Are all waterholes equal from a lion's view? Exploring the role of prey abundance and catchability in waterhole visitation patterns in a savannah ecosystem

Romain Dejeante[1*]; Andrew J. Loveridge[2,3]; David W. Macdonald[2]; Daphine Madhlamoto[4]; Simon Chamaillé-Jammes[1,5,6,#] ; Marion Valeix[1,5,7,#]

1. CEFE, Univ Montpellier, CNRS, EPHE, IRD, Montpellier, France

2. Wildlife Conservation Research Unit, Department of Biology, The Recanati-Kaplan Centre, University of Oxford, Oxford, UK

3. Panthera, 104 West 40th Street, New York, NY 10018, USA.

4. Zimbabwe Parks and Wildlife Management Authority, Main Camp Research, Hwange National Park, Zimbabwe

5. LTSER France, Zone Atelier 'Hwange', Hwange National Park, Zimbabwe

6. Mammal Research Institute, Department of Zoology and Entomology, University of Pretoria, Pretoria, South Africa

7. REHABS/SRU, International Research Lab, CNRS-NMU, George Campus, Nelson Mandela University, 6530, George, South Africa

* Corresponding author: dejeante.romain@laposte.net

shared last authorship

Abstract

Prey abundance and catchability shape the spatial ecology of predators. Predators can select habitats where prey are more abundant to maximize encounter rate with prey, or alternatively, habitats where prey are more catchable to maximize prey capture. These hypotheses are commonly referred to as prey-abundance and prey-catchability hypotheses. Although these hypotheses are often tested at the landscape scale, little is known about how between-patch variations in prey abundance and catchability determine the space use of predators. In particular, in savannah ecosystems, waterholes are hotspots of prey and their selection by predators is classically interpreted as supporting the prey-abundance hypothesis. Here, we investigated whether between-waterhole variations in prey abundance and catchability influence the frequency and duration of lion visits to waterholes, testing the prey-abundance and prey-catchability hypotheses at the resource-patch scale. We combined datasets (1) on lion movements recorded from GPS collars deployed on 20 adult males and 16 adult females between 2002 and 2015, (2) on prey abundance evaluated from long-term, regular monitoring of waterholes conducted by the Wildlife Environment Zimbabwe and (3) on prey catchability evaluated from remote-sensing satellite imagery of vegetation cover around waterholes in Hwange National Park (Zimbabwe). Although lions did not use equally all waterholes in their territory, surprisingly, between-waterholes variations in prey abundance and catchability only slightly explained the variations in frequency – and even less in duration – of lion visits to waterholes. We discuss the ecological mechanisms and the limits of our work that may explain these findings. First, lions and their prey are involved in a 'shell-game' that lead them to adopt unpredictable movement strategies. Second, lions have only access to a limited number of waterholes to spread out their hunting effort. Lastly, lions visit waterholes not only to hunt but also to interact with social mates and competitors. This work challenges the implicit assumption that all waterholes are the same from a lion's view and may call for further studies investigating the prey-abundance and catchability hypotheses at the resource-patch scale, especially in savannah ecosystems.

Introduction

The study of how predator-prey interactions shape the spatial behaviour of both predators and prey has a long history in the ecological literature (Creel et al., 2005; Fortin et al., 2005; Gaynor et al., 2019; Gilliam & Fraser, 1987; Palmer et al., 2022). Predators and prey are involved in a 'shell game' in which they endlessly adjust their behaviour in response to one another (Mitchell & Lima, 2002; Simon et al., 2019) with prey attempting to avoid predators and predators attempting to be unpredictable for prey (Patin et al., 2020). From a predation perspective, predators are classically expected to use more intensively habitats where prey are more abundant, seeking to maximize the encounter rate with prey (i.e., "prey-abundance hypothesis") (Davidson et al., 2012; Kittle et al., 2017; Zabihi-Seissan et al., 2022). Alternatively, they are also expected to use more intensively habitats where prey are more catchable, seeking to maximize prey capture (i.e., "prey-catchability hypothesis") (Balme et al., 2007; Hopcraft et al., 2005; Zabihi-Seissan et al., 2022). However, these hypotheses are not mutually exclusive and predators should adjust their spatial behaviour to variations in both prey abundance and catchability simultaneously (Smith et al., 2020). Despite this recognition, the roles of prey abundance and prey catchability on predator spatial behaviour have largely been examined separately, limiting our understanding of their relative importance or their possible interactive influence on spatial ecology of predators. For example, do predators consistently favour habitats where prey are more catchable, or do they seek to maximize prey-catchability solely where prey are locally scarce?

Unfortunately, ecologists often lack knowledge on prey abundance and catchability at the resource-patch scale to investigate the intricacies between prey abundance and catchability. Instead, prey abundance and catchability are commonly evaluated at the landscape scale by considering some habitat types as a proxy for prey abundance, and others as a proxy for prey catchability (see for example Balme et al. 2007; Kauffman et al. 2007; Smith et al. 2020). In savannah ecosystems, waterholes and open grasslands are intensively used by large herbivores (Courbin et al., 2019; Smit et al., 2007; Valeix, Loveridge, et al., 2009) and their selection by predators is interpreted as supporting the prey-abundance hypothesis (Davidson et al., 2012). Conversely, tall grasslands, woody vegetation, thickets or rocky outcrops offer numerous ambush sites for stalk-and-ambush predators and their selection by such predators is interpreted as supporting the prey-catchability hypothesis (Davidson et al.,

2012; Davies et al., 2016; Hopcraft et al., 2005). However, comparing a predator's selection for several habitat

types does not allow ecologists to address how predators respond to between-patch variation in prey abundance

and catchability. Indeed, prey abundance and catchability likely vary spatially from one site to another – see for

example Chamaillé-Jammes et al. (2016) for between-waterhole variations in herbivore distribution – and

predators may seek to maximize differently prey capture or prey encounter rate depending on these local variations.

An important gap in our understanding of the spatial ecology of predators is therefore predator responses

to between-patch variations in prey abundance and catchability. Here, to explore the prey-abundance and prey-

catchability hypotheses at the between-patch scale, we investigated the spatial responses of African lions

(*Panthera leo*) to between-waterhole variations in prey abundance and catchability in a savannah ecosystem,

Hwange National Park, Zimbabwe, where the role of waterholes on the spatial ecology of lions has been

extensively studied.

In this ecosystem, waterholes are known to attract large herbivores particularly in the dry season (Valeix,

2011; Valeix, Loveridge, et al., 2009) and thus can be considered as prey hotspots, and ultimately key hunting

sites for lions (Davidson et al., 2012; Valeix et al., 2010; Valeix, Loveridge, et al., 2009). They even shape lion

territory size (Loveridge et al., 2009; Valeix et al., 2012). Most of these studies did not consider heterogeneity in

the characteristics of waterholes. However, the abundance of herbivores varies, spatially, from one waterhole to

another, and between-waterhole variations in prey abundance are consistent over years and substantially explain

the spatial structure of herbivores communities (Chamaillé-Jammes et al., 2016). Further, whereas some

waterholes are located in large open grasslands, others are located in woodlands or bushlands, influencing the

catchability of herbivores near these waterholes, as lions are ambush hunters expected to have a higher hunting

success in environments where vegetation can provide concealment (Davies et al., 2016; Funston et al., 2001;

Schaller, 1972). How lions adjust their space use to between-waterholes variations in both prey abundance and

catchability remains unknown.

The intensity of use of a location – although commonly evaluated in movement ecology, for example to

estimate an animal's home range or an animal's habitat selection – has the limitation of aggregating how frequently

and for how long an individual visits a location which restricts our understanding of the animal's spatial ecology

(Bastille-Rousseau et al., 2024; Benhamou & Riotte-Lambert, 2012). Here, we therefore investigated how

between-waterhole differences in prey abundance and catchability influenced the frequencies and durations of lion

visits to the waterholes available in their territory. In particular, following the *prey-abundance hypothesis*, we

predicted that lions would use more frequently and over a longer time the waterholes with a higher prey abundance.

However, such a response may depend on the average prey abundance within lion territories: for example, lions

may spend more time close to a waterhole with a higher prey abundance when prey are generally scarce within

their territory, but may not do so if prey are abundant throughout their territory *(functional response in prey-*

abundance hypothesis). Following the *prey-catchability hypothesis*, because lions are predators that commonly

hunt by stalking and ambushing their prey (Funston et al., 2001; Schaller, 1972), we predicted that lions would

use more frequently or over longer time those waterholes where the vegetation cover is closer to the waterhole

bank and where the vegetation surrounding the waterhole is denser, i.e., where ambush sites are more numerous.

