Influence of Copper Dose on *Mycobacterium avium* and *Legionella*

pneumophila **Growth in Premise Plumbing**

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ABSTRACT

Copper can be released from pipes or added intentionally to drinking water as an antimicrobial,

possibly limiting biofilm and pathogen growth in building (i.e., premise) plumbing systems, but

longer-term effects of various doses are poorly defined. Here, triplicate 120-mL water heater

microcosms equipped with PEX-b pipes containing mature biofilms were fed influent dosed with

copper at 0, 4, 30, 250, or 2000 ppb (as total Cu) and monitored for 11 months. Effluent total cell

counts and *Mycobacterium avium* peaked at 250-ppb copper dosage, reflecting the dual role of

copper as a nutrient and antimicrobial. Total cell counts and *Mycobacterium avium* were

relatively consistent in replicate microcosms at each dose. While *Legionella pneumophila (Lp)*

responded similarly to copper and also peaked at 250 ppb, this dose also caused bifurcating

behavior of biological replicates, with one consistently producing high culturable *Lp* (average of

2.5 log (MPN/mL)) and the other two consistently producing low or non-detect levels over an

11-month dosing period. We cross-inoculated the microcosms ten times periodically throughout

the 6-month acclimation phase before copper dosing, and reinoculated the microcosms with *Lp*

one time ~ 3.5 months into dosing in an attempt to normalize the microbial community

composition across replicate microcosms. However, *Lp* levels across the three 250-ppb

microcosms remained bifurcated throughout the copper dosing period. 16S rRNA gene

sequencing revealed that the 250-ppb replicate with high *Lp* was characterized by a distinct

microbial community composition relative to the two replicates with low *Lp*. At the highest dose

of 2,000 ppb copper total cell counts initially decreased by about 78%, but then steadily

increased to a point where they were not statistically different from the 250-ppb microcosms.

Dosing 2000 ppb copper also reduced microbial diversity.

SYNOPSIS

A dose of 250 ppb copper caused peak growth of *Mycobacterium avium* and *Legionella*

pneumophila in drinking water microcosms simulating hot water plumbing lines.

INTRODUCTION
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Copper can be released into drinking water from corrosion of brass or copper pipe or can

be intentionally dosed to control microbial growth in premise (i.e., building) plumbing.[1]

Evidence suggests that copper can act as a nutrient to drinking water microbes at low

concentrations and an antimicrobial at high concentrations,[2] suggesting an optimal range for

growth of pathogens. Efficacy of copper as an antimicrobial can be dependent on water

chemistry, physiology of the target microbes, biofilm characteristics, water flow patterns, levels

of other disinfectants (e.g., chlorine), and other factors.[2,3]

Mycobacterium avium and *Legionella pneumophila* (*Lp*) are two key opportunistic

pathogens of concern that are prone to growth in premise plumbing.[2] Prior field surveys of *Lp*

occurrence in drinking water suggest antimicrobial thresholds for copper of > 50 μg/L,[4] > 400

μg/L,[5] or > 1055 μg/L,[6] whereas other studies have reported contradictory results suggesting

positive effects on growth at concentrations > 500 μg/L[7], or *Lp* persistence at concentrations >

1170 μg /L[8]. *M. avium* has been shown to be inactivated by high levels of copper in warm

premise plumbing environments,[9,10] but *M. avium* and other nontuberculous mycobacteria

(NTM) are generally considered more resilient to copper than *Lp*.[10,11] These prior studies suggest

that copper, if present at sufficiently high concentrations, could be an effective means to

simultaneously control both NTM and *Lp,* a goal that has remained elusive given that many

existing drinking water control strategies (such as chloramination or silver ions), dependent upon

disinfection concentration, may have no impact or even favor the growth of the other.[11,12]

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To gain insight into the effects of copper on *Lp* and NTM control in real-world simulations

of hot water plumbing systems, we conducted a series of studies using local Blacksburg

municipal tap water at both large scale pilot (spanning 3 years) and microcosm scale (spanning

198 days).[9,13,14] These studies demonstrated how higher rates of anode rod corrosion, pH shifts,

and hydrogen evolution can confound antimicrobial action of copper. Similar confounding

effects were likely at play but not considered in the interpretation of results derived from prior

field-scale research.[2] Comparatively, another relatively long-term study (12 months)

investigating opportunistic pathogens in water heaters was only able to detect NTM, and not

Legionella spp., *Lp*, or *P. aeruginosa* at any point in the experiment,[15] highlighting challenges in

simultaneous study of *Lp* and NTM at pilot scale, especially over extended time periods. In

another prior microcosm-scale study, the effect of copper dosed 0.8 – 5 mg/L total Cu to assess

the culturability of *Lp* over a 4-week period using synthetic water to mimic tap water at 36°C.[16]

That work found that different *Lp* strains reacted differently to copper. Furthermore,

environmental isolates derived from biofilm in a hot water system exhibited greater resistance to

copper compared to the clinical strains tested, suggesting that the environmental isolates (which

had a significantly higher expression of the copper resistance gene copA) were better adapted to

copper.[16] In a 76 week microcosm-scale study examining the impact of different materials

(including pipe coupons) on *Lp* and NTM, copper coupons released high concentrations of

copper (≈10 mg/L) when heated to 55 and 60°C, which in tandem, strongly suppressed

established NTM and *Lp* concentrations in the microcosms.[17]

The objective of this study was to assess the effects of a wide range of copper

concentrations on *M. avium* and *Lp* in premise plumbing systems colonized by both organisms.

