25 Mar 2025

**Environmental and temporal factors affecting record white-tailed deer antler characteristics in**

**Ontario, Canada**

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**ABSTRACT**

White-tailed deer (*Odocoileus virginianus*) are an ecologically and economically important

species in North America. Their antlers, one of their most recognizable features, are used for dominance

displays, mate attraction, and defense, with size and morphology being key determinants of success.

Antler characteristics are influenced by a combination of genetic and environmental factors, including

habitat quality and resource availability. In this study, we explored how diverse environmental factors,

including climate and land cover composition, impact antler size, morphology, and the distribution of

record scoring white-tailed deer across Ontario, Canada, using data from long-term antler scoring

records. We used conditional autoregressive (CAR) models to examine these relationships and found

that warmer temperatures the year of harvest were positively associated with larger antlers and more

record deer in a given county, while winter precipitation the year of harvest was negatively associated

with these characteristics, likely due to reduced forage availability or increased energy expenditure

during more severe winters. Rangeland and forest land cover types were positively associated with

increased antler size and tine number. We observed no temporal changes in antler size in Ontario,

contrasting with broader trends observed in North America. These results show how local

environmental conditions and land cover composition influence antler traits and the distribution of

record white-tailed deer, highlighting the complexity of environmental influences on trait variation.

**KEYWORDS** white-tailed deer, *Odocoileus virginianus,* antler size, antler morphology, environmental

factors, Ontario, hunters, harvest, climate effects, land cover

The Cervidae (deer) family emerged during the late Miocene where they developed the

secondary sex characteristics that we associate with modern deer (Gilbert et al. 2006). One such

secondary sex characteristic, antlers, are bony extensions of the skull that generally shed every year, and

apart from caribou (*Rangifer tarandus*) are only found on males (Rue, 1989; Hewitt, 2011). Antlers

have multiple purposes including, dominance displays and interactions, use as weapons against

predators and other male Cervidae, and as indicators to females that they are of good genetic quality

(Clutton-Brock, 1982; Goss, 2012). In white-tailed deer, males are the dominant sex where social

dominance is largely dictated by age and body mass, both of which are closely tied to antler size

(Townsend and Bailey 1981). Fully matured males tend to have larger antlers due to the lower energy

demands for body growth and thus their ability to allocate resources to antlers (Goss, 1983). Likewise,

in red deer, larger antlers are associated with increased reproductive success (Kruuk et al. 2002). That

being said, younger or smaller-antlered males can achieve mating success if they are more dominant and

in good overall health, particularly in populations with a female-skewed sex ratio or a younger male age

structure, or by avoiding direct competition with larger males and opportunistic mating (Townsend and

Bailey 1981, Turner et al. 2016, Partridge 2017).

Antler characteristics are influenced by both genetic and environmental factors. Jamieson et al.

(2020) found that the number of antler tines and antler beam length in white-tailed deer (*Odocoileus*

*virginianus*) were approximately 32% and 45% heritable, respectively. Cain et al. (2019) showed that

larger-antlered deer in the Midwest United States were associated with areas that had more fragmented

land cover and edge habitats, suggesting that habitat composition can shape resource availability and, in

turn, antler growth. Strickland and Demarais (2008) also observed a positive association between

agriculture and mean antler size in the Southern United States. Intergenerational maternal effects have

also been seen to influence antler and other physical traits (Freeman et al. 2013), with for example

maternal nutrition influencing offspring body mass and antler size in white-tailed deer (Monteith et al.

2009). Similarly, Peterson et al. (2019) found that extreme environmental conditions, such as drought,

had lasting effects on antler development in white-tailed deer: deer born during an extreme

environmental year had smaller antlers in later years, suggesting that early-life environmental stress can

have persistent effects on antler growth. Collectively, these studies highlight the need to consider multi-

year and spatial environment covariates when modelling trait variation.

