Model substrate particles uncover dynamics of microbial communities associated with

particulate organic matter decomposition in soil

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Running Title: Model particle-associated microbial communities

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Abstract

Soil organic matter is the largest terrestrial reservoir of organic carbon. Its particulate fraction, particulate organic matter (POM), serves as a resource and surface for microbial colonization. Degradation of complex biopolymers like cellulose and chitin requires extracellular enzymes produced by phylogenetically diverse microbes. Despite their importance for carbon cycling, the structure and spatio-temporal dynamics of POM-associated microbial communities in soil and how specific substrates influence them remain poorly understood. This study investigated whether microbial communities associated with POM change in composition and richness over time and whether chitin and cellulose select for distinct fungal and bacterial taxa. We incubated self-manufactured, millimetre-sized model substrate particles containing chitin or cellulose in soil under laboratory and field conditions. We assessed particle-associated communities at multiple time points over a 50-day-long incubation in the lab and after 47 days in the field. Our results show that community structure and temporal dynamics of particle-associated microbial communities were substrate-specific. While microbial biomass increased on both particle types, chitin-associated microbial communities exhibited stronger temporal changes. Communities on chitin and cellulose particles were enriched in specific bacterial and fungal genera compared to communities in the surrounding soil. We demonstrate that microbial communities associated with model chitin particles underwent notable temporal changes, including decreased microbial richness and shifts in community composition over the incubation period. This study shows the potential of model particles to advance our understanding of particle- and substrate-associated microbial communities in soil.

Keywords

Soil microbial communities, particulate organic matter, model particles, chitin, cellulose, temporal dynamics

Introduction

Soil organic matter is the largest terrestrial carbon reservoir [1, 2]. Conceptually, soil organic matter can be divided into mineral-associated organic matter and particulate organic matter (POM) [3]. POM serves as a surface for colonization [4] and particles are considered hotspots for microbial activity [5] as they represent spatially isolated, nutrient-rich microhabitats in the otherwise nutrient-poor soil environment.

Microbes decompose POM, as they can only take up small molecules (typically less than 600 Da) [1, 6]. To acquire resources, they produce and excrete extracellular enzymes [7, 8], which break down large molecules into subunits and ultimately monomers that can be taken up [9, 10]. By decomposing organic matter via extracellular enzymes, heterotrophic microorganisms mediate the recycling of carbon and nutrients [9, 11–13]. The complete breakdown of biopolymers requires multiple enzymes that act in synergy [14–16]. Extracellular enzymes are substrate-specific [7] and metabolically costly to produce [8, 16]. Genetic potential to produce extracellular enzymes differs between microbial taxa [6, 17–19]. Consequently, phylogenetically diverse microbial communities are necessary to complete biopolymer degradation.

Cellulose and chitin are the two most abundant polymers in nature [18, 20, 21]. Both must be enzymatically degraded prior to microbial uptake. While the potential to degrade these polysaccharides is widespread, the structure of degrader communities is driven by environmental conditions and substrate availability [18, 22]. Amongst fungi, the ability to degrade cellulose is more widespread than the ability to degrade chitin [12].

Spatial proximity of microbial cells to enzymatically degraded substrates is advantageous to enhance the return on investment. Microbial attachment to POM and formation of spatially structured communities should minimize losses of enzymatic breakdown products to the surrounding environment through diffusion. The presence of both bacteria and fungi has been detected on POM in soil. More specifically, it has been shown that bacteria occur in biofilms on the surface of POM and fungal hyphae pervade bacterial biofilms [5]. However, little is known about microbial communities associated with POM, despite the relevance of POM-degrading communities for carbon and nutrient cycling. We lack knowledge about the assembly and temporal dynamics of microbial communities on POM, and whether specific substrates select distinct microbial communities and particular microbial taxa in soil.

This knowledge gap is partly due to the complex nature of soil, making it challenging to study microbial communities directly associated with POM in soil. Soils are opaque and characterized by high substrate heterogeneity [4]. Organic matter exhibits diverse chemical composition and is spatially heterogeneously distributed throughout the soil matrix [3, 23–

26]. Investigation of microbial communities associated with natural POM requires extraction
with chemical agents, which may alter associated microbial communities [27, 28], possibly
resulting in erroneous findings. Other studies investigated substrate-associated communities
using isotopic labelling or metagenomics [29, 30], however, community dynamics at the
millimetre-spatial scale remain poorly understood.

In this study, we assess how the complex substrates, chitin and cellulose, drive successional
changes in microbial community assembly. Specifically, we address microbial colonization
and assembly at the particle interface in the heterogeneous soil environment using substrate-
containing model particles. We hypothesized that (i) microbial colonization of particles is
highly substrate-specific due to the degree of substrate complexity. Additionally, (ii) changes
in microbial communities are more pronounced at the particle-scale than at the soil-scale due
to the relevance of the micro-scale for microbial activity and interactions. Finally, (iii)
substrate-driven changes in community composition observed under controlled laboratory
conditions are reproducible in the field due to the fundamental influence of substrate in
driving community assembly. We applied self-manufactured model particles containing either
chitin or cellulose to study microbial communities at the particle interface in soil. While model
particles differ from natural POM in their physical and chemical properties, they provide a
simple and tractable experimental system which facilitates the study of microbial
communities associated with complex substrates in particulate form. The promising
applications of model substrate particles in microbial ecology were recently highlighted [31].
Model particles have been used to study microbial community assembly and dynamics on
POM surfaces in aquatic systems [32–35], but, to our knowledge, not in soils. We incubated
model substrate particles in forest soil and examined associated microbial communities
associated with particles and surrounding soil through a series of sampling time points over
50 days in the lab. In parallel, we incubated particles in soil in the field for 47 days to
investigate microbial community assembly under more natural conditions. At harvest time
points, we performed amplicon sequencing of model particle fractions and the surrounding
soil and assessed taxa richness and community composition. In addition, we determined
microbial biomass on particles and in soil via droplet digital PCR.

116 Materials and Methods

117 Model particle production

118 We manufactured three types of model particles: (i) particles containing chitin ("model chitin
119 particles"), (ii) cellulose ("model cellulose particles"), and (iii) particles containing neither
120 chitin nor cellulose ("no substrate"). We used Gelrite® gellan gum (Carl Roth, Karlsruhe, DE)
121 to produce model particles because of its physical properties (e.g., rapid and stable gelling

with ionic cross-linkers, high transparency) and its resistance to enzymatic degradation, making it a relatively biologically inert scaffold for embedding substrates compared to other frequently used gelling agents, like alginate and agarose [36, 37]. Following a protocol by Mussmann [38], we dissolved 1.5 g of Gelrite® gellan gum in 100 mL of MilliQ water while stirring. To facilitate polymer cross-linking, $CaCl_2$ (Sigma-Aldrich, Vienna, AT) was added at a final concentration of 8 mM. Additionally, we added 5 mL of 10% w/v iron (II, III) oxide (Sigma-Aldrich, Vienna, AT) to facilitate magnetic separation in downstream washing steps. As the only modification between particle types, we suspended 1.4 g of chitin from shrimp shells (Sigma-Aldrich, Vienna, AT) or cellulose powder (Sigma-Aldrich, Vienna, AT) in the gellan gum solution and applied an additional 0.5 g of respective substrate to the bottom and top of the poured medium in a petri plate prior to solidification. No substrate was supplemented to the "no substrate" control particles. Model particles were cylindrical and approximately 3 x 2 mm. Particles were UV-sterilized and stored in 20% ethanol at 4°C until use.

