Continental-scale integration of soil metagenomes and organic matter chemistry reveals ubiquitous microbial capacity for chemically-recalcitrant carbon decomposition

Young C. Song[1], Cheng Shi[1,2], Kelly G. Stratton[3], Christian Ayala-Ortiz[1], Izabel Stohel[1], Viviana Freire-Zapata[4], Malak M. Tfaily[4], Emiley Eloe-Fadrosh[5] and Emily B. Graham[3,6]

1. Environmental Molecular Sciences Laboratory, Pacific Northwest National Laboratory, Richland WA 99354, United States
2. Eawag, Swiss Institute of Aquatic Science and Technology, Überlandstrasse 133 CH-8600 Dübendorf, Switzerland
3. Biological Sciences Division, Pacific Northwest National Laboratory, Richland WA 99354, United States
4. Department of Environmental Science, University of Arizona, Tucson, AZ 85719, United States
5. US Department of Energy Joint Genome Institute, Lawrence Berkeley National Laboratory, Berkeley, CA 94720, United States
6. School of Biological Sciences, Washington State University, Richland, WA 99354

Abstract

Soil organic matter (SOM) decomposition by microorganisms represents one of the largest uncertainties in predicting terrestrial carbon-atmosphere feedbacks. Yet we lack systematic understanding of the microbial taxa that drive decomposition of chemically distinct carbon pools and of the metabolic pathways they employ across environmental gradients. Here, we address this critical gap using the first continental-scale soil dataset pairing shotgun metagenomes with high-resolution molecular characterization of SOM chemistry. We use data from 47 standardized soil cores – selected using microbial respiration rates from 106 soils across the United States – to assemble 0.76 terabases (Tbs) of prokaryotic metagenome-assembled genomes (828 MAGs) and identify 66,727 distinct SOM molecules. Integrating these datasets revealed widespread microbial potential for depolymerizing chemically-recalcitrant carbon compounds previously considered stable. Notably, we uncover complementary metabolic specialization between genera affiliated with two abundant bacterial orders, *Rhizobiales* and *Chthoniobacterales,* and an archaeal order, *Nitrososphaerales.* This metabolic partitioning is consistent across soil depths and activity levels, suggesting coordinated decomposition of chemically-complex carbon through distinct but complementary biochemical strategies. The metabolic potential for depolymerization of chemically-recalcitrant compounds is supported by the abundance of these molecules across the soils, as indicated by Fourier-Transform Ion Cyclotron Resonance Mass Spectrometry (FTICR-MS), and by flux balance analysis of metabolic models. Our continental-scale integration of microbial genotypes with their chemical substrates reveals that a substantial fraction of SOM that is considered to be relatively stable harbors decomposition potential that current Earth system models fail to capture. These findings provide mechanistic insight for incorporating microbial metabolic pathways into carbon cycle predictions and highlight an underappreciated vulnerability of soil carbon to environmental variability.

Main

Soil microbial respiration is a critical component of the global carbon cycle, yielding 98 ± 12 Petagram (Pg C/year) of carbon dioxide (CO_2) into the atmosphere [1]. Most soil carbon (C) is contained within chemically diverse soil organic matter (SOM) pools that, through their interactions with plants, mineral surfaces, water, and other biological and physical soil properties, have the potential to sequester carbon within soils for centuries [2]. Chemically-recalcitrant SOM is typically thought to have long residence times in soils [3,4], and the widespread potential for its decomposition suggests a possible missing link in the global carbon cycle. Microbial metabolism of SOM proceeds through myriad metabolic processes [5], yet how interconnected factors and environmental contexts regulate SOM decomposition is largely unknown.

SOM contains a wide variety of organic molecules derived from plant-microbe interactions [5,6], biomass decomposition, and biochemical transformation into chemically-recalcitrant components [7,8]. These small (<1,000 Dalton (Da)) compounds, including amino acids, lipids and organic phosphates, are transformed by soil microorganisms into CO_2 through respiration [5]. While phylogenetic and metabolic inferences of soil metagenome-assembled genomes (MAGs) have revealed potential microbial mechanisms for transforming various organic molecules [9–11], robust characterization and confirmation of these pathways have historically depended on isolation and cultivation the organisms represented by MAGs [12]. Alternatively, organic matter characterization approaches using Fourier-Transform Ion Cyclotron Resonance Mass Spectrometry (FTICR-MS) provide molecular-level information about SOM chemistry [5], which can be integrated with MAG, potentially enhancing our understanding of SOM transformations while circumventing cultivation bottlenecks.

Given that the predicted residence times of stabilized SOM span decades to centuries, current Earth system models consider large fractions of SOM as carbon sinks [13]. We add to a growing body of work challenging this notion, while addressing knowledge gaps in the metabolic pathways that regulate decomposition of SOM across environmental gradients, using a unique soil dataset that integrates standardized metagenomics with high-resolution molecular characterization of SOM chemistry collected from surface and subsoils across the Continental

United States (CONUS). Through the 1000 Soils Pilot of the Molecular Observation Network (MONet) [14], we assayed 106 surface and subsoils to uncover complementary roles in continental-scale soil C cycling processes conducted by genera within two predominant microbial clades, *Rhizobiales* and *Chthoniobacterales*. Our observations stress the vulnerability of the soil carbon stocks to increased microbial activity under continued environmental change, emphasizing the need for a tractable genetic framework for incorporating microbial mechanisms into Earth system models.

Results

We investigated 106 soils that spanned the continental United States, representing 17 of 20 ecoclimatic domains defined by the National Ecological Observatory Network (NEON) [15]. This design enabled us to capture a wide range of spatial variation in soils across the continental United States and is consistent with other published continental-scale works [16,17]. We then assessed variation in microbial communities and SOM chemistry across 47 soils at two depths (surface and subsoil). Soils spanned 37 sites across the CONUS, encompassing most major ecoregions (Supplementary Fig. 1 and Supplementary Table 1). High-resolution mass spectrometry (FTICR-MS) identified 66,727 distinct SOM molecules (average 7.6 x 10^3 molecules per sample), while metagenomic sequencing (average 1.1×10^8 reads per sample) yielded 828 metagenome-assembled genomes (MAGs). Analyses of microbial relative abundance in soils indicated that on average 73.1% of metagenomic reads were microbial. Additionally, the assembled MAGs captured 50.6% and 36.5% of the bacterial and archaeal communities at the genus level, respectively.

While we acknowledge that non-prokaryotic organisms such as fungi and viruses have large impacts on biogeochemistry, assembling their metagenomes from soil remains a challenging task for several reasons, including short-read sequencing biases, limited reference databases, and variability in GC content [18–21]. These challenges were evident in our results, as taxonomic classification using Kraken2 [22] revealed that only a small fraction (less than 1%) of the reads from all soil samples were fungal. Therefore, we focused on bacterial and archaeal communities owing to their substantial roles in soil carbon cycling [23]. Given the above, computational approaches for soil microbiomes allow for deeper investigation into bacteria and archaea than

other clades [18–21], and we expect that the results presented here will lay a foundation for

investigations into the role of interkingdom interactions on belowground biogeochemistry, as

new technological advances in identifying fungi and viruses are available. Detailed procedures of

both SOM measurements and MAG assemblies are highlighted in the **Methods** section.

Prevalence of *Rhizobiales*, *Chthoniobacterales*, and *Nitrososphaerales* across soil samples

We recovered 828 MAGs, which were dereplicated into 358 genomes. From the dereplicated

MAGs, 48 were high quality (completion > 90% with contamination < 5%) and 310 were

medium quality (completion ≥ 50% with contamination of < 10%) according to the Minimum

Information about Metagenome-Assembled Genome (MIMAG; Supplementary Table 2) [24].

Bacterial MAGs constituted 89.1% of the dereplicated genomes (n=319 MAGs), while the

remaining 39 MAGs were archaeal. Phylogenetic inference of the bacterial MAGs (143 from the

surface and 176 from the subsoil) revealed that bacterial diversity was dominated by members of

Pseudomonadota (21.0%; n=67), *Actinomycetota* (23.8%; n=76) and *Acidobacteriota* (17.2%;

n=55). Of the archaeal MAGs, 92.3% (n=36) were classified as the phylum *Thermoproteota*,

while the remaining three genomes were assigned to the phylum, *Thermoplasmatota*. Nearly

92% of the *Thermoplasmatota* genomes (n=34) were associated with the order *Nitrososphaerales*

(Fig. 1).