In recent years, hunters in Ontario (Canada) can purchase a white-tailed deer licence and an

antlered deer (i.e. Buck) tag valid for any open deer season, while antlerless deer (female/fawn buck)

tags are awarded via random draw for a specific Wildlife Management Unit (WMU) (Ontario Ministry

of Natural Resources and Forestry 2019). This system allows wildlife managers to regulate the

population through antlerless tag quotas as females are critical for population growth (Ontario Ministry

of Natural Resources and Forestry 2019). Hunters that target male white-tailed deer are often seeking

trophy bucks with higher scoring antlers, which are evaluated based on characteristics such as tine

length and number, beam length and circumference, and antler spread (D'Angelo and Grund 2015,

Boone and Crockett Club 2025). As a result of high-scoring bucks often experiencing heavier harvests,

there has been a broad negative trend in the mean antler size of trophy white-tailed deer over the last

century (Monteith et al. 2013). This decline is not only significant for hunters who might experience

reduced opportunities for harvesting large, trophy bucks, but the removal of larger males could also

disrupt the populations mating structure, potentially resulting in smaller less dominant males becoming

more prevalent (Langvatn and Loison 1999, Pigeon et al. 2016). This shift could also weaken so-called

indicators of overall genetic quality and potentially lead to reduced reproductive success and a

population that is less adaptable to environmental changes (Ditchkoff et al. 2001, Malo et al. 2005).

This study aims to determine which environmental factors influence antler size and morphology

in white-tailed deer across Ontario, Canada. Long-term antler scoring records provided us with a unique

opportunity to examine white-tailed deer antler characteristics and temporal patterns over time and

space in the province. We used conditional autoregressive (CAR) and generalized linear models to

examine the influence of environmental factors on the spatial distribution, tine number, antler

symmetry, and overall antler score of record white-tailed deer across Ontario. We predicted that there is

a relationship between antler size and environmental factors, such as climate and habitat composition,

with larger antlers expected in counties with cover types that allow for a balance of forage and cover

(Cain et al., 2019) and agriculture (Strickland and Demarais 2008). Additionally, we predicted that a

selective harvest pressure on record deer would have led to a decline in antler size across the province

over time (Monteith et al., 2013).

**STUDY AREA**

White-tailed deer in Ontario, Canada, occupy a diverse landscape with varying anthropogenic

influences. The climate across Ontario is highly variable with short summer growing seasons followed

by long winters where food is limited, often causing a decline in body condition (Ontario Ministry of

Natural Resources and Forestry 2019, Government of Ontario 2024). Deer in Ontario inhabit coniferous

and mixed woods forests, and agricultural landscapes. Coniferous forests offer shelter from excessive

snowfall and deciduous forest offer winter browse (Hewitt 2011). While browse is lower in nutritional

quality, it is abundant during Ontario's colder, making up 74-91% of the deer's diet seasons when other

vegetation is scarce (Mautz et al. 1976, Hewitt 2011). Further, Ontario's expanding infrastructure and

growing population, which has tripled in the last 50 years, have confined deer to smaller areas, leading

to an increase in deer density and a decline in the availability of deer habitat and food (Armstrong et al.

1983, Government of Canada 2016, 2022). However, the proliferation of agricultural landscapes might

increase food availability in some regions (Hewitt 2011).

## METHODS

### Data collection

We obtained Ontario antler data from The Foundation for the Recognition of Ontario Wildlife

(FROW) for the years 1941 to 2022. FROW is a non-profit organization that manages the records of big

game animals in Ontario. Antler measurements were submitted by hunters to FROW from every county

in Ontario, except for Temiskaming. The antler dataset included date of kill (DOK), antler type (typical

or atypical), tine numbers and multiple scores following FROW's scoring system, which assigns scores

based on additive measurements of the antler including tine length, beam circumference, and spread.

FROW has a minimum requirement that must be met in order for a deer to be entered into the record

(Foundation for the Recognition of Ontario Wildlife 2015), so we refer to these deer as "record deer"

hereafter. Because we were interested in assessing the influence of spatial and temporal variables on

records, we removed entries that did not contain a county or DOK. Lack of landcover data dating back

pre--1999 resulted in all deer entries prior to 1997 being removed only in the spatial models. We then

assessed the influence of a series of environmental and temporal variables on tine number, gross score,

net score, symmetry and the number of record deer per county.