Soil collection and processing

Soil was collected in a beech-dominated temperate forest at the long-term ecological research site (LTER) in Klausen-Leopoldsdorf, Austria (48.11925 N, 16.04343 E, 580 m altitude) on 26 April 2021.This soil is classified as a dystric cambisol over sandstone with a loam-loamy clay texture [39–41]. The litter layer was removed prior to sampling, where soil cores were taken from the upper 5 cm of mineral soil at five established 5 x 5 m field plots [42] using a 5 x 8 cm metal corer. For each plot, two soil cores were pooled and homogenized into a single representative sample using a 2 mm mesh sieve, producing five composite soil samples. We recorded a gravimetric soil water content of 0.52 ± 0.13 g water g^{-1} dry soil (median ± standard deviation) and a pH of 4.85 ± 0.18 for sieved soil samples used for lab and field incubations.

Set-up of laboratory and field incubation

We incubated model particles in sieved forest soil in the laboratory and the field. For the lab incubation, we prepared mesocosms using 50 mL plastic vials (Greiner, Kremsmuenster, AT) by cutting them at 5 cm height with their cap serving as the bottom of the mesocosm. For the field incubation, we sewed mesh bags from polyester screen-printing fabric with a mesh width of 40 μm (Siebdruckversand, Magdeburg, DE) to allow fungal hyphal ingrowth but exclude roots, mimicking lab conditions [43, 44].

We set up four incubation treatments: i) 2-mm-sieved soil plus model chitin particles, ii) soil plus model cellulose particles, iii) soil plus "no substrate" model particles containing neither chitin nor cellulose, and iv) soil without model particles (Fig. 1). We rinsed model particles,

which had been stored in 20% ethanol, three times in 15 mL pure MilliQ water to remove
residual ethanol. We added 1 g of model particles to 15 g of sieved soil to individual plastic
bags to homogenize and distribute model particles evenly. Additionally, 15 g sieved soil
which was incubated without model particles was homogenized. Soil (and model particles)
was filled in lab mesocosms and field mesh bags (28 April 2021). Mesocosms were
compacted to field density (1.0 g cm^{-3}). Mesocosms were covered with parafilm during
incubation to prevent desiccation. Mesh bags were sealed with a heat-sealer (Mercier
Corporation, Taiwan) and measured 7 x 10 cm in size. In total, we set up 80 mesocosms and
20 mesh bags (Fig. 1).

Mesocosms were incubated in an incubator at 11°C ± 2°C, matching the soil temperature at
5 cm soil depth on the day of soil sampling. Filled mesh bags were kept in closed plastic
bags at 11°C in the lab incubator until they were buried in the field at a depth of 5 cm (30
April 2021). Potential moisture loss was assessed by weighing mesocosms weekly and
adjusted by adding MilliQ water if weight deviated by ≥ 2%.

Destructive harvest

Lab mesocosms were destructively harvested one, 13, 29, and 50 days after the start of the
lab incubation (start on 28 April; harvest on 29 April, 11 May, 27 May, 17 June 2021). Mesh
bags were destructively harvested after 47 days of incubation in the field (filled on 28 April;
buried on 30 April; harvest on 14 June 2021). Upon destructive harvest, we extracted the
mesocosms from the vials and separated model particles from soil using tweezers.
Separated particles were rinsed three times with MilliQ water to remove lightly adhering soil
and were (i) weighed for DNA extraction and (ii) fixed with ethanol for downstream
microscopic analysis.

DNA extraction and sequencing

DNA extraction, library preparation, and amplicon sequencing were conducted as in Simon et
al (2024) [42]. In short, we extracted DNA from a minimum of 20 model particles (150 mg)
and 250 mg soil with the Qiagen PowerSoil kit (Qiagen, Valencia, CA, USA), following the
manufacturer's protocol with exception of an added bead beating step (30 sec). For amplicon
sequencing, we followed a two-step PCR barcoding approach [45] to amplify the V4 region of
the 16S rRNA genes targeting bacteria and the fungal internal transcribed spacer region 2
(ITS2) of the nuclear rRNA operon. We used primers 515F [46] and 806R [47] for 16S rRNA
genes. For the ITS2 region, a nested approach [48] was chosen, generating a PCR product
using primers ITSOF-T [49, 50] and ITS4 [51], which was used as a template for a second
PCR using primers gITS7 [52] and ITS4. Amplicons were sequenced on a MiSeq platform
(Illumina, V3 chemistry, 600 cycles) at the Joint Microbiome Facility (JMF) of the Medical

University of Vienna and the University of Vienna under project number JMF-2106-12. Raw
data processing was performed as described previously [45].

194 Droplet digital PCR (ddPCR)

We quantified the number of copies of the V4 hypervariable region of the 16S rRNA gene
and of the fungal ITS region of model particles and in soils with droplet digital PCR (ddPCR).
Gene copy numbers can be used as a proxy for microbial biomass [53]. Prior to performing
ddPCR, we measured DNA concentration of all DNA extracts using PicoGreen (Quant-iT
PicoGreen dsDNA Assay Kit; Thermo Fisher Scientific, Vienna, AT). Extracts were re-
measured using a Qubit high-sensitivity dsDNA assay (Thermo Fisher Scientific, Vienna, AT)
if the concentration was under the detection limit. For 16S ddPCR, DNA extracts were diluted
to 0.05 ng μL^{-1} and for ITS ddPCR to 0.25 ng μL^{-1}. 2 µL of diluted DNA template were used
as input for each ddPCR assay. Gene copy numbers per µL DNA extract were assessed for
a single replicate of each sample with the QX200 Droplet Digital PCR (Bio-Rad) and
analysed with QuantaSoft software (Bio-Rad, Vienna, AT).

206 Particle fixation and fluorescence microscopy

A fraction of incubated model particles (~ 12 model particles) were fixed with 96% ethanol in
2 mL Eppendorf tubes (Sigma-Aldrich, Vienna, AT) and stored overnight at 4°C. Supernatant
was removed using magnetic pulldown and model particles were resuspended in 1.5 mL of
1X PBS to wash a total of three times. Ethanol-fixed samples were suspended in 1.5 mL of
cold 96% ethanol. Following fixation, samples were stored at -20°C. Selected chitin particles
were stained with SYBR Green I (Invitrogen, Thermo Fisher Scientific, Vienna, AT) and
analysed using a Leica confocal laser scanning microscope, coupled with a Thunder imaging
system (Leica, Vienna, AT).

215 Data analyses

Amplicon sequence variants (ASVs) were inferred using the DADA2 R package v1.20
(https://www.ncbi.nlm.nih.gov/pubmed/27214047) applying the recommended workflow
(https://f1000research.com/articles/5-1492). 16S rRNA FASTQ reads were trimmed at 220
nucleotides (nt) and 150 nt respectively with allowed expected errors of 2. ITS FASTQ reads
were trimmed at 230 nt with an expected error of 2 and 4, respectively.