Examining soil microbial diversity at the order level revealed that *Rhizobiales* (phylum

Pseudomonadota) and *Chthoniobacterales* (phylum *Verrucomicrobiota*) were the two most

prevalent lineages, representing 12.0% (n = 43) and 6.1% (n = 22) of the dereplicated genomes,

respectively (Fig. 1). Relative abundances inferred from read mapping indicated that some

genera, such as the candidate group VAZQ01 (order *Rhizobiales*) and the *Udaeobacter* (order

Chthoniobacterales) were the dominant taxa in high-respiration soils. Both VAZQ01 and

Udaeobacter were also the two most prevalent genera overall, representing 3.6% (n=13) and

2.5% (n=9) of the total dereplicated MAGs, respectively. Genera within *Rhizobiales*, such as

Bradyrhizobium and *Methyloaceanibacter*, were abundant across all soil types except for

low-respiration subsoils.

Unlike *Rhizobiales* and *Chthoniobacterales*, some genera within the phyla *Actinomycetota* and
Chloroflexota were more prevalent in the subsoils compared to the surface soils. For instance,
the candidate genera VRUE01 and DASSJM01, both the members of *Actinomycetota*, had the
highest average number of mapped reads in high-respiration subsoils, while the candidate groups
CF-13 and 3−1−20CM−2−70−9 affiliated with *Chloroflexota* were the most prevalent genera in
low-respiration subsoils. The only exception to this observation was the genus *Rubrobacter_F*
within *Actinomycetota*, which was the most prevalent genus in low-respiration surface soils.
145
146 Of the genera within the archaeal order *Nitrosophaerales*, the candidate genus TA-21 and
147 *Nitrosphaera* were the most prevalent groups, represented by 9 and 7 dereplicated MAGs,
148 respectively. Both genera had the highest average number of mapped reads in low-respiration
149 surface soils, while TA-21 was also present in high-respiration subsoils.
150

Soil chemistry reveals depth-stratified carbon pools with distinct decomposition signatures
Pairwise comparisons of SOM molecular composition across depths and potential rates of
microbial respiration revealed systematic chemical differentiation with implications for carbon
cycling (G-test of uniqueness; Fig. 2 and Supplementary Table 3). While lignin- and condensed
hydrocarbon-like compounds were prevalent across all soil types, we observed distinct SOM
chemistry between surface and subsoils. For instance, comparison of chemistry in
low-respiration subsoils with surface soils (both high- and low-respiration) revealed unique
compounds in the subsoils (G-test p-value < 0.05; Fig. 2C and E), of which $> 70\%$ were
composed of lignin-, condensed-hydrocarbon- and tannin-like compounds.
160

Similarly, when comparing the chemistry of high-respiration subsoils against high-respiration
surface soils, we observed SOM molecules unique to the subsoils, mostly composed (~80%) of
lignin-, condensed hydrocarbon-, and tannin-like compounds (Fig. 2B). Comparison between
high-respiration subsoils and low-respiration surface soils (Fig. 2D) revealed that while 57% of
SOM molecules unique to the subsoils were lignin-, condensed hydrocarbon- and tannin-like
compounds, 32.4% were protein- (16.2%) and unsaturated hydrocarbon- (16.2%) like
compounds (Fig. 2 and Supplementary Table 3). Between subsoil environments (Fig. 2F),
high-respiration subsoils contained 2,322 unique compounds (>75% lignin- and condensed

hydrocarbon-like molecules), while low-respiration subsoils had only 18 unique compounds, half

of which were lipid-like—potentially reflecting microbial necromass under carbon-limited

conditions.

Microbial potential for depolymerizing chemically-recalcitrant SOMs is widespread

Using co-occurrence networks between annotated orthologs from MAGs and SOM molecules

acquired from each of the four depth/respiration partitions, we evaluated potential relationships

between microorganisms and metabolites. In particular, we identified densely connected modules

within a large SOM-KEGG Orthology (KO) network (for all soil layers and potential respiration

rates), revealing genes that showed significant relationships with chemically-recalcitrant classes

of SOM, such as lignin-, condensed hydrocarbon-, and tannin-like molecules. A case in point is a

set of 22 modules recovered in the high-respiration surface soils, with MCODE cluster scores

ranging from 2.7 to 19.9 (Fig. 3 and Supplementary Fig. 2). We also recovered a single module

from the low-respiration surface soil with a cluster score of 94.2. This module was composed of

189 nodes, 103 of which represented KO-annotated genes that were associated with metabolic

pathways such as amino acid metabolism/synthesis, metabolism of C5 sugars, carbon fixation

and cellular signalling and processing (Supplementary Table 4). Nearly 75% of the 86 SOM

nodes within this module represented lignin-like compounds, while condensed-hydrocarbon- and

tannin-like compounds were represented by 6 nodes each. This module also contained 5

amino-sugar-, 4 protein-like compounds and an unclassified compound.

The modules identified in high-respiration surface soils revealed the microbial potential for

interacting with chemically-recalcitrant carbohydrates, including molecules resembling

unsaturated hydrocarbons. Among the 22 modules detected in these soils, 15 contained between

2 and 5 KO-annotated genes linked to pathways such as amino acid catabolism, C5 sugar and

starch metabolism, and secondary metabolite biosynthesis (Fig. 3 and Supplementary Fig. 2).

Notable genes included *ligXa* and *graA* that are associated with the breakdown of lignin and

other chemically-recalcitrant compounds [25,26]. These genes showed positive correlations with

various combinations of metabolites, including one or multiple instances of lignin-, condensed

hydrocarbon-, protein-, and unsaturated hydrocarbon-like compounds.

At the opposite end of our spectrum of soils, the module describing low-respiration subsoils was consistent with its unique chemistry profile, which included lipid-like compounds that were absent in the surface soils and in high-respiration subsoils (Fig. 3 and Supplementary Table 4). In this module, we observed 306 separate interactions (all represented by edges with correlations of > 0.95) between a set of SOMs including two lignin-like compounds, 10 protein-like compounds, three lipid-like compounds, two amino-sugar like compounds and an unclassified molecule, and 17 KO-annotated genes (Supplementary Table 4). Five of these genes were associated with amino acid catabolism, while the remaining 12 were linked to cellular functions (e.g., secretion system and biosynthesis of secondary metabolites and cofactors).

Genes in the KEGG database that are associated with metabolites detected by FTICR-MS (see **Methods**) further supported the potential for widespread microbial interactions with the chemically-recalcitrant SOM pool and its intermediates [27,28]. For instance, we uncovered metabolites associated with the biosynthesis of cofactors (map01240), which constituted 9.7%, 11.3% and 12.2% of metabolite-associated pathways identified in high- and low-respiration surface soils and in high-respiration subsoil, respectively (Fig. 4). We also detected metabolites affiliated with the metabolism of xenobiotics (map00980) and degradation of aromatic compounds (map01220) — key pathways for processing complex organic polymers. These specialized metabolisms were distributed across all soil environments, constituting between 3.5% to 8.3% of metabolite-associated pathways in all groups of soils (Fig. 4).

Soil microbial communities universally encode complementary arsenals for depolymerizing chemically-recalcitrant SOM

Genomic analysis for carbohydrate-active enzymes (CAZymes) [29] revealed that the selected genera within the two widespread bacterial orders (*Rhizobiales* and *Chthoniobacterales*) and the archaeal order *Nitrososphaerales* possessed extensive and complementary enzymatic machinery for processing chemically-recalcitrant carbon substrates. Consistent with the prevalence of lignin-, tannin-, and/or hydrocarbon-associated SOM, these genera contained a variety of genes encoding complex carbon depolymerization (Fig. 5). For instance, we detected 92 CAZyme classes in ≥ 50% of the MAGs assigned to the archaeal genus, *Nitrososphaera*, while recovering 43, 60, 52 and 54 CAZyme classes in ≥ 50% of the *Methyloacenibacter*, *Bradyrhizobium*,

Pseudolabrys and VAZQ0, respectively. In addition, 28 CAZyme classes were recovered in ≥

50% of the genus *Udaeobacter* (Table 1). In all sets glycoside hydrolases (GH) and

glycosyltransferases (GT) were the most prevalent CAZyme classes. Unlike the bacterial genera,

Nitrosphaera and the *Candidatus* TA-21 also possessed polysaccharide lyases (PL).

