Activation dynamics and assembly of root zone soil bacterial communities in response to

stress-associated phytohormones

Running title

Activation of bacterial communities to phytohormones

Keywords

Plant microbiome, rhizosphere, salicylic acid, abscisic acid, mesocosm, 16S ratios, amplicon

sequencing, transcriptionally active populations, reactivation, before after control impact design

(BACI)

Authors

Sreejata Bandopadhyay[1,2], Oishi Bagchi[1,2], Ashley Shade[2#]

Affiliations

1. The Plant Resilience Institute, Michigan State University, 567 Wilson Rd, East Lansing MI

48824

2. The Great Lakes Bioenergy Research Center, Michigan State University, 567 Wilson Rd, East

Lansing MI 48824

3. Universite Claude Bernard Lyon 1, CNRS, INRAE, VetAgro Sup, Laboratoire d'Ecologie

Microbienne LEM, CNRS UMR5557, INRAE UMR1418, Villeurbanne, F-69100 France

Correspondance : **ashley.shade@cnrs.fr**

Abstract

Plants can "cry for help" to recruit supportive microbiome members during stressful conditions.

We evaluated the activation dynamics of root zone soil bacteria in response to phytohormones

produced when plants are stressed, hypothesizing that the activated taxa support plant resilience.

We conducted a 2-week laboratory experiment using mesocosms of root zone soil collected from

two different crops: the annual legume common bean (*Phaseolus vulgaris* L.) and the perennial

grass switchgrass (*Panicum virgatum*). We inactivated the microbiome by drying and then treated

the soils with either abscisic acid, salicylic acid, a carrier control (methanol), or water, and then

quantified the reactivation dynamics of bacterial populations over time, at one, 7, and 14 days after

phytohormone addition, using amplicon sequencing of 16S rRNA and rDNA. There were several

Actinobacterial taxa that switched from an average population-inactive to a population-active state

after exposure to abscisic acid and salicylic acid, with *Microbispora* lineages switching especially

noted. Some taxa were activated only in one crop's soil, and some were activated in both crops'

soils in response to the same phytohormone. This work suggests that different bacteria have

different specificities to phytohormones as plant stress signals and provides insights into

understanding the mechanisms by which stressed plants may "cry for help" to recruit bacteria from

the root zone to the rhizosphere.

Importance

Global food security is an urgent societal challenge that has been intensified by climate change

and other anthropogenic stressors placed on the environment. Microbial bioinoculants are a

promising solution to improve crop health and resilience, but ensuring their persistence and

activation in the field remains a challenge. This study examined how dormant root-zone associated

bacteria reactivate after exposure to the plant stress hormones abscisic acid and salicylic acid. The

experiment revealed that certain bacteria taxa could reactivate in response to these plant stress

signals and persist for at least two weeks. This work advances understanding of the potential cues

for reactivation of beneficial plant-associated microbes and supports goals toward developing

microbial solutions for sustainable agriculture.

Introduction

Plant hormones, called *phytohormones*, are essential for a plant's response to stress. Phytohormones improve germination, upregulate antioxidant systems to reduce reactive oxygen species (ROS) accumulation, induce systemic acquired resistance, and assist in cell signaling. They are also vital growth regulators that can impact plant metabolism and are often externally applied to plants to improve plant growth under stress conditions (1). Several microbes are known to produce phytohormones or phytohormone mimics (2-5) which influence phytohormone networks in plants and subsequently affect microbial metabolism. Thus, phytohormone regulation in plants is considered to be connected to microbial phytohormone production, together forming inter-organismal systems that can support plants (6).

The plant rhizosphere and proximal root zone are nutrient-rich compartments due to root exudation and thus are heavily colonized by microbes (7, 8). Sugars and amino acids within root exudates serve as nutrients for microbes (9, 10), which in turn can assist plants by improving their nutrient acquisition (11) and protecting against stress (12, 13). Specifically, microbial production of phytohormones (6) such as abscisic acid (ABA), auxins (indole 3-acetic acid), salicylic acid (SA), and cytokinins can improve a plant's ability to cope with stress. These well-known plant-microbe interactions have been extensively investigated (6, 14-16).

The phytohormones ABA and SA play an important role in plant abiotic stress tolerance. Abscisic acid belongs to a group of phytohormones called sesquiterpenoids, which regulate plant growth. ABA-induced signaling can regulate the expression of stress-responsive genes and contribute to stress tolerance (17). During drought, ABA induces stomatal closure and thus controls transpiration (18). Exogenous ABA application to plant seeds before sowing can protect a drought-sensitive wheat cultivar from drought-induced oxidative damage by increasing the activity

of the antioxidant enzyme peroxidase (19). Salicylic acid also modulates antioxidant enzyme

activities (20, 21). Similar to ABA, the application of SA to plant tissues has been shown to

relieve water stress (22), potentially enabling continued plant growth despite drought (21).

While much is understood about a plant's coping mechanisms during stress via

phytohormone regulation of cellular pathways, we are still learning how phytohormones impact

rhizosphere microbiome assembly during stress. It has been shown that SA signaling can increase

microbial community diversity and result in the colonization of specific taxa (23-25). Similarly,

low molecular weight organic acids such as malic acid, citric acid, and fumaric acid secreted by

roots have also been shown to serve as a source of carbon substrate and signaling molecule for

rhizobacteria recruitment (26-28). Direct evidence of phytohormones serving as a carbon source

for microbes is more limited; however, some studies demonstrated increased soil respiration when

treated with one mM abscisic acid (29) and higher soil respiration when soils were treated with

root exudates from drought exposed plants compared to exudates from non-droughted plants (30).

However, there is indirect evidence that rhizobacteria recruitment can be enhanced in response to

specific phytohormone precursors such as L-tryptophan, a precursor for auxin synthesis (31). This

study showed that auxin-producing *Pseudomonas fluorescens* stimulated the root growth of radish,

which exudes high levels of tryptophan. If microbes use phytohormones as a carbon source, it is

possible that they could cause a soil-priming effect that stimulates microbial respiration. This has

been shown in experiments where low levels of jasmonic acid (JA) and 1-aminocyclopropane-1-

carboxylic acid (ACC) addition to soils caused a relatively large increase in soil respiration (29).

Such priming effects have been documented for various organic substrates in root exudates (32-

35).

The rhizosphere microbiome members recruited during stress are expected to support a

plant's ability to recover from that stress or exploit the plant during stress. Thus, understanding the

role of stress-associated phytohormones in determining rhizosphere microbiome assembly is

important for fully understanding the complex dialogue and feedback between a plant and its

microbiome. Phytohormone release or leakage into the root zone is also one potential mechanism

by which plants may "cry for help" via root exudation to attract supportive microbial partners from

the root zone to the rhizosphere (36).

A complicating factor in understanding plant-soil-microbial interactions during stress is the

unclear contribution of those soil microbes that reactivate from dormancy. A large fraction of the

soil bacterial community can be dormant at a given time (80-90% of cells and 55% of taxa) with

dormant cells remaining inactive for prolonged periods (37). Even nutrient-rich rhizosphere soil

can have substantial microbial dormancy, with around 40-80% inactive cells reported in the

rhizosphere soil of bean plants (38). Thus, we reasoned that the rhizosphere and proximal root-

zone compartment may be highly dynamic in activity switching during stress, with some taxa

reactivating in response to plant signals (and possibly supporting the plant's stress response) while

others initiate dormancy to avoid the stress by protecting themselves (39).