16S rRNA ASV sequences were subsequently classified using DADA2 against the SILVA
database SSU Ref NR 99 release 138.1 (https://doi.org/10.5281/zenodo.4587955,
https://www.ncbi.nlm.nih.gov/pubmed/23193283), and ITS ASV sequences were classified
against the UNITE all eukaryotes general FASTA release version 8.2 [54] using a confidence
threshold of 0.5.

All data analyses were performed in R (v.4.4.0) [55]. We used packages
TreeSummarizedExperiment (v.2.12.0) [56], phyloseq (v.1.48.0) [57], and vegan (v.2.6-6.1)
[58]. We generated plots using the ggplot2 package (v.3.5.1) [59] and edited them in
Inkscape (v.0.92.4) [60]. We analysed samples that yielded at least 3000 reads (16S rRNA
genes: n = 172, ITS2 sequences: n = 179, see Table S1 and S2).

To compare the number of bacterial and fungal ASVs detected in the model particle fractions
and in soil, we randomly rarefied samples to 3015 (16S rRNA genes) and 5676 (ITS2
sequences) reads to establish equal sequencing depth across samples using the rrarefy
function from the vegan package [58]. We performed differential abundance analysis using
the ANCOMBC package (v.2.6.0) [61]. The analysis of microbiome compositions with bias
correction determined which genera were significantly enriched and depleted in microbial
communities associated with model particles compared to the "surrounding soil", in which
particles had been incubated. We examined temporal dynamics of particle-associated
microbial communities over the incubation period with principal coordinate analysis (PCoA).
Plots depict Aitchison distances, which are Euclidean distances calculated from centred-log-
ratio transformed counts matrices and account for the compositional nature of amplicon
sequencing data [62]. Additionally, we performed permutational analysis of variance
(PERMANOVA) to test the significance of sampling time points on community composition
using the adonis2 function of the vegan package.

245 Results

246 Microbial biomass increased on model particles over time, but not in the surrounding
247 soil

We assessed microbial colonization and biomass of model particles using ddPCR and
microscopic evaluation. Bacterial biomass increased on model particles over time as
evidenced by increasing gene copy numbers on model chitin and cellulose particles over 50
and 47 days of incubation for lab and field incubations, respectively (Fig. 2A). After 50 days,
the median number of 16S rRNA gene copies per gram of particles was highest on model
chitin particles compared to cellulose and "no substrate" particles (Fig. 2A). Bacterial gene
copies increased 52- and 31-fold on cellulose particles and 7- and 6-fold on chitin particles
between day one and the end of the incubation. Microbial colonisation was lower for "no
substrate" particles than chitin and cellulose particles, with gene copies increasing 3.6- and
4.5-fold more on chitin particles and 2.8-fold and 2.7-fold more on cellulose compared to "no
substrate" particles after 50 and 47 days of incubation.

The number of ITS gene copies was also the highest on chitin particles after 50 and 47 days
of incubation (Fig. 2B). Between one and 50 days of lab incubation, we observed 143-fold,

152-fold, and 72-fold increases in ITS gene copies on chitin, cellulose, and "no substrate" particles (Fig. 2B). Fungal gene copies were 3.9- and 16.4-fold higher on chitin particles than on "no substrate" particles after 50 and 47 incubation days, and 1.9- and 10.9-fold higher on cellulose particles. Microscopy visualization of chitin particles supports the finding that particles were more densely populated after several days of incubation than after one day (Fig. 2E). Particles incubated in the lab and the field were similarly densely populated by microbes after 47 and 50 days. In contrast, microbial biomass in the soil surrounding the model particles, and biomass in the soil incubated without model particles did not reveal a clear pattern over time (Fig. 2C, 2D).

ASV richness of model chitin particles decreased over the incubation period

To determine whether changes in microbial communities occurred on model particles and in soil across the different treatments, we investigated amplicon sequence variant (ASV) richness of rarefied data. Overall, the richness of initial particle-associated microbial communities was higher compared to communities associated with particles later in the incubation (Fig. 3). The number of bacterial taxa associated with model chitin particles dropped 1.2-fold between the start of the incubation and 50 and 47 days (Fig. 3A). Fungal richness of chitin particles was also approximately 1.2- and 1.5-fold lower after 50 and 47 days than after one day. In contrast, bacterial and fungal richness remained relatively unchanged on cellulose and "no substrate" particles between one and 50 days of incubation. Microbial richness of the incubated soil remained relatively constant across time points and treatments (Fig. 3).

Chitin particle-attached microbial communities underwent temporal changes

We investigated whether microbial community structure on model particles and the surrounding soil changed over the incubation period (Fig. 4). Differences in bacterial and fungal community composition between the time points were most pronounced for communities associated with chitin particles (Fig. 4A, 4D, Table S3). Communities detected on the first day of incubation were most similar to the communities of the soil used to set up incubations, while later communities (after 13, 29, and 50 days) differed more strongly from initial communities. PCoA plots suggest that bacterial and fungal communities associated with chitin particles, which had been incubated for 47 or 50 days, were very similar in their composition (Fig. 4A, 4D). In contrast, microbial communities associated with cellulose particles did not show pronounced changes over the course of the incubation (Fig. 4B, 4E). Bacterial communities associated with "no substrate" particles exhibited a shift in community composition after one day of incubation but remained similar over the remaining incubation period (Fig. 4C). In comparison, fungal communities associated with "no substrate" particles did not show a separation between the initial soil- and particle-associated communities after

one incubation day compared to particles incubated for multiple days (Fig. 4F). Over all
treatments, communities in soil did not exhibit strong changes over the incubation period
(Fig. 4G-4L). Results of PERMANOVA analyses support the observation that sampling time
point had a significant effect on composition of communities associated with chitin particles
(Table S4).

Specific microbial genera were enriched on chitin and cellulose particles

Analysis of the initial soil community composition revealed bacterial genera *Mucilaginibacter*,
Acidothermus, and *Bryobacter* and fungal genera *Mortierella*, *Russula*, and *Tomentella*
among the most abundant genera, which served as the "seed bank" for microbial
communities that could colonize model particles (Figure S1, S2). We compared relative
abundances of microbial genera between model substrate particles and the surrounding soil
to address whether substrates led to the enrichment of microbial genera. Differential
abundance analysis showed enrichment of bacterial and fungal genera on chitin and
cellulose particles compared to soil in which these particles had been incubated (Fig. 5, 6).
At the same time, other genera were significantly depleted on model chitin and cellulose
particles compared to the surrounding soil (Fig. S3, S4, S6, S7-). In the initial phase of
colonization, few bacterial genera, including *Paenibacillus* and *Bacillus* on chitin particles and
KD3-10, *Paenibacillus*, *Clostridium sensu stricto* 1, and *Micropepsis* on cellulose, were
significantly more abundant on particles compared to the surrounding soil (Fig. 5A, 5B).