Predictor variables in our model included temperature ($^o$C), precipitation (mm), and number of

growing degree days for each county corresponding to the year a record deer was harvested (Table S1,

available in Supporting Information). These variables were obtained from Natural Resources Canada

(Hutchinson et al., 2009; Hopkinson et al., 2011; McKenney et al., 2011). We extracted temperature and

precipitation data from the months March to August as these are the months when antler growth occurs

(Hewitt 2011). For precipitation we also extracted the months January to March to gauge winter severity

through snow fall. Precipitation data were averaged. For temperature, we followed Siegert *et al.* (2013)

and calculated the yearly temperature difference by subtracting each county's average temperature over

the time series from the temperature for each individual year. The temperature and precipitation data for

1 year prior to the DOK and 2 years prior to the DOK also extracted.

ArcGIS pro (version 3.1.3; Johnston and Environmental Systems Research Institute, 2004), was

used to extract landcover data. We used 2 land cover layers: the Ontario Land Cover Compilation v.2.0

(Land Information Ontario, 2023a) which captured the years 1999-2011 and Ontario Land Cover

Version 1.0 (Ontario Land Cover Version 1.0 - Overview n.d.) which captured the years 2015 to 2020.

Depending on their DOK the record deer were assigned to a landcover dataset (Table S1). For the years

not captured by the landcover databases (2012-2014), whichever layer encompassed the year closest to

the DOK was used. Reclassification of the land cover categories within these databases was completed

to ensure consistency between land cover types: reclassified categories were; agriculture, infrastructure,

deciduous forest, coniferous forest, mixed forest, other forest, rangeland, water, wetlands, and other

landcover types (Fig. 1). The proportion of each landcover type per county was then calculated, and

"other landcover" was dropped in order to interpret the relationship between the response variable and

the proportion covariates (see Valle *et al.*, 2024).

The final 3 predictor variables used were average county elevation (m), county size (km$^2$) and

deer abundance (Table S1). Average county elevation and county size are fixed variables, with a single

value used over the study period. A relative index of white-tailed deer abundance per county was

generated using the deer seen per day dataset collected by the province of Ontario and was averaged

across all available years (2008-2023) to provide a single estimate per county.

**Statistical analysis**

All variables were standardized, and we assessed pairwise correlations among them, removing

those with a correlation greater than 0.70 or less than -0.70. We then fit a series of models to each

response variable. All spatial models were fitted using the S.CARmultilevel function from the

CARBayes package (Table S2, available in Supporting Information; Lee, 2013). To assess the number

of record deer harvested in each county we used a Poisson regression with a county area offset to ensure

that the number of record deer was not an artifact of county size. All record deer despite year were

included in the frequency count, so DOK was not considered when assigning temperature, precipitation,

or land cover in this model. Instead, average values for each county over the specified time span were

used. This allowed us to observe relationships between environment and the all-time spatial distribution

of record deer harvests across Ontario counties.

To evaluate antler characteristics, we used DOK when assigning temperature, precipitation and

landcover values which will allow us to observe spatial-temporal relationships at the individual level. A

Poisson regression model was used to assess the number of antler tines, a binomial regression model

was used to assess antler symmetry (symmetrical = 0, asymmetrical = 1) and gaussian regression

models were used to assess antler score (Net and Gross). Finally, a standard generalized linear model

from the base r package (R Core Team 2023) was used to assess temporal trends in antler score and tine

number, with year as the predictor variable. Harvest data was sparse pre 1980 and so we used only

entries from 1980 onward. Additionally, minimum scoring requirements were increased in 2006, so all

records from 1980 to 2006 below this adjusted score were excluded (Atypical = 130 and Typical = 120;

Foundation for the Recognition of Ontario Wildlife, 2015).