The overarching objective of this study was to understand how the root-zone bacterial

microbiome activates and inactivates in response to phytohormones that signal plant stress. Our

null hypothesis was that the microbiome does not react to phytohormone exposure. We had three

alternate and non-mutually exclusive hypotheses:

H1: The microbiomes of root-zone soils from different plants have different

activation and assembly dynamics in response to phytohormones. The motivation for

this hypothesis is our previous work on these plant genotypes (40) and the literature, which

shows that different plant species harbor different bacterial microbiome structures. If

supported, this hypothesis suggests the importance of the plant's ability to locally assemble

and activate "custom" microbiome members to respond to its own phytohormones.

H2: Certain bacterial lineages generally activate in response to phytohormones

regardless of the plant. The rationale for this hypothesis is that the phytohormones could

serve as resources for microbes. Therefore, the bacteria that can utilize exogenous

phytohormones may generally activate if provided either SA or ABA from any plant

species. This activation pattern may indicate that taxa respond to resource availability, not

a stress-specific "cry for help."

H3: There is a hormone-specific activation of certain bacterial taxa. Each

phytohormone indicates a different type of stress for the plant: mainly pathogen infection

in the case of SA and drought in the case of ABA. Thus, if some bacterial lineages activate

after exposure to one of these phytohormones but not both, it would be consistent with a

bacterial taxon's response to a plant's "cry for help" given a specific stress.

We addressed these hypotheses using a mesocosm experiment in which we exposed root-

zone soils to either a stress-associated phytohormone or controls. We collected root-influenced soil

adjacent to plant roots (<20cm) from two different cropping systems: common bean (*Phasoelus*

vulagris) and switchgrass (*Panicum virgatum*). These two plants were chosen for their different

life strategies (annual and perennial) and families (Fabacaea and Poaceae) and, therefore, to

provide a contrast for understanding the potential generalizability of microbiome responses to

phytohormones. We used field soils to retain the legacy of the crop in assembling the bacterial

communities relevant to agricultural conditions. We applied the stress-associated phytohormones

abscisic acid (drought) and salicylic acid (pathogen infection) to the mesocosm soils. We used 16S

rRNA and rDNA amplicon sequencing to designate a taxon-level activity state (likely active,

likely inactive) before and after phytohormone application. We also related microbiome data to

soil nutrient concentrations to understand its impact on the microbial community shifts in activity

across treatments and over time.

Results

Experimental design.

There were four experimental conditions including controls: methanol carrier control (water and

methanol), abscisic acid (water, methanol, and abscisic acid), salicylic acid (water, methanol, and

salicylic acid), and a water control (water only) and three time points (Day 1, Day 7 and Day 14).

Each condition was replicated four times for 16 mesocosms (four conditions, four replicates) for

each soil type (See **Figure 1** for experimental design). This design was executed for two soils

(bean and switchgrass root zone), resulting in 32 mesocosms. Including the field soil, pre- and

post-dry samples, water acclimation, and the time series for each experimental condition, there

were 333 microbiome samples across the bean and switchgrass soils.

Sequencing summary

We sequenced a total of 333 DNA samples and 333 cDNA samples from this experiment. After

rarefying all read libraries to 12,000 reads per sample (**Figure S1A**), 324 DNA and 324 cDNA

samples met the minimum observation effort. After removing all contaminants from the OTU

table using package decontam (**Figure S1 B, C**), 13,138 OTUs were used to determine the active

community.

Immediate activation and persistence of the root zone bacterial microbiome to phytohormones

After determining the likely active taxa at Day 1 post-hormone addition, we assessed their relative

abundances (DNA counts) before and after phytohormone application. To focus on enrichments

observed only in the phytohormone treatments, we also compared the enrichment of specific taxa

after hormone addition (Day 1) to the water-acclimated sample and the water and methanol

controls. There were statistical differences in the active bacterial community structure between the

water-acclimated and Day 1 for bean ABA soil (PERMANOVA F=3.95, R2=0.15, p=0.002),

bean SA soil (PERMANOVA F=3.96, R2=0.15, p=0.006), switchgrass ABA soil

(PERMANOVA F=10.50, R2=0.32, p=0.001) and switchgrass SA soil (PERMANOVA F=2.31,

R2=0.09, p=0.002).

In the ABA-treated soils, three OTUs were enriched by log two-fold at Day 1 compared to

the water-acclimated sample and not enriched in water and methanol controls (**Figure S2A**).

These included Actinobacteriota OTUs from Family Myxococcaceae, genus *Microbispora*, and

genus *Streptomyces*. However, these enrichments were detected only in the switchgrass soils, and

there were no OTUs enriched in bean soils treated with ABA on Day 1. Furthermore, all three taxa

enriched in the switchgrass ABA soils switched from either an inactive or below detection state in

the water-acclimated sample to high abundance at Day 1 after hormone addition (**Figure S3B**).

Similarly, for the SA treatments, the two OTUs belonging to genus *Microbispora* were log

two-fold enriched at Day 1 compared to the water acclimated sample in both bean and switchgrass

and were enriched solely in the SA treatments to the exclusion of the methanol and water controls

(**Figure S2B**). Furthermore, the response pattern suggested a likely switch from inactive (detected

in DNA but not RNA) to active for both taxa. One was enriched on Day 1 in bean soils, and the

other OTU was enriched on Day 1 in both bean and switchgrass (**Figure S3 C, D**)

Several taxa belonging to classes Myxococcia, Actinobacteria, Alphaproteobacteria, and

Bacilli were resuscitated within one day after SA addition to the bean soil (**Figure S3C**), but most

of these taxa were also responsive to methanol (**Figure S3E**). Two Actinobacterial OTUs (both

belonging to the genus *Microbispora*) were activated in response to SA treatment in bean soils and

not in the methanol controls. These two OTUs were log two-fold enriched after Day 1 of SA

exposure compared to the water-acclimated soils. However, only one of these *Microbispora* OTUs

was represented in the top 50 abundant and active taxa for the bean SA treatment but both are

included for comparison (**Figure S3C**). Both taxa were detected as active in methanol controls but

not as consistently enriched as in the SA treatment. However, apart from the log two-fold

enrichments mentioned above, no substantial response to ABA or SA was noted among other

active OTUs for switchgrass beyond their similar response to the methanol on Day 1.

While several active taxa increased abundances after phytohormone addition (**Figure 2**),

many of these were also active in the methanol controls (**Figure S4**). However, some active taxa

persisted past Day 1, a pattern consistent with growth if the phytohormones were used as a

resource. We observed the persistence of active taxa over 14 days for four of the enriched taxa

(indicated in red in **Figure 2**), including a *Microbispora* for switchgrass and bean SA and

switchgrass ABA and a Myxococcaceae and *Streptomyces* in switchgrass ABA. Notably,

persistent activation over time could also be an indirect response to the phytohormone addition

(e.g., due to priming), which would not be distinguishable here.

Addressing H1: The microbiomes of root-zone soils from different plants have different activation

and assembly dynamics in response to phytohormones.

Bean and switchgrass soils differed distinctly in their active community structure across all time

points and treatments (PERMANOVA F=29.05, R2=0.08, p=0.001). For the ABA treatment, there

was a higher proportion of Bacilli and Verrucomicrobiae in bean soils on Day 1 after ABA

addition, but a higher proportion of Actinobacteria and Myxococcia in switchgrass soils. Then,

Alphaproteobacteria increased proportion 7 and 14 days after ABA exposure in switchgrass

compared to bean soils (**Figure 3**). For the SA treatment, there was also a relatively higher

abundance of active Actinobacteria on Day 1 in switchgrass soils than in bean soils. However,

Actinobacteria declined in switchgrass soils over time and was generally replaced with

Alphaproteobacteria. In bean soils, active Actinobacteria increased over time, with a concurrent

relative decline in several other bacterial classes (**Figure 3**).