Similarly, we tested whether lions respond more strongly to between-waterhole variations in prey catchability

when prey were scarce in their territories *(functional response in prey-catchability hypothesis)*. Finally, because

lions may adjust how often they visit a waterhole and the time they spend close to this waterhole to both the local

variations in prey abundance and catchability simultaneously, we also predicted that they should maximize the use

of waterholes with both high prey abundance and high prey catchability *(functional response in prey-abundance*

hypothesis).

Methods

Study area

The study was conducted in the north-eastern area of Hwange National Park, Zimbabwe. The park covers 14,600

km² of semi-arid savannah. In this study, we used data collected during the late dry season, from July to mid-

October, when natural rain-fed pans are dry and surface water remains available only in artificial permanent

waterholes in which underground water is pumped (Chamaillé-Jammes et al., 2007). A GIS layer of all artificial

permanent waterholes that retain water throughout the dry season was available.

Lion visits to waterholes

We investigated lion visits to waterholes from movement data collected on 20 adult male lions and 16 adult female lions between 2002 and 2015. Lions were monitored with GPS collars, collecting hourly or two-hourly locations. As lions were commonly equipped over more than one year, we used animal-year as our sampling unit, resulting in 38 animal-year units for males, and 35 animal-year units for females. Because prey-abundance data (see the following section) was only available during the dry season, only movement data from the late dry season, between July and mid-October, were analysed. Lion handling and care protocols were consistent with guidelines provided in the 'Code of Practice for Biologists using Animals', Department of Zoology, University of Oxford and approved by University of Oxford, Biomedical Sciences, Animal Welfare and Ethical Review Body. Lion handling procedures were carried out by project staff trained and certified by the Zimbabwe Veterinary Association, Wildlife Group (Certificate numbers: Davidson: 09/03; Elliot: 9/10; Hunt: A11/04, 005/09, 5/14; Loveridge: 6/2000, 20/36 (2007), 6/14; Stapelkamp: 34/2008) in accordance with Statutory Instrument 409 of 1999 (Clause 21A to 21J) amending the Regulations of 1975 to the Dangerous Drugs Act, Zimbabwe.

For each animal-year, we evaluated the frequency and the duration of visits to each waterhole of a lion territory using the R package *recurse* (Bracis et al., 2018) to count, respectively, the number of times a collared lion entered a circular buffer around a waterhole – over 100 days – and to measure the time spent within this buffer per visit. We performed our analyses using both a 1km and a 2km radius around each waterhole. These distances are commonly employed to investigate the behaviour of lions in the vicinity of waterholes in the study ecosystem (Courbin et al., 2016; Davidson et al., 2013; Dejeante et al., 2024; Valeix et al., 2010; Valeix, Fritz, et al., 2009). We used the two distance thresholds to test the sensitivity of our results to the chosen value. Results did not differ qualitatively, and we present results obtained with the 1km radius threshold in the main text. Results from the analyses with the 2km radius threshold are available in Table S1 and Table S2. Because the number of days of tracking during the dry season varied from one animal-year to another – from 60 to 105 days – we re-calculated the number of visits to a waterhole over a period of 100 days using a proportional relationship. This allowed us to compare the number of visits from one animal-year to another.

We estimated the frequency and the duration of lion visits to each waterhole within their territory, as delineated by the minimum convex polygon created from 100% of lion's movement data collected during the late

dry season. Although using kernel-based methods is commonly a more revealing approach to delineate an animal's home range (Börger et al., 2006; Nilsen et al., 2008), here we used a minimum convex polygon to consider as available all waterholes potentially reachable by an individual, even those rarely or never visited. Contrary to kernel-based methods, drawing a minimum convex polygon does not allow us to measure differences in the intensity of use. However, these relative differences in the intensity of use of each available waterhole within a lion territory were explored from the differences in the frequency and duration of visits. In the following analysis, we seek to explain differences in the frequency and duration of lion visits to each waterhole of their territory by between-waterhole variations in prey abundance and catchability.

Prey abundance data

To evaluate prey abundance at each waterhole, we used a long-term, regular monitoring of waterholes conducted by the Wildlife Environment Zimbabwe (WEZ) since 1972. Every year, at the end of the dry season, the WEZ records the animals coming to drink to each waterhole of the park over a 24h period. This monitoring provided information on the total number of herds and the average herd size that visited a waterhole for all large herbivore species. Importantly, Chamaillé-Jammes et al. (2016) showed that the spatial structure of the herbivore community is mostly explained by between-waterhole variations and that differences in prey abundance between waterholes are consistent over years. Although the park has a wide diversity of large herbivores, here, we focused only on the main prey species of lions (Davidson et al., 2013): African buffalo (*Syncerus caffer*), plains zebra (*Equus quagga*), blue wildebeest (*Connochaetes taurinus*), roan antelope (*Hippotragus equines*), sable antelope (*Hippotragus niger*), greater kudu (*Tragelaphus strepsiceros*) and eland (*Taurotragus oryx*). For species other than buffalo, because encounter rates between predators and prey are better predicted by prey herd density than by prey individual density (Fryxell et al., 2007), and because their herd size does not show a high variability in Hwange National Park (Table S3), we used the total number of herds drinking at a waterhole (pooling all species) as a proxy for medium-sized prey abundance at this waterhole. Hunting medium-sized prey or larger prey as buffalos is different, because large prey are harder and more dangerous to catch but provide more food (Funston et al., 1998, 2001) and because buffalos move in very large herds (~ 95 individuals). Therefore, we did not pool the number of buffalo herds with the number of herds of medium-sized prey at a waterhole, but instead we used a

second metric to describe buffalo abundance. As buffalo herds roam over stable areas over the years, we averaged

the number of buffalo herds that visited each waterhole using a 5-year temporal window to smooth between-year

fluctuations in buffalo counts.

Variables related to herd or buffalo abundance were centred and scaled per animal-year ('Herd' and

'Buffalo' variables) to test whether, within a territory, lions visit more waterholes with higher abundance in herds

or in buffalos ('prey-abundance hypothesis'). In addition, we also calculated the mean abundance of herd and

buffalo abundance per animal-year ('Herd Territory' and 'Buffalo Territory') and then centred and scaled this

variable across all animal-year. Adding an interaction term between the 'Herd' and 'Herd Territory' variables

allowed us to test whether lions visit more waterholes with higher herd abundance specifically when their territory

is poor or rich in herds, relative to territories of other lions in the study area ('functional response in prey-

abundance hypothesis').

Prey catchability data

In Hwange National Park, the vegetation is dominated by bushlands and woodlands interspersed with patches of

grasslands surrounding waterholes (Arraut et al., 2018). In this analysis, we considered two scales. First, we used

the distance between the waterhole bank and the vegetation cover (measured using Google Earth Pro © for each

cardinal direction (North, South, East and West) and then averaging these 4 distances for each waterhole) as a

proxy for the fine-scale prey catchability at a waterhole (Figure 1; 'Distance to cover' variable). Second, we used

the mean vegetation openness within the 1km surrounding each waterhole as a proxy for the large-scale prey

catchability around a waterhole ('vegetation openness' variable). To do so, we used the 30-m resolution vegetation

map produced by Arraut et al. (2018) to calculate, for each pixel, the proportion of open vegetation (category

'grassland' and 'open bushlands' in the original map) in a radius of 250 m. We then averaged the proportion of

open vegetation within a radius of 1km around each waterhole.

Data analysis

We investigated how between-waterhole variations in prey abundance and catchability influenced the differences

in the frequency and duration of lion visits to the different waterholes of their territory. Since the frequency and

duration of lion visits to waterholes exhibited overdispersion in Poisson models, we used negative binomial

regressions to model these response variables. For each response variable, we tested the influence of explanatory

variables related to prey abundance and prey catchability. Model sets are presented in Tables 1 and 2 and are

grouped according to the tested hypotheses (see Introduction). Because the frequency of visits to a waterhole and

the duration of these visits may not be related solely to prey availability but also to the number of waterholes

within the lion territory, we controlled for this effect by including the number of waterholes within the territory of

an animal-year as a simple effect in all tested models.

To determine the most parsimonious model explaining the differences of frequency and duration of lion

visits to a waterhole, we used the Akaike information criterion corrected for small sample size (AICc) (Burnham

& Anderson, 2004). In the following, we present results obtained from the most parsimonious models, i.e., with

both a ΔAICc < 2 and the lowest number of explanatory variables (Arnold, 2010). Selected models show no

multicollinearity between predictors, as indicated by variance inflation factors (VIF) lower than 2. All models

included a random intercept with animal-year identity; analyses were conducted using the R package *glmmTMB*

(Brooks et al., 2017). We used the R package *performance* (Lüdecke et al., 2021) to measure the VIF scores and

the goodness-of-fit measure of the models as estimated by the marginal pseudo-R^2 for generalized linear mixed

models (Nakagawa et al., 2017).

