Impact of Group Size and Habitat Disturbance on Parasitic Infection in Free-

ranging Proboscis Monkeys

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Abstract

Group-living primates experience the benefits and costs associated with sociality, including

an elevated risk of parasite transmission. However, the relative influence of group type (i.e.,

social structure), group size, and habitat disturbance on parasitic infection remains unclear,

particularly in Southeast Asian primates. In this study, the abundance of intestinal parasites in

proboscis monkeys (*Nasalis larvatus*) inhabiting a riverine forest along the Menanggul River,

Sabah, Malaysian Borneo, was investigated. Fecal samples (n = 160) were collected from

one-male–multifemale and all-male groups in areas with varying levels of anthropogenic

disturbance, with efforts made to ensure that each sample originated from a different

individual. In addition, the effects of group type, group size, and sampling location on

parasite abundance were evaluated using fecal egg counts and Bayesian models. Three

dominant   parasite   species   groups   (*Trichuris*   sp.,   *Strongyloides   fuelleborni*,   and

Oesophagostomum aculeatum) with an overall infection prevalence of 81.25% were

identified. Results showed that group type did not significantly affect parasite abundance.

However, group size showed a positive correlation with the abundance of *Trichuris* sp. and a

negative correlation with *S. fuelleborni* and *O. aculeatum*. In addition, our models revealed

that the infection loads of *Trichuris* sp. decreased with the increase of distance from the river

mouth (less disturbed upstream areas), whereas *O. aculeatum* exhibited higher infection loads

with greater distance, indicating a lower prevalence in more disturbed downstream areas.

Thus, parasite abundance in proboscis monkeys may be shaped by social and environmental

factors, with taxa-specific responses to these variables.

Introduction

Highly social animals such as primates exhibit complex group dynamics that influence their

health, behavior, and interactions with pathogens (Kappeler et al., 2015). Although group-

living promotes social support and cooperative behaviors that can mitigate some health risks

(Silk, 2007), it also imposes costs, including an increased potential for pathogen transmission

and heightened susceptibility to disease caused by social stress (Altizer et al., 2003). Thus,

the relationship between sociality and disease risk is complicated, as factors such as group

structure, contact patterns, and social hierarchies play key roles in determining infection

dynamics (Nunn et al., 2015). Notably, in primates, the relationship between parasites and

host social structures is bidirectional, with social factors such as group size and composition

influencing parasite transmission. Consequently, infection prevalence and intensity can affect

social interaction and group dynamics (Vitone et al., 2004; Nunn, 2012; Rushmore et al.,

2017).

Apart from social dynamics, anthropogenic environmental changes, such as

deforestation and provisioning, can also influence parasite transmission by altering host

movement patterns, population densities, and contact rates (Gillespie and Chapman, 2006;

Becker et al., 2018). These ecological modifications can create conditions that either facilitate

or hinder parasite spread, depending on the nature and extent of human-induced disturbances.

Therefore, understanding the mechanism by which environmental, social, and ecological

factors contribute to parasitic infections in social animals is crucial for developing effective

conservation and management strategies (Chapman et al., 2005; Gillespie and Chapman,

2006).

On the contrary, research on parasitic infection and sociality in wild primates has

yielded inconsistent findings, while the effects of anthropogenic environmental changes on

parasitic infections in primates have been relatively consistent (Gillespie and Chapman,

2006). The relationship between parasite load and group size is complex, as it depends on

various factors, such as host susceptibility and environmental conditions (e.g. Chapman et al.,

2009; Klaus et al., 2018). Although larger groups might increase the likelihood of infection

because of higher social interactions, group mobility and seasonal changes may reduce the

risk of parasite transmission (e.g., increased precipitation during the rainy season washing

away feces and parasites), thereby complicating the host–parasite dynamics (Nguyen et al.,

2015; Radespiel et al., 2015; Knauf and Jones-Engel, 2020).