Fifteen 120-mL water heater microcosms equipped with PEX-b pipes containing mature biofilms

(>3.5 years old and carrying both *Lp* and *M.* avium)[9] were acclimated to Blacksburg tap water

over a 6-month acclimation period and subsequently divided into biological triplicate

microcosms subject to copper dosing at 0, 4, 30, 250, or 2,000 ppb (as total copper) for 11-

months. Microbial total cell counts (TCC) were monitored via flow cytometry, *M. avium* was

measured via droplet digital polymerase chain reaction (ddPCR), and *Lp* was measured both via

ddPCR and Legiolert™. 16S rRNA gene sequencing was applied to profile microbial

communities and provide insight into microbial ecological factors mediating the effects of

copper. The goal was to assess doses at which copper was a nutrient and doses where it was

antimicrobial for opportunistic pathogens in warm water premise plumbing.

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98   **MATERIALS AND METHODS**
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100 *Premise Plumbing Microcosm Design and Establishment*
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Fifteen 120-mL glass microcosms, referred to as "simulated glass water heaters" in

previous works,[18] were employed in this study to provide replicable simulation of premise

plumbing with mature drinking water biofilm. Each replicate microcosm contained two new and

two conditioned PEX-b pipe coupons, with the conditioned coupons cut from the middle of ~4.6-

meter recirculating lines of pilot-scale hot water plumbing systems, one that was previously

dosed with phosphate and one that received no added phosphate.[9,13] All coupons were 2.5 cm

long with a ~ 1.7 cm inner diameter. The pilot-scale hot water plumbing systems had been

inoculated with two strains of *Lp* serogroup-1 that were isolated from the Quincy, Illinois

Veterans Home during a Legionnaires' Disease outbreak.[19] Prior to this experiment, the

recirculation line biofilms were confirmed to be colonized by both *M. avium* and *Lp*,[9] serving as

an inoculum to the microcosms.

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114 *Microcosm Influent & Water Changes: Acclimation Phase*
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The water heater microcosms were filled with 110 mL of local tap water (Blacksburg, VA)

that had been further treated by granular activated carbon (GAC), which was the same water fed

to the pilot-scale rig from which the pipe coupons were originally extracted. The influent water

was collected in 5-week intervals and stored at 4°C until the day before the water change. On the

day before a water change, 2L of the water was collected and heated to 37°C overnight then

adjusted to pH 6.70 $\pm$ 0.05 using H_2SO_4, prior to being used for the microcosm water changes.

Throughout the study, the microcosm bulk water was changed 2× weekly, by replacing 75% of

the bulk water volume (~82 mL) in each microcosm with new influent, to match the 3.2-day

hydraulic retention time of the pilot-scale plumbing system. 3.2-days was selected to simulate

common residential hot water usage in the United States.[13] The microcosms were maintained at

37°C, representing the low end of a water heater set point, and gently mixed on a shaker table at

100 rpm. The microcosms were acclimated for 6 months to these conditions prior to commencing

the copper dosing.

Cross-Inoculation of the Microcosms

A month into the acclimation phase, the effluent from each microcosm was cross

inoculated with the aim of establishing baseline *Lp* and microbial populations across the 15

replicates. This was achieved by collecting the effluent (75% bulk water volume) from the

microcosms into a common reservoir (an autoclaved 2-L polypropylene bottle), vigorously

shaking this mixture, returning 25% of the combined effluent to the microcosms, and replacing

the remaining 50% with the GAC-treated water. This procedure was repeated over the course of

ten sequential water changes, starting a month into the acclimation phase, for five weeks

thereafter.

Copper Dosing and Reinoculation

Nearing the end of the 6-month acclimation phase, a week before the copper dosing

commenced, the microcosms were sampled for culturable *Lp* levels measured by Legiolert[TM]

tests (IDEXX Laboratories, Westbrook, USA) and TCCs. Based on these measurements, the

microcosms were grouped into five sets of triplicate microcosms, so that there was no statistical

difference in mean values of TCC and *Lp* among experimental conditions (ANOVA, $p > 0.05$)

(Supplemental Figure 1). Triplicate microcosms subsequently received one of five copper

dosages of 0, 4, 30, 250, or 2000 ppb copper over an 11-month period. Copper was dosed

directly to the influent as total Cu using stock solutions prepared with $CuSO_4 \cdot 5H_2O$. Five

separate 500 mL glass containers were used for dosing: one contained 280 mL of GAC-treated

influent for the 0 ppb Cu control, while the 4, 30, and 250 ppb conditions were prepared by

diluting a 25 mg/L total Cu stock solution with GAC-treated influent to achieve the target

concentrations, resulting in 280 mL of copper-dosed influent to use per condition. The 2000 ppb

condition was prepared by mixing GAC-treated influent with a small volume of a 200 mg/L Cu

stock solution. Stock solutions (25 mg/L and 200 mg/L) were verified using a DR3900

Spectrophotometer (Hach, Loveland, USA) prior to each water change, and dosing was adjusted

accordingly based on spectrophotometric readings. After ~ 3.5 months of copper dosing, random

testing among the microcosms indicated that *Lp* still had not established in some replicates and,

therefore, all microcosms were reinoculated once more. This was achieved using water from the

same pilot-scale water heater from which the microcosm pipe coupons were originally derived.

