a common greenhouse environment. This was

necessary to provide sufficient time for each seedling's cyanotype to be confirmed through

Feigl-Anger assays before transplanting one individual of each cyanotype into each subplot

(Feigl and Anger, 1966). However, the seedling stage is critical to lifetime fitness due to strong

viability selection (Agrawal et al., 2013; Barton and Hanley, 2013; Warvell and Shaw, 2019),

which could be particularly true for seedling mortality caused by our variables of interest,

precipitation and herbivory. The current study experienced very low mortality overall, while a

separate experiment transplanting *T. repens* individuals into similar common garden sites after

only four weeks experienced substantially higher mortality of seedlings (Albano et al., 2024).

Future studies directly investigating the effects of precipitation and herbivory on white clover

seedlings across latitudes would be useful to corroborate our results with data across the entire

life cycle of *T. repens*.

*Conclusions*

We employed a manipulative experiment in natural conditions to investigate precipitation and

herbivory as potential agents of natural selection on cyanogenesis, driving the evolution of cline

formation at *Ac/ac* and *Li/li* in *T. repens*. Overall, there was limited evidence for either of these

manipulated environmental variables in imposing differential selection on *T. repens* cyanotypes

at either study site. Further, in isolated instances that did indicate differential selection,

cyanotypes were not affected in a manner consistent with the maintenance of cyanogenesis

polymorphisms within populations or clines across latitudinal gradients. Therefore, our results do

not support previous hypotheses surrounding precipitation and herbivory as influential factors in

the formation of latitudinal cyanogenesis clines. Instead, we found that plants that produce

CNglcs, exhibited higher fitness than other cyanotypes in general, regardless of cyanogenesis,

and that this effect was stronger at the higher latitude garden, which is again, contrary to

expectations based on observed cyanogenesis clines. This indicates that the ability to produce

HCN as an antiherbivore defense could still outweigh any potential costs to the plant, even in

environments that experience seasonal freezing temperatures that exacerbate costs (Kooyers et

al., 2018). Furthermore, the ability to produce CNglcs may provide additional benefits even in

the absence of plant-derived β-glycosidases, such as downstream antiherbivore effects through

the activity of herbivore-derived β-glycosidases or the ability to recycle N from CNglcs for use

in other plant functions. However, direct tests of the significance of these potential benefits are

needed in *T. repens* to provide insight into their relative impact on plant fitness. Additionally,

future experiments using QTL mapping or genome editing to more definitively isolate the effects

of *Ac*/*ac* and *Li*/*li* on plant fitness and environmental response can help elucidate their potentially

multifaceted roles in plant function, cline formation, and adaptation to environmental change.

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**Conflict of Interest**

We have no conflict of interest to declare.

**Data Availability Statement**

The dataset generated and analyzed during this study is published in Dryad Digital Repository:

https://doi.org/ 10.5061/dryad.4xgxd25n1.

**References**

Agrawal, A. A., Johnson, M. T. J., Hastings, A. P., Maron, J. L. (2013). A field experiment

demonstrating plant life-history evolution and its eco-evolutionary feedback to seed

predator populations. *The American Naturalist*, *181*, S35-S45.

Albano, L. J., Johnson, M. T. J. (2023). Interactions environmental factors drive selection on

cyanogenesis in *Trifolium repens*. *Oikos*, *2023*, e09629.

Albano, L. J., Bastias, C. C., Estarague, A., Hendrickson, B. T., Innes, S. G., King, N., … ,

Johnson, M. T. J. (2024). A transcontinental experiment elucidates (mal)adaptation of a

cosmopolitan plant to climate in space and time. *bioRxiv*. doi:

10.1101/2024.09.16.613311

Anderson, J. T., Eckhart, V. M., Geber, M. A. (2015). Experimental studies on adaptation in

*Clarkia xantiana*. III. Phenotypic selection across a subspecies border. *Evolution*, *69*,

2249-2261.

Angseesing, J. P. A. (1974). Selective eating of the acyanogenic form of *Trifolium*

*repens*. *Heredity*, *32*, 73.

Anstett, D. N., Naujokaitis-Lewis, I., Johnson, M. T. J. (2014). Latitudinal gradients in herbivory

on *Oenothera biennis* vary according to herbivore guild and specialization. *Ecology*, *95*,

2915-2923.

Anstett, D. N., Nunes, K. A., Baskett, C., Kotanen, P. M. (2016). Sources of controversy

surrounding latitudinal patterns in herbivory and defense. *Trends in Ecology & Evolution*,

*31*, 789-802.