Although knowledge of parasitism in wild primates is expanding, it remains limited,

particularly in Southeast Asia (Hopkins and Nunn, 2007; Cooper and Nunn, 2013). In Borneo,

which is a highly biodiverse island in Southeast Asia, numerous endemic species are

experiencing population declines (Wilcove et al., 2013), and the large arboreal proboscis

monkey (*Nasalis larvatus*) inhabiting mangroves, peat swamps, and riverine forests is no

exception to this trend (Lhota et al., 2019; Bernard et al., 2025). Several studies have

provided insights into primate–parasite interactions in Sabah, including the proboscis monkey.

This primate has unique parasite associations (Hasegawa et al., 2020) and infection dynamics

that reflect its specialized ecological adaptations (Klaus et al., 2017) and interactions with

sympatric primates (Frias et al., 2019; Frias et al., 2021). Previous studies have reported a

high prevalence of helminth eggs in fecal samples, identifying five major parasite groups,

with co-infections being common (Salgado-Lynn, 2010; Klaus et al., 2017; Frias et al., 2021).

In addition, parasite species richness has been reported to be significantly higher in semi-wild

populations than in wild ones. This increase likely results from the high levels of human

disturbance, which can alter the mechanism by which parasites are transmitted. By contrast,

group type and group size have been reported not to significantly affect parasite richness

(Klaus et al., 2018). Despite the increasing knowledge of parasitic infections in proboscis

monkeys, the fine-scale variation in infection patterns—across individuals or groups

inhabiting distinct microhabitats and experiencing different levels of anthropogenic

disturbance—remains unclear.

104    In this study, the abundance of intestinal parasites in proboscis monkeys was

105 investigated, with particular emphasis on the effects of group type, group size, and habitat

106 disturbance. Fecal samples collected over a 1-year period were analyzed to identify the

107 parasites present and to examine the mechanism by which group size and social structure (i.e.,

108 group type: one-male–multifemale or all-male groups) influence parasite loads in a forest

109 subjected to varying degrees of anthropogenic disturbance. Considering that proboscis

110 monkeys exhibit seasonal variation in foraging behavior, ranging patterns, and intergroup

111 spacing at our study site (Matsuda et al., 2009a; Matsuda et al., 2009b; Matsuda et al., 2010a),

112 a year-round sampling approach may provide a comprehensive assessment of parasite

113 dynamics. Thus, extended sampling could reveal a positive correlation between group size

114 and parasitic infection, which is contrary to previous findings (Klaus et al., 2018). In addition,

115 the study forest along the tributary river exhibits varying degrees of anthropogenic

116 disturbance between the upstream and downstream areas in terms of plant diversity (Jose et

117 al., 2024). The impact of habitat disturbance on parasite abundance can be assessed by

118 comparing the fecal samples collected from individuals across these areas. Given that

119 research in our study area has documented alterations in the gut microbiota of proboscis

120 monkeys living in disturbed versus undisturbed areas (Jose et al., 2024), we hypothesize that

121 parasite abundance would also differ between these environments because of human-induced

122 disturbances.

124 **Methods**

125 *Study site and subjects*

126 This study was conducted in a riverine forest along the Menanggul River, a tributary of the

127 Kinabatangan River, in Sabah, Malaysian Borneo (118°30′ E, 5°30′ N). The study site has a

128 mean daily temperature range of 24°C–30°C and an annual average precipitation of 2474 mm

129 (Matsuda et al., 2019). The river exhibits daily water-level fluctuations of approximately 1 m,

while seasonal inundation results in an average increase exceeding 3 m (Matsuda et al.,

2010b). The study area spans 4 km along the Menanggul River and supports a minimum

population of 200 proboscis monkeys, forming a multilevel society composed of one-male–

multifemale groups and all-male groups, which frequently assemble in the riverside (Matsuda

et al., 2024). The riparian zone primarily consists of secondary forest in the southern portion,

while much of the northern portion has been converted to oil palm plantations, except for a

narrow, protected riparian strip. Accordingly, in this tributary, forests extending upstream

from the river mouth have a higher plant diversity (Jose et al., 2024). The proboscis monkeys

in the study area are well habituated to the human presence, as the Menanggul River has been

a major tourist attraction for over a decade, and it serves as a frequent route for boats and

observers.