Several taxa were enriched on Day 1 after ABA addition in switchgrass soils compared to

the water-acclimated control. Specifically, OTUs belonging to phyla Actinobacteriota and

Myxococcota were enriched by log two-fold on Day 1 compared to the pre-hormone time point

(**Figure S2**). Bean soil, however, showed no taxa enriched on Day 1 in response to ABA. For SA

treatment, one OTU belonging to the Microbispora was enriched on Day 1 in both bean and

switchgrass. A different *Microbispora* OTU was also specifically enriched in bean soil in response

to SA treatment (**Figure S2**).

Addressing H2: Certain bacterial lineages generally activate in response to phytohormones

regardless of the plant.

One genus *Microbispora* OTU, was enriched in response to both ABA and SA in switchgrass soils

on Day 1 compared to the water-acclimated sample. This same OTU was also enriched in bean

soils in response to SA. Thus, this lineage responded consistently across different soils and

phytohormones.

We also related changes in the active bacterial community to measured soil properties to

consider any co-occurring factors (**Figure S5).** We found that consistent phytohormone responses

by microbial taxa were unlikely to be attributable to concurrent pH or organic matter changes

because the treatments generally had different properties (**Figure S5**). Nitrate (in bean) and

ammonium (in switchgrass) explained the variation in the soil communities after receiving water

(**Figure S5**). Overall, soil differences in pH, nitrate concentrations, and organic matter content

explained about 27% of the variation in the active bacterial community in bean soils. In contrast,

pH, ammonium concentration, and organic matter content explained about 22% of the variation in

the active bacterial community in switchgrass.

Addressing H3: There is a hormone-specific activation of certain bacterial taxa.

We found no strong evidence of any taxa common to both plants' root zone soils and similarly

responsive to either ABA or SA exposure, and thus, H3 was not supported. Two OTUs (family

Myxococcaceae and genus *Streptomyces*) were enriched solely in response to ABA in switchgrass.

Similarly, one *Microbispora* OTU in bean soil was exclusively enriched in response to SA,

indicating a hormone-specific and soil-specific response from these taxa.

However, there was a generally consistent response dynamic of the active bacterial

microbiome's structure (beta diversity) to phytohormone addition regardless of the soil origin

(**Figure 4, Table 2**). Phytohormone treatment and time both explained global variation in the

active community structure. When excluding the three baseline samples (field, pre-dry, post-dry),

we found that bacterial community richness was statistically different between crop type (ANOVA

F=36.73, SumSq=46434, p<0.001) and time point (water acclimated, Day 1, Day 7 and Day 14)

(ANOVA F=8.22, SumSq=31165, p<0.001). In contrast, Inverse Simpson's diversity, which

accounts for richness and evenness, differed only by time point (ANOVA F=3.04, SumSq=35840,

p=0.03). There was also a strong interaction between crop and treatment (abscisic acid, salicylic

acid, methanol control, and water control) (ANOVA F=12.78, SumSq= 32290, p<0.001). At the

same time, treatment strongly influenced Inverse Simpson's diversity (ANOVA F=4.81,

SumSq=47335, p=0.003). Overall, we found that richness decreased over time for abscisic and

salicylic acid treatments across both soils (**Figure S6**).

Discussion

This study assessed the activation and assembly dynamics of activated root zone bacterial taxa in

response to phytohormones produced by plants during stress. Thus, this study provides new

insights into which plant-associated bacterial populations immediately activate in response to SA

and ABA phytohormones, suggesting that these molecules could serve as reactivation signals.

Furthermore, this study also distinguished general and specific activation dynamics of bacterial

taxa for two different plants' root zone soils over two weeks after exposure, allowing improved

precision on the ecological underpinnings of these responses. Ultimately, we can weigh the

evidence that the activated taxa responded to the phytohormones as a "cry for help" from the plant.

In this study, crop-specific responses of the root zone microbiome to phytohormones

indicated a plant legacy on its associated soil that partially determines its microbiome's response,

providing evidence in support of H1. Given that most plant species harbor unique microbial

communities (41-43), it is not unreasonable that each plant species may have a different

microbiome response to stress indicators like phytohormones. While switchgrass responded to

ABA via immediate enrichment of three OTUs within 24 hours of phytohormone application, bean

soils show no enrichment of any taxa in response to ABA. Similarly, one Actinobacterial OTU

was enriched in response to SA treatment only in bean but not in switchgrass soils, suggesting a

legacy effect that impacts the recruitment of active members to the bean rhizobiome. This finding

agrees with the results observed in our previous study that assessed bacterial reactivation after

drought (40). Other studies have also shown that root exudate-induced soil respiration can be

observed nearly immediately within 6-21 h (30).

Actinobacterial genera are widely reported to be associated with drought stress in plants

(44), so it was not surprising that we detected several Actinobacterial taxa that activated in

response to ABA and SA exposure, showing support for H2 among these lineages. Of the most

abundant active members in bean soils exposed to ABA and SA, we found an

Alphaproteobacterial OTU (*Microvirga*) that activated after hormone addition within 24 hours but

did not activate in the methanol control mesocosms within the same time frame. This

Alphaproteobacterial genus *Microvirga* has also been shown to alleviate drought stress in plants

such as cowpeas, such that cow peas inoculated with *M. vignae* had no differences between water

deficit and well-watered conditions concerning stomatal conductance, shot dry mass, or N

accumulation (45). This same study found that ABA biosynthesis and ABA-dependent genes were

less upregulated under drought when plants were inoculated with a *Microvirga sp.* compared to a

Bradyrhizobium sp, suggesting that the presence of *Microvirga* relatively decreased the plant's

typical ABA production during stress.

We found a specific Actinobacterial genus, *Microbispora,* that consistently responded to

both ABA and SA across bean and switchgrass. This response pattern indicates that this microbial

group could be a generic responder to phytohormones as stress signals. Actinobacteria have been

widely explored as targets for developing bioinoculants. *Microbispora* and *Streptomyces* have

been suggested to be both soil-inhabiting and endophytic genera, though only a limited number of

species are likely to have this dual habitat preference (46). *Microbispora* and other Actinobacterial

species belonging to *Actinoplanes, Streptomyces, Rhodococcus*, and *Micromonospora* have also

been explored for direct/indirect plant growth-promoting capabilities via secretion of

phytohormones, siderophores, and antimicrobial agents (47, 48).

Across the two soils used in this study (bean, switchgrass), we found no taxa responding

consistently to either ABA or SA, indicating that hormone-specific responses may not be

conserved across different plants, so H3 was not supported. However, we did find hormone-

specific responses to either ABA or SA within each soil type. For example, we found one

Actinobacterial species (*Microbispora*) that uniquely responded in bean to SA, while one

Actinobacterial (*Streptomyces*) and one Myxoccoccal OTU responded uniquely to ABA in

switchgrass. While we found signatures of Gram-positive bacteria activating to phytohormones,

other studies have found a declining abundance of Gram-positive biomarkers following

phytohormone addition (29). However, these other studies did not discriminate activity states but

rather assessed the microbial communities using phospholipid-derived fatty acid profiles, and a

negative response to phytohormones also may be due to an investment in complementary survival

strategies such as osmolyte production and spore production as opposed to growth and turnover

(49-51).

The literature shows that phytohormone signaling can play a major role in rhizobiome

assembly (25). For instance, using natural selection T. Kalachova et al. (52) showed a reduction in

disease severity of Arabidopsis plants infected with foliar pathogen *Pseudomonas syringae*. This

was associated with SA-mediated induction of defense mechanisms and a shift in the soil bacterial

community, but not the fungal. While one study showed that ABA and polyacrylamide treatments

to soil coupled with foliar application of ABA reduced the relative abundances of Actinobacteria

in the rhizosphere using amplicon DNA sequencing and increased drought resistance in forage

grass (53), other studies have shown that Actinobacteria increase in plant roots and rhizosphere

soil during drought likely due to expansion of C allocation in root exudates (44).