The inoculation water was mixed with the normal influent water to target ~50 MPN/mL

culturable *Lp* in each microcosm.

The influent water preparation protocol underwent various adjustments, throughout the 6-

month acclimation (months 0-6 overall) and 11 month copper dosing (months 6-17 overall)

phases of the experiment, in an attempt to establish a replicable response of *Lp* to copper.

Modifications included: 1) passing GAC-filtered water through a ferric oxide filter to decrease

phosphate levels to < 0.05 mg/L followed by filter-sterilizing the water using a 0.22-μm pore

size mixed nitrocellulose ester membrane (Whatman, Maidstone, United Kingdom) throughout

the acclimation phase and for the first two months of the copper dosing phase (months 0-8), 2)

dosing GAC-treated influent with phosphate throughout the copper dosing phase to achieve 5

μg/L as P (months 6-17), 3) no longer sterilizing the influent or treating it with a ferric oxide

filter after 8 weeks into the copper dosing phase of the experiment (months 8-17), and 4) dosing

ferric pyrophosphate (100 μg as Fe) after week 35 of the copper dosing phase to provide

bioavailable iron (months 14-17).

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173   *Microbial Analysis*
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Effluent collected from the microcosms during bi-weekly water changes was subject to

microbial analysis. TCC was measured on a BD Accuri C6 (BD Bioscience, Franklin Lakes,

USA) using SYBR Green I dye to stain total (intact + damaged) cells.[20,21,22] Gating used the

Eawag FL1-A (emission filter 533/30) vs FL3-A (emission filter 670 LP) template for drinking

water, with 50 μL of each sample analyzed at medium speed (35 μL/min) and with an acquisition

threshold set to 800.[22] Legiolert[TM] was used to enumerate culturable *Lp*, subjecting 1.0 mL of

sample bulk water to the nonpotable procedure per the manufacturer's instructions.

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During regular water changes, ~82 mL of influent or effluent water from each microcosm

was filtered through a 0.22-μm mixed-cellulose ester membrane filter (MilliporeSigma,

Burlington, USA) and the filter was subject to DNA extraction using a FastDNA SPIN kit (MP

Biomedicals Inc., Solon, USA). DNA extracts were analyzed for opportunistic pathogens using a

QX200 ddPCR (Bio-Rad, Hercules, USA) (Supplemental Table 1). These assays had previously

been optimized for qPCR[23] and were adapted for ddPCR in this study. Thermal gradients, ranging

from 50-60°C, were conducted on each assay to determine the optimal annealing temperature

(Supplemental Table 1) to optimally separate negative and positive droplets. Samples for all

targets were analyzed in triplicate and deemed as positive if at least 3 positive droplets were

found per reaction.[24]

16S rRNA Sequencing and Data Analysis

To analyze the microbial composition of each microcosm, 16S rRNA sequencing was

carried out on the DNA extracts. Amplicon sequence reads targeted the V4-V5 hypervariable

region of the 16S rRNA gene using the 515f/926r primer set. Samples were sequenced on a

MiSeq V3 600 cycle run, yielding an average of 119,910 reads per sample (SD = 69,156). Reads

and all subsequent analysis were carried out using DADA2 (v1.18.0)[25] in R (v 4.0.3).[26] Reads

were filtered, trimmed, and paired before exact amplicon sequence variants (ASVs) were

generated and chimeras removed. Taxonomy of ASVs was assigned using RDP Train Set 18.[27]

ASVs were rarefied to the minimum sequencing depth of 10,544.[28] Permutational multivariate

analysis of variance (Permanova) using Bray-Curtis distance matrices was conducted using the

vegan R package (v 2.6.4).[29]

Data Availability

The 16S rRNA Sequencing gathered from the microcosms in this experiment has been

uploaded to the National Center for Biotechnology Information Sequence Read Archive database

under Accession number PRJNA1229645.

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210   **RESULTS**
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213   *Total Cell Counts*
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The effluent mean TCC for all the microcosms during the end of the acclimation phase

(month 6), before copper addition, was 3.18×10^5 events/mL (SD = 4.16×10^3) (Figure 1,

Supplemental Figure 1b). In the short-term after copper dosing began (1-3.5 months post copper

dosing), the total cell counts indicated optimal growth at 250-ppb, and lower growth at lower and

higher copper levels. As the experiment progressed, the three replicates dosed with 250 ppb

copper consistently produced the highest cell counts. Replicates dosed with 2000 ppb copper

initially had a 0.66 log reduction (a 78% decrease from its final pre-dosing value), causing the

lowest cell counts observed in the study. But by the end of the 11- month experiment the mean

cell counts at 2000 ppb copper gradually recovered and eventually surpassed those at doses ≤ 30

ppb and were not significantly different than at 250 ppb (Tukey's HSD p<.05) (Figure 1,

Supplemental Table 2). In fact, the TCC steadily increased with time in the 2000 ppb

microcosms both pre-and post-re-inoculation attempt 3.5 months into the copper dosing phase

(Figure 1, Supplemental Figure 2).