Asticcacaulis and *Mucilaginibacter* were two genera that were enriched on chitin particles at
all later harvest time points (13d, 29d, 50d, 47d) (Fig. 5). Most genera were enriched
sporadically at single time points. On chitin particles, bacterial genera *Alkanindiges*,
Silvimonas, *Paucibacter*, *Roseatales*, and *Sphaerotilus* were amongst the most enriched at
the later time points of the lab incubation (13d, 29d, 50d) (Fig. 5A). We show that a larger
number of bacterial genera was significantly increased on chitin particles compared to soil
than in the case of cellulose particles (Fig. 5). While the number of bacterial genera enriched
on model chitin particles was similar for lab and field incubations after 47 and 50 days of
incubation (Fig. 5A), more bacterial genera were enriched on cellulose particles incubated in
the field than in the lab (Fig. 5B). Most bacterial genera enriched on chitin particles compared
to soil were consistent in lab and field incubations (Fig. 5A, 5B).

We detected sporadic enrichment of fungal genera on particles compared to the surrounding
soil (Fig. 6). Chitin particles were significantly enriched in *Umbelopsis* compared to
surrounding soil (13d, 29d) (Fig. 6A). *Pseudogymnascus* was relatively more abundant on
cellulose particles than in the surrounding soil after 13 and 47 days of incubation (Fig. 6B).
We found a larger number of fungal genera enriched on chitin and cellulose particles
incubated in the field (after 47 days) than in the lab (after 50 days) (Fig. 6A). For example,

Metapochonia, *Pochonia*, and *Pseudogymnascus* were only significantly enriched on chitin

particles, as well as *Ceratobisidium* and *Leotiomycetes* on cellulose particles in the field.

Overall, fewer fungal than bacterial genera were significantly enriched on model chitin

particles compared to soil (Fig. 6A, 5A).

Discussion

Our study applied an innovative approach toward investigating microbial community

assemblages directly at the physical interface of POM. We studied temporal changes in

microbial communities colonizing model substrate particles under both lab and field

conditions, revealing enrichment of specific taxa at the particle-scale over time. Our study

demonstrates model substrate particles as a suitable tool for studying substrate-specific

microbial communities on POM in soil. We demonstrate that microbial colonization and

temporal change in microbial community structure on the surface of model particles are

substrate-specific. The higher number of gene copies on substrate particles compared to "no

substrate" particles for bacteria and fungi indicates that presence of substrates enhanced

microbial colonization and growth on particles. Our results suggest that microbes were

physically attached to particles, as they would otherwise have been removed during the

washing steps. This is in line with a previous study which proposed that bacteria are widely

surface-attached in soil to improve enzymatic return on investment and promote microbial

growth [8].

Analysis of changes in ASV number revealed a decrease in bacterial and fungal richness on

model chitin particles over time. In contrast, bacterial richness remained unchanged on

cellulose and "no substrate" particles during the incubation period, while fungal richness

increased on both model particle types. Microbial communities associated with chitin

particles underwent temporal changes, whereas cellulose particles did not exhibit clear

temporal patterns. Communities associated with chitin particles became more chitin-specific

over time, supporting the occurrence of temporal succession on chitin-containing POM [63].

Cellulose did not select for specific microbial communities as strongly as chitin, potentially

due to the greater chemical complexity of chitin, which may necessitate a specific suite of

enzymes or microbial interactions for degradation. In contrast, we did not detect temporal

patterns of microbial community composition for the soil incubated with chitin particles,

suggesting that the effects of complex particulate substrates on microbial communities may

not be captured with composite soil samples.

Fewer bacterial genera were enriched on model cellulose particles than on chitin particles in

the lab incubation, supporting our claim that chitin exerts a stronger selective pressure on

microbial communities than cellulose. Genera enriched on chitin and cellulose were

potentially well-adapted to utilize or degrade chitin, cellulose or their degradation by-
products. Members of some genera that were enriched on model substrate particles have
been described to possess the genetic potential to produce chitin- or cellulose-degrading
enzymes. For example, the genus *Asticcacaulis* contains many glycoside hydrolases, key
enzymes targeting chitin [15], while members of *Mucilaginibacter* may possess certain
glycoside hydrolases capable of degrading chitin and cellulose [64, 65]. Species of other
significantly enriched genera on chitin and/or cellulose particles, *Silvimonas* and *Collimonas,*
have been found to possess chitinases [66–68] and a member of *Roseateles,* which was
enhanced on cellulose particles in the field, is known to degrade cellulose [69].

However, not all taxa detected on model substrate particles necessarily degraded the
substrate themselves. Some taxa might have profited from degradation products released by
other microbes or might have fed on metabolites of degraders. Importantly, degradation
potential is not necessarily conserved at the genus level, and assigning specific taxa to
substrate degradation is limited based on amplicon sequencing results. Some genera
enriched on chitin and cellulose particles were also enriched on "no-substrate" particles (Fig.
5), indicating non-specific colonization or a broader enzymatic repertoire. For instance,
Mucilaginibacter was enriched on chitin particles, cellulose particles, and "no substrate"
particles (Fig. 5), suggesting that members of this genus were not directly favoured by chitin
or cellulose but may have utilized other particle components, or stochastically colonized
particle surfaces. Similarly, enrichment of *Paenibacillus* on all particle types after one
incubation day suggests efficient colonization of particles irrespective of substrate, although
the genus includes known chitin and cellulose degraders [70–72].

Comparison of chitin particles to "no substrate" particles (Fig. S9) supports our observation
that bacterial genera like *Silvimonas, Asticcacaulis, Alkanindiges, Collimonas*, and
Mucilaginibacter were favoured by the presence of chitin rather than particle surfaces or
other particle components. To a lesser extent, enrichment of *Asticcacaulis* and *Collimonas*
could be attributed to the presence of cellulose, as these were also significantly enhanced on
cellulose particles compared to "no substrate" particles (Fig. S10). We recommend including
"no substrate" particles as controls for particles containing specific substrates. Overall, it
would be advantageous for future studies to investigate genes and extracellular enzyme
activity of taxa associated with model substrate particles or perform isotopic tracing
experiments in combination with model particles to uncover taxa that degrade specific
substrates.

Our results demonstrate that changes in microbial communities are indeed more pronounced
at the particle interface than at the soil scale. Only a small number of fungal genera,
including *Umbelopsis* and *Pseudogymnascus*, significantly increased on model chitin and

cellulose substrate particles compared to the surrounding soil. Both genera were significantly enriched on substrate particles compared to "no substrate" controls (Fig. S11, S12). This suggests that members of *Umbelopsis* may degrade chitin or utilize its degradation products, while *Pseudogymnascus* may similarly use cellulose-derived compounds. Nevertheless, our results support that the ability to degrade chitin or cellulose is potentially more widespread among fungal than bacterial taxa [12, 73]. Consequently, these specific substrates might exhibit a lower selective pressure on fungi than on bacteria. Comparisons between model substrate particles and "no substrate" particles underline that differences in relative abundances of microbial genera were driven by the specific substrate and likely not other particle components (Fig. S9-S12). We hypothesized that microbial communities in lab and field incubations would follow similar community assembly trajectories driven by substrate type. This was partially supported, as a subset of taxa was shared between the lab and field experiments and consistently showed substrate-specific enrichment. However, enrichment patterns differed between lab and field. In the lab, more bacterial genera were enriched on chitin particles than on cellulose, whereas equal numbers were enriched on both substrates in the field. For fungi, substantially more genera were enriched on model chitin and cellulose particles in the field than the lab. For example, *Metapochonia* and *Pochonia* were both only significantly enriched on chitin particles in the field (Fig. 6A), suggesting either migration to field mesocosms during the 47-day incubation or enhanced growth under field conditions. For bacteria and fungi, log fold changes were larger for field-incubated particles than for lab-incubated particles. While mesocosms containing model particles and soil were incubated under constant temperature and moisture conditions, the contents of mesh bags in the field were exposed to natural fluctuations and potential influxes from plant exudates over the incubation period. Interpretation of substrate-specific enrichment is further constrained by differences in sampling design, as lab incubation captured multiple harvests while the field experiment represents a single point in time. These differences highlight the need for caution when translating lab findings to nature.