Furthermore, certain Actinobacteria, such as *Streptomyces,* have been shown to increase

the ABA content of wheat leaves to upregulate the expression of drought-resistance genes (54).

Thus, there could be implications for a positive feedback loop in Actinobacterial responses to a

plant's cry for help. The role of ABA as a carbon substrate for Actinobacteria is also indicated. For

example, *Rhodococcus* sp. utilized ABA as the sole carbon source in batch culture (55). These

findings support the idea that phytohormone substrates can be used to select rhizosphere microbes

(56) and corroborate our finding of active Actinobacterial in this study. Overall, there are limited

experiments that consider microbiome reactivation with plant signaling.

A limitation of this study is that we did not assess whether microbes utilized the

phytohormones ABA and SA. A previous study reported that increasing ABA concentrations

caused increased soil respiration, suggesting microbial utilization of ABA as a substrate. However,

they also observed a disproportionate increase in respiration relative to the minimal phytohormone

inputs of JA and ACC, suggesting that these instead stimulated microbial mineralization of

existing soil carbon (29). Here, we observed a continued persistence of activated taxa in response

to phytohormones over 14 days of a few taxa (specifically the log two-fold enriched taxa in red –

Figure 2), which could be an indirect response due to a potential priming effect. However, the

immediate activation of bacterial taxa within 24 hours suggests that the phytohormones can dually

serve as a signal, potentially pointing to a plant's cry for help. Another limitation of this study is

that the 16S rRNA and rDNA sequencing method provides a population-level average of whether

a taxon is active (yes or no) and no indication of relative activity levels. Thus, we cannot know

whether all individuals within that taxon consistently responded or if some cells remained inactive

while others became disproportionately active.

Overall, this experiment's results show that microbial reactivation can be used to identify

immediate and then persistent bacterial responders among the root zone soil microbiome to plant

stress signals. The Actinobacterial members, such as *Microbispora* sp., were consistently noted

among the responsive bacterial populations across the different phytohormones and soils. The next

steps should investigate the molecular mechanisms and the dynamic host-microbiome feedback

with *Microbispora* sp. given phytohormones as stress signals from the plant.

Methods

Soil collection, processing, and storage. Five kilograms of root-zone soil was collected from four

replicated switchgrass plots at Kellogg Biological Station's Great Lakes Bioenergy Research

Center Bioenergy Cropping Systems Experiment in November 2019. Approximately 4 kg of soil

was also collected from four soil core replicates for common bean (*Phaseolus vulgaris* L.) from a

rotation plot at the Michigan State University's Agronomy Farm in East Lansing, also in

November 2019. Approximately 1000 g of soil from each plot was collected from the top 10

centimeters for each replicate using an ethanol-sterilized soil auger of 10 cm diameter and placed

in Whirlpak bags. Soils were transported to the lab in a cooler with ice and then stored at 4°C until

further processing.

The following day after sample collection, soils from replicate plots were composited into

ethanol-sterilized buckets and hand-mixed to homogenize. Composited soils were sieved through

4mm mesh to remove roots and debris. Three 0.5g subsamples of each composited soil were flash-

frozen in liquid nitrogen for active microbiome assessment and designated as the "field-soil"

samples. The remaining sieved soil was stored at 4°C until the start of the experiment. All flash-

frozen soil samples were stored at -80°C.

Soil organic matter, pH, lime index, phosphorous, potassium, calcium, magnesium, nitrate,

and ammonium levels of the "field-soil" soils were analyzed by the Michigan State Soil Plant and

Nutrient Laboratory according to their standard protocols listed here

https://www.canr.msu.edu/fertrec/about/detailed-instructions. (**Table 1**).

Gravimetric soil moisture was determined by comparing the mass difference between the

initial field soil sample and a sample dried at 92°C for 4 days. Water holding capacity (WHC) was

assessed by taking 10g of dry soil and placing it in a funnel lined with filter paper (57). The dry

soil was saturated with water and allowed to drain for 2 hours. The mass of the original soil, the

mass of water added, and the mass of water drained were used to calculate WHC as percent

gravimetric soil moisture. Before setting up the mesocosms, 0.5 g of stored soil was flash-frozen

denoting "pre-dry" condition accounting for any shifts in microbial community due to storage,

thus capturing a true baseline reference community. The first step of the experiment was drying

the soils to induce dormancy and use it as a reference for subsequent treatments. Soils were dried

in a 55°C incubator for 3 days and after drying, soil subsamples were flash-frozen as the "post-

dry" treatment before adding water and phytohormones.

Mesocosm set-up, phytohormone treatment, and maintenance. To create each mesocosm,

100g dried soil was adjusted to 25% WHC and placed into 473 mL sterile glass jars (**Figure 1**).

Mesocosms were maintained at 27°C with lids secured loosely to ensure aerobic conditions.

Mesocosm manipulations and samplings were performed inside a sterilized biosafety cabinet.

Every 2-3 days, mesocosms were massed, and any water loss was replaced with sterile water to

maintain soil moisture.

Before hormone addition, mesocosms were incubated for two weeks to acclimatize to the

baseline water content of 25% WHC. This water acclimatization step was included to ensure that

changes in activation could be attributed to the experimental treatment of hormone addition and

not exclusively to water addition from a dry state, as it has been well-reported that soil

microorganisms activate in response to water (58-61) After two weeks, three 0.5-g subsamples

from four replicate mesocosms designated as "post 25% WHC" were flash frozen in liquid

nitrogen and designated as "water acclimated" samples. After sample collection at two weeks, the

"post 25%" mesocosms were supplemented with the required water to reach 50% WHC. They

were designated as "water control" samples, while the rest of the mesocosms received

phytohormones dissolved in methanol and methanol alone (carrier control) mixed with enough

water to reach 50% WHC.

Phytohormones were dissolved in methanol to create a 0.3 M solution, and 1 mL of the

phytohormone-methanol solution was filter sterilized (0.2 micron) before adding to the

mesocosms in combination with enough water to achieve 50% WHC. We normalized the hormone

addition based on the amount of carbon added per g of dry soil, and a 3M solution for ABA and

SA resulted in approximately 250-500 ug C g-1 dry soil for SA and ABA. Phytohormone

concentrations reported in soil are typically in the nanomolar range (62). However, we chose a

concentration that was 4-7 times the reported values used in the literature (which ranges anywhere

between 5-64 ug C g-1 dry soil (30, 63)), to elicit a measurable response to the phytohormone such

that comparisons across treatments would be possible. Methanol carrier controls received 1 ml of

methanol supplemented with water to reach 50% WHC. After adding the aqueous solution,

mesocosm soil was mixed using a sterile spatula. The soil was non-destructively sampled from

each mesocosm and immediately flash-frozen on days 1, 7, and 14 after hormone addition.

Thus, there were four conditions in total – water control, methanol control, ABA, and SA.

With four replicate mesocosms for four treatments and three time points collected in three

replicates, n = 144 total samples for each soil type after water acclimation, totaling 288 samples

across both soils.

RNA/DNA coextractions. A manual phenol-chloroform nucleic acid extraction was performed to

obtain RNA and DNA from the same cell lysis pool (64) with minor modifications. For the

modifications, we used Qiagen Powerbead Garnet Tube (0.70mm), and after a 30 s lysis in a bead

beater, we used a 10-minute centrifugation at 12,000 g at 4°C. After adding the chloroform-

isoamyl alcohol (24:1), the tubes were inverted several times to form an emulsion, followed by a

5-minute centrifugation at 16,000 g at 4°C. After adding 30% PEG6000-1.6 M NaCl, the tubes

were inverted several times to mix the contents and then placed on ice for a 2-hour incubation.