Anstett, D. N., Ahern, J. R., Johnson, M. T. J., Salminen, J.-P. (2018). Testing for latitudinal

gradients in defense at the macroevolutionary scale. *Evolution*, *72*, 2129-2143.

Atwood, S. S., Sullivan, J. T. (1943). Inheritance of a cyanogenetic glucoside and its hydrolyzing

enzyme in *Trifolium repens*. *Journal of Heredity*, *34*, 311-320.

Barton, K. E., Hanley, M. E. (2013). Seedling-herbivore interactions: insights into plant defence

and regeneration patterns. *Annals of Botany*, *112*, 643-650.

Briscoe Runquist, R. D., Gorton, A. J., Yoder, J. B., Deacon, N. J., Grossman, J. J., Kothari, S.,

… , Moeller, D. A. (2020). Context dependence of local adaptation to abiotic and biotic

environments: A quantitative and qualitative synthesis. *The American Naturalist*, *195*,

412-431.

Brooks, M. E., Kristensen, K., van Benthem, K. J., Magnusson, A., Berg, C. W., Nielsen, A., … ,

Bolker, B. M. (2014) glmmTMB balances speed and flexibility among packages for zero-

inflated generalized linear mixed modeling. *The R Journal*, *9*, 378-400.

Burdon, J. J. (1983). Trifolium repens L. *Journal of Ecology*, *71*, 307-330.

Coley, P. D., Bryant, J. P., Chapin, F. S. (1985). Resource availability and plant antiherbivore

defense. *Science*, *230*, 895-899.

Coley, P. D. (1991). A comparison of herbivory and plant defenses in temperate and tropical

broad-leaved forests. In Price, P. W., Lewinsohn, T. M., Fernandes, G. W., Benson, W.

W. (editors). Plant-animal interactions: Evolutionary ecology in tropical and temperate

regions. Wiley & Sons, New York, NY, USA.

Coley, P. D., Barone, J. A. (1996). Herbivory and plant defenses in tropical forests. *Annual*

*Review of Ecology and Systematics*, *27*, 305-335.

Corkill, L. D. (1940). Cyanogenesis in White Clover (*Trifolium repens* L.). I. Cyanogenesis in

Single Plants. *New Zealand Journal of Science and Technology, 22*, 65-67.

Daday, H. (1954a). Gene frequencies in wild populations of *Trifolium repens.* I. Distribution by

latitude. *Heredity*, *8*, 61-78.

Daday, H. (1954b). Gene frequencies in wild populations of *Trifolium repens.* II. Distribution by

altitude. *Heredity*, *8*, 377-384.

Daday, H. (1958). Gene frequencies in wild populations of *Trifolium repens.* III. World

distribution. *Heredity*, *12*, 169-184.

Daday, H. (1965). Gene frequencies in wild populations of *Trifolium repens.* IV. Mechanism of

natural selection. *Heredity*, *20*, 355-365.

Desroches, P., Mandon, N., Baehr, J. C., Huignard, J. (1997). Mediation of host-plant use by a

glucoside in *Callosobruchus maculatus* F. (Coleoptera: Bruchidae). *Journal of Insect*

*Physiology*, *43*, 439-446.

Dirzo, R., Harper, J. L. (1982). Experimental studies on slug-plant interactions: III. Differences

in the acceptability of individual plants of *Trifolium repens* to slugs and snails. *Journal of*

*Ecology*, *70*, 101-117.

Easlon, H. M., Bloom, A. J. (2014). Easy Leaf Area: Automated digital image analysis for rapid

and accurate measurement of leaf area. *Applications in Plant Sciences*, *2*, 1400033.

Endler, J. A. (1977). Geographic variation, speciation, and clines. Princeton University Press.

Fadoul, H. E., Albano, L. J., Bergman, M. E., Phillips, M. A., Johnson, M. T. J. (2023).

Assessing the benefits and costs of the hydrogen cyanide antiherbivore defense in

*Trifolium repens*. *Plants*, *12*, 1213.

Feigl, F., Anger, V. (1966). Replacement of benzadine by copper ethylacetoacetate and tetra base

as spot-test reagent for hydrogen cyanide and cyanogen. *Analyst*, 91, 282-284.

Ferris, K. G., Willis, J. H. (2018). Differential adaptation to a harsh granite outcrop habitat

between sympatric *Mimulus* species. *Evolution*, *72*, 1225-1241.