Results

Influence of prey abundance and catchability on the frequency of lion visits to a waterhole

On average, male lions visited each waterhole of their territory about 9 times over a 100 day period of the late dry

season (ranging from 1 to 50 visits) and, similarly, female lions visited each waterhole about 8 times over the same

period (ranging from 1 to 64 visits). For male lions, the most parsimonious model explaining the variability of the

frequency of visits to a waterhole included the number of waterholes in the territory, the number of buffalo herds

visiting the waterhole and the vegetation openness in the surroundings of the waterhole (Table 1; *Buffalo +*

Vegetation openness). Male lions visited more frequently waterholes surrounded by open vegetation and those

used more frequently by buffalo herds, relative to other waterholes available in their territories (Figure 2).

Unsurprisingly, they also visited each waterhole less frequently when their territory encompassed a larger number

of waterholes. Overall, this model explained 28% of the variability in the number of visits to a waterhole. The

vegetation openness was the main variable explaining this variability. Indeed, the null model – that solely included

the number of waterholes – explained only 6% of this variability and only 8% when also accounting for buffalo

abundance against 25% when accounting for the vegetation openness (and not accounting for buffalo abundance)

(Table 1).

For female lions, the most parsimonious model explaining the variability of the frequency of visits to a

waterhole included the number of waterholes in the territory, the number of herds of medium-sized prey and the

vegetation openness in the surroundings of the waterhole (Table 1; *Herd + Vegetation openness*). Female lions

visited more frequently waterholes surrounded by open vegetation and those used more frequently by herds of

medium-sized prey (Figure 2). As male lions, they also visited less frequently each waterhole when the number of

waterholes in their territory increased. Although this model explained 36% of the variability in the frequency of

visit to a waterhole by female lions, the metrics related to prey abundance and catchability explained only a small

amount of this variability. Indeed, the null model – that included the number of waterholes only – explained 30%

of this variability. This result suggests that the vegetation openness and the number of medium-sized prey herds

only slightly affect the frequency of visits to a waterhole by female lions.

Influence of prey abundance and catchability on the duration of lion visits to a waterhole

On average, male lions visited a waterhole within their territory for about 6 hours per visit (ranging from 1 to 76

hours) and, similarly, female lions visited a waterhole for about 7 hours per visit (ranging from 1 to 48 hours). For

both male and female lions, all models tested only slightly explained the variability in the duration of visits to a

waterhole. For female lions, the most parsimonious model included solely the number of waterholes in the territory

(Table 2; *null hypothesis*) and only explained 3% of the variability in the duration of visits (Figure 3). For male

lions, the most parsimonious model, which included the number of waterholes in the territory and the vegetation

openness in the surroundings of the waterhole (Table 2; *Vegetation openness*) – only explained 9% of this

variability. Although male lions stayed longer times at a waterhole where the vegetation openness was greater in

the surroundings (Figure 3), the low pseudo-R² of this model suggests that this variable only slightly affect the

duration of visits.

Discussion

In savannah ecosystems, waterhole areas are key foraging patches for predators and their prey (Davidson et al.,

2012; Valeix et al., 2010; Valeix, Loveridge, et al., 2009). However, whether all waterholes are similarly attractive

from a predator's viewpoint or whether predators favour some waterhole areas rather than others remains

unknown. Here, we investigated how between-waterhole variations in prey abundance and catchability influenced

the frequency and duration of lion visits to a waterhole. Surprisingly, we found that the prey-related variables

considered in this work only slightly affected the frequency – and even less the duration – of visits to a waterhole

by male and female lions. Our main findings are that (1) lions did not use equally (different frequencies and

durations of visits) the waterholes in their territory, (2) the duration of visits of waterholes by lions is not influenced

by the associated prey abundance and vegetation characteristics, (3) lions of both sexes tend to visit more

frequently waterholes in their territory when there are fewer waterholes, (4) prey abundance (as measured in this

work) does not affect the frequency of visits at a waterhole by lions, and (5) male lions tend to visit more often

waterholes surrounded by open vegetation. The ecological mechanisms and the limits of our work that may explain

these findings are discussed below.

A dynamic predator-prey space race

Although predators can rely on cues of higher prey abundance or higher prey catchability to move in the landscape

(Balme et al., 2007; Davidson et al., 2012; Hopcraft et al., 2005), predator and prey are involved in a 'shell game'

in which they endlessly adjust their behaviour in relation to one another, and if predators are too predictable, prey

can proactively increase their vigilance when foraging close to well-known ambush site or can avoid foraging

nearby (Gaynor et al., 2019; Mitchell, 2009; Mitchell & Lima, 2002). As demonstrated by Patin et al. (2020),

predators should therefore seek to be unpredictable to optimize their hunting success, integrating some level of

randomness in their movement behaviour to maximize their chances of encountering and capturing prey.

Regarding this predator-prey space race, lions would therefore benefit from not visiting too regularly the same

waterholes in order to be less predictable by prey. Lions in Hwange National Park are known to make most of their kills close to a waterhole and to commonly depart from a kill site to visit another waterhole (Valeix et al., 2011). Altogether, this and our findings suggest that lions may spread out their hunting grounds in an unpredictable manner. Even in the absence of kills, lions need to frequently depart from their hunting patch to counter balance the anti-predator responses adopted by prey once they detect their presence (Charnov et al. 1976; Kotler 1992). One of the anti-predator responses adopted by prey is to move away from the site of an encounter with lions (Courbin et al., 2016; Martin & Owen-Smith, 2016; Say-Sallaz et al., 2023), possibly leading to a depression in prey abundance in the vicinity of a waterhole following a lion visit. Here, unfortunately, our prey-related variables do not allow us to draw inference about these short-term variations in prey abundance and may therefore not totally reflect prey abundance from a lion's viewpoint. Indeed, although some waterholes have consistently greater herbivore abundance than others in Hwange National Park (Chamaillé-Jammes et al., 2016), lions may use other cues to infer instantaneous prey abundance. Because of a short-term depression in prey abundance, one waterhole having consistently greater herbivore abundance than others may offer – at a given time, for example following an attempted hunt – less opportunities for lions to encounter prey than an undisturbed waterhole having on average a lower number of herbivores in its vicinity. Being territorial animals (Schaller, 1972), lions have access to only a limited number of waterholes to spread out their hunting effort. Therefore, even if they seek to preferentially visit undisturbed feeding patches where prey are the most available, lions may be constrained to visit patches with low prey abundance or poor prey catchability if those patches are the only ones undisturbed by their previous presence.

A territorial and social species

Although this study focuses on the importance of prey-related variables to explain between-waterhole variations in the duration and frequency of visits to a waterhole, lions are not solely predators but also territorial and social animals (Schaller, 1972). Other variables, unrelated to prey, may therefore partly explain differences in the frequency and duration of visits to a waterhole. In particular, beyond being key hunting patches, waterholes are key marking places for lions (Dejeante, Loveridge, Macdonald, Madlamoto, et al., 2025). For example, lions vocalize more repeatedly near water sources (Wijers et al. 2021), or male lions encounter more frequently their rivals near waterholes (Dejeante et al. 2024). The distribution of these waterholes greatly structures the size and

shape of lion territories and drives their territorial movements in the core and in the periphery of the territory (Valeix et al., 2012). We could therefore expect that lions' territorial behaviours – and not solely predator-prey interactions – influence the duration and the frequency of lion visits to a waterhole. As demonstrated for another territorial species (wolves that integrate space and time to visit territorial borders; Schlägel et al. 2017), lions could, for example, be more likely to visit a waterhole not visited for a long time to renew their marking signals at this place. Interestingly, here, we found a great influence of the vegetation openness around a waterhole: male lions were more likely to visit a waterhole frequently if it was surrounded by open habitats. This result is clearly not supporting the prey-catchability hypothesis. However, in these large open areas, prey catchability may still matter, but at a finer scales than the ones investigated here. In addition, because a more open habitat is likely associated with a higher probability of encountering large herds of large herbivores in Hwange National Park, this result could be shaped by a long-term cue of high prey abundance, not captured by the data available from waterhole counts. However, other explanations are possible. In particular, a territorial-related explanation is that male lions spend more time with their pride in open areas because the visibility is greater in these habitats, allowing them to detect more easily intruders and so to patrol less frequently their territory (Dejeante et al., 2024; Funston et al., 1998). If male lions favour visiting waterhole surrounded by open habitats for territorial reasons not related to prey, these prey should however become reluctant to use these waterholes. This would lead to a mismatch between prey abundance and the patterns of lion visits to a waterhole.