After incubation, the samples were centrifuged for 20 minutes at 16,000 g at 4°C. After

centrifugation, the supernatant was removed, and 1.0mL of 70% EtOH (stored at -20°C overnight)

was added. The samples were centrifuged for 20 minutes at 16,000g at 4°C, and the ethanol wash

was removed. The samples were centrifuged for an additional 20 seconds. Residual ethanol was

removed with a pipette, and the pellet was left to air-dry before they were resuspended in 30 μL of

nuclease-free water. Negative extraction controls (tubes with reagents and beads) were processed

alongside experimental samples. Nucleic acids were visualized using agarose gel electrophoresis

and validated with a band for DNA and RNA (64). All DNA and RNA samples were quantified

using a Qubit® dsDNA BR assay kit and RNA HS assay kit on a Qubit 2.0 fluorometer

(Invitrogen, Carlsbad, CA, USA). Samples were stored at -80°C.

DNase treatment and cDNA synthesis. RNA samples were purified using an Invitrogen TURBO

DNA-free™ Kit with a slightly modified protocol. A six μL aliquot of the nucleic acid

coextraction was treated with one μL of 10X TURBO DNase Buffer and three μL of Turbo DNase

enzyme and then incubated at 37°C for 30 minutes. After this incubation period, two μL of DNA

inactivation reagent was added to each tube before incubating at room temperature for five

minutes. Then, the sample was centrifuged at 2,000 g for five minutes at room temperature. The

purified RNA was used as a template to form cDNA using Invitrogen's SuperScript™ III First-

Strand Synthesis System using random hexamers per kit directions and with negative controls to

assess for reagent contamination. Final cDNA samples were stored at -80°C.

PCR check and gel electrophoresis.16S rRNA gene V4 PCR was performed for every DNA and

cDNA sample to ensure the presence of the 16S rRNA gene and the RNA sample to ensure

effective DNase treatment. The PCR reaction used a 2X GoTaq Green Master Mix and the primers

515F (GTGYCAGCMGCCGCGGTAA) and 806R (GGACTACNVGGGTWTCTAAT). PCR

cycle parameters were: 95°C for 3 minutes; 30 cycles of 95°C for 45 seconds, 50°C for 60

seconds, 72°C for 90 seconds; and a final step at 72°C for 10 minutes. The PCR product was

examined via gel electrophoresis. All PCR reactions included a no-template negative control and

an *E. coli* DNA template as the positive control.

Illumina sequencing of the 16SrRNA gene and 16SrRNA. DNA and cDNA samples were sent

for 16S rRNA gene V4 amplicon sequencing at the Genomics Research Technology Support

Facility at Michigan State University. The V4 hypervariable region of the 16S rRNA gene was

amplified using dual-indexed Illumina compatible primers 515f/806r, as described by J. J. Kozich

et al. (65). PCR products were batch normalized using an Invitrogen SequalPrep DNA

Normalization Plate, and the normalized products recovered were pooled. The pool was cleaned

and concentrated using a QIAquick PCR Purification column followed by AMPureXP magnetic

beads; it was then quality controlled and quantified using a combination of Qubit dsDNA HS,

Agilent 4200 TapeStation HS DNA1000, and Invitrogen Collibri Library Quantification qPCR

assays. The pool was loaded onto an Illumina MiSeq v2 standard flow cell, and sequencing was

performed in a 2x250bp paired-end format using a MiSeq v2 500 cycle reagent cartridge. Base

calling was done by Illumina Real Time Analysis (RTA) v1.18.54, and the output of RTA was

demultiplexed and converted to FastQ format with Illumina Bcl2fastq v2.20.0.

Sequence data were analyzed using QIIME2 (66). Briefly, all paired-end sequences with

quality scores were compressed and denoised using the DADA2 plugin (67). The denoising step

dereplicated sequences, filtered chimeras, and merged paired-end reads. The truncation parameters

for DADA2 were determined using FIGARO (68) developed by Zymo Research Corporation. All

truncation was performed from the 3' end for consistent final read lengths. The DNA and cDNA

datasets were separately quality controlled and denoised because the additional cDNA

amplification step could have introduced different errors for the cDNA than for the DNA

amplicons. The quality-controlled DNA and cDNA count tables were merged into a single

QIIME2 artifact using the feature-table merge command. Similarly, the DNA and cDNA

representative sequences were merged into a single QIIME2 artifact using the feature-table merge-

seqs command. The representative sequences from the combined count tables were clustered at

99% identity de-novo, and the clustered representative sequences were classified using SILVA

v138 (69) to generate the taxonomy file. Again, 99% sequence identity was used to conservatively

account for possible errors in the cDNA amplicons resulting from the additional amplification

(RT-PCR). The resulting OTU table and taxonomy files were exported to R for ecological

analysis.

Designating the active bacterial populations. All downstream analyses were performed in R

version 4.0. The R package decontam (70) was used to determine the number and identity of

contaminants in the dataset and remove them using the prevalence method. Contaminating taxa,

mitochondria, and chloroplast sequences were filtered from the datasets. A subsampling depth of

12,000 reads per sample was selected based on rarefaction curves. After subsampling, 16S rRNA

to rRNA gene ratios were computed from the cDNA and DNA datasets described in A. W.

Bowsher et al. (38). We chose the method that applied a 16S rRNA:rRNA gene ratio threshold

>=1. We also included "phantom taxa" and changed all DNA counts = 0 corresponding to RNA

counts >0 to DNA counts =1. The DNA counts table was then filtered to include only taxa

meeting the criteria to be likely active at the population average (e.g., 16S rRNA:rRNA >= 1).

The filtered DNA OTU table was used for all ecological analyses, and we refer to this as the

"active community."

Ecological statistics. Microbiome analysis was performed in R using packages phyloseq (71) and

vegan (72). Permutational analysis of variance (PERMANOVA) was used for multivariate

statistics. For data visualization, we used a constrained analysis of principal coordinates (CAP)

and partitioned out variances contributed by selected variables. Specifically, switchgrass had a

significant mesocosm effect (PERMANOVA F=2.84, R2=0.05, p=0.002), so to account for that,

we partitioned out the mesocosm variance using a constrained ordination. We also used soil

chemistry data to conduct a constrained analysis of principal coordinates to understand how soil

nutrient status affected microbiome community changes across treatments using package

phyloseq. Edaphic factors included in the final models were selected with function ordistep from

the vegan package with the initial full model incorporating pH, percent soil organic matter, nitrate,

and ammonium concentrations. We used differential expression analysis (DESeq2, (73)) to assess

the log two-fold changes between treatments and time points. We looked at each taxa's relative

abundance separately to determine the reactivation of taxa from an inactive state. Activity

dynamics were evaluated using the methods described in S. Bandopadhyay et al. (40). In this

study, we did not set a threshold for detecting phantom taxa across samples. To generate the

heatmaps and bubble plots looking at relative abundances for each taxon, we used a max

standardization approach using the decostand function in the vegan package in R (72). For alpha

diversity, we used the estimate_richness function in package phyloseq with ANOVA type III

results reported using R package car (74).

Data and code availability statement. Sequence data is submitted to the Sequence Read Archive

under BioProject PRJNA932434. The scripts to analyze the data and generate figures are on

GitHub

(https://github.com/ShadeLab/PAPER_Dormancy_resuscitation_phytohormone_mesocosm_Band

opadhyay2025.git.)