A cross-comparison of the CAZymes identified in the MAGs from each selected genus further

highlighted *Nitrosphaera*'s metabolic potential for degrading a wide variety of

chemically-recalcitrant carbohydrates. Of the 92 CAZyme classes recovered from the genomes

belonging to this genus, 25 classes were unique to *Nitrosphaera*. Of them, 52% (13) were GHs,

with the rest classified as GTs, auxiliary activity enzymes (AAs), carbohydrate esterases (CEs),

carbohydrate binding molecules (CBMs) or PL. The composition of the CAZyme classes in the

other selected bacterial and archaeal genera were not as diverse as *Nitrososphaera*. For instance,

of the 6 enzymes unique to the candidate archaeal genus JAFAQB01, 66.7% (4) were GHs, while

the remaining two were GTs. Similar observations were made in the genera belonging to both

Rhizobiales and *Chthoniobacterales*, where the majority of the CAZymes unique to each of these

lineages were dominated by GHs, GTs and CBMs.

While the recovery of fungal MAGs from metagenomic sequencing was limited [18,21], we

uncovered a set of potential fungal pathways for depolymerizing SOM from annotated contigs.

In total, we detected 11 CAZymes that were present in fungal contigs in ≥ 50% of the soils,

largely classified as GH, GT, AA or CBMs (Supplementary Fig. 3). Of these CAZymes, there

was a single AA unique to the high-respiration surface soils and two GTs unique to the

low-respiration subsoils. We also detected a GH unique to subsoils across both high- and

low-respiration.

Complementary amino acid and inorganic nitrogen acquisition mechanisms are enriched in

***Rhizobiales* and *Nitrososphaerales*-associated genera**

Flux balance analysis (FBA) of the selected genera belonging to *Rhizobiales*,

Chthoniobacterales and *Nitrososphaerales* revealed that despite the observed metabolic

potentials for amino acid catabolism in all prokaryotic MAGs, substrate preferences for these

amino acids appeared to differ by genera (Fig. 6). A closer examination of the selected bacterial

and archaeal MAGs revealed a set of genes encoding the transformation of L-aspartate,

L-glutamate, L-arginine and L-serine to various intermediates of the TCA cycle (Supplementary

Table 5). For instance, *Nitrosphaera* and *Udaeobacter* MAGs possessed at least a single copy of

most genes encoding transformation of L-aspartate to L-asparagine or oxaloacetate. Meanwhile,

the presence of the glutamine synthetase gene (*glnA*), which facilitates the amidation of

L-glutamate to synthesize L-glutamine, was more prevalent in *Bradyrhizobium* compared to

other genera. FBA results were supported by the distribution of KO-annotated genes, which

showed metabolic potential for transforming both L-aspartate and L-glutamate in

Methyloaceanibacter and *Pseudolabrys*. The annotation profiles for the bacterial and archaeal

genera aligned with the FBA, which indicated a high average uptake rate of L-aspartate and/or

L-glutamate for for *Nitrosphaera*, *Udaeobacter* and most genera in *Rhizobiales*.

The transformation of L-arginine was notably prevalent among *Methyloaceanibacter*,

Pseudolabris, candidate genus VAZQ01, and *Udaeobacter*, with these four bacterial genera

exhibiting a higher average uptake of arginine compared to other lineages under study (Fig. 6).

Interestingly, MAGs associated with *Bradyrhizobium* and *Udaeobacter* displayed a high average

copy number of the gene encoding aliphatic amidase (*amiE*), an enzyme that catalyzes the

hydrolysis of short-chain amides generated from the activity of arginine monooxygenase [30].

However, the average arginine uptake in *Bradyrhizobium* was close to zero, unlike in

Udaeobacter. Additionally, most *Bradyrhizobium* MAGs lacked genes associated with the

conversion of arginine to putrescine (e.g., *speB*, *speA*, *aguA*, and *aguB*), distinguishing them

from other *Rhizobiales*-associated genera. Similarly, the same genes were absent in JAFAQB01

and *Nitrososphaera*. These archaeal genera, along with TA-21, exhibited an almost negligible

average arginine uptake rate. All bacterial and archaeal MAGs contained at least one copy of

ilvA, which encodes L-serine/L-threonine dehydratase, an enzyme responsible for converting

serine into pyruvate. Among these genera, only the three archaeal genera and *Bradyrhizobium*

demonstrated a comparable average serine uptake rate. These patterns suggest complementary

amino acid utilization strategies between bacterial and archaeal lineages, whereby arginine

catabolism is concentrated in genera such as *Methyloaceanibacter*, *Pseudolabrys*, candidate

genus VAZQ01, and *Udaeobacter*, while archaeal taxa and *Bradyrhizobium* rely more

consistently on serine, aspartate and glutamate rather than arginine-based pathways.

In addition to complementary amino acid catabolism, bacterial and archaeal lineages also
displayed partitioned pathways for inorganic nitrogen transformations, with *Nitrosphaerales*
specializing in ammonia oxidation and key bacterial genera encoding dissimilatory nitrate
reduction. A nearly complete set of genes involved in ammonia oxidation to nitric oxide (NO),
including *pmoABC* and *nirK*, was found across all three *Nitrosphaerales* genera (Supplementary
Table 5). However, the *hao* gene, which encodes hydroxylamine dehydrogenase, was notably
absent in all *Nitrosphaerales* MAGs. In contrast, ammonia oxidation genes were entirely absent
from *Rhizobiales* and *Chthoniobacterales* MAGs. Instead, a partial set of genes involved in
dissimilatory nitrate reduction, such as *narGHI*, *napAB*, *nirBD*, and *nrfAH*, were detected,
particularly in *Methyloaceanibacter* and *Udaeobacter* (Supplementary Table 5).
303

Lipid metabolism reveals contrasting survival strategies under carbon limitation
The distribution of SOM molecular classes across soils revealed that while microbial potential
for depolymerizing chemically-recalcitrant organic matter appeared to be widespread in surface
soils and in high-respiration subsoils, biomass synthesis pathways tended to dominate
low-respiration subsoils. A complete or near-complete set of lipopolysaccharide biosynthesis
genes (e.g., *lpxD*, *lpxL*, and *gmhC*) were identified in MAGs across nearly all bacterial and
archaeal genera studied, with the exception of *Bradyrhizobium* (Supplementary Table 6).
Notably, these genes were present in ≥ 50% of the MAGs affiliated with *Nitrososphaera* and
Udaeobacter, highlighting fundamentally different strategies in these microbial lineages for
surviving carbon limitation [31,32].
314

We also investigated the occurrence of lipid biosynthesis genes in MAGs associated with two
candidate genera, CF-13 and 3-1-20CM-2-70-9, both classified under the phylum *Chloroflexota*
and the most prevalent taxa in low-respiration subsoils (Fig. 1). Compared to MAGs of dominant
bacterial and archaeal lineages, the genomes of CF-13 and 3-1-20CM-2-70-9 exhibited lower
CheckM completion rates (between 50% and 80%). Only a partial set of lipopolysaccharide
synthesis genes was detected in a small fraction of these MAGs, consistent with their relatively
lower genome completion rates relative to those of more prevalent bacterial and archaeal
lineages.
323

Discussion

Our study provides evidence that soil microbiomes across the continental United States possess genetic potential associated with the depolymerization of compounds traditionally considered to be chemically stable, including complementary metabolic specialization among abundant bacterial (*Rhizobiales*, *Chthoniobacterales*) and archaeal (*Nitrososphaerales*) lineages. This widespread microbial capacity for chemically-recalcitrant SOM turnover underscores the need for further investigations into the specific enzymatic pathways (e.g., CT-depolymerization, C_{15} metabolism and phenolic metabolism [33]) that appear to support the breakdown of chemically-recalcitrant compounds in most soils. By analyzing 828 MAGs - dereplicated to 358 -dereplicated MAGs and 66,727 molecular-level SOM measurements, we identified genera within dominant bacterial and archaeal orders—*Rhizobiales*,*Chthoniobacterales* and *Nitrosospaherales*—that use complementary enzymatic strategies to process chemically complex carbon across diverse soils. These findings suggest that slow carbon pools in current models may be more vulnerable to decomposition than previously thought, exposing a critical gap in carbon-climate feedback predictions.