Conclusion

Many studies investigating the role of waterholes, i.e. prey patches, in the spatial behaviour of lions implicitly assumed that all waterholes are the same (Davidson et al., 2012, 2013; Valeix et al., 2010; Valeix, Fritz, et al., 2009). Here – to our knowledge for the first time – we investigated whether this assumption was true. Clearly there is an heterogeneity in the frequency and duration of lion visits at waterholes, suggesting that not all waterholes are the same from a lion's viewpoint. Although lions did not use equally all waterholes within their territory, we were able to provide little evidence that the variability of the frequency and duration of lion visits to a waterhole was determined by between-waterhole variations in prey abundance and catchability. Three main points can explain this result: (1) lions and their prey are involved in a "shell-game" that leads them to adopt unpredictable movement

324 strategies, (2) lions have access to a limited number of waterholes amongst which to spread out their hunting effort,

325 and this leads them to visit some waterholes with lower prey abundance or catchability solely because those are

326 the only ones undisturbed, (3) lions do not visit waterholes solely to hunt but also to interact with conspecifics

327 (both social mates or rivals) for example to renew frequently their territorial cues at these places, whatever the

328 abundance and catchability of prey is there.

Declaration of interest

330 The authors declare that they have no conflict of interest.

Acknowledgments

The Hwange Lion Project was funded by grants from the Robertson Foundation, the Recanati-Kaplan Foundation,

the Darwin Initiative for Biodiversity grant 162/09/015, a CV Starr Scholarship, The Eppley Foundation, Disney

Foundation, Marwell Preservation Trust, Regina B. Frankenburg Foundation, Rufford Maurice Laing Foundation,

Panthera Foundation and the generosity of Joan and Riv Winant. We are deeply grateful to Jane Hunt, Zeke

Davidson, Nicholas Elliot, Brent Stapelkamp, Dan Parker, Agrippa Moyo, Lovemore Sibanda, Moreangels

Mbizah, Andrea Sibanda and Liomba Mathe for their roles in the collection of lion GPS data.

Data Availability

Data and codes used in this study are publicly available in a figshare repository (Dejeante, Loveridge, Macdonald,

Madhlamoto, et al., 2025)

REFERENCES

Arnold, T. W. (2010). Uninformative Parameters and Model Selection Using Akaike's Information Criterion. *The*

Journal of Wildlife Management, *74*(6), 1175–1178. https://doi.org/10.1111/j.1937-2817.2010.tb01236.x

Arraut, E. M., Loveridge, A. J., Chamaillé-Jammes, S., Valls-Fox, H., & Macdonald, D. W. (2018). The 2013-

2014 vegetation structure map of Hwange National Park, Zimbabwe, produced using free satellite images

and software. *Koedoe*, *60*(1), 1–10. https://doi.org/10.4102/koedoe.v60i1.1497

Balme, G., Hunter, L., & Slotow, R. (2007). Feeding habitat selection by hunting leopards Panthera pardus in a

woodland savanna: Prey catchability versus abundance. *Animal Behaviour*, *74*, 589–598.

https://doi.org/10.1016/j.anbehav.2006.12.014

Bastille-Rousseau, G., Crews, S. A., Donovan, E. B., Egan, M. E., Gorman, N. T., Pitman, J. B., Weber, A. M.,

Audia, E. M., Larreur, M. R., Manninen, H., Blake, S., Eichholz, M. W., Bergman, E., & Rayl, N. D.

(2024). A multi-property assessment of intensity of use provides a functional understanding of animal

movement. *Methods in Ecology and Evolution*, *15*(2), 345–357. https://doi.org/10.1111/2041-210X.14274

Benhamou, S., & Riotte-Lambert, L. (2012). Beyond the Utilization Distribution: Identifying home range areas

that are intensively exploited or repeatedly visited. *Ecological Modelling*, *227*, 112–116.

https://doi.org/10.1016/j.ecolmodel.2011.12.015

Börger, L., Franconi, N., De Michele, G., Gantz, A., Meschi, F., Manica, A., Lovari, S., & Coulson, T. (2006).

Effects of Sampling Regime on the Mean and Variance of Home Range Size Estimates. *Journal of Animal*

Ecology, *75*(6), 1393–1405.

Bracis, C., Bildstein, K. L., & Mueller, T. (2018). Revisitation analysis uncovers spatio-temporal patterns in animal

movement data. *Ecography*, *41*(11), 1801–1811. https://doi.org/10.1111/ecog.03618

Brooks, M. E., Kristensen, K., van Benthem, K. J., Magnusson, A., Berg, C. W., Nielsen, A., Skaug, H. J., Mächler,

M., & Bolker, B. M. (2017). *glmmTMB Balances Speed and Flexibility Among Packages for Zero-inflated*

Generalized Linear Mixed Modeling. *9*.

Burnham, K. P., & Anderson, D. R. (2004). Multimodel Inference: Understanding AIC and BIC in Model

Selection. *Sociological Methods & Research*, *33*(2), 261–304.

https://doi.org/10.1177/0049124104268644

Chamaillé-Jammes, S., Charbonnel, A., Dray, S., Madzikanda, H., & Fritz, H. (2016). Spatial Distribution of a

Large Herbivore Community at Waterholes: An Assessment of Its Stability over Years in Hwange

National Park, Zimbabwe. *PLOS ONE*, *11*(4), e0153639. https://doi.org/10.1371/journal.pone.0153639

Chamaillé-Jammes, S., Fritz, H., & Murindagomo, F. (2007). Climate-driven fluctuations in surface-water

availability and the buffering role of artificial pumping in an African savanna: Potential implication for

herbivore dynamics. *Austral Ecology*, *32*(7), 740–748. https://doi.org/10.1111/j.1442-9993.2007.01761.x

Charnov, E. L., Orians, G. H., & Hyatt, K. (1976). Ecological Implications of Resource Depression. *The American*

Naturalist, *110*(972), 247–259.

Courbin, N., Loveridge, A. J., Fritz, H., Macdonald, D. W., Patin, R., Valeix, M., & Chamaillé-Jammes, S. (2019).

Zebra diel migrations reduce encounter risk with lions at night. *Journal of Animal Ecology*, *88*(1), 92–

101. https://doi.org/10.1111/1365-2656.12910

Courbin, N., Loveridge, A. J., Macdonald, D. W., Fritz, H., Valeix, M., Makuwe, E. T., & Chamaillé-Jammes, S.

(2015). Reactive responses of zebras to lion encounters shape their predator-prey space game at large

scale. *Oikos*, *125*(6). https://ora.ox.ac.uk/objects/uuid:63cebb0e-a530-45a0-8234-cd4464f77328

Courbin, N., Loveridge, A. J., Macdonald, D. W., Fritz, H., Valeix, M., Makuwe, E. T., & Chamaillé-Jammes, S.

(2016). Reactive responses of zebras to lion encounters shape their predator–prey space game at large

scale. *Oikos*, *125*(6), 829–838. https://doi.org/10.1111/oik.02555

Creel, S., Winnie Jr., J., Maxwell, B., Hamlin, K., & Creel, M. (2005). Elk Alter Habitat Selection as an

Antipredator Response to Wolves. *Ecology*, *86*(12), 3387–3397. https://doi.org/10.1890/05-0032

Davidson, Z., Valeix, M., Kesteren, F. V., Loveridge, A. J., Hunt, J. E., Murindagomo, F., & Macdonald, D. W.

(2013). Seasonal Diet and Prey Preference of the African Lion in a Waterhole-Driven Semi-Arid Savanna.

PLOS ONE, *8*(2).

Davidson, Z., Valeix, M., Loveridge, A. J., Hunt, J. E., Johnson, P. J., Madzikanda, H., & Macdonald, D. W.

392   (2012). Environmental determinants of habitat and kill site selection in a large carnivore: Scale matters.

393   *Journal of Mammalogy*, *93*(3), 677–685. https://doi.org/10.1644/10-MAMM-A-424.1

394   Davies, A. B., Tambling, C. J., Kerley, G. I. H., & Asner, G. P. (2016). Effects of Vegetation Structure on the

395   Location of Lion Kill Sites in African Thicket. *PLOS ONE*, *11*(2), e0149098.

396   https://doi.org/10.1371/journal.pone.0149098

397   Dejeante, R., Loveridge, A. J., Macdonald, D. W., Madhlamoto, D., Chamaillé, S., & Valeix, M. (2025). *Data*

398   *from: Are all waterholes equal from a lion's view? Exploring the role of prey abundance and catchability*

399   *in waterhole visitation patterns in a savannah ecosystem* (p. 503717 Bytes) [Dataset]. figshare.