Acknowledgments

Support for this research was provided by the United States National Science Foundation under

Grant No. MCB #1817377 to A.S. Additional support was provided by the Great Lakes Bioenergy

Research Center, U.S. Department of Energy, Office of Science, Office of Biological and

Environmental Research under Award Number DE-SC0018409 and by the National Science

Foundation Long-Term Ecological Research Program (DEB #1832042). Additional support was

provided by the Michigan State University Plant Resilience Institute by, the USDA National

Institute of Food and Agriculture and Michigan State University AgBioResearch. AS

acknowledges project support from the European Union (ERC, MicroRescue, 101087042). Views

and opinions expressed are however those of the author(s) only and do not necessarily reflect those

of the European Union or the European Research Council. Neither the European Union nor the

granting authority can be held responsible for them.

References

1.   Sonkar S, Sharma L, Singh RK, Pandey B, Rathore SS, Singh AK, Porwal P, Singh SP. 2021.
Plant Stress Hormones Nanobiotechnology, p 349-373. *In* Al-Khayri JM, Ansari MI, Singh
AK (ed), Nanobiotechnology : Mitigation of Abiotic Stress in Plants doi:10.1007/978-3-
030-73606-4_15. Springer International Publishing, Cham.
2.   Sgroy V, Cassán F, Masciarelli O, Del Papa MF, Lagares A, Luna V. 2009. Isolation and
characterization of endophytic plant growth-promoting (PGPB) or stress homeostasis-
regulating (PSHB) bacteria associated to the halophyte Prosopis strombulifera. ApplIed
microbiology and Biotechnology 85:371-381.

3. Egamberdieva D, Kucharova Z, Davranov K, Berg G, Makarova N, Azarova T, Chebotar V,
Tikhonovich I, Kamilova F, Validov SZ. 2011. Bacteria able to control foot and root rot and
to promote growth of cucumber in salinated soils. Biology and fertility of soils 47:197-205.

4. Egamberdieva D, Wirth S, Jabborova D, Räsänen LA, Liao H. 2017. Coordination between
Bradyrhizobium and Pseudomonas alleviates salt stress in soybean through altering root
system architecture. Journal of Plant Interactions 12:100-107.

5. Liu Y, Shi Z, Yao L, Yue H, Li H, Li C. 2013. Effect of IAA produced by Klebsiella oxytoca Rs-5
on cotton growth under salt stress. The Journal of General and Applied Microbiology
59:59-65.

6. Nakano M, Omae N, Tsuda K. 2022. Inter-organismal phytohormone networks in plant-
microbe interactions. Current Opinion in Plant Biology 68:102258.

7. Schlaeppi K, Bulgarelli D. 2015. The plant microbiome at work. Molecular Plant-microbe
interactions 28:212-217.

8. Hashem A, Abd_Allah EF, Alqarawi AA, Al-Huqail AA, Wirth S, Egamberdieva D. 2016. The
interaction between arbuscular mycorrhizal fungi and endophytic bacteria enhances plant
growth of Acacia gerrardii under salt stress. Frontiers in microbiology 7:1089.

9. Badri DV, Chaparro JM, Manter DK, Martinoia E, Vivanco JM. 2012. Influence of ATP-
binding cassette transporters in root exudation of phytoalexins, signals, and in disease
resistance. Frontiers in plant science 3:24907.

10. Jones DL, Nguyen C, Finlay RD. 2009. Carbon flow in the rhizosphere: carbon trading at
the soil–root interface. Springer.

11. Marschner H, Römheld V, Horst W, Martin P. 1986. Root-induced changes in the
rhizosphere: Importance for the mineral nutrition of plants. Zeitschrift für
Pflanzenernährung und Bodenkunde 149:441-456.

12. Couillerot O, Prigent-Combaret C, Caballero-Mellado J, Moënne-Loccoz Y. 2009.
Pseudomonas fluorescens and closely-related fluorescent pseudomonads as biocontrol
agents of soil-borne phytopathogens. Letters in applied microbiology 48:505-512.

13. Rodriguez RJ, Henson J, Van Volkenburgh E, Hoy M, Wright L, Beckwith F, Kim Y-O,
Redman RS. 2008. Stress tolerance in plants via habitat-adapted symbiosis. The ISME
journal 2:404-416.

14. Zhalnina K, Louie KB, Hao Z, Mansoori N, da Rocha UN, Shi S, Cho H, Karaoz U, Loqué D,
Bowen BP, Firestone MK, Northen TR, Brodie EL. 2018. Dynamic root exudate chemistry
and microbial substrate preferences drive patterns in rhizosphere microbial community
assembly. Nature Microbiology 3:470-480.

15. Mitra D, Mondal R, Khoshru B, Senapati A, Radha T, Mahakur B, Uniyal N, Myo EM, Boutaj
H, Sierra BEG. 2022. Actinobacteria-enhanced plant growth, nutrient acquisition, and crop
protection: Advances in soil, plant, and microbial multifactorial interactions. Pedosphere
32:149-170.

16. Haney CH, Samuel BS, Bush J, Ausubel FM. 2015. Associations with rhizosphere bacteria
can confer an adaptive advantage to plants. Nature Plants 1:15051.

17. Sah SK, Reddy KR, Li J. 2016. Abscisic acid and abiotic stress tolerance in crop plants.
Frontiers in plant science 7:571.

18. Cutler SR, Rodriguez PL, Finkelstein RR, Abrams SR. 2010. Abscisic acid: emergence of a
core signaling network. Annual review of plant biology 61:651-679.

19.   Bano A, Ullah F, Nosheen A. 2012. Role of abscisic acid and drought stress on the activities
601   of antioxidant enzymes in wheat. Plant, Soil and Environment 58:181-185.
602   20.   Ahmad P, Nabi G, Ashraf M. 2011. Cadmium-induced oxidative damage in mustard
603   [Brassica juncea (L.) Czern. & Coss.] plants can be alleviated by salicylic acid. South African
604   journal of botany 77:36-44.
605   21.   Silva ACd, Suassuna JF, Melo ASd, Costa RR, Andrade WLd, Silva DCd. 2017. Salicylic acid
606   as attenuator of drought stress on germination and initial development of sesame.
607   Revista Brasileira de Engenharia Agrícola e Ambiental 21:156-162.
608   22.   Senaratna T, Touchell D, Bunn E, Dixon K. 2000. Acetyl salicylic acid (Aspirin) and salicylic
609   acid induce multiple stress tolerance in bean and tomato plants. Plant Growth Regulation
610   30:157-161.
611   23.   Vincent SA, Ebertz A, Spanu PD, Devlin PF. 2022. Salicylic Acid-Mediated Disturbance
612   Increases Bacterial Diversity in the Phyllosphere but Is Overcome by a Dominant Core
613   Community. Frontiers in Microbiology 13.
614   24.   Su Y, Wang J, Gao W, Wang R, Yang W, Zhang H, Huang L, Guo L. 2023. Dynamic
615   metabolites: A bridge between plants and microbes. Science of The Total Environment
616   899:165612.
617   25.   Lebeis SL, Paredes SH, Lundberg DS, Breakfield N, Gehring J, McDonald M, Malfatti S,
618   Glavina del Rio T, Jones CD, Tringe SG, Dangl JL. 2015. Salicylic acid modulates
619   colonization of the root microbiome by specific bacterial taxa. Science 349:860-864.
620   26.   Yuan J, Zhang N, Huang Q, Raza W, Li R, Vivanco JM, Shen Q. 2015. Organic acids from
621   root exudates of banana help root colonization of PGPR strain Bacillus amyloliquefaciens
622   NJN-6. Scientific Reports 5:13438.
623   27.   Zhou Y, Wei Y, Zhao Z, Li J, Li H, Yang P, Tian S, Ryder M, Toh R, Yang H, Denton MD. 2022.
624   Microbial communities along the soil-root continuum are determined by root anatomical
625   boundaries, soil properties, and root exudation. Soil Biology and Biochemistry
626   171:108721.
627   28.   Zhang F, Hou Y, Zed R, Mauchline TH, Shen J, Zhang F, Jin K. 2023. Root exudation of
628   organic acid anions and recruitment of beneficial actinobacteria facilitate phosphorus
629   uptake by maize in compacted silt loam soil. Soil Biology and Biochemistry 184:109074.
630   29.   Sayer EJ, Crawford JA, Edgerley J, Askew AP, Hahn CZ, Whitlock R, Dodd IC. 2021.
631   Adaptation to chronic drought modifies soil microbial community responses to
632   phytohormones. Communications Biology 4:516.
633   30.   de Vries FT, Williams A, Stringer F, Willcocks R, McEwing R, Langridge H, Straathof AL.
634   2019. Changes in root-exudate-induced respiration reveal a novel mechanism through
635   which drought affects ecosystem carbon cycling. New Phytologist 224:132-145.
636   31.   Kamilova F, Kravchenko LV, Shaposhnikov AI, Azarova T, Makarova N, Lugtenberg B. 2006.
637   Organic Acids, Sugars, and l-Tryptophane in Exudates of Vegetables Growing on
638   Stonewool and Their Effects on Activities of Rhizosphere Bacteria. Molecular Plant-
639   Microbe Interactions® 19:250-256.
640   32.   Hamer U, Marschner B. 2005. Priming effects in different soil types induced by fructose,
641   alanine, oxalic acid and catechol additions. Soil Biology and Biochemistry 37:445-454.