Our study underpins that model substrate particles provide a highly flexible, standardized experimental system for the investigation of microbial degrader communities [31, 38, 74]. Model particles allow control over particle size, shape, and substrate composition [38], and remain stable and retrievable through incubation unlike natural substrates that become indistinguishable from soil over time. Particles can also be investigated using microscopy due to the high transparency. Future experiments could be conducted using particles to isolate substrate-specific consortia for downstream microbiological applications.

A potential caveat of this method is the use of gellan gum, which may be metabolized by a small subset of microorganisms, despite its resistance to enzymatic degradation [36, 75]. The low increase in gene copies on "no substrate" particles indicates non-specific

colonization. Only few genera including *Paenibacillus*, *Alkanindiges, Trichoderma,* and *Ilyonectria* were enriched across all particle types, whereas most genera enriched on model chitin and cellulose particles were not enriched on "no substrate" particle controls. It is essential to bear in mind that a certain fraction of model particle-associated microbes was possibly lost due to the washing steps or that passively attached communities were captured during sequencing. It might be advisable to shorten the here applied incubation duration in future experiments to capture and study more fine-scale temporal dynamics of microbial communities associated with POM in soil.

Conclusion

We found that complex substrates, chitin and cellulose, selected for specific microbial communities. Microbial communities associated with chitin underwent stronger temporal changes than those associated with cellulose, demonstrating substrate-specificity of communities in the here investigated forest soil. This approach has the potential to advance our knowledge of microbial taxa involved in degradation of complex biopolymers in soil by enabling isolation and characterization of substrate-specific microbial communities. We emphasize that further studies are necessary to determine whether these patterns are generalisable to other soil types and ecosystems. This study demonstrates the applicability of substrate particles as a model system for investigating substrate-specific microbial communities associated with particulate organic matter and their temporal dynamics in soil.

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Credit authorship contribution statement

Eva Simon: Formal analysis, Investigation, Visualization, Writing – original draft, Writing – review & editing. **Lauren Alteio**: Conceptualization, Formal analysis, Investigation, Methodology, Project administration, Writing – original draft, Writing – review & editing. **Alexander König:** Investigation, Writing – review & editing. **Bruna Imai**: Investigation, Writing – review & editing. **Julia Horak:** Investigation, Writing – review & editing. **Julia Wiesenbauer**: Investigation, Writing – review & editing. **Joana Séneca**: Data curation, Writing – review & editing. **Bela Hausmann**: Data curation, Writing – review & editing. **Marc Mussmann**: Methodology, Writing – review & editing. **Barbara Kitzler:** Resources, Writing – review & editing. **Christina Kaiser**: Conceptualization, Funding acquisition, Project administration, Supervision, Writing – review & editing.

Conflicts of Interest

The authors declare no conflict of interest.

Data Availability

Raw 16S rRNA genes and ITS2 amplicon sequencing data have been deposited in the NCBI Sequence Read Archive under the BioProject accession number PRJNA1208268. Bacterial, archaeal, and fungal gene copy numbers are available on the Zenodo repository (10.5281/zenodo.17901929).

Figure captions

Fig. 1 Schematic of the experimental set-up. (**A**) We set up four incubation treatments, incubating (i) self-manufactured model chitin (CH), (ii) cellulose (CE), (iii) "no substrate" (NS) particles and (iv) no model particles in 2-mm-sieved soil. We investigated five biological replicates by using soil sampled from five field plots. We set up four technical replicates per treatment and plot for the lab incubation, which we harvested at four time points, after one, 13, 29 or 50 days of incubation in the laboratory, and one technical replicate per treatment and plot for incubation in the field which we harvested after 47 days. (**B**) We mixed 1 g of model particles and 15 g 2-mm-sieved soil and filled it in lab mesocosms that were produced from 50 mL Greiner tubes or mesh bags for the field. Upon harvest of mesocosms and mesh bags, we manually separated model particles from the soil for subsequent analysis. (**C**) Example of self-manufactured model chitin substrate particles prior to incubation. (**D**) Laboratory mesocosm filled with soil and model particles. (**E**) Mesh bag used for the field incubation.

Fig. 2 Bacterial, archaeal, and fungal biomass increased on model particles over the incubation period, but not in the surrounding soil. Boxplots show (**A, C**) 16S rRNA (16S) and (**B, D**) ITS gene copy numbers (per gram of particles or soil) (**A, B**) on model particles and (**C, D**) in the soil after zero, one, 13, 29, and 50 days of lab incubation, and 47 days of field incubation. Gene copy numbers can be interpreted as a proxy for microbial biomass associated with model particles or soil. (**E**) Fluorescence microscopy images show microbial DNA at model chitin particle surfaces after one, 13, 29, and 50 days of lab incubation. Yellow points represent DNA stained with SYBR Green I double stranded nucleic acid stain. Microscopy was performed using a Leica confocal laser scanning microscope coupled with a Thunder imaging system. Scale bars indicate 1 µm. (**A, B, C, D**) Different letters indicate significant pairwise differences across time points for each treatment and sample type based on Dunn's post hoc tests with Bonferroni adjustment following Kruskal–Wallis tests (adjusted P-value < 0.05). Supplementary Table S3 gives the exact number of measurements that boxplots are based on.

Fig. 3 Bacterial and archaeal, and fungal richness of model chitin particles significantly decreased over the incubation period. Boxplots (from left to right) show (**A**) bacterial and archaeal richness (number of ASVs), and (**B**) fungal richness that was detected in (i) the soil, which was used to set up incubations, (ii) on model particles, and (iii) in the surrounding soil after one, 13, 29, and 50 days of lab incubation and 47 days field incubation. Sequencing data was randomly rarefied at the minimum read count for bacteria and archaea (3015 reads), and for fungi (5676 reads) to guarantee equal sequencing depth of samples. Different letters indicate significant pairwise differences across time points for each treatment and sample type based on Dunn's post hoc tests with Bonferroni adjustment following Kruskal–Wallis tests (adjusted P-value < 0.05). S – sieved soil that was used to set up incubations, CH – chitin, CE – cellulose, NS – no substrate (control to account for the effect of the scaffold material of model particles, gellan gum), NP – soil without particles.

Fig. 4 Chitin particle-associated bacterial and archaeal communities underwent temporal changes. PCoA plots show potential dissimilarity of (**A-C, G-I**) bacterial and archaeal, and (**D-F, J-L**) fungal communities (i) associated with model chitin, cellulose or "no substrate" particles or (ii) in the surrounding soil across harvest time points (day one to 50 or 47). Plots also include communities of the five sieved soils (at incubation start, 0d) that were used to set up mesocosms and mesh bags. PCoA plots are based on Aitchison distances (Euclidean distance of clr-transformed sequencing data). The colour of data points indicates the harvest time point, the shape of points depicts the type of incubation (lab or field). CH – chitin, CE – cellulose, NS – "no substrate".