400   https://doi.org/10.6084/M9.FIGSHARE.28825691

401   Dejeante, R., Loveridge, A. J., Macdonald, D. W., Madhlamoto, D., Valeix, M., & Chamaillé-Jammes, S. (2024).

402   Counter-strategies to infanticide: The importance of cubs in determining lion habitat selection and social

403   interactions. *Journal of Animal Ecology*, *93*(2), 159–170. https://doi.org/10.1111/1365-2656.14045

404   Dejeante, R., Loveridge, A. J., Macdonald, D. W., Madlamoto, D., Chamaille-James, S., & Valeix, M. (2025).

405   Territorial movements of African lions: Can waterholes shape multiple central-place territories in a

406   savanna ecosystem? *Animal Behaviour*. https://ora.ox.ac.uk/objects/uuid:e2674ff2-4050-489d-9f1d-

407   3e86cb118e80

408   Fortin, D., Beyer, H. L., Boyce, M. S., Smith, D. W., Duchesne, T., & Mao, J. S. (2005). Wolves Influence Elk

409   Movements: Behavior Shapes a Trophic Cascade in Yellowstone National Park. *Ecology*, *86*(5), 1320–

410   1330. https://doi.org/10.1890/04-0953

411   Fryxell, J. M., Mosser, A., Sinclair, A. R. E., & Packer, C. (2007). Group formation stabilizes predator–prey

412   dynamics. *Nature*, *449*(7165), 1041–1043. https://doi.org/10.1038/nature06177

413   Funston, P. J., Mills, M. G. L., & Biggs, H. C. (2001). Factors affecting the hunting success of male and female

414   lions in the Kruger National Park. *Journal of Zoology*, *253*(4), 419–431.

415   https://doi.org/10.1017/S0952836901000395

Funston, P. J., Mills, M. G. L., Biggs, H. C., & Richardson, P. R. K. (1998). Hunting by male lions: Ecological

influences and socioecological implications. *Animal Behaviour*, *56*(6), 1333–1345.

https://doi.org/10.1006/anbe.1998.0884

Gaynor, K. M., Brown, J. S., Middleton, A. D., Power, M. E., & Brashares, J. S. (2019). Landscapes of Fear:

Spatial Patterns of Risk Perception and Response. *Trends in Ecology & Evolution*, *34*(4), 355–368.

https://doi.org/10.1016/j.tree.2019.01.004

Gilliam, J. F., & Fraser, D. F. (1987). Habitat Selection Under Predation Hazard: Test of a Model with Foraging

Minnows. *Ecology*, *68*(6), 1856–1862. https://doi.org/10.2307/1939877

Hopcraft, J. G. C., Sinclair, A. R. E., & Packer, C. (2005). Planning for success: Serengeti lions seek prey

accessibility rather than abundance. *Journal of Animal Ecology*, *74*(3), 559–566.

https://doi.org/10.1111/j.1365-2656.2005.00955.x

Kauffman, M. J., Varley, N., Smith, D. W., Stahler, D. R., MacNulty, D. R., & Boyce, M. S. (2007). Landscape

heterogeneity shapes predation in a newly restored predator–prey system. *Ecology Letters*, *10*(8), 690–

700. https://doi.org/10.1111/j.1461-0248.2007.01059.x

Kittle, A. M., Anderson, M., Avgar, T., Baker, J. A., Brown, G. S., Hagens, J., Iwachewski, E., Moffatt, S., Mosser,

A., Patterson, B. R., Reid, D. E. B., Rodgers, A. R., Shuter, J., Street, G. M., Thompson, I. D., Vander

Vennen, L. M., & Fryxell, J. M. (2017). Landscape-level wolf space use is correlated with prey abundance,

ease of mobility, and the distribution of prey habitat. *Ecosphere*, *8*(4), e01783.

https://doi.org/10.1002/ecs2.1783

Kotler, B. P. (1992). Behavioral resource depression and decaying perceived risk of predation in two species of

coexisting gerbils. *Behavioral Ecology and Sociobiology*, *30*(3), 239–244.

https://doi.org/10.1007/BF00166708

Loveridge, A. J., Valeix, M., Davidson, Z., Murindagomo, F., Fritz, H., & Macdonald, D. W. (2009). Changes in

home range size of African lions in relation to pride size and prey biomass in a semi-arid savanna.

Ecography, *32*(6), 953–962.

Lüdecke, D., Ben Shachar, M., Patil, I., Waggoner, P., & Makowski, D. (2021). performance: An R Package for

Assessment, Comparison and Testing of Statistical Models. *The Journal of Open Source Software*, *6*,

3139. https://doi.org/10.21105/joss.03139

Martin, J., & Owen-Smith, N. (2016). Habitat selectivity influences the reactive responses of African ungulates to

encounters with lions. *Animal Behaviour*, *116*, 163–170. https://doi.org/10.1016/j.anbehav.2016.04.003

Mitchell, W. A. (2009). Multi-behavioral strategies in a predator–prey game: An evolutionary algorithm analysis.

Oikos, *118*(7), 1073–1083. https://doi.org/10.1111/j.1600-0706.2009.17204.x

Mitchell, W. A., & Lima, S. L. (2002). Predator-prey shell games: Large-scale movement and its implications for

decision-making by prey. *Oikos*, *99*(2), 249–259. https://doi.org/10.1034/j.1600-0706.2002.990205.x

Nakagawa, S., Johnson, P. C. D., & Schielzeth, H. (2017). The coefficient of determination R2 and intra-class

correlation coefficient from generalized linear mixed-effects models revisited and expanded. *Journal of*

The Royal Society Interface, *14*(134), 20170213. https://doi.org/10.1098/rsif.2017.0213

Nilsen, E. B., Pedersen, S., & Linnell, J. D. C. (2008). Can minimum convex polygon home ranges be used to

draw   biologically   meaningful   conclusions?   *Ecological   Research*,   *23*(3),   635–639.

https://doi.org/10.1007/s11284-007-0421-9

Palmer, M. S., Gaynor, K. M., Becker, J. A., Abraham, J. O., Mumma, M. A., & Pringle, R. M. (2022). Dynamic

landscapes of fear: Understanding spatiotemporal risk. *Trends in Ecology & Evolution*, *37*(10), 911–925.

https://doi.org/10.1016/j.tree.2022.06.007

Patin, R., Fortin, D., & Chamaillé-Jammes, S. (2020). *A theory of the use of information by enemies in the*

predator-prey space race. https://doi.org/10.1101/2020.01.25.919324

Say-Sallaz, E., Chamaillé-Jammes, S., Périquet, S., Loveridge, A. J., Macdonald, D. W., Antonio, A., Fritz, H., &

Valeix, M. (2023). Large carnivore dangerousness affects the reactive spatial response of prey. *Animal*

Behaviour, *202*, 149–162. https://doi.org/10.1016/j.anbehav.2023.05.014

Schaller, G. (1972). *Serengeti Lion: A Study of Predator-Prey Relations (Wildlife behavior and ecology)*.

University   of   Chicago   Press.   https://www.biblio.com/book/serengeti-lion-study-predator-prey-

relations/d/460478448

Schlägel, U. E., Merrill, E. H., & Lewis, M. A. (2017). Territory surveillance and prey management: Wolves keep

track of space and time. *Ecology and Evolution*, *7*(20), 8388–8405. https://doi.org/10.1002/ece3.3176

Simon, R. N., Cherry, S. G., & Fortin, D. (2019). Complex tactics in a dynamic large herbivore–carnivore

spatiotemporal game. *Oikos*, *128*(9), 1318–1328. https://doi.org/10.1111/oik.06166

Smit, I. P. J., Grant, C. C., & Devereux, B. J. (2007). Do artificial waterholes influence the way herbivores use the

landscape? Herbivore distribution patterns around rivers and artificial surface water sources in a large

African   savanna   park.   *Biological   Conservation*,   *136*(1),   85–99.

https://doi.org/10.1016/j.biocon.2006.11.009

Smith, J. A., Donadio, E., Bidder, O. R., Pauli, J. N., Sheriff, M. J., Perrig, P. L., & Middleton, A. D. (2020).

Where and when to hunt? Decomposing predation success of an ambush carnivore. *Ecology*, *101*(12),

e03172. https://doi.org/10.1002/ecy.3172

Valeix, M. (2011). Temporal dynamics of dry-season water-hole use by large African herbivores in two years of

contrasting rainfall in Hwange National Park, Zimbabwe. *Journal of Tropical Ecology*, *27*(2), 163–170.

https://doi.org/10.1017/S0266467410000647

Valeix, M., Chamaillé-Jammes, S., Loveridge, A. J., Davidson, Z., Hunt, J. E., Madzikanda, H., & Macdonald, D.