Sample collection

Between June 2015 and April 2016, boat surveys were conducted in the late afternoon on

approximately 2–3 days per week to locate the proboscis monkeys as they typically return to

the riverside trees for sleeping (Matsuda et al., 2011) and to record their group composition,

along with the GPS coordinates of their sleeping locations at the riverside trees. In the early

morning, before the monkeys awakened, these sleeping locations were revisited. Considering

that proboscis monkeys often defecate just before moving into the forest, the area was

carefully searched around their sleeping sites for fresh feces after the monkeys had departed

from their sleeping trees. Multiple one-male–multifemale groups often remained close to one

another along the river, making it challenging to determine to which group the feces found on

the ground belonged. Therefore, we inferred the group associated with the feces based on the

location of the feces and the group observed during the survey from the previous day. Only

fecal samples presumed to originate from adult individuals were included, based on fecal size,

which allows for differentiation between adults and immature individuals despite the

challenges in distinguishing feces from neighboring groups (Matsuda et al., 2014).

Consequently, between June 2015 and April 2016, a total of 160 fecal samples were collected

genetically distinct individuals (Matsuda et al., 2024). At least nine groups were identified in

the study area, although the number of groups these fecal samples were collected from

remains unclear, as group identification was not complete at the time the samples were

collected (Matsuda et al., 2024). The fecal samples were preserved in tubes with 70% ethanol.

Parasite detection and identification

A modified protocol was used for concentrating parasite eggs using formalin-ethyl acetate

sedimentation, and fecal samples were analyzed using the sequential sedimentation–flotation

method (Frias et al., 2021). The final pellet was resuspended in saline, and 1 mL of aliquot

was transferred into a vial and placed on a magnetic stirrer to ensure thorough

homogenization during analysis. In estimating the number of eggs per gram of feces (EPG) in

each sample, an aliquot was collected from the homogeneous suspension and placed in a

McMaster counting chamber for examination at $100\times$ magnification. EPG was calculated on

the basis of the average of five replicate counts of parasite eggs observed within the

chamber's grid, using the known weight of fecal sediment in a 0.15 mL volume. After

quantifying the eggs, each sample was centrifuged again; the supernatant was discarded, and

the concentrated pellet was resuspended in Sheather's solution with a specific gravity of 1.27.

Two slides were examined to reduce the likelihood of overlooking less abundant helminth

eggs. Parasite identification was performed using standard identification keys (Modrý et al.,

2018).

Parasite prevalence and abundance

The prevalence of a given parasite was calculated as the percentage of hosts infected by that

parasite, while parasite abundance was calculated as the number of individuals of a given

parasite in a given sample, using the EPG as a proxy measure.

Factors influencing parasite abundance

A Bayesian model was developed using the "brms" package (Bürkner, 2018) to evaluate the

mechanism by which group type (one-male–multifemale or all-male groups), group size, and

fecal sampling locations, represented as the geographical distance from the river mouth to

upstream, affect parasite abundance. In our analysis, parasite abundance was used as a

dependent variable, rounded to the nearest integer. Group type was incorporated as a group-

level effect, whereas group size and the distance of the sampling location to the river mouth

were included as continuous variables. All models were controlled for host group identity.

Distinct groups were not identified throughout the study period; thus, samples were assigned

to the same group based on their spatial and temporal proximity along the river. This

approach was guided by observations, that is, the mean distance between consecutive

sleeping sites for each group was approximately 350 m in the study area (Matsuda et al.,

2024). In addition, changes in food availability over time can influence spatial patterns along

the river (Matsuda et al., 2010a). Consequently, two samples collected within 350 meters of

each other and during the same month were classified as originating from the same group.

Groups with fewer than three observations were excluded from the analysis.

Weakly informative priors were assigned to population-level effects, following a

Student's t-distribution. The posterior distributions of the model parameters were estimated

using Markov chain Monte Carlo methods. For this purpose, four chains were constructed,

each consisting of 2000 steps, with 1000-step warm-up periods. Consequently, a total of 4000

steps were retained to estimate the posterior distributions. The independent continuous

variables were scaled to aid model fitting and coefficient interpretation. Chain convergence

was assessed visually, and the potential scale reduction factor (Rhat) values were examined to

ensure they were less than or equal to 1 for all parameters (Bürkner, 2018). This modeling

process was repeated for the three parasites identified within the host population.