Fig. 5 Model chitin and cellulose particles favoured certain bacterial genera compared to the soil, in which particles had been incubated. Plots show genera that were significantly enriched in their relative abundance on (**A**) model chitin, (**B**) model cellulose particles, and (**C**) "no substrate" model particles (controls) compared to the surrounding soil at one day, 13, 29, and 50 days after the start of the lab incubation and 47 days after the start of the field incubation. Bars represent the log fold change, a measure of effect size, in the relative abundance of significantly enriched genera (FDR-adjusted P value < 0.05). Differentially abundant genera were identified using the ANCOM-BC package [61]. Error bars represent the standard error of the log fold change. We controlled false discovery rates with the Benjamini-Hochberg multiple test correction. *Aquaspirillum** refers to the *[Aquaspirillum] articum group*, *Rhizobiaceae spp.*** refers to the *Allorhizobium-Neorhizobium-Pararhizobium-Rhizobium* clade. Note that we did not detect any genera that were significantly enriched on model "no substrate" particles after 47 days of incubation in the field. See Fig. S3-S5 for positive and negative log fold changes of genera on model particles compared to the surrounding soil.

Fig. 6 Model chitin and cellulose particles favoured certain fungal genera compared to the soil, in which particles had been incubated. Plots show genera which were significantly enriched in their relative abundance on (**A**) model chitin, (**B**) model cellulose particles, and (**C**) model "no substrate" particles (controls) compared to the surrounding soil at one day, 13, 29, and 50 days after the start of the lab incubation and 47 days after the start of the field incubation. Bars represent the log fold change, a measure of effect size, in the relative abundance of significantly enriched genera (FDR-adjusted P value < 0.05). Differentially abundant genera were identified using the ANCOM-BC package [61]. Error bars represent the standard error of the log fold change. We controlled false discovery rates with the Benjamini-Hochberg multiple test correction. Note that we did not detect any genera which were significantly enriched on model "no substrate" particles after 13 and 50 days of incubation in the lab, and 47 days of incubation in the field. See Fig. S6-S8 for positive and negative log fold changes of genera on model particles compared to the surrounding soil.

References

1. Lehmann J, Kleber M. The contentious nature of soil organic matter. *Nature* 2015;**528**:60–68. https://doi.org/10.1038/nature16069

2. Angst G, Mueller KE, Nierop KGJ, et al. Plant- or microbial-derived? A review on the molecular composition of stabilized soil organic matter. *Soil Biology and Biochemistry* 2021;**156**:108189. https://doi.org/10.1016/j.soilbio.2021.108189

3. Lavallee JM, Soong JL, Cotrufo MF. Conceptualizing soil organic matter into particulate

and mineral-associated forms to address global change in the 21st century. *Global*

Change Biology 2020;**26**:261–273. https://doi.org/10.1111/gcb.14859

4. Young IM, Crawford JW, Nunan N, et al. Chapter 4 Microbial Distribution in Soils.

Physics and Scaling. Advances in Agronomy. Elsevier Inc., 2008, 81–121.

5. Witzgall K, Vidal A, Schubert DI, et al. Particulate organic matter as a functional soil

component for persistent soil organic carbon. *Nature Communications* 2021;**12**:4115.

https://doi.org/10.1038/s41467-021-24192-8

7. Allison SD, Weintraub MN, Gartner TB, et al. Evolutionary-Economic Principles as

Regulators of Soil Enzyme Production and Ecosystem Function. In: Shukla G, Varma A

(eds), Soil Enzymology. Berlin, Heidelberg: Springer Berlin Heidelberg, 2010, 229–243.

8. Nunan N, Schmidt H, Raynaud X. The ecology of heterogeneity: soil bacterial

communities and C dynamics. *Philosphical Transactions B* 2020;**375**.

https://doi.org/10.1007/s00248-002-1044-x

9. Sinsabaugh RL, Lauber CL, Weintraub MN, et al. Stoichiometry of soil enzyme activity

at global scale. *Ecology Letters* 2008;**11**:1252–1264. https://doi.org/10.1111/j.1461-

0248.2008.01245.x

10. Gómez EJ, Delgado JA, González JM. Persistence of microbial extracellular enzymes

in soils under different temperatures and water availabilities. *Ecol Evol* 2020;**10**:10167–

10176. https://doi.org/10.1002/ece3.6677

11. Bardgett RD, Van Der Putten WH. Belowground biodiversity and ecosystem functioning.

Nature 2014;**515**:505–511. https://doi.org/10.1038/nature13855

12. Berlemont R. Distribution and diversity of enzymes for polysaccharide degradation in

fungi. *Sci Rep* 2017;**7**:222. https://doi.org/10.1038/s41598-017-00258-w

13. Wu C, Ren H, Liu Z, et al. Spatial heterogeneity of resource availability drives soil

bacterial community assembly along the sandy coast of Southern China. *Global*

Ecology and Conservation 2024;**54**:e03171.

https://doi.org/10.1016/j.gecco.2024.e03171

14.   Sinsabaugh RL, Hill BH, Follstad Shah JJ. Ecoenzymatic stoichiometry of microbial
organic nutrient acquisition in soil and sediment. *Nature* 2009;**462**:795–798.
https://doi.org/10.1038/nature08632

15.   Talamantes D, Biabini N, Dang H, et al. Natural diversity of cellulases, xylanases, and
chitinases in bacteria. *Biotechnol Biofuels* 2016;**9**:133. https://doi.org/10.1186/s13068-
016-0538-6

16.   Guseva K, Mohrlok M, Alteio L, et al. Bacteria face trade-offs in the decomposition of
complex   biopolymers.   *PLoS   computational   biology*   2024;**20**:e1012320.
https://doi.org/10.1371/journal.pcbi.1012320

17.   Berlemont R, Martiny AC. Phylogenetic Distribution of Potential Cellulases in Bacteria.
Applied   and   Environmental   Microbiology   2013;**79**:1545–1554.
https://doi.org/10.1128/AEM.03305-12

18.   Berlemont R, Martiny AC. Genomic Potential for Polysaccharide Deconstruction in
Bacteria. *Applied   and   Environmental   Microbiology*   2015;**81**:1513–1519.
https://doi.org/10.1128/AEM.03718-14

19.   Zimmerman A, Martiny A, Allison S. Microdiversity of extracellular enzyme genes
among sequenced prokaryotic genomes. *The ISME journal* 2013;**7**:1187–1199.
https://doi.org/10.1038/ismej.2012.176

20.   Das SN, Sarma PVSRN, Neeraja C, et al. Members of Gammaproteobacteria and
Bacilli represent the culturable diversity of chitinolytic bacteria in chitin-enriched soils.
World J Microbiol Biotechnol 2010;**26**:1875–1881. https://doi.org/10.1007/s11274-010-
0369-8

21.   Wieczorek AS, Hetz SA, Kolb S. Microbial responses to chitin and chitosan in oxic and
anoxic   agricultural   soil   slurries.   *Biogeosciences*   2014;**11**:3339–3352.
https://doi.org/10.5194/bg-11-3339-2014