W. (2011). Understanding Patch Departure Rules for Large Carnivores: Lion Movements Support a Patch-

Disturbance Hypothesis. *The American Naturalist*, *178*(2), 269–275. https://doi.org/10.1086/660824

Valeix, M., Fritz, H., Loveridge, A. J., Davidson, Z., Hunt, J. E., Murindagomo, F., & Macdonald, D. W. (2009).

Does the risk of encountering lions influence African herbivore behaviour at waterholes? *Behavioral*

Ecology and Sociobiology, *63*(10), 1483–1494. https://doi.org/10.1007/s00265-009-0760-3

Valeix, M., Loveridge, A. J., Chamaillé-Jammes, S., Davidson, Z., Murindagomo, F., Fritz, H., & Macdonald, D.

W. (2009). Behavioral adjustments of African herbivores to predation risk by lions: Spatiotemporal

variations influence habitat use. *Ecology*, *90*(1), 23–30. https://doi.org/10.1890/08-0606.1

Valeix, M., Loveridge, A. J., Davidson, Z., Madzikanda, H., Fritz, H., & Macdonald, D. W. (2010). How key

habitat features influence large terrestrial carnivore movements: Waterholes and African lions in a semi-

arid   savanna   of   north-western   Zimbabwe.   *Landscape   Ecology*,   *25*(3),   337–351.

https://doi.org/10.1007/s10980-009-9425-x

Valeix, M., Loveridge, A. J., & Macdonald, D. W. (2012). Influence of prey dispersion on territory and group size

of African lions: A test of the resource dispersion hypothesis. *Ecology*, *93*(11), 2490–2496.

https://doi.org/10.1890/12-0018.1

Wijers, M., Trethowan, P., du Preez, B., Chamaillé-Jammes, S., Loveridge, A. J., Macdonald, D. W., & Markham,

A. (2021). The influence of spatial features and atmospheric conditions on African lion vocal behaviour.

Animal Behaviour, *174*, 63–76. https://doi.org/10.1016/j.anbehav.2021.01.027

Zabihi-Seissan, S., Prokopenko, C. M., & Vander Wal, E. (2022). Wolf spatial behavior promotes encounters and

kills of abundant prey. *Oecologia*, *200*(1), 11–22. https://doi.org/10.1007/s00442-022-05218-4

Table 1. Summary of the candidate models explaining the variability of the frequency (i.e., the number of visits

over 100 days) of lion visits to a waterhole. All models included the number of waterholes within the territory of

each animal-year and a random effect with the identity of each animal-year. The most parsimonious models

(ΔAICc < 2 and lowest number of parameter *K*) are shown in bold. For these models, we used a distance threshold

of 1km to consider that lions visited a waterhole.

Frequency of visits to a waterhole – 1km		Female lion			Male lion		
Candidate models	**K**	**AICc**	**ΔAICc**	**R²**	**AICc**	**ΔAICc**	**R²**
Null hypothesis							
Number of waterholes	4	1752,1	16,5	0,30	2413,8	71,4	0,06
Prey abundance hypothesis							
Herd	5	1738,8	3,2	0,35	2391,4	49,0	0,14
Buffalo	5	1747,1	11,5	0,32	2410,3	67,9	0,08
Herd + Buffalo	6	1739,1	3,4	0,35	2393,1	50,7	0,14
Functional response in prey-abundance hypothesis							
Herd × Herd Territory	7	1741,5	5,9	0,35	2391,9	49,6	0,15
Buffalo × Buffalo Territory	7	1742,7	7,1	0,35	2412,1	69,7	0,09
Herd × Herd Territory + Buffalo × Buffalo Territory	10	1742,9	7,3	0,37	2396,3	54,0	0,16
Prey catchability hypothesis							
Distance to cover	5	1753,4	17,8	0,30	2398,3	55,9	0,13
vegetation openness	5	1742,9	7,2	0,32	2350,9	8,5	0,25
Distance to cover + vegetation openness	6	1744,0	8,4	0,33	2352,8	10,4	0,26
Functional response in prey-catchability hypothesis							
Distance to cover × Herd Territory	7	1757,3	21,7	0,30	2398,6	56,2	0,14
vegetation openness × Herd Territory	7	1741,7	6,0	0,34	2351,3	9,0	0,26
Distance to cover × Buffalo Territory	7	1754,3	18,7	0,31	2393,8	51,5	0,17
vegetation openness × Buffalo Territory	7	1738,4	2,8	0,36	2346,3	3,9	0,28
Distance to cover × (Herd Territory + Buffalo Territory)	9	1757,9	22,2	0,32	2395,1	52,8	0,18
vegetation openness × (Herd Territory + Buffalo Territory)	9	1740,1	4,4	0,36	2347,2	4,9	0,28

Model						
vegetation openness + Distance to cover × (Herd Territory + Buffalo Territory)	10	1746,4	10,7	0,35	2347,7	5,3	0,29
Distance to cover + vegetation openness × (Herd Territory + Buffalo Territory)	10	1741,4	5,7	0,37	2349,3	7,0	0,28
Prey-abundance + Prey-catchability hypothesis							
Herd + Distance to cover	6	1740,7	5,0	0,35	2380,8	38,4	0,18
Herd + vegetation openness	6	**1736,3**	**0,6**	**0,36**	2346,0	3,6	0,27
Herd + Distance to cover + vegetation openness	7	1737,8	2,2	0,36	2348,0	5,6	0,27
Buffalo + Distance to cover	6	1748,9	13,3	0,32	2395,1	52,7	0,16
Buffalo + vegetation openness	6	1738,6	2,9	0,35	**2344,2**	**1,8**	**0,28**
Buffalo + Distance to cover + vegetation openness	7	1739,0	3,3	0,35	2346,0	3,7	0,29
Herd + Buffalo + Distance to cover	7	1741,0	5,4	0,35	2382,3	39,9	0,18
Herd + Buffalo + vegetation openness	7	1735,9	0,2	0,36	2343,8	1,5	0,28
Herd + Buffalo + Distance to cover + vegetation openness	8	1737,0	1,4	0,37	2345,8	3,4	0,28
Prey abundance × Prey catchability hypothesis							
Herd × Distance to cover	7	1740,1	4,4	0,35	2377,8	35,4	0,20
Herd × vegetation openness	7	1735,6	0,0	0,36	2347,3	4,9	0,27
Herd × (Distance to cover + vegetation openness)	9	1737,8	3,1	0,36	2351,0	8,6	0,27
Buffalo × Distance to cover	7	1750,2	14,5	0,33	2393,6	51,3	0,17
Buffalo × vegetation openness	7	1740,5	4,9	0,35	2343,6	1,2	0,29
Buffalo × (Distance to cover + vegetation openness)	9	1741,7	6,1	0,36	2345,7	3,3	0,29
Herd × Distance to cover + vegetation openness	8	1737,8	2,2	0,36	2349,1	6,7	0,27
Herd × vegetation openness + Distance to cover	8	1737,2	1,6	0,36	2349,4	7,0	0,27
Buffalo × Distance to cover + vegetation openness	8	1739,6	4,0	0,36	2344,3	2,0	0,29
Buffalo × vegetation openness + Distance to cover	8	1741,0	5,3	0,35	2345,5	3,1	0,29
(Herd + Buffalo) × Distance to cover + vegetation openness	10	1736.3	0.6	0.38	2343,6	1,2	0,30
(Herd + Buffalo) × vegetation openness + Distance to cover	10	1736.6	0.9	0.38	2342,4	0,0	0,30
(Herd + Buffalo) × (vegetation openness + Distance to cover)	12	1738.7	3.1	0.38	2343,9	1,5	0,30

510　**Table 2.** Summary of the candidate models explaining the variability of the duration of lion visits to a waterhole.

511　All models included the number of waterholes within the territory of each animal-year and a random effect with

512　the identity of each animal-year. The most parsimonious model (ΔAICc < 2 and lowest number of parameter K)

513　is shown in bold. For these models, we used a distance threshold of 1km to consider that lions visited a waterhole.