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Figure 1. Mean Total Cell Counts (events/mL) in microcosm bulk water effluent over the course
of the experiment. Error bars are the standard deviation of biological triplicate microcosms.
Lines connect the data points to guide comparison and are colored by copper concentration. The
first measurement (time -0.25) occurred at the end of the acclimation phase, before copper
dosing began (time = 0) represented by the red dashed vertical line. The black dashed vertical
line at ~ 3.5 months represents the point at which all reactors were reinoculated with *Lp*. Time
points have a set jitter for visual purposes.

Mycobacterium avium response to copper in the microcosms

When tested over three sequential samplings in the final (11th) month of the experiment,

significant differences in *M. avium* gene copies/mL were found in the microcosm bulk water as a

function of copper concentration (Linear Mixed-Effect Model ANOVA, $p < 0.05$) (Figure 2). The

250-ppb condition had higher concentrations of *M. avium* compared to other copper dosages

(0.55 – 0.85 logs higher) (Tukey's HSD, $p < 0.05$, Supplementary Table 3). *M. avium* was not

detectable in the influent water using ddPCR so the increase in the effluent represents growth in

the microcosms.

Figure 2. Mean gene copies/mL counts of (A) *M. avium* detected through ddPCR and (B) *Lp*
detected via ddPCR and Legiolert™ across the fifteen microcosms over three consecutive
samplings in the final (11th) month of the experiment.

L. pneumophila response to copper in the microcosms

Lp response to copper dosing was confirmed both by ddPCR and Legiolert[TM]. On dates

when both Legiolert[TM] and ddPCR tests were conducted on aliquots of the same sample (n=45),

results were strongly correlated ($R^2 = 0.86$) (t-test, $p<0.05$), but the log ddPCR gc/mL values

were typically 0.51 times the log of Legiolert[TM] MPN/mL measurements (Supplemental Figure

3). *Lp* concentrations did not significantly differ across the microcosms as a function of copper

condition (Linear Mixed-Effect Model ANOVA, $p > 0.05$), but both Legiolert[TM] and ddPCR

measurements indicated a consistent trend ranking as follows: 2000 ppb < 4 ppb < 30 ppb < 0

ppb < 250 ppb. *Lp* was never detectable in the influent used for the microcosm water changes

using ddPCR or Legiolert[TM] so the increase in the effluent represents growth in the microcosm. It

was noted that there was a much wider variance in *Lp* numbers within replicates than there were

for *M. avium* (Figure 2). This was most evident for microcosms dosed with 250 ppb copper, for

which one microcosm consistently displayed the highest concentrations of culturable *Lp*, while

the second and third replicates produced non-detectable levels of culturable *Lp* for 87% and 73%

of the sampling events over the course of the copper-dosing phase of the experiment (Figure 3).

Lp levels quantified via Legiolert[TM] ranged from 8 MPN/mL to 1460 MPN/mL for the first

replicate microcosm in the 250-ppb Cu condition (Figure 3). The stark differences between these

microcosms were maintained even though they had been cross-inoculated ten times during

acclimation to establish the same baseline microbial community prior to copper dosing as well as

in addition to a post-copper dosing reinoculation event ~ 3.5 months into the copper dosing

phase and four adjustments to the water chemistry.

Figure 3. *Lp* MPN/mL, determined via Legiolert™, among the three replicate microcosms dosed with 250 ppb copper that displayed divergent behavior over the 11-month experimental period. The vertical, dashed line at ~3.5 months represents the point at which all microcosms were reinoculated with *Lp*. Time 0, shown with the red dashed vertical line represents when the copper dosing began. Points before 0 months represents culturable *Lp* in the microcosms prior to copper dosing, during the acclimation phase. Culturable *Lp* concentration measured with time for the microcosms receiving other copper concentrations can be found in Supplemental Figure 4. Legiolert™ detection limit was 1 MPN/mL.

Consistent with the Legiolert™ results, replicates 2 and 3 of the 250-ppb copper

condition consistently yielded undetectable gene copies/mL of *Lp*, while up to 22 gene

copies/mL were measured in replicate 1 (Figure 2).

16S rRNA Sequencing

16S rRNA sequencing allowed comparison of the microbial community profiles across the

microcosms and influent water (Figure 4). The microbial communities were significantly

different from each other as a function of copper concentration (ANOSIM, p < 0.05) (Figure 5).