It is important to note that SOM degradation not only depends on the chemical composition of SOM, but also on factors that control how easily microbes can encounter and use that material [16,17,34,35]. The rate at which SOM turns over is governed both by the chemical properties of carbon inputs and by physical and spatial constraints that limit contact between microorganisms and organic substrates. Organic compounds that are chemically bioavailable could be protected from microbial decomposition by physical barriers that restrict access or create micro-environmental constraints on decomposer activity and movement—for example, protection within soil aggregates or nanopores, or isolation in ice or anaerobic microsites [36,37]. These abiotic factors themselves could also be influenced by chemistry (e.g., chemical binding). Together, these approaches emphasize that accurate representation of soil carbon dynamics will require coupling molecular-level descriptions of SOM chemistry with explicit consideration of microbial physiology and the spatially heterogeneous soil environment. Although we focus on dissolved SOM, particulate organic carbon also plays key roles in nutrient cycling, carbon storage, and the formation of stable organic matter fractions that underpin soil health and ecosystem sustainability [38,39]. Together, these perspectives emphasize that accurate

representation of soil carbon dynamics will require coupling molecular-level descriptions of SOM chemistry with explicit consideration of microbial physiology and the spatially heterogeneous soil environment.

Integrating chemical profiling of SOM and functional annotation of soil MAGs showcases widespread microbial potential for lignin-, condensed hydrocarbon-, and tannin-like molecules that are typically thought to have long residence times in soils (Figs. 3 and 4) [40]. These dynamics may arise from microbial interactions with plant residues and their decomposition products, which are shaped by the molecular composition and spatial arrangement of SOM, as well as the functional potential of the co-located microbiome [40–42]. The prevalence of lignin- and condensed hydrocarbon-like molecules across soil depths in particular, regardless of microbial activity level, and their association with genes encoding SOM-degrading enzymes suggest current soil carbon stocks may be vulnerable to microbial activity in response to changing environmental conditions (Figs. 2, 3, and 4) [43,44]. The apparent persistence of the soil carbon pool therefore likely reflects limits on microbial access or activity, and the shifts in factors such as temperature, moisture, or redox conditions could relax these constraints and stimulate additional decomposition of existing soil carbon stocks. Identifying genes, such as *ligXa* and *graA*, directly linked to chemically-recalcitrant carbon processing highlights potential microbial mechanisms in decomposing chemically-recalcitrant carbohydrates, such as lignin to simpler molecules [25,26,45]. Additionally, genes for polyketide synthesis and other secondary metabolites may contribute to SOM depolymerization through mechanisms like metal mobilization [46,47] and the activation of quorum sensing activation [48]. The detection of complex networks in the low-respiration surface soils, together with prevalence of *Methyloceanibacter*, *Pseudolabrys*, and *Nitrososphaera* at the same soil depth and respiration level further underscores the increased reliance of these microorganisms on accessible carbohydrates during periods of microbial low activity (Figs. 1 and 3). While these bacterial and archaeal lineages are primarily known for their C1 and nitrogen metabolism (e.g., nitrogen fixation or nitrate reduction), presence of CAZymes for depolymerizing complex carbon polymers in some members of these clades have also been previously described [49–53]. Together, this suggests that under low microbial activity, communities in surface soils are supported by taxa such as *Methyloceanibacter* and *Pseudolabrys*, which encode diverse

CAZymes and central carbohydrate pathways alongside aspartate-, glutamate-, and

arginine-catabolic routes, and by *Nitrososphaera* with its capacity for ammonia oxidation,

enabling continued turnover of accessible complex SOM even when bulk respiration is low.

The widespread presence of *Rhizobiales*, *Chthoniobacterales*, and *Nitrososphaerales* across all

soils, combined with the abundance of CAZymes in the MAGs associated with genera within

these orders, suggests mechanisms for the depolymerization of SOM during both high and low

microbial activity in surface and subsurface soils (Fig. 5). Notably, the prevalence of GH and GT

highlights their ability to interact with glycans, monosaccharides, and polysaccharides, in

particular, and to construct extracellular structures that support adaptation to varying levels of

carbon availability [54,55]. Additionally, the detection of carbohydrate-binding modules (CBMs)

unique to genus-level lineages within *Rhizobiales*, such as *Methyloceanibacter* and VAZQ01,

further emphasizes the potential of these microbes to break down chemically-recalcitrant

compounds, particularly under low microbial activities.

Nitrosphaera in particular contained a distinct set of CAZymes, exhibiting greater diversity

compared to those found in other bacterial or archaeal MAGs (Fig. 5). While members of this

genus have traditionally been identified as ammonia-oxidizing chemolithotrophs, recent studies

indicate metabolic versatility within certain *Nitrosphaera* sub-clades [49]. In support of this

emerging idea, we detected both intracellular and extracellular CAZymes, such as CBMs and

GHs, alongside genes involved in ammonia oxidation (Table 1 and Supplementary Table 5).

Additional evidence for metabolic diversity in *Nitrosophaera* includes the presence of amino

acid catabolic pathways, particularly for L-aspartate and L-serine (Fig. 6). Given the prevalence

of *Nitrosophaera* in low-respiration surface soils, as well as in subsoils with limited respiration,

we hypothesize that these archaea are equipped with mechanisms utilize amino acids as a source

of nitrogen required to depolymerize complex carbon compounds, enabling them to persist under

energy- and carbon-limited conditions.

In *Nitrosphaera* and other genera within *Nitrosphaerales*, pathways involving amino acids

appear to be coupled with ammonia oxidation, given that none of the MAGs belonging to these

groups possessed the *hao* gene, which is thought to be an essential component of ammonia

oxidation [56–58]. The absence of *hao* may reflect sequencing limitations or suggest the

involvement of uncharacterized Cu-protein complexes in ammonia oxidation, a mechanism

observed in other ammonia-oxidizing archaea (Supplementary Table 9) [57]. Ammonia that is

oxidized by *Nitrososphaerales* could be generated from catabolism of available amino acids in

soil. These findings underscore the role of *Nitrososphaerales* in linking soil carbon and nitrogen

cycles and highlight the need for further investigation into their genetic machinery for ammonia

oxidation.

Amino acid and inorganic nitrogen acquisition mechanisms were also enriched in *Rhizobiales*

and *Chthoniobacterales*, as observed in flux balance analyses of the MAGs, indicating a vital

role of these pathways in nitrogen acquisition from SOM and energy generation in *Rhizobiales*

and *Chthoniobacterales* in particular (Fig. 6, Supplementary Fig. 2) [59,60]. In *Rhizobiales*- and

Chthoniobacterales-affiliated genera, particularly in *Methyloceanibacter* and *Udaeobacter*,

intermediates generated through amino acid catabolism (e.g., pyruvate, NADH, and FADH2)

could serve as sources of reducing power to drive dissimilatory nitrate reduction, which was also

detected in these lineages (Supplementary Table 5). Therefore, the degradation of amino acids

like L-arginine, observed in *Methyloceanibacter* and *Udaeobacter*, along with their prevalence in

soils with low microbial activity, indicates that amino acid catabolism is crucial for maintaining

energy balance and supporting nitrogen biosynthesis required for fundamental cellular functions

(Figs.1 and 6) [59,61–63].

The broad spectrum of CAZyme classes identified in *Rhizobiales*-affiliated genera versus other

MAGs highlights a putative advantage for them to break down chemically-recalcitrant SOM in

soils with high microbial activity. This is especially evident for *Bradyrhizobium*, *Pseudolabrys*,

and VAZQ01, which were prevalent in both surface and subsoils with elevated respiration rates

(Figs. 1 and 5). While the potential to decompose complex carbohydrates has been documented

for both *Bradyrhizobium* and *Pseudolabrys*, the metabolic characteristics of VAZQ01 remain

poorly understood [64,65]. Although the *Chthoniobacterales*-affiliated lineage *Udaeobacter* was

also present in both surface and subsoils with high respiration, the limited set of CAZymes

detected in its MAGs indicates a dependence on mono- and polysaccharides for carbon and

energy metabolism, further reinforcing its auxotrophy for various amino acids. (Figs. 5 and 6;

[66]). The amino acid catabolism observed in *Udaeobacter*, combined with its previously

identified ability to sustain the respiratory chain using atmospheric H_2 under nutrient-deficient

conditions, aligns with its prevalence in high-respiration subsoils, where carbon and nitrogen

sources are scarce and competition for these resources is intense (Fig. 1; [67,68]). These findings

suggest contrasting carbon usage strategies in *Rhizobiales* and *Chthoniobacterales*, with

Rhizobiales utilizing a broader range of carbohydrate polymers and amino acids as sources of

both carbon and nitrogen, whereas *Chthoniobacterales* relies on scavenging amino acids, likely

derived as byproducts of microbial activities in soil, highlighting how taxon-specific metabolic

traits can shape nutrient cycling and microbial interactions in terrestrial ecosystems.