Foulds, W., Grime, J. P. (1972a). The response of cyanogenic and acyanogenic phenotypes of

*Trifolium repens* to soil moisture supply. *Heredity*, *28*, 181-187.

Foulds, W., Grime, J. P. (1972b). The influence of soil moisture on the frequency of cyanogenic

plants in populations of *Trifolium repens* and *Lotus corniculatus*. *Heredity*, *28*, 143-147.

Fox, J., Weisberg, S. (2019). An R companion to applied regression, Third edition. Sage,

Thousand Oaks, CA, USA.

Ganders, F. R. (1990). Altitudinal clines for cyanogenesis in introduced populations of white

clover near Vancouver, Canada. *Heredity*, *64*, 387-390.

Gillespie, J. H., Turelli, M. (1989). Genotype-environment interactions and the maintenance of

polygenic variation. *Genetics*, *121*, 129-138.

Grant, P. R., Grant, B. R., Huey, R. B., Johnson, M. T. J., Knoll, A. H., Schmitt, J. (2017).

Evolution cause by extreme events. *Philosophical Transactions of the Royal Society B:*

*Biological Sciences*, *372*, 20160146.

Hargreaves, A. L., Germain, R. M., Bontrager, M., Persi, J., Angert, A. L. (2020). Local

adaptation to biotic interactions: A meta-analysis across latitudes. *The American*

*Naturalist*, *195*, 395-411.

Hendrickson, B. T., Stamps, C., Patterson, C. M., Strickland, H., Foster, M., Albano, L., … ,

Kooyers, N. J. (2025). Evolution of drought resistant strategies following the introduction

of white clover (*Trifolium repens* L.). *Annals of Botany*, *in press*.

Hereford, J. (2009). A quantitative survey of local adaptation and fitness trade-offs. *The*

*American Naturalist*, *173*, 579-588.

Johnson, M. T. J., Rasmann, S. (2011). The latitudinal herbivory-defence hypothesis takes a

detour on the map. *New Phytologist*, *191*, 589-592.

Johnson, M. T. J., Bertrand, J. A., Turcotte, M. M. (2016). Precision and accuracy in quantifying

herbivory. *Ecological Entomology*, *41*, 112-121.

Kakes, P. (1989). An analysis of the costs and benefits of the cyanogenic system in *Trifolium*

*repens* L. *Theoretical and Applied Genetics*, *77*, 111-11.

Kawecki, T. J., Ebert, D. (2004). Conceptual issues in local adaptation. *Ecology Letters*, *7*, 1225-

1241.

Kjærgaard, T. (2003). A plant that changed the world: The rise and fall of clover 1000-2000.

*Landscape Research*, *28*, 41-49.

Kooyers, N. J., Olsen, K. M. (2012). Rapid evolution of an adaptive cyanogenesis cline in

introduced North American white clover (*Trifolium repens* L.). *Molecular Ecology*, *21*,

2455-2468.

Kooyers, N. J., Gage, L. R., Al-Lozi, A., Olsen, K. M. (2014). Aridity shapes cyanogenesis cline

evolution in white clover (*Trifolium repens* L.). *Molecular Ecology*, *23*, 1053-1070.

Kooyers, N. J., Blackman, B. K., Holeski, L. M. (2017). Optimal defense theory explains

deviations from latitudinal herbivory defense hypothesis. *Ecology*, *98*, 1036-1048.

Kooyers, N. J., Hartman Bakken, B., Ungerer, M. C., Olsen, K. M. (2018). Freeze-induced

cyanide toxicity does not maintain the cyanogenesis polymorphism in white clover

(*Trifolium repens*). *American Journal of Botany*, *105*, 1224-1231.

Kundel, D., Meyer, S., Birkhofer, H., Fliessbach, A., Mäder, P., Scheu, S., … , Birkhofer, K.

(2018). Design and manual to construct rainout-shelters from climate change experiments

in agroecosystems. *Frontiers in Environmental Science*, *6*, 14.

Kuo, W.-H., Zhong, L., Wright, S. J., Goad, D. M., Olsen, K. M. (2024a). Beyond cyanogenesis:

Temperature gradients drive environmental adaptation in North American white clover

(*Trifolium repens* L.). *Molecular Ecology*, *33*, e17484.

Kuo, W.-H., Wright, S. J., Small, L. L., Olsen, K. M. (2024b). De novo genome assembly of

white clover (*Trifolium repens* L.) reveals the role of copy number variation in rapid

environmental adaptation. *BMC Biology*, *22*, 165.