Duration of visit to a waterhole – 1km		Female lion			Male lion		
Candidate models	**K**	**AICc**	**ΔAICc**	**R²**	**AICc**	**ΔAICc**	**R²**
Null hypothesis							
Number of waterholes	4	**4026,6**	**0,0**	**0,03**	5020,0	6,1	0,06
Prey abundance hypothesis							
Herd	5	4028,5	1,9	0,03	5019,1	5,1	0,07
Buffalo	5	4028,4	1,8	0,03	5018,0	4,1	0,07
Herd + Buffalo	6	4030,0	3,4	0,03	5018,8	4,9	0,08
Functional response in prey-abundance hypothesis							
Herd × Herd Territory	7	4030,1	3,5	0,04	5020,4	6,5	0,08
Buffalo × Buffalo Territory	7	4031,8	5,2	0,03	5021,5	7,6	0,07
Herd × Herd Territory + Buffalo × Buffalo Territory	10	4035,6	9,0	0,05	5024,6	10,7	0,09
Prey catchability hypothesis							
Distance to cover	5	4028,2	1,6	0,03	5020,3	6,3	0,06
vegetation openness	5	4028,0	1,4	0,03	**5015,8**	**1,9**	**0,08**
Distance to cover + vegetation openness	6	4030,0	3,4	0,03	5017,9	4,0	0,08
Functional response in prey-catchability hypothesis							
Distance to cover × Herd Territory	7	4030,3	3,8	0,04	5022,1	8,2	0,07
vegetation openness × Herd Territory	7	4028,4	1,8	0,05	5017,5	3,5	0,08
Distance to cover × Buffalo Territory	7	4031,7	5,1	0,03	5021,6	7,7	0,07
vegetation openness × Buffalo Territory	7	4027,5	0,9	0,05	5016,0	2,1	0,09
Distance to cover × (Herd Territory + Buffalo Territory)	9	4034,2	7,6	0,04	5023,3	9,4	0,08
vegetation openness × (Herd Territory + Buffalo Territory)	9	4030,0	3,4	0,06	5016,2	2,3	0,10
vegetation openness + Distance to cover × (Herd Territory + Buffalo Territory)	10	4036,2	9,6	0,04	5021,1	7,2	0,10

Distance to cover + vegetation openness × (Herd Territory + Buffalo Territory)	10	4032,1	5,5	0,06	5018,3	4,4	0,10
Prey abundance + Prey catchability hypothesis							
Herd + Distance to cover	6	4030,2	3,6	0,03	5019,6	5,6	0,07
Herd + vegetation openness	6	4030,0	3,4	0,03	5016,8	2,8	0,08
Herd + Distance to cover + vegetation openness	7	4032,0	5,4	0,03	5018,8	4,9	0,08
Buffalo + Distance to cover	6	4029,9	3,3	0,03	5018,6	4,6	0,08
Buffalo + vegetation openness	6	4029,7	3,1	0,03	5013,9	0,0	0,09
Buffalo + Distance to cover + vegetation openness	7	4031,7	5,1	0,03	5016,0	2,1	0,09
Herd + Buffalo + Distance to cover	7	4031,5	4,9	0,03	5019,5	5,5	0,08
Herd + Buffalo + vegetation openness	7	4031,6	5,0	0,03	5015,8	1,9	0,09
Herd + Buffalo + Distance to cover + vegetation openness	8	4033,6	7,0	0,03	5017,9	4,0	0,09
Prey abundance × Prey catchability hypothesis							
Herd × Distance to cover	7	4032,1	5,5	0,03	5021,2	7,3	0,07
Herd × vegetation openness	**7**	4030,1	3,5	0,04	5018,8	4,9	0,08
Herd × (Distance to cover + vegetation openness)	9	4034,2	7,6	0,04	5021,6	7,6	0,08
Buffalo × Distance to cover	7	4031,8	5,2	0,03	5019,5	5,6	0,08
Buffalo × vegetation openness	7	4030,2	3,6	0,04	5016,0	2,1	0,09
Buffalo × (Distance to cover + vegetation openness)	9	4033,2	6,6	0,04	5018,7	4,8	0,09
Herd × Distance to cover + vegetation openness	8	4034,0	7,4	0,03	5019,9	5,9	0,08
Herd × vegetation openness + Distance to cover	8	4032,1	5,5	0,04	5020,9	7,0	0,08
Buffalo × Distance to cover + vegetation openness	8	4033,6	7,0	0,03	5016,8	2,9	0,09
Buffalo × vegetation openness + Distance to cover	8	4032,2	5,6	0,04	5018,1	4,1	0,09
(Herd + Buffalo) × Distance to cover + vegetation openness	10	4036,9	10,3	0,04	5020,3	6,4	0,10
(Herd + Buffalo) × vegetation openness + Distance to cover	10	4035,5	8,9	0,04	5022,1	8,1	0,09
(Herd + Buffalo) × (vegetation openness + Distance to cover)	12	4038,7	12,1	0,05	5023,5	9,6	0,10

516 **Figure 1.** A patch of grassland surrounding a waterhole in Hwange National Park, Zimbabwe. Photo © Google

517 Earth

Figure 2. Influence of prey abundance and catchability on the frequency of visits of a waterhole by female (yellow) and male (red) lions. Ribbon extremities show 95% confidence interval, whereas lines show the mean value of the response variable. Predictions are made from models using a 1km distance threshold to consider that lions visit a

522   waterhole. The dotted horizontal line shows the frequency of visits to a waterhole expected from the null

523   hypothesis (i.e., the frequency of lion visits to a waterhole solely depends on the number of waterholes within their

524   territory). Prey-related metrics are centered and scaled per animal-year territory to compare the relative abundance

525   and the relative catchability of prey from one waterhole to another waterhole available in the animal-year territory.

Figure 3. Influence of the relative vegetation openness in the surroundings of a waterhole on the durations of visits to a waterhole by male lions. Ribbon extremities show 95% confidence interval, whereas lines show the mean value of the response variable. Predictions are made from models using a 1km distance threshold to consider that lions visit a waterhole. The dotted horizontal line shows the frequency of visits to a waterhole expected from the null hypothesis (i.e., the frequency of lion visits to a waterhole solely depends on the number of waterholes within their territory). For male lions, one visit to a waterhole lasting 76h is not shown in the plot.

 Supplementary Information

Table S1. Results using a 2km distance threshold to define lion visits to a waterhole. Summary of the candidate

models explaining the variability in the frequency of lion visits to a waterhole. The frequency of lion visits to a

waterhole is calculated as the number of visits over 100 days. All models included the number of waterholes within

the territory of each animal-year and a random effect with the identity of each animal-year. The most parsimonious

model (ΔAICc < 2 and lowest number of parameter K) is shown in bold. For these models, we used a distance

threshold of 2km to consider that lions visited a waterhole.

Frequency of visits to a waterhole – 2km		Female lion			Male lion		
Candidate models	**K**	**AICc**	**ΔAICc**	**R²**	**AICc**	**ΔAICc**	**R²**
Null hypothesis							
Number of waterholes	4	1939,6	17,7	0,32	2652,8	64,8	0,06
Prey abundance hypothesis							
Herd	5	1933,9	12,0	0,35	2643,5	55,4	0,10
Buffalo	5	1937,9	16,0	0,33	2651,3	63,2	0,07
Herd + Buffalo	6	1935,0	13,1	0,35	2645,0	56,9	0,10
Functional response in prey-abundance hypothesis							
Herd × Absolute Herd	7	1936,4	14,5	0,35	2644,8	56,7	0,11
Buffalo × Absolute Buffalo	7	1937,2	15,3	0,34	2653,9	65,8	0,08
Herd × Absolute Herd + Buffalo × Absolute Buffalo	10	1939,6	17,7	0,36	2649,6	61,6	0,11
Prey catchability hypothesis							
Distance to cover	5	1939,1	17,1	0,32	2647,3	59,3	0,09
Habitat openness	5	1924,5	2,6	0,35	2593,7	5,7	0,23
Distance to cover + Habitat openness	6	1926,6	4,6	0,35	2594,5	6,4	0,23
Functional response in prey-catchability hypothesis							
Distance to cover × Absolute Herd	7	1941,6	19,7	0,33	2648,9	60,9	0,10
Habitat openness × Absolute Herd	7	1927,0	5,0	0,36	2595,0	6,9	0,23
Distance to cover × Absolute Buffalo	7	1942,1	20,2	0,33	2645,1	57,0	0,11
Habitat openness × Absolute Buffalo	7	1928,0	6,1	0,35	2591,2	3,1	0,25