22.   Li D, Zhang Z, Wang J, et al. Estimate of the degradation potentials of cellulose, xylan,
and chitin across global prokaryotic communities. *Environmental Microbiology*
2023;**25**:397–409. https://doi.org/10.1111/1462-2920.16290

23.   Lehmann J, Solomon D, Kinyangi J, et al. Spatial complexity of soil organic matter

forms   at   nanometre   scales.   *Nature   Geosc*   2008;**1**:238–242.

https://doi.org/10.1038/ngeo155

24.   Lehmann J, Bossio DA, Kögel-Knabner I, et al. The concept and future prospects of soil

health.   *Nature   Reviews   Earth   and   Environment*   2020;**1**:544–553.

https://doi.org/10.1038/s43017-020-0080-8

25.   Cotrufo MF, Wallenstein MD, Boot CM, et al. The Microbial Efficiency-Matrix

Stabilization (MEMS) framework integrates plant litter decomposition with soil organic

matter stabilization: Do labile plant inputs form stable soil organic matter? *Global*

Change Biology 2013;**19**:988–995. https://doi.org/10.1111/gcb.12113

26.   Schlüter S, Leuther F, Albrecht L, et al. Microscale carbon distribution around pores and

particulate organic matter varies with soil moisture regime. *Nature Communications*

2022;**13**:2098. https://doi.org/10.1038/s41467-022-29605-w

27.   Fu Y, Xu Y, Wang Q, et al. Deciphering the microbial players driving straw

decomposition and accumulation in soil components of particulate and mineral-

associated   organic   matter.   *Soil   Biology   and   Biochemistry*   2025;**209**:109871.

https://doi.org/10.1016/j.soilbio.2025.109871

28.   Martens J, Mueller CW, Joshi P, et al. Stabilization of mineral-associated organic

carbon   in   Pleistocene   permafrost.   *Nat   Commun*   2023;**14**:2120.

https://doi.org/10.1038/s41467-023-37766-5

29.   Wieczorek AS, Schmidt O, Chatzinotas A, et al. Ecological Functions of Agricultural Soil

Bacteria and Microeukaryotes in Chitin Degradation: A Case Study. *Front Microbiol*

2019;**10**:1293. https://doi.org/10.3389/fmicb.2019.01293

30.   Jacquiod S, Franqueville L, Cécillon S, et al. Soil Bacterial Community Shifts after Chitin

Enrichment: An Integrative Metagenomic Approach. *PLoS One* 2013;**8**:e79699.

https://doi.org/10.1371/journal.pone.0079699

31.   Cordero OX, Stocker R. A particularly useful system to study the ecology of microbes.

656   *Environmental Microbiology Reports* 2017;**9**:16–17. https://doi.org/10.1111/1758-

657   2229.12490

658   32.   Cordero OX, Datta MS. Microbial interactions and community assembly at microscales.

659   *Current Opinion in Microbiology* 2016;**31**:227–234.

660   https://doi.org/10.1016/j.mib.2016.03.015

661   33.   Datta MS, Sliwerska E, Gore J, et al. Microbial interactions lead to rapid micro-scale

662   successions on model marine particles. *Nat Commun* 2016;**7**:11965.

663   https://doi.org/10.1038/ncomms11965

664   34.   Ebrahimi A, Schwartzman J, Cordero OX. Cooperation and spatial self-organization

665   determine rate and efficiency of particulate organic matter degradation in marine

666   bacteria. *Proceedings of the National Academy of Sciences of the United States of*

667   *America* 2019;**116**:23309–23316. https://doi.org/10.1073/pnas.1908512116

668   35.   Enke TN, Datta MS, Schwartzman J, et al. Modular Assembly of Polysaccharide-

669   Degrading Marine Microbial Communities. *Curr Biol* 2019;**29**:1528-1535.e6.

670   https://doi.org/10.1016/j.cub.2019.03.047

671   36.   Das N, Triparthi N, Basu S, et al. Progress in the development of gelling agents for

672   improved culturability of microorganisms. *Front Microbiol* 2015;**6**.

673   https://doi.org/10.3389/fmicb.2015.00698

674   37.   Wang X, Zohar-Perez C, Zeng Y, et al. Assessment of the environmental impact of

675   agar, alginate, and gellan gum carbohydrate gum macro beads biodegradation in a

676   simulated agricultural field system. *Environmental Technology & Innovation*

677   2023;**30**:103034. https://doi.org/10.1016/j.eti.2023.103034

678   38.   Mussmann M. A simple method to prepare substrate-coated model particles for in situ

679   enrichment of macromolecule-degrading microorganisms. Zenodo. *Zenodo* 2026.

680   https://doi.org/10.5281/zenodo.18020354

39. Kitzler B, Zechmeister-Boltenstern S, Holtermann C, et al. Nitrogen oxides emission
from two beech forests subjected to different nitrogen loads. *Biogeosciences*
2006;**3**:293–310. https://doi.org/10.5194/bgd-2-1381-2005

40. Kaiser C, Koranda M, Kitzler B, et al. Belowground carbon allocation by trees drives
seasonal patterns of extracellular enzyme activities by altering microbial community
composition in a beech forest soil. *New Phytologist* 2010;**187**:843–858.
https://doi.org/10.1111/j.1469-8137.2010.03321.x

41. Eichorst SA, Strasser F, Woyke T, et al. Advancements in the application of NanoSIMS
and Raman microspectroscopy to investigate the activity of microbial cells in soils.
FEMS Microbiology Ecology 2015;**91**:fiv106. https://doi.org/10.1093/femsec/fiv106

42. Simon E, Guseva K, Darcy S, et al. Distinct microbial communities are linked to organic
matter properties in millimetre-sized soil aggregates. *The ISME Journal*
2024;**18**:wrae156. https://doi.org/10.1093/ismejo/wrae156

43. Fahey TJ, Fahey C, Goebel M, et al. Fine root decomposition and macronutrient
release in temperate forests: Examining the litter bag and intact soil core approaches
and their relationship with fungal community composition. *Applied Soil Ecology*
2024;**193**:105147. https://doi.org/10.1016/j.apsoil.2023.105147

44. Moravcová A, Barbi F, Algora C, et al. Contrasting stability of fungal and bacterial
communities during long-term decomposition of fungal necromass in Arctic tundra.
Environ Microbiome 2025;**20**:75. https://doi.org/10.1186/s40793-025-00730-5

45. Pjevac P, Hausmann B, Schwarz J, et al. An Economical and Flexible Dual Barcoding,
Two-Step PCR Approach for Highly Multiplexed Amplicon Sequencing. *Frontiers in*
Microbiology 2021;**12**:669776. https://doi.org/10.3389/fmicb.2021.669776

46. Parada AE, Needham DM, Fuhrman JA. Every base matters: assessing small subunit
rRNA primers for marine microbiomes with mock communities, time series and global
field samples. *Environmental Microbiology* 2016;**18**:1403–1414.
https://doi.org/10.1111/1462-2920.13023

47. Apprill A, McNally S, Parsons R, et al. Minor revision to V4 region SSU rRNA 806R

gene primer greatly increases detection of SAR11 bacterioplankton. *Aquat Microb Ecol*