Lenth R (2024). emmeans: Estimated Marginal Means, aka Least-Squares Means. R package

version 1.10.4.

Macel, M., Lawson, C. S., Mortimer, S. R., Šmilauerova, M., Bischoff, A., Crémieux, L., … ,

Steinger, T. (2007). Climate vs. soil factors in local adaptation of two common plant

species. *Ecology*, *88*, 424-433.

Machingura, M., Salomon, E., Jez, J. M., Ebbs, S. D. (2016). The β-cyanoalanine synthase

pathway: beyond cyanide detoxification. *Plant, Cell and Environment*, *39*, 2329-2341.

Moles, A. T., Bonser, S. P., Poore, A. G. B., Wallis, I. R., Foley, W. J. (2011). Assessing the

evidence for latitudinal gradients in plant defence and herbivory. *Functional Ecology*, *25*,

380-388.

Møller, B. L. (2010). Functional diversifications of cyanogenic glucosides. *Current Opinion in*

*Plant Biology*, *13*, 338-347.

Moreira, X., Abdala-Roberts, L., Parra-Tabla, V., Mooney, K. A. (2015). Latitudinal variation in

herbivory: influences of climatic drivers, herbivore identity and natural enemies. *Oikos*,

*124*, 1444-1452.

Moreira, X., Castagneyrol, B., Abdala-Roberts, L., Berny-Mier y Teran, J. C., Timmermans, B.

G. H., Bruun, H. H., … , Tack, A. J. M. (2018). Latitudinal variation in plant chemical

defences drives latitudinal patterns of leaf herbivory. *Ecography*, *41*, 1124-1134.

Mundim, F. M., Pringle, E. G. (2018). Whole-plant metabolic allocation under water stress.

*Frontiers in Plant Science*, *9*, 852.

Nuismer, S. L., Thompson, J. N., Gomulkiewicz, R. (2000). Coevolutionary clines across

selection mosaics. *Evolution*, *54*, 1102-1115.

Olsen, K. M., Sutherland, B. L., Small, L. L. (2007). Molecular evolution of the *Li/li* chemical

defence polymorphism in white clover (*Trifolium repens* L.). *Molecular Ecology*, *16*,

4180-4193.

Olsen, K. M., Hsu, S.-C., Small, L. L. (2008). Evidence of the molecular basis of the *Ac/ac*

adaptive cyanogenesis polymorphism in white clover (*Trifolium repens* L.). *Genetics*,

*179*, 517-526.

Olsen, K. M., Ungerer, M. C. (2008). Freezing tolerance and cyanogenesis in white clover

(*Trifolium repens* L. Fabaceae). *International Journal of Plant Sciences*, *169*, 1141-1147.

Olsen, K. M., Small, L. L. (2018). Micro- and macroevolutionary adaptation through repeated

loss of a complete metabolic pathway. *New Phytologist*, *219*, 757-766.

Pankoke, H., Bowers, M. D., Dobler, S. (2012). The interplay between toxin-releasing β-

glucosidase and plant iridoid glycosides impairs larval development in a generalist

caterpillar, *Grammia incorrupta* (Arctiidae). *Insect Biochemistry and Molecular Biology*,

*42*, 426-434.

Poulton, J. E. (1990). Cyanogenesis in plants. *Plant Physiology*, *94*, 401-405.

Santangelo, J. S., Thompson, K. A., Johnson, M. T. J. (2019). Herbivores and plant defences

affect selection on plant reproductive traits more strongly than pollinators. *Journal of*

*Evolutionary Biology*, *32*, 4-18.

Scharnagl, A., Genung, M. A., Holeski, L. M., Kooyers, N. J. (2023). Life history and chemical

defense interact to drive patterns of adaptation in an annual monkeyflower. *Evolution*, *77*,

370-383.

Siepielski, A. M., Morrissey, M. B., Buoro, M., Carlson, S. M., Caruso, C. M., Clegg, S. M., … ,

MacColl, A. D. C. (2017). Precipitation drives global variation in natural selection.

*Science*, *355*, 959-962.

Tattersall, D. B., Bak, S., Jones, P. R., Olsen, C. E., Nielsen, J. K., Hansen, M. L., Høj, P. B.,

Møller, B. L. (2001). Resistance to an herbivore through engineered cyanogenic

glucoside synthesis. *Science*, *293*, 1826-1828.

Thompson, K. A., Renaudin, M., Johnson, M. T. J. (2016). Urbanization drives the evolution of

parallel clines in plant populations. *Proceedings of the Royal Society B: Biological*

*Sciences*, *283*, 20162180.