Lastly, the bioavailability of soil lipids has been hotly debated, because they may dually serve as

a high-energy substrate and as a reservoir of sequestered carbon derived from microbial

necromass [69–71]. We found a diversity of lipid-like molecules in low-respiration subsoils that

may represent sequestered plant and/or microbial residues from prior periods of higher metabolic

activity [72,73] (Fig. 2). In subsoils, nutrient limitations and/or mineral protection likely promote

the disproportionate accumulation of lipid-like molecules, contributing significantly to the stable

fraction of SOM. While lipid persistence has been linked to soil carbon stabilization, the

universality of soil microbial potential for lipid metabolism — even in soils with low rates of

measured metabolic activity — suggest that they may become susceptible to decomposition as

environmental conditions change. Notably, interactions between lipid-like molecules and

microbial genes in low-respiration soils also indicate that lipid-like necromass components may

be vulnerable to microbial decomposition during periods of reduced carbon availability and

respiration rates [73,74] (Fig. 3). Thus, we highlight potential lipid turnover as a key

consideration in soil carbon dynamics [51].

Strong network correlations between lipid-like compounds and microbial secretion systems,

cofactors, and secondary metabolites in low-respiration subsoils may also elucidate strategies for

microbial adaptation to carbon-limited environments (Figs. 3 and 4; Supplementary Table 6).

Secretion systems enable microbial interactions by exporting compounds to facilitate nutrient

acquisition, particularly labile substrates [75,76], while cofactors (e.g., molybdopterin) and

secondary metabolites support energy metabolism, redox reactions, and oxidative damage

mitigation under stress [77,78]. The functional annotations of the MAGs linked to *Rhizobiales*, *Chthoniobacterales*, and *Nitrososphaerales* suggest that these microbes may possess the ability to enhance lipopolysaccharide (LPS) synthesis, thereby stabilizing their cell envelope and boosting their resistance to environmental stress [79,80]. These characteristics were notably prevalent in *Udaeobacter* and *Nitrososphaera* (**Supplementary Table 6**). Gram-negative bacteria, including *Udaeobacter* and *Nitrososphaera*, contribute significantly to organic carbon storage through the chemical persistence of LPS, particularly its chemically-recalcitrant lipid A component [81,82]. On cell death, released LPS acts as a binding agent between soil particles, enhancing soil stability and water retention [83]. These adaptations highlight how microbial communities persist and exploit resources under carbon deficiency while shaping key soil ecosystem processes like organic carbon cycling and storage (Figs. 1 and 2).

While we focus on bacterial and archaeal regulation of SOM cycling, our study also highlights the need for investigating abiotic processes and interactions with fungi and viruses, which are harder to disentangle using current metagenomic methods [18–21]. Assembling and annotating fungal metagenomes presents clear challenges, as evidenced by the smaller fraction of reads and the relatively low number of assembled contigs compared to those of bacterial or archaeal origin; however, the annotation of these contigs revealed the potential role of soil fungal communities in contributing to soil carbon flux. At the same time, the multi-omics integration strategies employed in this study have facilitated the recovery of bacterial and archaeal lineages with unique yet complementary carbon sequestration strategies. However, FBA-derived metabolic inferences assume microbes operate in a steady state, and the incomplete recovery of MAGs may limit interpretation, warranting validation of these inferred traits using transcriptomics and/or proteomics; in particular, validation through cultivation methods and emerging computational tools for co-assemblies [84,85] would enhance the robustness of the findings. The recovery of 828 MAGs—though indicative of substantial prokaryotic diversity—may be biased toward ubiquitous, abundant taxa favored by continental-scale sampling across geographic gradients, and site-specific specialists and rare taxa involved in specialized carbon transformations [86,87] are likely underrepresented, suggesting that the results primarily reflect the "core" microbial carbon cycling machinery rather than the full functional diversity present in soil communities.

Nonetheless, our study provides a comprehensive, genome-resolved view of microbial

decomposition potential across diverse soil environments, illuminating the metabolic

underpinnings of carbon turnover at continental scale. By integrating thousands of chemically

distinct SOM molecules with hundreds of metagenome-assembled genomes, we uncover

widespread microbial capacities to degrade carbon compounds long thought to be

chemically-resistant to decay. This work helps fill a critical knowledge gap in linking microbial

taxonomy and metabolism to specific carbon pools, offering a mechanistic foundation for

improving predictions of soil carbon dynamics under shifting environmental conditions. As

global change accelerates, incorporating these microbial processes into Earth system models will

be essential to more accurately project the fate of soil carbon reservoirs.

Methods

Data Collection

Through the 1000 Soils Pilot of the Molecular Observation Network (MONet) [14], 109 soil

samples were collected from 60 sites across the CONUS. Full methodological details are

available in refs. ([14] and [88]), and at https://github.com/EMSL-MONet/MONet-Protocols-.

With a goal to investigate the microbial pathways that underscore variation in microbial activity

across geography and soil depth profiles, we selected the subset of 47 soil samples, which

exhibited the highest and lowest (30%) rates of respiration within either surface or subsoils for

analysis. These soil samples spanned 37 geographical locations (Table 1) and were divided into

four groups based on depth and respiration: surface-high (n=14), surface-low (n=12), suboil-high

(n=11), and subsoil-low (n=10).

Fourier Transform Ion Cyclotron Resonance Mass Spectrometry (FTICR-MS)

Water-soluble SOM composition was analyzed by Fourier Transform Ion Cyclotron Resonance

Mass Spectrometry (FTICR-MS) [88,89]. Briefly, 6 g of dried soil in triplicate were resuspended

in 30 mL DI water, and shaken for 2 hours at 800 rpm. The samples were then centrifuged at

6,000 rpm for 8 minutes and 5 mL of supernatant was transferred for solid phase extraction

(SPE) with Agilent Bond Elut PPL cartridges [89] on Gilson ASPEC® SPE system. We used a

Bruker 7 Tesla scimaX FTICR-MS at the Environmental Molecular Sciences Laboratory

(EMSL) in Richland, WA to analyze the extracted SOM, with a negative ionization mode and ion

accumulation time of 0.01 or 0.025 seconds. The mass accuracy was below 1 ppm, as confirmed

by the Suwannee River Fulvic Acid (SRFA) samples. We used CoreMS (v. 2.0.0) [90] to process

the raw FTICR-MS data, with noise thresholding of 10 (*noise_threshold_log_nsigma*) and

minimum peak prominence of 0.1 (*peak_min_prominence_percent*). Mass calibration was

performed by setting a threshold that uses the most calibration points (usually >200). Molecular

formulae were annotated by both accurate mass and isotopologues, with a confidence score

calculated for each formula. We then filtered the assigned peaks by *m/z* between 200 to 1,000,

present in at least 2 out of 3 replicates, not present in two or more lab blanks, and with formulae

confidence scores (combines *m/z* error and isotopic pattern) above 0.5 [91].

Distribution and uniqueness of SOMs across the soil samples

Using the customized Python script, we assigned the 66,727 FTICR-derived molecular formulas

into SOM compound classes, based on thresholds of O:C and H:C ratios on the van Krevelen

diagram [92]. The matrix of the SOM assignment and the presence/absence matrix of the

molecules across the 47 samples were used in a statistical analysis using the R package

ftmsRanalysis [93] and associated scripts, to first determine the distribution of SOM types across

the four soil sample types, and then to conduct a series of pairwise G-tests of uniqueness [94]

with unadjusted p-value threshold of 0.05, generating profiles of SOMs that are differentially

present in one soil type versus the other. The results of the G-tests were plotted as shown in Fig.

2.

DNA extraction and sequencing

563 DNA was extracted using *Quick*-DNA Fecal/Soil Microbe Miniprep Kit D6010 (ZymoResearch)

564 and then cleaned and concentrated with DNA Clean and Concentrator-25 (ZymoResearch).

565 Samples were sequenced at Azenta Life Sciences using the Illumina Whole Genome

566 Metagenomics pipeline or at the Joint Genome Institute, following identical protocols.

Quality control, assembly and binning of the metagenome reads

Metagenome construction, including quality-control/read filtering, contig assembly and binning

processes, were performed using a suite of tools available in the MetaGEM pipeline [95].

Specifically, short read quality filtering and adapter trimming were applied to the set of reads

(de-interleaved) using fastp v0.20.0 [96] with default settings. The filtered reads from each

sample were assembled into contigs using MEGAHIT v1.2.9 [97], with *--presets meta-sensitive*

and *--min-contig-len 1000* flags.