Results

Parasite prevalence and abundance

Parasites were detected in 130 out of the 160 analyzed samples, resulting in an overall

parasite prevalence of 81.25%. The eggs of *Trichuris* sp. (78.12%, n = 125), *Strongyloides*

fuelleborni (10.62%, n = 17), *Oesophagostomum aculeatum* (5%, n = 8), *Enterobius*

(*Colobenterobius*) *serratus* (1.25%, n = 2), and Spirurida sp. (0.63%, n = 1) were identified.

The mean abundance of the top three prevalent parasites was recorded at 168.2 EPG (range:

0–984), 9.41 EPG (range: 0–650), and 2.83 EPG (range: 0–109), respectively.

Factors influencing parasite abundance

A total of 124 samples were included in the models. The type of host group did not have a

statistically significant effect on parasite abundance for any of the parasite groups (Fig. 1A).

However, the size of the host group emerged as an important predictor of parasite abundance,

with varying effects observed for specific parasite groups. The abundance of *Trichuris* sp.

was positively correlated with group size, whereas that of *S. fuelleborni* and *O. aculeatum*

was negatively correlated with group size (Fig. 1B). In addition, our models indicated that the

distance from the river mouth to upstream locations markedly influenced the parasite

abundance for *Trichuris* sp. and *O. aculeatum*. In particular, as the distance from the river

increased, the abundance of *Trichuris* sp. decreased, whereas *O. aculeatum* exhibited an

opposite trend, with higher counts linked to greater distances from the river mouth (Fig. 1C).

A summary of the results from the models is provided in Table 1.

Discussion

In this study, the relationship among social structure (i.e., group type: one-male–multifemale

and all-male groups), the degree of anthropogenic habitat disturbances within a 4-km stretch

along the river, and parasite load in free-ranging proboscis monkeys was examined. A high

prevalence of intestinal parasites was observed (81.25%), which is consistent with the

previous prevalences recorded for proboscis monkeys along the Kinabatangan River in Sabah

[i.e., 96.6% in Salgado-Lynn (2010), 92.3% in Klaus et al. (2017), and 80.2% in Frias et al.

(2021)] and in Bako National Park in Sarawak [i.e., 95.4% in Adrus et al. (2019)]. The

predominant parasite groups in this study included *Trichuris* sp., *Strongyloides fuelleborni*,

and *Oesophagostomum aculeatum*, which have also been identified as dominant in other fecal

analyses of proboscis monkeys, in wild and captive settings (Table 2), reinforcing a

consistent pattern of parasitic infections across different populations. Nonetheless, given the

substantial differences in proboscis monkey diets between riverine and mangrove forests

(Matsuda et al., 2019), investigating the mechanism by which such habitat-linked dietary

variation may contribute to individual- or population-level differences in parasitic infections

across their range is necessary. Habitat type not only determines available food resources but

also varies in forest height and ground use frequency (Feilen and Marshall, 2014, 2020;

Atmoko et al., 2021), which may influence parasite transmission dynamics. Thus, whether

similar trends of parasites are observed for proboscis monkeys on the island of Borneo with

different habitat types should be examined in the future.

However, a notable link was not found between parasite abundance and group type

(i.e., one-male–multifemale and all-male groups) in proboscis monkeys in this study, which is

consistent with previous findings (Klaus et al., 2018), indicating that sex differences may

have a limited influence on parasitic infection. This interpretation is further supported by a

study conducted at the same site that examined the gut microbial composition from fecal

samples and found no significant sex-based differences (Jose et al., 2024). Given that gut

microbiota, such as intestinal parasites, would be transmitted through direct or indirect social

contact and shared use of space and food resources, the absence of sex-specific patterns in

both types of organisms indicates that behavioral and spatial overlap between males and

females minimizes sex-related divergence in microbial and parasite exposure.