**RESULTS**

The final dataset after cleaning consisted of 2,413 entries between 1997 and 2020. Number of

growing days was removed due to its high correlation with temperature. Temperature was kept as it

directly reflects seasonal climate variation and is easier to interpret. Abundance was also removed due

to its high correlation with multiple landcover variables. In this analysis, we used 2 different confidence

intervals (CI) to assess the significance of each parameter. A 95% CI was used to identify parameters

that had a high level of certainty in their effect, as the CI not overlapping zero indicates strong evidence

of influence. For parameters with an 80% CI, we interpreted the effect as moderate, meaning the

evidence was less definitive but still suggested a potential influence

Average temperature had a positive effect on the number of record deer per county (Fig. 2; $\beta =$

1.24, 95% CI = 0.53 $\square$ 1.95). Elevation also showed a positive but less certain effect (Fig. 2; $\beta = 0.31$,

80% CI = -0.07 $\square$ 0.72). This model had strong spatial variance ($\tau^2$; 0.98) and moderate spatial

autocorrelation ($\rho$; 0.46). For tine number, the temperature 1 year prior to the year of harvest had a

positive effect  (Fig 2; $\beta = 0.02$, 95% CI = 0.003 $\square$ 0.032). The temperature of the year of harvest and

the % of Rangeland per county also showed positive but less certain effects (Fig 2 & 3; $\beta = 0.01$, 80%

CI = -0.005 $\square$ 0.028 and $\beta = 0.94$, 80% CI = -0.40 $\square$ 2.26, respectfully). This model had a low $\tau^2$

(0.003) and moderate $\rho$ (0.63). For antler symmetry, precipitation during antler growth 2 years prior to

the year of harvest had a negative effect (Fig 2; $\beta = -0.12$, 95% CI = -0.23 $\square$ -0.01), while the

temperature of the year of harvest had a positive but less certain effect (Fig 2; $\beta = 0.10$, 80% CI = -0.02

$\square$ 0.21. This model had a low $\tau^2$ (0.03) and moderate $\rho$ (0.38).

For Gross Score, we found the temperature the year of harvest had a positive effect (Fig 2; $\beta =$

0.01, 95% CI = 0.004 $\square$ 0.02), while precipitation during the winter the year of harvest had a negative

effect (Fig 2; $\beta = -0.01$, 95% CI = -0.01 $\square$ -0.0001). Further, precipitation during antler growth the year

of harvest, precipitation during the winter 1 year prior to the year of harvest, temperature 2 years prior

to harvest and the % of Deciduous Forest per county all showed positive but less certain effects (Fig 2

& 3; $\beta = 0.01$, 80% CI = -0.0003 $\square$ 0.01, $\beta = 0.01$, 80% CI = -0.001 $\square$ 0.01, $\beta = 0.003$, 80% CI = -0.001

$\square$ 0.01 and $\beta = 0.22$, 80% CI = -0.07 $\square$ 0.50, respectfully). This model had $\tau^2$ and residual variance ($v^2$),

both of which were relatively low (0.001 and 0.01 respectively), and $\rho$ was moderately high (0.70).

For Net Score, temperature the year of harvest, temperature 2 years prior to the year of harvest

and the % of Rangeland per county had positive effects (Fig 2 & 3; $\beta$ = 0.01, 95% CI = 0.003 ☐ 0.02, $\beta$

= 0.01, 95% CI = 0.001 ☐ 0.01 and $\beta$ = 0.58, 95% CI = 0.07 ☐ 1.09 respectively). Further, precipitation

during antler growth the year of harvest, the % of Deciduous Forest per county and the % of Coniferous

Forest per county showed positive but less certain effects (Fig 2 & 3; $\beta$ = 0.004, 80% CI = -0.001 ☐

0.01, $\beta$ = 0.27, 80% CI = -0.04 ☐ 0.57 and $\beta$ = 0.28, 80% CI = -0.05 ☐ 0.61, respectfully), while the

precipitation during the winter the year of harvest had a less certain negative effect (Fig 2; $\beta$ = -0.004

and a CI of -0.01 ☐ 0.002. $\tau^2$ and $v^2$ were relatively low (0.002 and 0.01 respectively), and $\rho$ was

moderately high ($\beta$ =0.74). Finally, there was no temporal trend in size and tine number over time

(Table S3, available in Supporting Information).