Warvell, M. V., Shaw, R. G. (2019). Phenotypic selection on ponderosa pine seed and seedling

traits in the field under three experimentally manipulated drought treatments.

*Evolutionary Applications*, *12*, 159-174.

Wright, S. J., Cui Zhou, D., Kuhle, A., Olsen, K. M. (2018). Continent-wide climatic variation

drives local adaptation in North American white clover. *Journal of Heredity*, *109*, 78-89.

Yadav, M., Singh, I. K., Singh, A. (2023). Dhurrin: A naturally occurring phytochemical as a

weapon against insect herbivores. *Phytochemistry*, *205*, 113483.

**Table 1** Results of mixed effects models assessing how cyanotype, manipulation of precipitation, manipulation of herbivore pressure,
and their two-way interactions affected the proportion of *T. repens* plants producing seeds and the seed set mass in Ontario and
Louisiana study sites in the 2021 and 2022 growing seasons. Custom contrasts were performed for significant main effects of
cyanotype to test the specific effects of cyanogenesis (AcLi vs. Acli + acLi + acli), presence of the dominant *Ac* allele (AcLi + Acli vs.
acLi + acli), and presence of the dominant *Li* allele (AcLi + acLi vs. Acli + acli) on the fitness of *T. repens*. Signs of > and < are with
respect to the first listed side of the custom contrast (cyanogenic, *Ac*, or *Li*). Significant effects ($P < 0.05$) are listed in bold.
812

| **Ontario Study Site** | | 2021 | | | | | | 2022 | | | | | |
| | Prop. Producing Seeds | | | Seed Set Mass | | | Prop. Producing Seeds | | | Seed Set Mass | | | |
| Predictor | $\chi^2$ | df | *P* | $\chi^2$ | df | *P* | $\chi^2$ | df | *P* | $\chi^2$ | df | *P* |
| Cyanotype (C) | 7.431 | 3 | 0.059 | **16.516** | **3** | **0.001** | **15.635** | **3** | **0.001** | **9.024** | **3** | **0.029** |
| Precipitation (P) | 0.051 | 1 | 0.822 | 0.266 | 1 | 0.606 | 0.893 | 1 | 0.345 | **6.070** | **1** | **0.014** |
| Herbivores (H) | 1.642 | 1 | 0.200 | 0.000 | 1 | 0.992 | **4.507** | **1** | **0.034** | 2.811 | 1 | 0.094 |
| C × P | 5.644 | 3 | 0.130 | 4.414 | 3 | 0.220 | **8.456** | **3** | **0.037** | 3.348 | 3 | 0.341 |
| C × H | 2.275 | 3 | 0.517 | 1.342 | 3 | 0.719 | 2.743 | 3 | 0.433 | 0.204 | 3 | 0.977 |
| P × H | 0.001 | 1 | 0.974 | 0.924 | 1 | 0.336 | 2.873 | 1 | 0.090 | 0.650 | 1 | 0.420 |

| Contrast | Sign | | *P* | Sign | | *P* | Sign | | *P* | Sign | | *P* |
| *Cyanotype* | | | | | | | | | | | | |
| Cyanogenic vs. Acyanogenic | — | | — | > | | **0.009** | = | | 0.062 | = | | 0.057 |
| *Ac* vs. *ac* | — | | — | > | | **0.002** | > | | **< 0.001** | > | | **0.026** |
| *Li* vs. *li* | — | | — | = | | 0.403 | = | | 0.170 | = | | 0.393 |

| **Louisiana Study Site** | | 2021 | | | | | | 2022 | | | | | |
| | Prop. Producing Seeds | | | Seed Set Mass | | | Prop. Producing Seeds | | | Seed Set Mass | | | |
| Predictor | $\chi^2$ | df | *P* | $\chi^2$ | df | *P* | $\chi^2$ | df | *P* | $\chi^2$ | df | *P* |
| Cyanotype (C) | **11.694** | **3** | **0.009** | 1.345 | 3 | 0.719 | 0.919 | 3 | 0.821 | **10.522** | **3** | **0.015** |
| Precipitation (P) | 2.386 | 1 | 0.122 | 0.074 | 1 | 0.786 | 0.790 | 1 | 0.374 | 0.155 | 1 | 0.694 |
| Herbivores (H) | **25.264** | **1** | **< 0.001** | **13.484** | **1** | **< 0.001** | 0.352 | 1 | 0.553 | 0.338 | 1 | 0.561 |
| C × P | 4.468 | 3 | 0.215 | 0.291 | 3 | 0.962 | 1.503 | 3 | 0.682 | 0.717 | 3 | 0.869 |
| C × H | 1.785 | 3 | 0.618 | **10.453** | **3** | **0.015** | 0.088 | 3 | 0.993 | 1.370 | 3 | 0.713 |
| P × H | 0.574 | 1 | 0.449 | 0.333 | 1 | 0.564 | 2.449 | 1 | 0.118 | 0.480 | 1 | 0.488 |