Contrary to the finding of Klaus et al. (2018), our study revealed that group size might

be a potential factor influencing parasite abundance for particular parasite species, that is,

Trichuris sp. This finding supports the hypothesis that larger groups may exhibit higher

infection rates probably because of increased contact rates and overlapping space use among

individuals (Snaith et al., 2008; Rifkin et al., 2012; Kappeler et al., 2015). Although the exact

mechanism remains unclear, a comparable trend was observed in the gut microbial diversity

in the same population. A larger group size was associated with greater alpha diversity (Jose

et al., 2024), indicating that higher host density may enhance intestinal parasites and

microbial exposure. This association may reflect the increased direct interactions among

individuals and result from enhanced environmental contamination in frequently reused areas.

Given that all the parasites identified in this study could be transmitted via fecal–oral routes,

the accumulation of feces in shared sleeping or foraging sites could be a key factor driving

infection. Limited territoriality and habitual reuse of sleeping trees by proboscis monkeys—

often shared across one-male–multifemale and all-male groups in multilevel societies—likely

facilitates such transmission (Matsuda et al., 2024). Furthermore, the differences in habitat

structure and spatial dynamics among the study sites may explain the contrasting results

Klaus et al. (2018). They focused on groups inhabiting the main channel of the Kinabatangan

River, where the river's width (>200 m) limits cross-group movement and shared space use.

By contrast, our study site along narrow tributaries (15–20 m wide) may allow more frequent

crossings and higher spatial overlap among groups (Matsuda et al., 2008; Matsuda et al.,

2010a), thereby promoting indirect transmission through contaminated environments.

On the contrary, *S. fuelleborni* and *O. aculeatum* were negatively correlated with

group size because of their differing ecological tolerances. *S. fuelleborni*, which is a parasitic

nematode, may be more influenced by habitat disturbance or seasonal changes, with larger

groups potentially reducing parasite exposure through increased host mobility and spatial

separation from contaminated areas (Gillespie and Chapman, 2006; Radespiel et al., 2015).

Similarly, *O. aculeatum* may require more stable environmental substrates for its fragile eggs,

and pristine upstream habitats may provide better conditions for egg survival. By contrast,

larger groups in disturbed downstream areas could hinder transmission (Kappeler et al., 2015).

These findings indicate that host social dynamics and environmental factors must be

considered, as species-specific responses shape infection patterns (Rifkin et al., 2012;

Nguyen et al., 2015).

Our findings revealed a more complex relationship between parasite abundance and

habitat disturbance than initially anticipated. Contrary to expectations that anthropogenic

habitat disturbance would universally elevate parasite loads (Gillespie and Chapman, 2006),

we observed that the impact of disturbance varied across different parasite taxa. In particular,

the abundance of *Trichuris* sp. decreased with the increase of distance from the river mouth,

indicating that higher parasite loads occur in downstream areas with more intense human

activity and tourism. Conversely, *O. aculeatum* displayed the opposite trend, with greater

abundance in less disturbed upstream forests. These contrasting patterns indicate that

different parasite species may respond to habitat disturbance through distinct ecological

mechanisms. Considering the elevated *Trichuris* abundance in downstream areas, more

disturbed habitats may not be attributable to the increased host density or spatial overlap, as

proboscis monkeys in these areas do not necessarily aggregate at higher densities compared

with the upstream areas. Aggregation likely occurs in upstream areas, where the river width is

narrower, providing a safer environment for proboscis monkeys because of their ability to

escape terrestrial predators, such as clouded leopards, by jumping to the opposite bank

(Matsuda et al., 2010a). Given its greater safety from predators, this environment likely

facilitates aggregation in upstream areas, making them a more favorable location for

proboscis monkey groups to congregate. This interpretation is consistent with the species'

highly resilient egg morphology, which supports environmental persistence even under

degraded conditions (Gillespie and Chapman, 2006; Snaith et al., 2008). By contrast, the

fragile eggs of *O. aculeatum* may require relatively intact environmental substrates, such as

undisturbed soil, to ensure survival and successful transmission, as the eggs of many parasites

are dependent on specific environmental conditions for persistence, including substrate type

and humidity (Radespiel et al., 2015). Accordingly, the relatively pristine conditions in

upstream sites likely provide a favorable microenvironment for their development.

Interestingly, these spatial patterns in parasite abundance are contrary to prior findings

from the same population, which identified clear site-based differences in gut microbial

composition, with lower microbial diversity in the downstream areas (Jose et al., 2024).