**DISCUSSION**

The recreational hunting of white-tailed deer brings in millions of dollars to the Canadian

economy annually (Fish and Wildlife Ontario 2016). To the Indigenous peoples of North America,

white-tailed deer are a food source but their antlers and hides provide material for clothing, tools, and

shelter, while also playing an important role in spiritual ceremonies (Brown 2011). Despite their

economic and cultural importance, the size of trophy antlers appears to be declining across North

America (Monteith et al. 2013). This decline showcases the need to better understand the factors

influencing antler growth on a regional level, with for example studies suggesting that edge habitats

containing forest cover could increase antler size (Cain et al. 2019). Accordingly, we investigated

spatial drivers and temporal patterns of harvested record deer in Ontario, Canada. We observed that

increases in temperature during the month's antler growth is occurring (March-August) positively

influenced all antler characteristics (Fig 2), and that rangeland and forest makeup are key landscape

components correlated to overall increase in antler size. No change in antler size was detected over time

in Ontario (Fig. 4).

This data set is representative of adult antlered white-tailed deer within Ontario and although not

a random sample, the data captures a large range of variation in antler size across the population (Fig

S1, available in Supporting Information) and temporal, spatial and environmental influences will be

reflected in this large sample of male white-tailed deer (see also Strickland and Demarais, 2008;

Monteith *et al.*, 2013; Cain *et al.*, 2019). All models exhibited moderate to high spatial autocorrelation

($\rho$) suggesting that nearby counties tend to have similar conditions (i.e., temperature). However, a

moderate value indicates that some unaccounted-for spatial covariates are still present but not explicitly

modeled. Possible missing covariates such as land productivity, landcover structure, or distribution of

certain minerals like phosphorus and calcium could be factors worth further exploration (Grasman and

Hellgren 1993, Cain et al. 2019), though the appropriate GIS layers currently do not exist for Ontario.

We observed low $\tau^2$ in models assessing antler characteristics indicating that, although the predictors

often have similar conditions in neighboring counties, antler size does not exhibit strong spatial

structure. However, a strong $\tau^2$ when assessing the count of record deer per county does suggest that

counties with higher numbers of record deer are more likely to have neighboring counties with similarly

high counts, likely reflective of variation regional hunting interest and behaviour. Notably, hunter

behaviour and success which will vary by roads, dwellings, and accessibility (Kilgo et al. 1998, Bonnot

et al. 2013).

**Environmental drivers of record antlers**

Temperature was the most consistent environmental driver of antler traits: higher temperature

was related to an increase in antler size metrics in all models (Fig 2).  During the year of harvest a 1°C

increase in temperature was associated with a ~1% increase in tine number, a ~10% increase in

symmetry, and a ~1% increase in score. Warmer temperatures have been linked to larger antlers and

greater reproductive success in white-tailed deer and other Cervidae (Clements et al. 2010, Simard et al.

2010). Increased plant growth from warmer conditions means more forage availability for the deer and

therefore more energy is able to be allocated to antler growth (French et al. 1956, Hatfield and Prueger

2015). Here, the unique annual regrowth of antlers makes them strongly affected by yearly temperature

changes (Foley et al. 2012). When looking at tine number and score, the temperature one year prior to

harvest and 2 years prior to harvest respectively, also had positive effects on antler size and morphology

(Fig. 2). Antler growth depends on the accumulation of resources stored in the animal's body, warmer

environments, and in turn higher nutritional gain, in prior years can enhance fat reserves which are then

used in subsequent antler cycles when that energy is needed (Hewitt 2011).

Similar inference can be made from the influence of precipitation during antler growth the year

of harvest on antler size. Wetter conditions can lead to increased plant growth which could again affect

the annual regrowth of antlers (Foley et al. 2012, Scharwies and Dinneny 2019). A 1mm increase in

precipitation was associated with a ~1% increase in net score. Similarly, we saw negative effects from

precipitation during the winter the year of harvest on antler size (Fig 2). More snow in winter can make

it more difficult for the deer to move around and find forage, which causes them to congregate in deer

yards (Government of Ontario 2019). Deer yards are an area made up of coniferous trees that have a

large canopy for blocking excessive snowfall (Government of Ontario 2019). With deer being confined

to these yards deer density and the availability of forage for each individual animal decreases (Patterson

and Power 2002). This could then put individuals with poor body conditions at a disadvantage and

minimal fat reserves when antler regrowth commences (Lesage et al. 2001).