| Contrast | Sign | | *P* | Sign | | *P* | Sign | | *P* | Sign | | *P* |
| *Cyanotype* | | | | | | | | | | | | |
| Cyanogenic vs. Acyanogenic | > | | **0.009** | — | | — | — | | — | = | | 0.893 |
| *Ac* vs. *ac* | = | | 0.114 | — | | — | — | | — | > | | **0.023** |
| *Li* vs. *li* | > | | **0.001** | — | | — | — | | — | = | | 0.125 |

**Figure 1** Map identifying study sites, sampled *T. repens* populations, and experiment
predictions. The Ontario and Louisiana study sites are identified by green diamonds. The 50
sampled populations used to investigate the presence of a latitudinal cyanogenesis clines in
North American *T. repens* are identified by grey circles. Each study site is associated with
predictions of relative fitness of *T. repens* cyanotypes under natural conditions and when
exposed to precipitation and herbivore manipulations. In Ontario, costs of CNglcs and
linamarase are exacerbated in the cooler climate, selecting against Acli, acLi, and AcLi plants
under natural/control conditions. Precipitation manipulation reduces fitness to a lesser degree in
*Ac* plants due to nitrogen recycling advantages under drought. Herbivore manipulation
exacerbates fitness advantages of acyanogenic cyanotypes due to limited metabolic investment
into defense compounds when plants are already experimentally defended. In Louisiana, AcLi
are selected for under natural conditions due to greater herbivore pressure, while Acli and acLi
are selected against due to a lack of defense but increased investment in defense precursors.
Precipitation manipulation again reduces fitness to a lesser degree in *Ac* plants due to advantages
under drought. Herbivore manipulation again provides relative fitness advantages to acyanogenic
cyanotypes due to limited metabolic investment into defense compounds when plants are already
experimentally defended, while the relative costs of *Ac* and *Li* are diminished compared to the
colder, northern climate.

**Figure 2** Clines in cyanogenesis and herbivory in the North American range of *T. repens* (A)
The relationship between latitude and mean annual temperature for 50 sampling locations across
a north-south transect in North America. (B) The effect of mean annual temperature on the
proportion of cyanogenic *T. repens* individuals (N = 140) from 47 of the sampling locations. (C)
The effect of mean annual temperature on the mean percentage of leaf area consumed for 50
individuals (1 trifoliate leaf per individual) assessed at each of the 50 sampled locations.

839
**Figure 3** The effect of cyanotype on plant fitness. We show the proportion of *T. repens* plants of each cyanotype (acli, acLi, Acli,
AcLi) that produced seeds and their total seed mass in Ontario in 2021 (top left) and 2022 (top right) and in Louisiana in 2021 (bottom
left) and 2022 (bottom right). Error bars are ± one standard error for each mean value.

**Figure 4** Herbivore manipulation imposes differential selection on cyanotypes in Louisiana in
2021. We show the effect of cyanotype (acli, acLi, Acli, AcLi) and herbivore manipulation
(control vs. reduced) on seed set mass produced by *T. repens* plants at the Louisiana study site in
2021 and 2022. Error bars are ± one standard error for each mean value.
847

848    **Figure 5** Precipitation manipulation imposes differential selection on cyanotypes in Ontario in
849    2022. We show the effect of cyanotype (acli, acLi, Acli, AcLi) and precipitation manipulation
850    (control vs. reduced) on the proportion of *T. repens* plants that produced seeds at the Ontario
851    study site in 2021 and 2022. Error bars are ± one standard error for each mean value.
852

853    **Figure 6** The effect of cyanotype on herbivory. We show the percent of leaf area consumed for
854    *T. repens* cyanotypes at the Ontario (acli, acLi, Acli, AcLi) study site (top left) and the Louisiana
855    study site (top right) in the 2021 growing season and at the Ontario study site (bottom left) and
856    the Louisiana study site (bottom right) in the 2022 growing season. Error bars are ± one standard
857    error for each mean value.