33. Mondini C, Cayuela ML, Sanchez-Monedero M, Roig A, Brookes P. 2006. Soil microbial
biomass activation by trace amounts of readily available substrate. Biology and Fertility of
Soils 42:542-549.
34. Kuzyakov Y, Friedel J, Stahr K. 2000. Review of mechanisms and quantification of priming
effects. Soil Biology and Biochemistry 32:1485-1498.
35. Keiluweit M, Bougoure JJ, Nico PS, Pett-Ridge J, Weber PK, Kleber M. 2015. Mineral
protection of soil carbon counteracted by root exudates. Nature Climate Change 5:588-
595.
36. Rolfe SA, Griffiths J, Ton J. 2019. Crying out for help with root exudates: adaptive
mechanisms by which stressed plants assemble health-promoting soil microbiomes.
Current Opinion in Microbiology 49:73-82.
37. Lennon JT, Jones SE. 2011. Microbial seed banks: the ecological and evolutionary
implications of dormancy. Nature Reviews Microbiology 9:119-130.
38. Bowsher AW, Kearns PJ, Shade A. 2019. 16S rRNA/rRNA Gene Ratios and Cell Activity
Staining Reveal Consistent Patterns of Microbial Activity in Plant-Associated Soil.
mSystems 4:e00003-19.
39. Lennon JT, den Hollander F, Wilke-Berenguer M, Blath J. 2021. Principles of seed banks
and the emergence of complexity from dormancy. Nature Communications 12:4807.
40. Bandopadhyay S, Li X, Bowsher AW, Last RL, Shade A. 2024. Disentangling plant-and
environment-mediated drivers of active rhizosphere bacterial community dynamics
during short-term drought. Nature Communications 15:6347.
41. Garbeva P, van Elsas JD, van Veen JA. 2008. Rhizosphere microbial community and its
response to plant species and soil history. Plant and Soil 302:19-32.
42. Berg G, Smalla K. 2009. Plant species and soil type cooperatively shape the structure and
function of microbial communities in the rhizosphere. FEMS Microbiology Ecology 68:1-
13.
43. Wieland G, Neumann R, Backhaus H. 2001. Variation of Microbial Communities in Soil,
Rhizosphere, and Rhizoplane in Response to Crop Species, Soil Type, and Crop
Development. Applied and Environmental Microbiology 67:5849-5854.
44. Naylor D, DeGraaf S, Purdom E, Coleman-Derr D. 2017. Drought and host selection
influence bacterial community dynamics in the grass root microbiome. The ISME Journal
11:2691-2704.
45. Correa SS, Pacheco RS, Viana GC, Vidal MS, Xavier GR, de Araújo JLS. 2024. Ability of
nitrogen-fixing bacteria to alleviate drought stress in cowpea varies depending on the
origin of the inoculated strain. Plant and Soil 498:391-408.
46. Shimizu M. 2011. Endophytic Actinomycetes: Biocontrol Agents and Growth Promoters, p
201-220. *In* Maheshwari DK (ed), Bacteria in Agrobiology: Plant Growth Responses
doi:10.1007/978-3-642-20332-9_10. Springer Berlin Heidelberg, Berlin, Heidelberg.
47. Menendez E, Carro L. 2019. Actinobacteria and their role as plant probiotics. Biofertilizers
for Sustainable Agriculture and Environment:333-351.
48. Oyedoh OP, Yang W, Dhanasekaran D, Santoyo G, Glick BR, Babalola OO. 2023. Rare rhizo-
Actinomycetes: A new source of agroactive metabolites. Biotechnology Advances
67:108205.

49. Schimel J, Balser TC, Wallenstein M. 2007. Microbial stress-response physiology and its
implications for ecosystem function. Ecology 88:1386-1394.
50. Schimel JP. 2018. Life in dry soils: effects of drought on soil microbial communities and
processes. Annual review of ecology, evolution, and systematics 49:409-432.
51. Kakumanu ML, Ma L, Williams MA. 2019. Drought-induced soil microbial amino acid and
polysaccharide change and their implications for CN cycles in a climate change world.
Scientific reports 9:10968.
52. Kalachova T, Jindřichová B, Burketová L, Monard C, Blouin M, Jacquiod S, Ruelland E,
Puga-Freitas R. 2023. Controlled natural selection of soil microbiome through plant-soil
feedback confers resistance to a foliar pathogen. Plant and Soil 485:181-195.
53. Tang X, Fei X, Sun Y, Shao H, Zhu J, He X, Wang X, Yong B, Tao X. 2022. Abscisic acid-
polyacrylamide (ABA-PAM) treatment enhances forage grass growth and soil microbial
diversity under drought stress. Front Plant Sci 13:973665.
54. Liu Y, Shi A, Chen Y, Xu Z, Liu Y, Yao Y, Wang Y, Jia B. 2024. Beneficial microorganisms:
Regulating growth and defense for plant welfare. Plant Biotechnology Journal n/a.
55. Belimov AA, Dodd IC, Safronova VI, Dumova VA, Shaposhnikov AI, Ladatko AG, Davies WJ.
2014. Abscisic acid metabolizing rhizobacteria decrease ABA concentrations in planta and
alter plant growth. Plant Physiology and Biochemistry 74:84-91.
56. Eichmann R, Richards L, Schäfer P. 2021. Hormones as go-betweens in plant microbiome
assembly. The Plant Journal 105:518-541.
57. Nelson JT, Adjuik TA, Moore EB, VanLoocke AD, Ramirez Reyes A, McDaniel MD. 2024. A
Simple, Affordable, Do-It-Yourself Method for Measuring Soil Maximum Water Holding
Capacity. Communications in Soil Science and Plant Analysis 55:1190-1204.
58. Aanderud ZT, Jones SE, Fierer N, Lennon JT. 2015. Resuscitation of the rare biosphere
contributes to pulses of ecosystem activity. Frontiers in Microbiology 6.
59. Zhao J, Chen D, Gao W, Guo Z, Jia Z, Hernández M. 2021. Resuscitation of soil microbiota
after > 70-years of desiccation. European Journal of Soil Biology 103:103290.
60. Angel R, Conrad R. 2013. Elucidating the microbial resuscitation cascade in biological soil
crusts following a simulated rain event. Environmental Microbiology 15:2799-2815.
61. Zhu G, Wang S, Wang C, Zhou L, Zhao S, Li Y, Li F, Jetten MSM, Lu Y, Schwark L. 2019.
Resuscitation of anammox bacteria after >10,000 years of dormancy. The ISME Journal
13:1098-1109.
62. Hartung W, Sauter A, Turner NC, Fillery I, Heilmeier H. 1996. Abscisic acid in soils: What is
its function and which factors and mechanisms influence its concentration? Plant and Soil
184:105-110.
63. Baumert VL, Vasilyeva NA, Vladimirov AA, Meier IC, Kögel-Knabner I, Mueller CW. 2018.
Root Exudates Induce Soil Macroaggregation Facilitated by Fungi in Subsoil. Frontiers in
Environmental Science 6.
64. Griffiths RI, Whiteley AS, O'Donnell AG, Bailey MJ. 2000. Rapid method for coextraction of
DNA and RNA from natural environments for analysis of ribosomal DNA- and rRNA-based
microbial community composition. Appl Environ Microbiol 66:5488-91.
65. Kozich JJ, Westcott SL, Baxter NT, Highlander SK, Schloss PD. 2013. Development of a
dual-index sequencing strategy and curation pipeline for analyzing amplicon sequence