Rangeland and Forest landcover had positive effects on antler characteristics, meaning

that as the percentages of these types of landcover in each county rose so too did antler score and

tine number (Fig 5). Rangeland and Forest together have been found to increase the probability

of harvesting a record deer and are ideal prime edge habitats (Cain et al. 2019). Forests will

block excessive snowfall and contain large amounts of northern deer's primary food source,

browse (Hewitt 2011). Further, rangeland allows for greater visibility of large antlered bucks

from further away, potentially increasing hunter success or improving the effectiveness of

dominance displays and mate attraction, while also providing food such as grasses (Gilbert et al.

2006, Hewitt 2011, Houde et al. 2020). Collectively, these combined edge habitats are preferred

by white-tailed deer as it allows them to have a larger variety of food sources and provides both

more space for movement (rangeland) while still having adequate cover (forest) (Halls 1984,

Alverson et al. 1988).

**Number of harvested deer and antler symmetry**

We observed that increased average temperatures were associated with more record white-tailed

deer being harvested in a given Ontario county with a 1°C increase in temperature during the year of

harvest being associated with an increase in the number of record-sized deer by 1.24 deer in a given

county. Similar to antler size, warmer temperatures allow for more forage availability assuming deer

densities are not overwhelming the landscape. With more food available, more energy is able to be

allocated to individual deer in a county, presumably allowing more to reach record deer status. The

number of record deer harvested were not significantly affected by any landcover covariates (Fig 3); so,

while landcover type positively impacts antler size, the overall effect is minor such that it does not

increase the net number of large animals harvested in a given year. Curiously, record deer numbers were

positively affected by elevation (Fig. 2). It is tempting to suggest this being reflective of mature deer

finding refuge from human activity via fewer roadways at these elevations. Less trafficked zones allow

deer to move more freely and access critical resources with minimal human interference (Hamilton et al.

2024) and behaviourally deer learn to avoid humans (Kilgo et al. 1998, Marantz et al. 2016). However,

the elevational range is small in the province (0 m – 700 m) and obfuscated when averaged across the

county as done here (66 m – 415 m), and thus finer scale analysis is warranted to better understand this

effect.

The symmetry of antlers appears to be impacted by precipitation, with higher precipitation levels

2 years prior to antler growth resulting in less symmetrical antlers. Under uniform environmental

conditions, white-tailed deer exhibit a genetic predisposition for antlers to grow symmetrically

(Ditchkoff et al. 2001). White-tailed deer can develop antler asymmetry when exposed to environmental

stressors affecting their energy allocation before and during antler growth (Goss 2012). The mechanism

by which precipitation 2 years prior impacts asymmetry is not clear. Maternal effects on antler size are

documented in cervids (Freeman et al. 2013), so early-in-life conditions can impact adult traits, with the

impact to antlers seemingly long-lasting (Freeman et al. 2013). Asymmetry has also been seen to

increase as trait size increases (De Coster et al. 2013), which we observed in our data set (Fig S2,

available in Supporting Information); however, we see no strong relationship between precipitation and

antler size, making this explanation unlikely. It is thus conceivable that high environmental stressors

might have long-term impacts on symmetry, even for antlers that are shed annually.

**Temporal patterns of record deer**

Over the last 40 years, there has been no significant change in antler size (or score) and

the tine number of harvested record white-tailed deer in Ontario (Table S3). This contrasts the

findings of Monteith et al. (2013) who reported a negative trend across North America in antler

size in white-tailed deer from 1950 to 2008. These negative trends were hypothesized to be a

combination of harvest-induced reduction in age structure, sociological effects and density

dependence (Monteith et al. 2013). Our finding suggests that antler trends in white-tailed deer

are region specific and might change depending on the environment and pressures present.