Although gut microbiota and intestinal parasites share common transmission pathways,

namely, social contact and environmental exposure, their ecological sensitivities appear to

diverge. In general, gut microbiota are more responsive to dietary shifts and subtle ecological

changes (e.g., Amato et al., 2013; Hayakawa et al., 2018; Lee et al., 2019; Barelli et al., 2020),

whereas parasite abundance in our study is influenced by a combination of host social

dynamics (e.g., group size and space use) and species-specific environmental tolerances. This

distinction emphasizes the importance of adopting a taxon-specific approach when

interpreting host–environment–parasite relationships and indicates that parasites and

microbiota can influence the ecological and health dynamics of primates (Rifkin et al., 2012;

Nguyen et al., 2015).

Understanding the interplay among social organization, habitat characteristics, and

parasitic infection in wild primates deepens our ecological insights and informs broader

questions in evolutionary biology and conservation. The findings of this study underscore the

complex, context-dependent nature of host–parasite dynamics, highlighting the mechanism

by which intrinsic factors such as group size and behavior can mediate exposure risks even in

the face of external environmental change. At present, given the escalating habitat

disturbance, such nuanced understanding is crucial for predicting disease risks in wildlife

populations and designing conservation strategies that account for social and ecological

variability. Integrative approaches that consider behavioral ecology alongside environmental

gradients may also be essential for safeguarding the health and persistence of socially

complex yet increasingly vulnerable species such as the proboscis monkey.

Finally, the zoonotic potential of the helminths identified in this study, future research

should assess cross-species transmission risks, especially in ecotourism corridors like our

study site, where frequent contact occurs among humans, livestock, and wild primates.

Trichuris sp. may include *T. trichiura*, a known human pathogen capable of infecting non-

human primates (Rivero et al., 2020), while *S. fuelleborni* and *O. aculeatum* have been

reported in both human and non-human hosts in Southeast Asia (Yalcindag et al., 2021; Chan

et al., 2023). The presence of these helminths in areas accessible to humans highlights the

need to understand their transmission dynamics and public health implications in the context

of increasing anthropogenic pressure (Gillespie and Chapman, 2006).

Declarations

Ethics approval

The animal study was reviewed and approved by the Sabah Biodiversity Centre and the

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Author contributions

MNF: conceptualization, writing—original draft, methodology; LF: conceptualization,

writing—review and editing, methodology, visualization, formal analysis, investigation; EZ:

conceptualization, methodology; PL: resources, writing—review and editing; JT: resources,

writing—review and editing; HB: resources, writing—review and editing; VK: resources,

writing—review and editing; IM: conceptualization, data curation, funding acquisition,

investigation, methodology, project administration, resources, supervision, validation,

visualization, writing—original draft. All authors (except deceased EZ) gave final approval

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Table and figure legends

Table 1. Summary of the fitted models.

Table 2. Comparison between the parasite species detected and parasite prevalence (%) in

fecal samples of proboscis monkeys in each study.

Fig. 1. Posterior distributions of the differences in parasite abundance for *Trichuris* sp. (top),

S. fuelleborni (middle), and *O. aculeatum* (bottom) in relation to host group type (A). The

green shaded areas represent the posterior distributions, with the black dot indicating the

median estimate and the black horizontal line shown in boldface representing the 99%

credible interval. The vertical dashed line at zero represents no difference in infection rates

among the groups. The predicted probability of abundance for the same parasite groups in

relation to host group size (B) and sampling location represented by the distance from the

river mouth (C). Credible intervals are set at 89%.