2015;**75**:129–137. https://doi.org/10.3354/ame01753

48. Nilsson RH, Anslan S, Bahram M, et al. Mycobiome diversity: high-throughput

sequencing and identification of fungi. *Nature Reviews Microbiology* 2019;**17**:95–109.

https://doi.org/10.1038/s41579-018-0116-y

49. Taylor DL, McCormick MK. Internal transcribed spacer primers and sequences for

improved characterization of basidiomycetous orchid mycorrhizas. *New Phytologist*

2008;**177**:1020–1033. https://doi.org/10.1111/j.1469-8137.2007.02320.x

50. Tedersoo L, Jairus T, Horton BM, et al. Strong host preference of ectomycorrhizal fungi

in a Tasmanian wet sclerophyll forest as revealed by DNA barcoding and taxon-specific

primers. *New Phytologist* 2008;**180**:479–490. https://doi.org/10.1111/j.1469-

8137.2008.02561.x

51. White TJ, Bruns T, Lee S, et al. Amplification and direct sequencing of fungal ribosomal

RNA genes for phylogenetics. PCR Protocols. Elsevier, 1990, 315–322.

52. Ihrmark K, Bödeker ITM, Cruz-Martinez K, et al. New primers to amplify the fungal ITS2

region – evaluation by 454-sequencing of artificial and natural communities. *FEMS*

Microbiology Ecology 2012;**82**:666–677. https://doi.org/10.1111/j.1574-

6941.2012.01437.x

53. Alteio LV, Séneca J, Canarini A, et al. A critical perspective on interpreting amplicon

sequencing data in soil ecological research. *Soil Biology and Biochemistry*

2021;**160**:108357. https://doi.org/10.1016/j.soilbio.2021.108357

54. Abarenkov K, Allan Zirk, Piirmann, Timo, et al. Full UNITE+INSD dataset for Fungi.

2020. 2020.

55. R Core Team. R: A Language and Environment for Statistical Computing. Vienna,

Austria: R Foundation for Statistical Computing, 2021.

56. Huang R, Soneson C, Ernst FGM. TreeSummarizedExperiment: a S4 class for data
with   hierarchical   structure.   *F1000Res*   2021;**9**:1246.   https://doi-
org.uaccess.univie.ac.at/10.12688/f1000research.26669.2

57. McMurdie PJ, Holmes S. Phyloseq: An R Package for Reproducible Interactive Analysis
and   Graphics   of   Microbiome   Census   Data.   *PLoS   ONE*   2013;**8**:e61217.
https://doi.org/10.1371/journal.pone.0061217

58. Oksanen J, Simpson GL, Blanchet FG. vegan: Community Ecology Package. 2024.
2024.

59. Wickham H. ggplot2: Elegant Graphics for Data Analysis. Springer Cham, 2016.

60. Inkscape Project. Inkscape. 2021. 2021.

61. Lin H, Peddada SD. Analysis of compositions of microbiomes with bias correction. *Nat*
Commun 2020;**11**:3514. https://doi.org/10.1038/s41467-020-17041-7

62. Gloor GB, Macklaim JM, Pawlowsky-Glahn V, et al. Microbiome datasets are
compositional: And this is not optional. *Frontiers in Microbiology* 2017;**8**:2224.
https://doi.org/10.3389/fmicb.2017.02224

63. Enke TN, Leventhal GE, Metzger M, et al. Microscale ecology regulates particulate
organic matter turnover in model marine microbial communities. *Nat Commun*
2018;**9**:2743. https://doi.org/10.1038/s41467-018-05159-8

64. Wang ZY, Wang RX, Zhou JS, et al. An assessment of the genomics, comparative
genomics and cellulose degradation potential of *Mucilaginibacter polytrichastri* strain
RG4-7.   *Bioresour   Technol*   2020;**297**:122389.
https://doi.org/10.1016/j.biortech.2019.122389

65. Kim J, Lee B, Chhetri G, et al. Identification of Mucilaginibacter conchicola sp. nov.,
Mucilaginibacter achroorhodeus sp. nov. and Mucilaginibacter pallidiroseus sp. nov.
and emended description of the genus Mucilaginibacter. *Int J Syst Evol Microbiol*
2022;**72**. https://doi.org/10.1099/ijsem.0.005431

66. Yang H-C, Im W-T, An D-S, et al. Silvimonas terrae gen. nov., sp. nov., a novel chitin-
degrading facultative anaerobe belonging to the 'Betaproteobacteria'. *Int J Syst Evol*
Microbiol 2005;**55**:2329–2332. https://doi.org/10.1099/ijs.0.63837-0

67. Song C, Schmidt R, de Jager V, et al. Exploring the genomic traits of fungus-feeding
bacterial genus Collimonas. *BMC Genomics* 2015;**16**:1103.
https://doi.org/10.1186/s12864-015-2289-3

68. Ballhausen M-B, Vandamme P, Boer W de. Trait Differentiation within the Fungus-
Feeding (Mycophagous) Bacterial Genus Collimonas. *PLOS ONE* 2016;**11**:e0157552.
https://doi.org/10.1371/journal.pone.0157552

69. So Y, Chhetri G, Kim I, et al. Roseateles caseinilyticus sp. nov. and Roseateles
cellulosilyticus sp. nov., isolated from rice paddy field soil. *Antonie Van Leeuwenhoek*
2024;**117**:87. https://doi.org/10.1007/s10482-024-01988-4

70. Itoh T, Hibi T, Fujii Y, et al. Cooperative Degradation of Chitin by Extracellular and Cell
Surface-Expressed Chitinases from Paenibacillus sp. Strain FPU-7. *Appl Environ*
Microbiol 2013;**79**:7482–7490. https://doi.org/10.1128/AEM.02483-13

71. Yadav S, Dubey SK. Cellulose degradation potential of *Paenibacillus lautus* strain
BHU3 and its whole genome sequence. *Bioresource Technology* 2018;**262**:124–131.
https://doi.org/10.1016/j.biortech.2018.04.067

72. Kim ES, Kim B-S, Kim K-Y, et al. Aerobic and anaerobic cellulose utilization by
Paenibacillus sp. CAA11 and enhancement of its cellulolytic ability by expressing a
heterologous endoglucanase. *Journal of Biotechnology* 2018;**268**:21–27.
https://doi.org/10.1016/j.jbiotec.2018.01.007

73. Rytioja J, Hildén K, Yuzon J, et al. Plant-Polysaccharide-Degrading Enzymes from
Basidiomycetes. *Microbiol Mol Biol Rev* 2014;**78**:614–649.
https://doi.org/10.1128/MMBR.00035-14

74. Enke TN, Datta MS, Schwartzman J, et al. Modular Assembly of Polysaccharide-
Degrading Marine Microbial Communities Report Modular Assembly of Polysaccharide-

Degrading   Marine   Microbial   Communities.   2019;1528–1535.

https://doi.org/10.1016/j.cub.2019.03.047

75. Kang KS, Veeder GT, Mirrasoul PJ, et al. Agar-Like Polysaccharide Produced by a

Pseudomonas Species: Production and Basic Properties. *Appl Environ Microbiol*

1982;**43**:1086–1091. https://doi.org/10.1128/aem.43.5.1086-1091.1982