The ten most abundant phyla from each replicate microcosm produced a similar profile for

copper doses of 0, 4, and 30 ppb, with a marked shift in composition in the 250 ppb and 2000

ppb conditions (Figure 4). Of note, the microcosms dosed with 0, 4, and 30 ppb copper were

mostly comprised of *Proteobacteria*, while the reactors dosed with 250 ppb copper contained

relatively equal proportions of *Proteobacteria* and *Gemmatimonadetes*. Microcosms dosed with

2000 ppb were primarily dominated by *Bacteriodetes* and *Proteobacteria* (Figure 4). Within

each condition from 0-250 ppb, the microbial communities were relatively consistent among the

three replicates. However, there was marked variation in the microbial community among the

2000 ppb condition replicates (Figure 5). Copper concentration had a significant effect on

Shannon diversity and Simpson diversity as measured by analysis of variance (ANOVA, $p <$

0.0001) (Figure 6).

The microbial community profile of the 250-ppb microcosm that consistently had the

highest level of *Lp* was dominated by the same top ten phyla, but differences in relative

abundance stood out (Figure 4, highlighted in orange boxes). Specifically, this microcosm was

statistically enriched in *Ignavibacteriae* and depleted in *Chlamydiae* and *Bacteroidetes* (Wald's

test, $p<0.05$) (Figure 4, highlighted in orange boxes). Beta diversity analysis, however, did not

reveal an obvious difference in the microbial community composition of this replicate beyond

the variation that was seen for replicates of other copper concentrations (Figure 5).

Figure 4. Relative abundances of each of the top ten most abundant phyla over all ASVs in each
replicate microcosm grouped by copper concentration and sampling date. The orange boxes
indicate the 250-ppb copper replicate that consistently displayed the highest concentrations of
Lp. The influent [collected at the start of month 10 for water changes (n=1)] represents the water
used throughout three sequential sampling events during water changes carried out in the final
(11th) month of the experiment. 11-A/B/C, represent the first (A), second (B), and third (C)
sequential water changes.

325
326 **Figure 5.** Bray-Curtis-based nonmetric multidimensional scaling of taxa from effluent water of
327 microcosms dosed with 0, 4, 30, 250, and 2000 ppb of copper over an 11-month period (n=3
328 (Stress = 0.08836)

333
334 **Figure 6.** Shannon and Simpson diversity index of replicate microcosms on the final three
335 sampling dates grouped by copper condition (n = 3 microcosms × 3 sampling events = 9 for each
336 copper condition).
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Relative abundances of individual classes of bacteria were further compared across

microcosm conditions revealing some notable differences as a function of copper concentration

(Figure 7). For example, *Proteobacteria* was higher in the 2000 ppb than in the 250 ppb

condition, while *Firmicutes* indicated little change across copper dosages (Figure 7).

Gemmatimonadetes were enriched from 0 to 250 ppb, but dramatically decreased at 2000 ppb

(Figure 7). *Deinococcus-Thermus* indicated a slight decrease in relative abundance from 250 ppb

to 2000 ppb (Figure 7). *Chloroflexi*, *Acidobacteria*, and *Actinobacteria* markedly decreased with

increasing copper (Figure 7).

347
348 **Figure 7.** Relative abundance of select phyla across copper dosages in each replicate over the
349 three sequential sampling events carried out in the final (11th) month of the experiment.
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351 **DISCUSSION**
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This study provided new insights into the effects of copper in premise plumbing as a

possible nutrient and antimicrobial. TCCs provided an indicator of broad-scale effects on

bacterial communities at-large, revealing some surprising long-term patterns. While TCCs

initially decreased markedly when the highest dose of 2000 ppb was added as expected, over the

course of subsequent months, the microbial community adapted such that TCCs were as high as

250 ppb by the end of the 11-month experiment. The rise in total cell counts at 2000 ppb did not

plateau by 11 months, suggesting it might eventually rise above the 250 ppb dose after years or

decades in premise plumbing.

Interestingly, TCCs initially increased in the 250-ppb condition and remained elevated

throughout the experiment relative to the un-dosed control, suggesting that copper acted largely

as a nutrient at this concentration. Doses of 4 and 30 ppb Cu were initially associated with a

significant decrease in TCCs (paired t-test, $p < 0.05$), but remained comparable to the control for

the majority of the experiment.

M. avium concentrations, which were only measured in the 11th month, were also highest in

the 250-ppb condition, but were lowest in the 2000 ppb condition. This suggests that very high

doses of copper are likely needed to effectively control *M. avium*, and lower doses serve as a

nutrient or selector for the organism. Overall, the findings for TCC and *M. avium* were consistent

with expectations that copper has a dual role of nutrient and antimicrobial,[2] or that a level of

copper that is antimicrobial in the short-term can behave as a nutrient over a period of months to

years as microbial resistance to copper develops.[16]

Support for this hypothesis was provided in the microbial community analysis. Even

though TCCs were elevated in the 250 ppb and 2000 ppb conditions, microbial diversity was

diminished. As richness and evenness dropped, a subset of the remaining species asserted

dominance over the community (Figures 4 and 6). With the significant decrease in diversity, a

few taxa began dominating the community, causing stochasticity as per the wide variance in

microbial community composition among the 2000 ppb replicates (Figure 5). Prior research

using the same municipal tap water supply also assessed the effect of dosed copper on the

microbial community in a full-scale water system, but the experimental design differed from the

present study in that the dose was incrementally increased from 0, 50, 100, 300, 600, 1200, to

2000 ppb, while maintaining each concentration for only periods of 4 to 16 weeks.[9] Song and

colleagues observed similar trends as those reported in this work for *Proteobacteria*,

Deinococcus-Thermus, *Chloroflexi*, *Acidobacteria*, and *Actinobacteria*. However, Song observed

enrichment over 6 weeks at 1200 ppb in *Firmicutes*, and enrichment of *Gemmatimonadetes* up

through 2000 ppb. The differences in *Firmicutes* and *Gemmatimonadetes* behavior when

comparing the results of this study could potentially be attributed to differences in duration of

copper dosing, as Song and colleagues maintained the high copper dose for 12 weeks,[9] whereas

the microcosms in this study were acclimated at this dose for 11 months.