Table 1. Summary of fitted models

	Parameter	Est.	Est. error	95% CI (Lower)	95% CI (Upper)	Rhat	Bulk ESS	Tail ESS
	Random Effects							
	sd (Intercept)	0.92	0.22	0.6	1.44	1	1600	2336
	Fixed Effects							
Trichuris sp.	Intercept	4.32	0.51	3.29	5.32	1	3515	4008
	group type	0.81	0.55	-0.27	1.93	1	3151	3511
	group size (scaled)	0.16	0.01	0.14	0.19	1	6937	4494
	distance to river (scaled)	-0.7	0.14	-0.97	-0.44	1	2949	3362
	Random Effects							
	sd (Intercept)	5.74	1.67	3.37	9.71	1	2691	3424
	Fixed Effects							
S. fuelleborni	Intercept	-1.68	3.25	-8.21	4.77	1	3101	3670
	group type	-1.09	3.7	-8.69	5.88	1	2916	3602
	group size (scaled)	-1.62	0.12	-1.86	-1.38	1	9301	5682
	distance to river (scaled)	1.03	0.71	-0.35	2.44	1	5057	4881
	Random Effects							
	sd (Intercept)	10.54	3.96	5	20.28	1	2116	3425
	Fixed Effects							
O. aculeatum	Intercept	-5.12	6.63	-18.87	7.67	1	3150	3920
	group type	-0.47	7.58	-16.08	14.32	1	2693	3620
	group size (scaled)	-1.48	0.19	-1.87	-1.11	1	7942	5445
	distance to river (scaled)	3.46	1.4	0.71	6.27	1	3953	4708

576 **Table 2.** Comparison of the parasite species detected and parasite prevalence (%) in fecal samples of proboscis monkeys in each study

Study site	Bekantan Rescue Center Sahabat Bekantan Foundation	Surabaya Zoo	Labuk Bay Proboscis Monkey Sanctuary	Bako National Park	Lower Kinabatangan (Lot. 1-7 areas)	Lower Kinabatangan (Lot. 6 area)	Lower Kinabatangan (Lot. 1-4 areas)	Lower Kinabatangan (Lot. 4 area)
Forest type	N/A	N/A	Mangrove	Mangrove	Riverine forest	Riverine forest	Riverine forest	Riverine forest
Living-condition	Captive	Captive	Semi-free ranging	Free-ranging	Free-ranging	Free-ranging	Free-ranging	Free-ranging
No. of individuals	3	< 27	< 195	Unknown	Unknown	< ca. 200	Unknown	160
No. of samples	60	20	90	65	91	652	146	160

Parasite species

Parasite species								
Anatrichosoma spp.			2.8			1.4	2.1	
Ascaris sp.		5.0		30.8			67.1	
Ascaris lumbricoides			2.2			8.6		
Dipylidium-like morphs							9.6	
Diphyllobotrium sp.				1.5				
Enterobius (Colobenterobius) serratus					1.25			1.25

Enterobius spp. (genus within Oxyurida)				5.5				
Fasciola sp.				1.5				
Hymenolepis sp.				3.1				
Hymenolepis nana		5.0						
Oesophagostomum sp.	5.0		31.4	33.8		22.8		
Oesophagostomum aculeatum								5.0
Physaloptera sp.				3.1				
Schistosoma sp.				1.5				
Spirurida sp.					2.1			
Strongyloides spp.		5.0	65.6	21.5	20.8	32.7	5.5	
Strongyloides fuelleborni								10.6
Strongyloides stercoralis				4.6				
Taenia sp.							28.8	
Trichostrongylus spp.	10.0		36.0	9.2		48.5		
Trichuris spp.		85.0	21.8–47.1	35.4	80.2	82.1	91.8	78.1

Trichuris trichiura	15.0			44.6				
Strongylida			43.3		29.6	3.1		
Spirurida sp.					0.63			0.63
strongylid			4.6				45.2	
oxyurid							4.1	
Ref.	Rahman et al. (2023)	Ramadhani et al. (2024)	Klaus et al. (2017); Klaus et al. (2018)	Adrus et al. 2019	Frias et al. (2021)	Klaus et al. (2017); Klaus et al. (2018)	Salgado-Lynn (2010)	This study

580 **Fig. 1.** Posterior distributions of the differences in parasite abundance for *Trichuris* sp. (top), *S. fuelleborni* (middle), and *O. aculeatum* (bottom)
581 in relation to host group type (A). The green shaded areas represent the posterior distributions, with the black dot indicating the median estimate
582 and the black bold horizontal line representing the 99% credible interval. The vertical dashed line at zero represents no difference in infection
583 rates between groups. Predicted probability of abundance for *the same parasite groups* in relation to host group size (B) and sampling location
584 represented by the distance from the river mouth (C). Credible intervals are set at 89%.

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