In Ontario, the number of active hunters has stayed relatively consistent since 2008 while the

number of records submitted to FROW has declined (Fig 6). This could be indicative of a possible shift

in hunting priorities (Monteith et al. 2013), but multiple tough winters in 07/08 and 10/11, have

coincided with apparent population declines (Ontario Outdoors Magazine Game Forecast 2009, 2011)

which would explain a decrease in entries. Still, the number of white-tailed deer in Ontario are relatively

low especially compared to other regions of North America where population numbers are substantially

higher. For example, Michigan with similar environmental conditions to Ontario and an area of ~

250,493 $km^2$ has a population of ~2 million white-tailed deer, compared to Ontario with an area of over

1 million $km^2$ and a white-tailed deer population of ~400 thousand (Mason 2016, Conner 2021). In

much of North America, warmer climates lead to longer growing seasons and an abundance of high-

quality food, which allows white-tailed deer to reach their maximum physical size during antler growth

(Sontheimer et al. 2024). In contrast, Ontario's colder winters limit access to nutritious food, which

could be preventing deer from achieving their full antler growth potential; this limitation could mean

there is still room for growth. Our models indicate that warming temperatures in Ontario are positively

influencing antler development, as milder winters and longer growing seasons improve food

availability. With ongoing climate changes, we would predict a positive trend in overall antler size if

environmental conditions became more favourable for deer and harvest pressures remain like those

during our time series.

**MANAGEMENT IMPLICATIONS**

Our findings highlight how variation in habitat conditions across Ontario counties are related to

differences in antler size and morphology as well as record white-tailed deer distributions.

Understanding these patterns and relationships can help identify conditions more likely to produce

larger antlered deer, and by communicating these findings, hunters can better manage their expectations

and recognize aspects that could cause certain regions to have higher large-antlered white-tailed deer

concentrations, leading to greater satisfaction (D'Angelo and Grund 2015). An increase in hunter

satisfaction should, in turn, generate revenue that supports the funding of conservation efforts, wildlife

management programs, and local economies. With no significant temporal trend in antler size, the

current harvest practices are likely effective, but we note increased harvest pressure on males in other

regions has led to a reduction in age structure (Langvatn and Loison 1999, Mysterud 2011). Density

dependence also disproportionately affects large males, as they often enter the winter months with low

fat reserves from energy expenditure during the mating season (Lesage et al. 2001, Keyser et al. 2005),

which could have implications for survival and future antler growth. This reinforces the need to

continue to monitor harvest practices to ensure that population dynamics and habitat limitations do not

negatively impact the long-term sustainability of Ontario's deer population.

**ACKNOWLEDGMENTS**

This work was supported by NSERC Discovery Grant (ABAS; 410 SDC); Canadian Foundation

for Innovation (ABAS): John R. Evans Leaders Fund (ABAS) and the Ontario Early Researcher Award

(ABAS). Trent University is located on the traditional territory of the Michi Saagiig Anishnaabeg.

**ETHICS STATEMENT**

No approval of research ethics committees was required to accomplish the goals of this study.

**DATA AVAILABILITY**

The antler data that support the findings of this study are available from The Foundation for the

Recognition of Ontario Wildlife, Ontario, Canada and were acquired under research agreements for this

study. All other model covariates are provided.

**CONFLICT OF INTEREST**

Authors declare no competing interests.

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**FIGURES**

Figure 1. Approximate Ontario land cover distribution (Left) and elevation (Right) divided by county.

Figure 2. Parameter estimates of 80% and 95% confidence intervals (CI) for all models, predicting the effect of elevation, temperature, and precipitation on the response variables.

Figure 3. Parameter estimates of 80% and 95% confidence intervals (CI) for all models, predicting the effect of landcover proportion on the response variables.

Figure 4. Change in antler score and tine number over time. Cutoffs are representative of Boone and Crockett data, depicting the higher cutoff size for record antlers in other parts of North America.

Figure 5. Percent of significant landcover per predicted antler score and tine number

Figure 6. Number of records submitted to the Foundation for the Recognition of Ontario Wildlife (FROW) from 1997 to 2020 (blue) and Number of active hunters from 2008 to 2022 (pink)

**APPENDIX A. SUPPLEMENTAL**