Before the binning process, contig coverages across the samples (required by some of the

binning tools as described below) were generated by cross-mapping the filtered short reads to the

set of assemblies. This was accomplished by generating an index for each contig using the 'bwa

index' command with default settings. The set of indices were used as part of the input for

bwa-mem v0.7.17 [98], generating SAM files, which were subsequently converted to BAM

format using the 'samtools view' command with the *-Sb* flag, and then were sorted using the

'samtools sort' command with the default settings. The sorted BAM files were then used to

generate the contig coverages for the binning tools, MetaBAT2 v.2.12.1 [99] and MaxBin2

v2.2.5 [100], using the script, 'jgi_summarize_bam_contig_depths' available as part of the

MetaBAT2 package. The contig coverage files were then used as input data for the MetaBAT2

and MaxBin2, with default settings for each tool.

Binning was performed using the tool CONCOCT v1.1.0 [101] and the scripts available in the

MetaGEM pipeline. The coverage files were generated using the script,

'concoct_coverage_table.py' (default settings), with a set of BED files generated for each contig

as input. The BED files were created using the 'cut_up_fasta.py' script, with the parameters *-c*

10000 -o 0 -m b that split the contigs into 10 kb fragments. For each fragment, CONCOCT was

run using the coverage file as input and with the *-c 800* parameter, yielding a set of bins for each

sample, based on the split contigs. The script 'extract_fasta_bins.py' was then run to generate

bins using the original unfragmented contigs.

Refinement and reassembly of the MAGs

To extract the best version of bins from the outputs of the three binning tools, the

'bin_refinement' script available in metaWRAP v1.2.3 [102] was used with *-x 10 -c 50*

parameters. This command de-replicates the results of the three binning methods by selecting the

bin with the highest CheckM2 [103] completeness and the lowest contamination, with the

completeness taking precedence over contamination. The parameters in this script ensure that

only the binning results with contamination and completeness thresholds of $\leq 10\%$ and $\geq 50\%$,

respectively, are considered for the bin refinement process.

To improve the quality of the bins selected from the refinement step, the 'reassemble_bins'

command (with the same parameters used for the 'bin_refinement' script) was used. This script

extracts the reads that map to the contigs belonging to the refined bins and reassembles them

using the SPAdes assembly. This process is done at both a 'strict' (no mismatches) or

'permissive' (< 5 mismatches) level. CheckM2 completeness and contamination between the

three versions of bins (i.e., strict, permissive or original derived from the refinement) are

compared, and the bins with the best score are selected as the finalized MAGs, yielding 344

genomes.

To determine the distribution of the bacterial and archaeal communities that were represented by

the MAGs, we first used the 'pipe' subcommand in SingleM [104], generating a summary of

operational taxonomic units (OTUs) detected based on 35 bacterial and 37 single-copy marker

genes identified from MAGs. The OTU summary table was used as input for the 'appraise'

function in SingleM, employing the parameters *--imperfect* and *--sequence-identity 0.87* to

account for genomes that are similar to those found in the metagenome that were classified at the

genus level (i.e. average nucleotide identity of 87%) [105].

According to the SingleM appraise result, the 344 MAGs represented only 44.4% and 32.3% of

the bacterial and archaea diversity. Thus, we made an attempt to improve the MAG recovery

through coassembly of the reads, as implemented in Bin Chicken [84]. Specifically, we used this

tool to identify MAGs through co-assembly of the reads from the soils associated with each of

the four depth- and respiration-level classification. We first executed Bin Chicken with the single

subcommand to generate MAGs from each sample, using forward and reverse reads and

SingleM-generated files (e.g., protein-coding genes and two OTU tables—one from contigs and

one from MAGs) as input, with the parameters *--run-aviary* and *--assembly-strategy* megahit..

The output MAGs from the 'single' subcommand were used to execute two rounds of iterative

coassemblies via the 'iterate' subcommand. For each run of the 'iterate' function, we used the

same set of parameters as we did with the 'single' subcommand. By the end of the second

execution of 'iterate,' the total number of MAGs increased to 828, improving the SingleM

appraisal value to 50.6% and 36.5% for bacteria and archaea. Dereplication of the MAGs and

calculating genome abundances were performed using the methods described below.

Genome dereplication

639 The 828 bins that were generated were dereplicated using galah (v0.4.0) [106] at 95% average

640 nucleotide identity (ANI) similarity, which can be used to represent species-level similarities

641 [107], resulting in 358 clusters. Representative, dereplicated MAGs from each cluster were

642 selected based on the quality metric derived from CheckM2.

Genome abundances

645 The abundance of the dereplicated MAGs in each of the samples was calculated via mapping of

646 the filtered reads to the dereplicated MAG dataset using CoverM [108] with the following flags:

647 *-p minimap2-sr -m trimmed_mean --min-read-percent-identity 0.80 --min-read-aligned-percent*

648 *0.75 --min-covered-fraction 20*. Read mapping indicated that the MAGs accounted for 1.3% to

649 22.2% of the reads of each sample.

Phylogenetic assignments of dereplicated MAGs

The dereplicated MAGs generated using the methods described above were taxonomically

classified using the 'identify', 'align' and 'classify' commands available in GTDB-Tk v2.5.2

[109]. This set of commands identifies 120 ubiquitous bacterial marker genes (and 53 archaeal

genes) from the input MAGs, which are then concatenated and aligned together. The alignment

of the concatenated marker genes is then used as input for Pplacer [110] to assign each MAG

into the optimal loci in the bacterial or archaeal tree of life. The output of the process includes

tabular files containing taxonomic classifications of the MAGs. For each of the depth and

respiration-level classifications, the average abundances of the taxonomic groups were calculated

from the total abundances of the dereplicated MAGs assigned to each lineage.

Assembly and analysis of fungal metagenomes

We used Kraken2 [22] to recruit the DNA reads against the fungal taxonomic marker genes

present in the RefSeq Fungi [111] and the 18S SSU rRNA sequences from the SILVA database

[112]. On average, 1.33×10^5 fungal reads were recovered, making up < 1% (i.e., 0.24%) of the

total reads from each soil. The reads were assembled using MEGAHIT v1.2.9 with parameters

--presets meta-sensitive --min-contig-len 1000, generating an average of 86 contigs per sample.

The presence of the carbohydrate-active enzymes (CAZymes) in the fungal contigs was assessed

by searching them against the Carbohydrate Active Enzymes (CAZy) database [29], using

dbCAN2 [113].

672 KEGG pathway analysis of FTICR-derived molecules using KO identifications

The ftmsRanalysis package and associated custom scripts mentioned above were also used to

generate a summary of KO identifications that were mapped to the FTICR-derived molecules.

These summary tables were parsed to extract the KOs present in 50% or more of the samples

belonging to each partition. The KEGG pathway representation analysis, which employs the

enrichKEGG algorithm in the clusterProfiler R package [114], was conducted using this set of

KOs. This method calculates the ratio of KOs affiliated with specific pathways versus the total

number of KOs in each sample set. This analysis generated a summary visualization of

significant KEGG pathways present in each layer/respiration partition (Fig. 4).

682 Network analysis of the relationship between KO derived from functional annotation of

683 MAGs and FTICR-MS derived molecules

For each of the four layer/respiration partitions, co-occurrence networks were constructed using

Spearman's correlation to explore the relationship between the KOs detected from the MAGs

and the FTICR-MS-derived compounds. To generate the KO abundance profile, functional genes

were first detected from the nucleotide sequences of all dereplicated MAGs using Prokka [115],

and were searched against KEGG Orthology (KO) [116] using the 'annotate' function

implemented in EnrichM [117]. For the bacterial MAGs, Prokka was executed with *--kingdom*

Bacteria and *--addgenes* parameters, while for the archaeal MAGs, the *--kingdom Archaea* flag

was used instead. The abundance tables of KO and the compounds were used as input for

Spearman correlation implemented in R. Each correlation matrix was filtered to obtain

relationships with *rho* > 0.6 and FDR-corrected P < 0.01 [118]. The output matrices were then

imported into Cytoscape [119] version 3.10.3. Clusters in each partition were identified using

the MCODE application, with default parameters, producing a set of clusters with scores

representing their densities and interconnectedness [120].

Assessment of microbial carbohydrate depolymerization in abundant prokaryotes using

CAZy annotation of MAGs

We selected two order-level bacterial lineages, *Rhizobiales* and *Chthoniobacterales*, each of

which is represented by > 5.0% of the 319 bacterial MAGs recovered from all soil samples.