At 250 ppb, stochasticity was observed in the concentrations of culturable Lp, which was

relatively consistent in all replicates at lower copper doses (0, 4, and 30 ppb total Cu) and was

almost always undetectable at 2000 ppb copper (Supplementary Figure 4). In a prior study

investigating the effect of copper pipe in warm premise plumbing systems, Lp numbers were

initially lowest in the biofilms on copper pipes, but eventually increased to the same

concentrations as PEX and stainless-steel pipes after two years.[30] In such studies, it is important

to note that copper pipe aging leads to decreasing release of copper into solution, and thus it is

difficult to know if the observed increase in Lp is due to lower levels of copper or microbial

adaptation.

A synthesis of our results over the last decade using Blacksburg tap water reveals two

general responses of experimental plumbing systems with respect to Lp inoculation. As one

would expect, there are many circumstances in which the system response to inoculated Lp is

deterministic, in which replicate reactors behave as true replicates. But here, we report a situation

in which stochastic behavior clearly occurred, because in some replicate reactors Lp consistently

died off and in other reactors it thrived. With benefit of hindsight, we now recognize similar

stochastic behavior for Lp amongst replicate reactors during at least two prior experiments in our

lab.[31,32] We were reluctant to accept the conclusion that these systems were stochastic, preferring

to believe there were subtle uncontrolled differences between replicates, but we consider the

evidence gathered herein to be conclusive. This study demonstrates that even in a relatively

simple simulation of premise plumbing, such as a glass microcosm containing pipe sections, that

there are complexities at play that can determine if effluent *Lp* is extremely high or very low.

Perhaps even more interestingly, the replicate reactors described herein were deterministic

and replicable for measured dimensions of cell counts, mycobacteria and community analysis,

even as they were stochastic for *Lp*. We speculate that a specific biological factor profoundly

affects the predator, prey and parasitic relationships of the *Lp* life cycle. If such stochastic

behavior for *Lp* is commonplace, as it is in certain other areas of microbial ecology,[33,34] this

discovery has profound implications for experimental design and the interpretation of past data

from laboratory and field sampling.[35]

ASSOCIATED CONTENT
Supporting Information

Information regarding ddPCR assays used, statistical analysis (on the total cell count and *M.*

avium.), initial *Lp* and TCC of the microcosm groups at the end of the acclimation phase,

additional total cell count information for the 2000 ppb Cu condition, a comparison between

Legiolert[TM] and ddPCR *Lp* results, and the culturable *Lp* for all the copper conditions throughout

the study, can be found in the attached supplemental document (PDF).

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NOTES
The authors declare no conflict of interest.

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REFERENCES

1. LeChevallier, M. Examining the Efficacy of Copper-Silver Ionization for Management of Legionella: Recommendations for Optimal Use. *AWWA Water Science* **2023**, *5* (2). DOI:10.1002/aws2.1327.

2. Cullom, A. C.; Martin, R. L.; Song, Y.; Williams, K.; Williams, A.; Pruden, A.; Edwards, M. A. Critical Review: Propensity of Premise Plumbing Pipe Materials to Enhance or Diminish Growth of Legionella and Other Opportunistic Pathogens. Pathogens **2020**, 9 (11), 957. DOI:10.3390/pathogens9110957.

3. Zhang, Y. Nitrification in Premise Plumbing and Its Effect on Corrosion and Water Quality Degradation; Ph.D. Dissertation, Virginia Tech, **2008**. https://vtechworks.lib.vt.edu/items/d57ec849-d055-4852-b34e-744a636f46ce

4. Bargellini, A.; Marchesi, I.; Righi, E.; Ferrari, A.; Cencetti, S.; Borella, P.; Rovesti, S. Parameters Predictive of Legionella Contamination in Hot Water Systems: Association

with Trace Elements and Heterotrophic Plate Counts. Water Research **2011**, 45 (6), 2315–2321. DOI:10.1016/j.watres.2011.01.009.

5. Liu, Z.; Stout, J. E.; Tedesco, L.; Boldin, M.; Hwang, C.; Diven, W. F.; Yu, V. L. Controlled Evaluation of Copper-Silver Ionization in Eradicating Legionella Pneumophila from a Hospital Water Distribution System. Journal of Infectious Diseases **1994**, 169 (4), 919–922. DOI:10.1093/infdis/169.4.919.