data on the MiSeq Illumina sequencing platform. Applied and environmental microbiology
79:5112-5120.

66.   Bolyen E, Rideout JR, Dillon MR, Bokulich NA, Abnet CC, Al-Ghalith GA, Alexander H, Alm
EJ, Arumugam M, Asnicar F, Bai Y, Bisanz JE, Bittinger K, Brejnrod A, Brislawn CJ, Brown
CT, Callahan BJ, Caraballo-Rodríguez AM, Chase J, Cope EK, Da Silva R, Diener C,
Dorrestein PC, Douglas GM, Durall DM, Duvallet C, Edwardson CF, Ernst M, Estaki M,
Fouquier J, Gauglitz JM, Gibbons SM, Gibson DL, Gonzalez A, Gorlick K, Guo J, Hillmann B,
Holmes S, Holste H, Huttenhower C, Huttley GA, Janssen S, Jarmusch AK, Jiang L, Kaehler
BD, Kang KB, Keefe CR, Keim P, Kelley ST, Knights D, et al. 2019. Reproducible, interactive,
scalable and extensible microbiome data science using QIIME 2. Nat Biotechnol 37:852-
857.

67.   Callahan BJ, McMurdie PJ, Rosen MJ, Han AW, Johnson AJA, Holmes SP. 2016. DADA2:
High-resolution sample inference from Illumina amplicon data. Nature methods 13:581-
583.

68.   Weinstein MM, Prem A, Jin M, Tang S, Bhasin JM. 2019. FIGARO: An efficient and
objective tool for optimizing microbiome rRNA gene trimming parameters.
BioRxiv:610394.

69.   Quast C, Pruesse E, Yilmaz P, Gerken J, Schweer T, Yarza P, Peplies J, Glöckner FO. 2012.
The SILVA ribosomal RNA gene database project: improved data processing and web-
based tools. Nucleic acids research 41:D590-D596.

70.   Davis NM, Proctor DM, Holmes SP, Relman DA, Callahan BJ. 2018. Simple statistical
identification and removal of contaminant sequences in marker-gene and metagenomics
data. Microbiome 6:226.

71.   McMurdie PJ, Holmes S. 2013. phyloseq: An R Package for Reproducible Interactive
Analysis and Graphics of Microbiome Census Data. PLOS ONE 8:e61217.

72.   Oksanen J, Blanchet FG, Kindt R, Legendre P, Minchin PR, O'hara R, Simpson GL, Solymos
P, Stevens MHH, Wagner H. 2013. Package 'vegan'. Community ecology package, version
2:1-295.

73.   Love MI, Huber W, Anders S. 2014. Moderated estimation of fold change and dispersion
for RNA-seq data with DESeq2. Genome Biology 15:550.

74.   Fox J, Weisberg S. 2018. An R companion to applied regression. Sage publications.

Tables

Table 1. Nutrient analysis of bean and switchgrass root zone soil before the experiment (pre-dry

sample). P-Phosphorus, K-Potassium, Ca- Calcium, Mg-Magnesium, OM-organic matter, NO_3-N-

Nitrate-N, NH_4-N-Nitrate.

Soil	pH	Lime Index	BrayP1 P ppm	K ppm	Ca ppm	Mg ppm	OM %	NO_3-N ppm	NH_4-N ppm
Switchgrass	6.0	69	27	96	930	146	2.5	1.5	2.6
Common bean	6.2	69	97	151	486	82	1.9	0.7	1.6

767　**Table 2. Permuted analysis of variance (PERMANOVA) to assess global differences in active**

768　**microbial community structure across experimental treatments and time points.**

	Degrees of Freedom	R2	Pseudo-F value	P value	Degrees of Freedom	R2	Pseudo-F value	P value
	Bean				Switchgrass			
Treatment	4	0.19	10.28	0.001	4	0.16	9.24	0.001
Time point	2	0.06	6.63	0.001	2	0.11	12.79	0.001
Treatment: Time point	6	0.10	3.66	0.001	6	0.11	3.98	0.001

Figure Legends

Figure 1. Schematic of the soil mesocosm experimental design and sampling approach to

assess how root zone bacterial communities activate and assemble in response to

phytohormones that indicate plant stress. The design was a time series of field-collected root

zone soils that were first deactivated by drying and then acclimated to water before the

experimental addition of phytohormones. The water acclimation step was used to partition the

response to moisture increase from the response to the phytohormones. Soils were amended with

either a stress-associated phytohormone (salicylic acid, SA, or abscisic acid, ABA) or a methanol

carrier or water control. WHC is water holding capacity.

Figure 2. Bacterial microbiome community activation, persistence, and assembly before and

after phytohormone addition to bean and switchgrass root zone soil. The relative abundances

of active taxa that were log two-fold enriched (DESeq analysis) at Day 1 as compared to the

water-acclimated sample are shown in red. All other taxa include the 50 most abundant active

members in each soil/phytohormone condition. All responses shown were specific to a

phytohormone exposure (salicylic acid, SA, or abscisic acid, ABA) and were not observed in

controls. Inactive taxa have a relative abundance of zero, so they do not contain a symbol on the

graph and are left blank for that time point/treatment. Taxa not detected in either DNA or cDNA

data set are denoted by a + shape. Color gradient denotes increased relative abundances as color

changes from blue to yellow. The shape size of each bubble denotes the consistency of detection

across experimental and technical replicates of soil mesocosms (e.g. with 1 denoting activity in all

replicates and 0.5 denoting activity in half). Empty rows indicate that the taxon was not detected

within the top 50 most abundant members for that condition. Empty spaces in rows that also have

symbols indicate that the taxon was not classified as active within those samples (columns) but

was detected in DNA.

Figure 3. Active bacterial community structures differed between bean and switchgrass.

Barplots show broad changes in the bacterial community composition over time and across

treatments in bean and switchgrass root zone soils for **A)** Abscisic acid and **B)** Salicylic acid.

Figure 4: Active microbiome community structure differs across time points for A) Bean soil

amended with abscisic acid (ABA), B) Bean soil amended with salicylic acid (SA), C)

Switchgrass soil amended with abscisic acid (ABA) and D) Switchgrass soil amended with

salicylic acid (SA) treatments. Ordination shows a constrained analysis of principal coordinates

(CAP) after excluding the variance attributed to mesocosm. The increasing symbol sizes denote

sampling time and the colors denote the treatments.

Full factorial design with 32 experimental units = 2 soil types * 4 treatments * 4 replicates (Treatments: ABA, SA, control with methanol + water, control only water)

Microbial Composition of treatments as compared to baseline and controls for
A. Abscisic acid

B. Salicylic acid