Only one order-level archaeal lineage (i.e., *Nitrososphaerales*) was selected, as > 90% of the 39

MAGs were assigned to this group. Functional genes detected from MAGs of these bacterial and

archaeal clades using Prokka (see above) were searched against the Carbohydrate-Active

Enzymes (CAZy) database [29], using dbCAN2 [113] to predict the capacities of these lineages

to depolymerize lignin and other chemically recalcitrant carbohydrates. .

Metabolic model prediction for *Rhizobiales*, *Chthoniobacterales* and *Nitrososphaerales*

Using the functional annotation profiles generated from the method described in the previous

section, we inferred metabolic model predictions for the two dominant bacterial orders,

Rhizobiales and *Chthoniobacterales* and the archaeal order *Nitrosophaerales*. To accomplish

this, we built a metabolic prediction narrative using a selected suite of applications available in

the DOE Systems Biology Knowledgebase (KBase) [121] platform. In this narrative, GFF tables

generated using Prokka containing genomic feature information (i.e., gene names, enzyme

commission identifications and protein names; see above section) for each detected coding

region and the nucleotide FASTA files of the MAGs are used to re-annotate the genomes using

RAST-Tk v1.73 (Rapid Annotation using Subsystems Technology - Toolkit) [122]. This process

generates the output files compatible with the KBase metabolic modeling applications.

The RAST-Tk annotation objects were used as input for the ModelSEED [123] application,

generating models representing the broader metabolic properties in the bacterial and archaeal

lineages under investigation. The models were generated by compiling the i metabolic pathways

information encoded by genes detected in the MAGs. The output generated by ModelSEED

includes a matrix of metabolic reactions, their associated biochemical compounds, and the

gene-protein-reaction (GPR) that encompass the dependencies between the responses and the

related genes. The GPR can help differentiate between cases where a reaction is catalyzed by a

protein complex encoded by multiple genes and where several individual proteins independently

orchestrate a reaction. The ModelSEED output also includes a 'biomass objective function' file

that indicates the abundance of the metabolites generated by the microbes (e.g., amino acids,

lipids and quinones) as bi-products of their growth metabolism.

Finally, the Flux Balance Analysis (FBA) [124] application was used to read the

ModelSEED-derived metabolic models and a default media file provided by KBase that

specified the set of > 500 compounds an organism can use for its growth, generating tables of

metrics that allowed estimation of growth rates of microbes and/or output of specific metabolites

on a specified media. The FBA estimates the growth rate of an organism by calculating

growth-optimal fluxes through all reactions in the metabolic network. The uptake values for the

compounds specified in the output file, measured in mmol per gram cell dry weight per hour,

indicate whether the microbes use the metabolites for growth or excrete them as a by-product.

The list of the output uptake values was examined to determine the spectrum of compounds

likely to be metabolized by the members of the prevalent order-level clades detected in the soils.

Data availability

The sequencing data associated with this study are available in the NCBI Sequence Read

Archive (SRA) under accession number **PRJNA1260013**. Data in the MONet open science

database can be found at https://sc-data.emsl.pnnl.gov/monet and MONet Zenodo DOI (https://

doi.org/10.5281/zenodo.7406532).

Code availability

The R and Python code used for analyzing the FTICR-MS data and for preparing the main and

supplementary figures are available upon request at the following GitHub repository:

https://github.com/EMSL-MONet/The-1000-Soil-ICR-Metagenome. The KBase narrative

developed to analyze the *Rhizobiales* and *Chthoniobacterales* MAGs can be accessed at

https://narrative.kbase.us/narrative/186750.

Acknowledgements

This research was performed on a project award DOI https://doi.org/10.46936/intm.proj.2021.60141/60000423 from the Environmental Molecular Sciences Laboratory (EMSL), a DOE Office of Science user facility sponsored by the Biological and Environmental Research program under Contract No. DE-AC05-76RL01830. Soil data was provided by the Molecular Observation Network (MONet) at the EMSL (DOI: https://doi.org/10.46936/10.25585/60008970). MONet is conducted in collaboration with Joint Genome Institute (https://ror.org/04xm1d337), a DOE Office of Science user facility supported by the Office of Science of the U.S. Department of Energy operated under Contract DE-AC02-05CH11231. The MONet database is an open, FAIR, and publicly available compilation of the molecular and microstructural properties of soil. The National Ecological Observatory Network is a program sponsored by the U.S. National Science Foundation and operated under cooperative agreement by Battelle. A subset of samples used in this research were obtained through NEON Research Support Services.

Competing interests

The authors declare no competing interests.

References

1. Saifuddin, M., Bhatnagar, J. M., Segrè, D. & Finzi, A. C. Microbial carbon use efficiency predicted from genome-scale metabolic models. *Nat Commun* **10**, 3568 (2019).

2. Bhattacharyya, S. S., Ros, G. H., Furtak, K., Iqbal, H. M. N. & Parra-Saldívar, R. Soil carbon sequestration – An interplay between soil microbial community and soil organic matter dynamics. *Science of The Total Environment* **815**, 152928 (2022).

3. Kleber, M. What is recalcitrant soil organic matter? *Environ. Chem.* **7**, 320–332 (2010).

4. Marschner, B. *et al.* How relevant is recalcitrance for the stabilization of organic matter in soils? *Journal of Plant Nutrition and Soil Science* **171**, 91–110 (2008).

5. Song, Y. *et al.* Soil metabolomics: Deciphering underground metabolic webs in terrestrial ecosystems. *Eco-Environment & Health* **3**, 227–237 (2024).

6. Swenson, T. L., Bowen, B. P., Nico, P. S. & Northen, T. R. Competitive sorption of microbial metabolites on an iron oxide mineral. *Soil Biology and Biochemistry* **90**, 34–41 (2015).

7. Feng, X. & Simpson, M. J. Temperature and substrate controls on microbial phospholipid fatty acid composition during incubation of grassland soils contrasting in organic matter quality. *Soil Biology and Biochemistry* **41**, 804–812 (2009).

8. Rillig, M. C., Caldwell, B. A., Wösten, H. A. B. & Sollins, P. Role of proteins in soil carbon and nitrogen storage: controls on persistence. *Biogeochemistry* **85**, 25–44 (2007).

9. Brown, R. W. *et al.* Soil metabolomics - current challenges and future perspectives. *Soil Biology and Biochemistry* **193**, 109382 (2024).

10. Roth, H. K. *et al.* Effects of burn severity on organic nitrogen and carbon chemistry in high-elevation forest soils. *Soil & Environmental Health* **1**, 100023 (2023).

11. Nelson, A. R. *et al.* Wildfire-dependent changes in soil microbiome diversity and function. *Nat Microbiol* **7**, 1419–1430 (2022).

12. Lewis, W. H., Tahon, G., Geesink, P., Sousa, D. Z. & Ettema, T. J. G. Innovations to culturing the uncultured microbial majority. *Nat Rev Microbiol* **19**, 225–240 (2021).

13. Luo, Y. *et al.* Toward more realistic projections of soil carbon dynamics by Earth system models. *Global Biogeochemical Cycles* **30**, 40–56 (2016).

14. Bowman, M. M. *et al.* One thousand soils for molecular understanding of belowground carbon cycling. *Front. Soil Sci.* **3**, (2023).

15. Thibault, K. M., Laney, C. M., Yule, K. M., Franz, N. M. & Mabee, P. M. The US National

Ecological Observatory Network and the Global Biodiversity Framework: national research infrastructure with a global reach. *Journal of Ecology and Environment* **47**, (2023).

16. Hu, J. *et al.* A Continental-Scale Estimate of Soil Organic Carbon Change at NEON Sites and Their Environmental and Edaphic Controls. *Journal of Geophysical Research: Biogeosciences* **128**, e2022JG006981 (2023).

17. Weiglein, T. L. *et al.* Key predictors of soil organic matter vulnerability to mineralization differ with depth at a continental scale. *Biogeochemistry* **157**, 87–107 (2022).

18. Avershina, E., Qureshi, A. I., Winther-Larsen, H. C. & Rounge, T. B. Challenges in capturing the mycobiome from shotgun metagenome data: lack of software and databases. *Microbiome* **13**, 66 (2025).

19. Graham, E. B. *et al.* A global atlas of soil viruses reveals unexplored biodiversity and potential biogeochemical impacts. *Nat Microbiol* **9**, 1873–1883 (2024).