6. Miuetzner, S.; Schwille, R. C.; Farley, A.; Wald, E. R.; Ge, J. H.; States, S. J.; Libert, T.; Wadowsky, R. M. Efficacy of Thermal Treatment and Copper-Silver Ionization for Controlling Legionella Pneumophila in High-Volume Hot Water Plumbing Systems in Hospitals. American Journal of Infection Control **1997**, 25 (6), 452–457. DOI:10.1016/s0196-6553(97)90066-3.

7. Mathys, W.; Stanke, J.; Harmuth, M.; Junge-Mathys, E. Occurrence of Legionella in Hot Water Systems of Single-Family Residences in Suburbs of Two German Cities with Special Reference to Solar and District Heating. International Journal of Hygiene and Environmental Health **2008**, 211 (1–2), 179–185. DOI:10.1016/j.ijheh.2007.02.004.

8. Mathys, W.; Palma Hohmann, C.; Junge-Mathys, E. Efficacy of Copper-Silver Ionization in Controlling Legionella in a Hospital Hot Water Distribution System: A German Experience. Wiley Online Library **2001**. DOI:10.1128/9781555817985.ch84.

9. Song, Y.; Finkelstein, R.; Rhoads, W.; Edwards, M. A.; Pruden, A. Shotgun Metagenomics Reveals Impacts of Copper and Water Heater Anodes on Pathogens and Microbiomes in Hot Water Plumbing Systems. Environmental Science & Technology **2023**, 57 (36), 13612–13624. DOI:10.1021/acs.est.3c03568.

10. Lin, Y. E.; Vidic, R. D.; Stout, J. E.; McCartney, C. A.; Yu, V. L. Inactivation of Mycobacterium Avium by Copper and Silver Ions. Water Research **1998**, 32 (7), 1997–2000. DOI:10.1016/s0043-1354(97)00460-0.

11. Kusnetsov, J.; Iivanainen, E.; Elomaa, N.; Zacheus, O.; Martikainen, P. J. Copper and Silver Ions More Effective against Legionellae than against Mycobacteria in a Hospital Warm Water System. Water Research **2001**, 35 (17), 4217–4225. DOI:10.1016/s0043-1354(01)00124-5.

12. Donohue, M. J.; Vesper, S.; Mistry, J.; Donohue, J. M. Impact of Chlorine and Chloramine on the Detection and Quantification of Legionella Pneumophila and Mycobacterium Species. Applied and Environmental Microbiology **2019**, 85 (24). DOI:10.1128/aem.01942-19.

13. Song, Y.; Pruden, A.; Rhoads, W. J.; Edwards, M. A. Pilot-Scale Assessment Reveals Effects of Anode Type and Orthophosphate in Governing Antimicrobial Capacity of Copper for Legionella Pneumophila Control. Water Research **2023**, 242, 120178. DOI:10.1016/j.watres.2023.120178.

14. Rhoads, W. J.; Pruden, A.; Edwards, M. A. Interactive Effects of Corrosion, Copper, and Chloramines on Legionella and Mycobacteria in Hot Water Plumbing. Environmental Science & Technology **2017**, 51 (12), 7065–7075. DOI:10.1021/acs.est.6b05616.

15. Tolofari, D. L.; Bartrand, T.; Masters, S. V.; Duarte Batista, M.; Haas, C. N.; Olson, M.; Gurian, P. L. Influence of Hot Water Temperature and Use Patterns on Microbial Water Quality in Building Plumbing Systems. Environmental Engineering Science **2022**, 39 (4), 309–319. DOI:10.1089/ees.2021.0272.

16. Bédard, E.; Trigui, H.; Liang, J.; Doberva, M.; Paranjape, K.; Lalancette, C.; Allegra, S.; Faucher, S. P.; Prévost, M. Local Adaptation of Legionella Pneumophila within a Hospital Hot Water System Increases Tolerance to Copper. Applied and Environmental Microbiology **2021**, 87 (10). DOI:10.1128/aem.00242-21.

17. Cazals, M.; Bédard, E.; Faucher, S. P.; Prévost, M. Factors Affecting the Dynamics of Legionella Pneumophila, Nontuberculous Mycobacteria, and Their Host Vermamoeba Vermiformis in Premise Plumbing. ACS ES&T Water **2023**, 3 (12), 3874–3883. DOI:10.1021/acsestwater.3c00288.

18. Martin, R. L.; Strom, O.; Song, Y.; Mena-Aguilar, D.; Rhoads, W. J.; Pruden, A.; Edwards, M. A. Copper Pipe, Lack of Corrosion Control, and Very Low pH May Have Influenced the Trajectory of the Flint Legionnaires' Disease Outbreak. ACS ES&T Water **2022**, 2 (8), 1440–1450. DOI:10.1021/acsestwater.2c00182.

19. Rhoads, W. J.; Keane, T.; Spencer, M. S.; Pruden, A.; Edwards, M. A. Did Municipal Water Distribution System Deficiencies Contribute to a Legionnaires' Disease Outbreak in Quincy, IL? Environmental Science & Technology Letters **2020**, 7 (12), 896–902. DOI:10.1021/acs.estlett.0c00637.