Distance to cover × (Absolute Herd + Absolute Buffalo)	9	1944,0	22,1	0,34	2645,4	57,3	0,13
Habitat openness × (Absolute Herd + Absolute Buffalo)	9	1930,0	8,1	0,36	2591,1	3,0	0,26
Habitat openness + Distance to cover × (Absolute Herd + Absolute Buffalo)	10	1930,2	8,3	0,37	2589,7	1,6	0,27
Distance to cover + Habitat openness × (Absolute Herd + Absolute Buffalo)	10	1932,1	10,2	0,36	2592,4	4,3	0,26
Prey abundance + Prey catchability hypothesis							
Herd + Distance to cover	6	1933,9	12,0	0,35	2640,6	52,5	0,11
Herd + Habitat openness	6	**1923,6**	**1,7**	**0,36**	2593,5	5,4	0,24
Herd + Distance to cover + Habitat openness	7	1925,7	3,8	0,36	2594,3	6,2	0,24
Buffalo + Distance to cover	6	1938,1	16,1	0,33	2646,2	58,2	0,10
Buffalo + Habitat openness	6	**1923,6**	**1,7**	**0,36**	**2589,8**	**1,8**	**0,25**
Buffalo + Distance to cover + Habitat openness	7	1925,5	3,6	0,36	2590,2	2,1	0,25
Herd + Buffalo + Distance to cover	7	1935,3	13,4	0,35	2642,0	54,0	0,12
Herd + Buffalo + Habitat openness	7	1924,3	2,4	0,36	2591,5	3,4	0,25
Herd + Buffalo + Distance to cover + Habitat openness	8	1926,3	4,4	0,36	2591,9	3,8	0,25
Prey abundance × Prey catchability hypothesis							
Herd × Distance to cover	7	1935,4	13,5	0,35	2637,4	49,3	0,13
Herd × Habitat openness	7	1921,9	0	0,37	2595,4	7,4	0,24
Herd × (Distance to cover + Habitat openness)	9	1925,9	3,9	0,37	2598,1	10,0	0,24
Buffalo × Distance to cover	7	1939,0	17,1	0,34	2646,1	58,1	0,11
Buffalo × Habitat openness	7	1925,5	3,6	0,35	2591,0	2,9	0,25
Buffalo × (Distance to cover + Habitat openness)	9	1927,1	5,2	0,36	2590,1	2,0	0,26
Herd × Distance to cover + Habitat openness	8	1927,7	5,7	0,36	2596,1	8,0	0,24
Herd × Habitat openness + Distance to cover	8	1924,0	2,1	0,37	2596,1	8,1	0,24
Buffalo × Distance to cover + Habitat openness	8	1925,6	3,6	0,36	2588,1	0	0,26
Buffalo × Habitat openness + Distance to cover	8	1927,4	5,5	0,35	2591,4	3,4	0,25
(Herd + Buffalo) × Distance to cover + Habitat openness	10	1927,7	5,8	0,37	2590,8	2,7	0,27
(Herd + Buffalo) × Habitat openness + Distance to cover	10	1926,4	4,5	0,37	2593,7	5,6	0,26
(Herd + Buffalo) × (Habitat openness + Distance to cover)	12	1928,5	6,6	0,38	2594,4	6,3	0,27

544 **Table S3.** Results using a 2km distance threshold to define lion visits to a waterhole. Summary of the candidate

545 models explaining the variability in the duration of lion visits to a waterhole. All models included the number of

546 waterholes within the territory of each animal-year and a random effect with the identity of each animal-year. The

547 most parsimonious model (ΔAICc < 2 and lowest number of parameter K) is shown in bold. For these models, we

548 used a distance threshold of 2km to consider that lions visited a waterhole.

Duration of visit to a waterhole – 2km		Female lion			Male lion		
Candidate models	**K**	**AICc**	**ΔAICc**	**R²**	**AICc**	**ΔAICc**	**R²**
Null hypothesis							
Number of waterholes	4	**4684,2**	**5,0**	**0,0**	5742,3	6,5	0,07
Prey abundance hypothesis							
Herd	5	4684,6	5,5	0,01	5740,2	4,3	0,08
Buffalo	5	4685,4	6,3	0,0	5744,4	8,5	0,07
Herd + Buffalo	6	4686,5	7,3	0,01	5742	6,1	0,08
Functional response in prey-abundance hypothesis							
Herd × Absolute Herd	7	4684,3	5,2	0,04	5741,2	5,4	0,10
Buffalo × Absolute Buffalo	7	4687,1	7,9	0,02	5747,8	12,0	0,07
Herd × Absolute Herd + Buffalo × Absolute Buffalo	10	4689,5	10,3	0,04	5747,2	11,3	0,10
Prey catchability hypothesis							
Distance to cover	5	4684,3	5,2	0,01	5743,5	7,6	0,07
Habitat openness	5	4684,6	5,5	0,01	**5736,0**	**0,1**	**0,09**
Distance to cover + Habitat openness	6	4681,4	2,2	0,03	5737,7	1,8	0,09
Functional response in prey-catchability hypothesis							
Distance to cover × Absolute Herd	7	4684,8	5,6	0,03	5744,2	8,3	0,09
Habitat openness × Absolute Herd	7	4684,2	5,1	0,04	5737,0	1,1	0,11
Distance to cover × Absolute Buffalo	7	4687,0	7,8	0,01	5746,8	10,9	0,08
Habitat openness × Absolute Buffalo	7	4683,1	3,9	0,03	5739,0	3,1	0,10
Distance to cover × (Absolute Herd + Absolute Buffalo)	9	4688,2	9,0	0,04	5748	12,1	0,09
Habitat openness × (Absolute Herd + Absolute Buffalo)	9	4684,6	5,4	0,05	5740,6	4,7	0,11

Habitat openness + Distance to cover × (Absolute Herd + Absolute Buffalo)	10	4684,1	5,0	0,06	5742,1	6,2	0,11
Distance to cover + Habitat openness × (Absolute Herd + Absolute Buffalo)	10	4680,0	0,8	0,08	5742,4	6,5	0,11
Prey abundance + Prey catchability hypothesis							
Herd + Distance to cover	6	4684,7	5,5	0,01	5741,7	5,8	0,08
Herd + Habitat openness	6	4685,5	6,4	0,01	5735,9	0,0	0,10
Herd + Distance to cover + Habitat openness	7	4682,5	3,4	0,03	5737,6	1,7	0,10
Buffalo + Distance to cover	6	4685,2	6,0	0,01	5745,5	9,6	0,07
Buffalo + Habitat openness	6	4685,8	6,7	0,01	5738,0	2,1	0,09
Buffalo + Distance to cover + Habitat openness	7	4681,5	2,3	0,03	5739,7	3,8	0,09
Herd + Buffalo + Distance to cover	7	4686,3	7,1	0,01	5743,4	7,5	0,08
Herd + Buffalo + Habitat openness	7	4687,3	8,1	0,01	5737,9	2,0	0,10
Herd + Buffalo + Distance to cover + Habitat openness	8	4683,4	4,3	0,03	5739,7	3,8	0,10
Prey abundance × Prey catchability hypothesis							
Herd × Distance to cover	7	4685,9	6,7	0,02	5743,7	7,8	0,08
Herd × Habitat openness	**7**	4683,1	3,9	0,03	5737,6	1,7	0,10
Herd × (Distance to cover + Habitat openness)	9	4682,7	3,5	0,04	5741,5	5,6	0,10
Buffalo × Distance to cover	7	4687,3	8,1	0,01	5747,6	11,7	0,07
Buffalo × Habitat openness	7	4683,9	4,7	0,02	5739,9	4,0	0,09
Buffalo × (Distance to cover + Habitat openness)	9	4681,2	2,0	0,05	5743,7	7,8	0,09
Herd × Distance to cover + Habitat openness	8	4683,4	4,2	0,03	5739,6	3,7	0,10
Herd × Habitat openness + Distance to cover	8	4680,6	1,4	0,04	5739,4	3,5	0,10
Buffalo × Distance to cover + Habitat openness	8	4683,6	4,4	0,03	5741,7	5,8	0,09
Buffalo × Habitat openness + Distance to cover	8	4679,2	0	0,05	5741,6	5,7	0,09
(Herd + Buffalo) × Distance to cover + Habitat openness	10	4686,1	7,0	0,04	5743,8	7,9	0,10
(Herd + Buffalo) × Habitat openness + Distance to cover	10	4681,7	2,5	0,05	5743,5	7,6	0,10
(Herd + Buffalo) × (Habitat openness + Distance to cover)	12	4685,4	6,2	0,06	5747,7	11,8	0,10

549
550
551

552   **T**able S3. Herd size of medium-sized prey in Hwange National Park. These sizes were calculated with both adult

553   and juvenile individuals.

Species	Mean	Lower CI 95%	Upper CI 95%
Eland	6.3	4.5	8.0
Greater kudu	4.4	4.3	4.5
Roan antelope	3.8	3.5	4.1
Sable antelope	4.7	4.3	5.1
Blue wildebeest	18.2	16.1	20.3
Plains zebra	7.3	7.1	7.5

555
556