20. Cazals, M.; Bédard, E.; Doberva, M.; Faucher, S.; Prévost, M. Compromised Effectiveness of Thermal Inactivation of Legionella Pneumophila in Water Heater Sediments and Water, and Influence of the Presence of Vermamoeba Vermiformis. Microorganisms **2022**, 10 (2), 443. DOI:10.3390/microorganisms10020443.

21. Nescerecka, A.; Hammes, F.; Juhna, T. A Pipeline for Developing and Testing Staining Protocols for Flow Cytometry, Demonstrated with SYBR Green I and Propidium Iodide Viability Staining. Journal of Microbiological Methods **2016**, 131, 172–180. DOI:10.1016/j.mimet.2016.10.022.

22. Singh, A.; Yelverton, C. J.; Barnard, T. G. Rapid Quantification of the Total Viable Bacterial Population on Human Hands Using Flow Cytometry with SYBR® Green I. Cytometry Part B: Clinical Cytometry **2019**, 96 (5), 397–403. DOI:10.1002/cyto.b.21776.

23. Wang, H.; Edwards, M.; Falkinham, J. O.; Pruden, A. Molecular Survey of the Occurrence of Legionella Spp., Mycobacterium Spp., Pseudomonas Aeruginosa, and Amoeba Hosts in Two Chloraminated Drinking Water Distribution Systems. Applied and Environmental Microbiology **2012**, 78 (17). DOI:10.1128/aem.01492-12.

24. Rački, N.; Morisset, D.; Gutierrez-Aguirre, I.; Ravnikar, M. One-Step RT-Droplet Digital PCR: A Breakthrough in the Quantification of Waterborne RNA Viruses. Analytical and Bioanalytical Chemistry **2013**, 406 (3), 661–667. DOI:10.1007/s00216-013-7476-y.

25. Callahan, B. J.; McMurdie, P. J.; Rosen, M. J.; Han, A. W.; Johnson, A. J.; Holmes, S. P. DADA2: High-Resolution Sample Inference from Illumina Amplicon Data. *Nature Methods* **2016**, *13* (7), 581–583. DOI:10.1038/nmeth.3869.

26. R Core Team. R: A Language and Environment for Statistical Computing; R Foundation for Statistical Computing: Vienna, Austria, **2021**. https://www.R-project.org/.

27. Callahan, B. RDP Taxonomic Training Data Formatted for DADA2 (RDP Trainset 18/Release 11.5). Zenodo (CERN European Organization for Nuclear Research) **2020**. DOI: 10.5281/zenodo.4310151.

28. Schloss, P. D. Rarefaction Is Currently the Best Approach to Control for Uneven Sequencing Effort in Amplicon Sequence Analyses. mSphere **2024**, 9 (2). DOI:10.1128/msphere.00354-23.

29. Oksanen, J.; Guillaume Blanchet, F.; Friendly, M.; Kindt, R.; Legendre, P.; McGlinn, D.; Minchin, P. R.; O'Hara, R. B.; Simpson, G. L.; Solymos, P.; Stevens, M. H. H.; Szoecs, E.; Wagner, H. vegan: Community Ecology Package; R package version 2.6-4, **2022**. https://CRAN.R-project.org/package=vegan

30. van der Kooij, D.; Veenendaal, H. R.; Scheffer, W. J. H. Biofilm Formation and Multiplication of Legionella in a Model Warm Water System with Pipes of Copper, Stainless Steel and Cross-Linked Polyethylene. Water Research **2005**, 39 (13), 2789–2798. DOI:10.1016/j.watres.2005.04.075.

31. Edwards, M.; Pruden, A.; Falkinham, J. O., III; Brazeau, R.; Williams, K.; Wang, H.; Martin, A.; Rhoads, W. Relationship Between Biodegradable Organic Matter and Pathogen Concentrations in Premise Plumbing. The Water Research Foundation, **2013**. https://www.waterrf.org/research/projects/relationship-between-biodegradable-organic-matter-and-pathogen-concentrations

32. Martin, R. L.; Harrison, K.; Proctor, C. R.; Martin, A.; Williams, K.; Pruden, A.; Edwards, M. A. Chlorine Disinfection of Legionella Spp., L. Pneumophila, and Acanthamoeba under Warm Water Premise Plumbing Conditions. Microorganisms **2020**, 8 (9), 1452. DOI:10.3390/microorganisms8091452.

33. Goss, P. J.; Peccoud, J. Quantitative Modeling of Stochastic Systems in Molecular Biology by Using Stochastic Petri Nets. Proceedings of the National Academy of Sciences 1998, 95 (12), 6750–6755. DOI:10.1073/pnas.95.12.6750.

34. De Vrieze, J.; De Mulder, T.; Matassa, S.; Zhou, J.; Angenent, L. T.; Boon, N.; Verstraete, W. Stochasticity in Microbiology: Managing Unpredictability to Reach the Sustainable Development Goals. Microbial Biotechnology **2020**, 13 (4), 829–843. DOI:10.1111/1751-7915.13575.

35. Zhou, J.; Ning, D. Stochastic Community Assembly: Does It Matter in Microbial Ecology? Microbiology and Molecular Biology Reviews 2017, 81 (4). DOI:10.1128/mmbr.00002-17.