20. Zimmerman, A. E., Graham, E. B., McDermott, J. & Hofmockel, K. S. Estimating the Importance of Viral Contributions to Soil Carbon Dynamics. *Global Change Biology* **30**, e17524 (2024).

21. Penton, C. R., Gupta, V. V. S. R., Yu, J. & Tiedje, J. M. Size Matters: Assessing Optimum Soil Sample Size for Fungal and Bacterial Community Structure Analyses Using High Throughput Sequencing of rRNA Gene Amplicons. *Front. Microbiol.* **7**, (2016).

22. Wood, D. E., Lu, J. & Langmead, B. Improved metagenomic analysis with Kraken 2. *Genome Biol* **20**, 257 (2019).

23. Fierer, N. Embracing the unknown: disentangling the complexities of the soil microbiome. *Nat Rev Microbiol* **15**, 579–590 (2017).

24. Bowers, R. M. *et al.* Minimum information about a single amplified genome (MISAG) and a metagenome-assembled genome (MIMAG) of bacteria and archaea. *Nat Biotechnol* **35**, 725–731 (2017).

25. Yoshida, M. *et al.* Biochemical and Genetic Analysis of the γ-Resorcylate (2,6-Dihydroxybenzoate) Catabolic Pathway in Rhizobium sp. Strain MTP-10005: Identification and Functional Analysis of Its Gene Cluster. *Journal of Bacteriology* **189**, 1573–1581 (2007).

26. Sonoki, T. *et al.* Coexistence of Two Different O Demethylation Systems in Lignin Metabolism by Sphingomonas paucimobilis SYK-6: Cloning and Sequencing of the Lignin

Biphenyl-Specific O-Demethylase (LigX) Gene. *Applied and Environmental Microbiology* **66**, 2125–2132 (2000).

27. Rocci, K. S. *et al.* Bridging 20 Years of Soil Organic Matter Frameworks: Empirical Support, Model Representation, and Next Steps. *Journal of Geophysical Research: Biogeosciences* **129**, e2023JG007964 (2024).

28. Lou, X. *et al.* The Biodegradation of Soil Organic Matter in Soil-Dwelling Humivorous Fauna. *Front. Bioeng. Biotechnol.* **9**, (2022).

29. Cantarel, B. L. *et al.* The Carbohydrate-Active EnZymes database (CAZy): an expert resource for Glycogenomics. *Nucleic Acids Research* **37**, D233–D238 (2009).

30. Skouloubris, S., Labigne, A. & De Reuse, H. The AmiE aliphatic amidase and AmiF formamidase of Helicobacter pylori: natural evolution of two enzyme paralogues. *Molecular Microbiology* **40**, 596–609 (2001).

31. Nagy, L. *et al.* The Effect of Mutation in Lipopolysaccharide Biosynthesis on Bacterial Fitness. *Cells* **11**, 3249 (2022).

32. Clark, M. M., Paxhia, M. D., Young, J. M., Manzella, M. P. & Reguera, G. Adaptive Synthesis of a Rough Lipopolysaccharide in Geobacter sulfurreducens for Metal Reduction and Detoxification. *Applied and Environmental Microbiology* https://doi.org/10.1128/AEM.00964-21 (2021) doi:10.1128/AEM.00964-21.

33. McGivern, B. B. *et al.* Decrypting bacterial polyphenol metabolism in an anoxic wetland soil. *Nat Commun* **12**, 2466 (2021).

34. Graham, E. B. & Hofmockel, K. S. Ecological stoichiometry as a foundation for omics-enabled biogeochemical models of soil organic matter decomposition. *Biogeochemistry* **157**, 31–50 (2022).

35. Waring, B. G. *et al.* From pools to flow: The PROMISE framework for new insights on soil carbon cycling in a changing world. *Global Change Biology* **26**, 6631–6643 (2020).

36. Janzen, H. H. Beyond carbon sequestration: soil as conduit of solar energy. *European Journal of Soil Science* **66**, 19–32 (2015).

37. Jastrow, J. D., Amonette, J. E. & Bailey, V. L. Mechanisms controlling soil carbon turnover and their potential application for enhancing carbon sequestration. *Climatic Change* **80**, 5–23 (2007).

38. Hansen, P. M. *et al.* Distinct, direct and climate-mediated environmental controls on global

particulate and mineral-associated organic carbon storage. *Global Change Biology* **30**,
e17080 (2024).

39. Tang, B., Rocci, K. S., Lehmann, A. & Rillig, M. C. Nitrogen increases soil organic carbon
accrual and alters its functionality. *Global Change Biology* **29**, 1971–1983 (2023).

40. Thevenot, M., Dignac, M.-F. & Rumpel, C. Fate of lignins in soils: A review. *Soil Biology*
and Biochemistry **42**, 1200–1211 (2010).

41. Wang, X. *et al.* Microbial keystone taxa drive succession of plant residue chemistry. *The*
ISME Journal **17**, 748–757 (2023).

42. Brown, M. E. & Chang, M. C. Exploring bacterial lignin degradation. *Current Opinion in*
Chemical Biology **19**, 1–7 (2014).

43. Qin, S., Zhang, D., Wei, B. & Yang, Y. Dual roles of microbes in mediating soil carbon
dynamics in response to warming. *Nat Commun* **15**, 6439 (2024).

44. Nottingham, A. T., Gloor, E., Bååth, E. & Meir, P. Soil carbon and microbes in the warming
tropics. *Functional Ecology* **36**, 1338–1354 (2022).

45. Peng, X. *et al.* Characterization of the 5-Carboxyvanillate Decarboxylase Gene and Its Role
in Lignin-Related Biphenyl Catabolism in Sphingomonas paucimobilis SYK-6. *Applied and*
Environmental Microbiology **68**, 4407–4415 (2002).

46. Anderson, C. G. *et al.* Seasonal Controls on Microbial Depolymerization and Oxidation of
Organic Matter in Floodplain Soils. *Environ. Sci. Technol.* **58**, 16815–16823 (2024).

47. Hall, S. J. & Silver, W. L. Iron oxidation stimulates organic matter decomposition in humid
tropical forest soils. *Global Change Biology* **19**, 2804–2813 (2013).

48. Yadav, S., Tripathi, S., Purchase, D. & Chandra, R. Development of a biofilm-forming
bacterial consortium and quorum sensing molecules for the degradation of lignin-containing
organic pollutants. *Environmental Research* **226**, 115618 (2023).

49. Bei, Q. *et al.* Metabolic potential of Nitrososphaera-associated clades. *ISME J* **18**, wrae086
(2024).

50. López-Sánchez, R. *et al.* Metagenomic analysis of carbohydrate-active enzymes and their
contribution to marine sediment biodiversity. *World J Microbiol Biotechnol* **40**, 95 (2024).

51. Albright, M. B. N. *et al.* Differences in substrate use linked to divergent carbon flow during
litter decomposition. *FEMS Microbiol Ecol* **96**, fiaa135 (2020).

52. Farhan Ul Haque, M., Xu, H.-J., Murrell, J. C. & Crombie, A. Facultative methanotrophs –

diversity, genetics, molecular ecology and biotechnological potential: a mini-review.

Microbiology **166**, 894–908 (2020).

53. Kämpfer, P. *et al.* Pseudolabrys taiwanensis gen. nov., sp. nov., an alphaproteobacterium

isolated from soil. *International Journal of Systematic and Evolutionary Microbiology* **56**,

2469–2472 (2006).

54. Berlemont, R. & Martiny, A. C. Glycoside Hydrolases across Environmental Microbial

Communities. *PLOS Computational Biology* **12**, e1005300 (2016).

55. Henrissat, B. & Coutinho, P. M. Classification of glycoside hydrolases and

glycosyltransferases from hyperthermophiles. in *Methods in Enzymology* vol. 330 183–201

(Academic Press, 2001).

56. Kerou, M. *et al.* Proteomics and comparative genomics of Nitrososphaera viennensis reveal

the core genome and adaptations of archaeal ammonia oxidizers. *Proceedings of the National*

Academy of Sciences **113**, E7937–E7946 (2016).

57. Simon, J. & Klotz, M. G. Diversity and evolution of bioenergetic systems involved in

microbial nitrogen compound transformations. *Biochimica et Biophysica Acta (BBA) -*

Bioenergetics **1827**, 114–135 (2013).

58. Ouverney, C. C. & Fuhrman, J. A. Marine Planktonic Archaea Take Up Amino Acids.

Applied and Environmental Microbiology **66**, 4829–4833 (2000).

59. Ramoneda, J., Jensen, T. B. N., Price, M. N., Casamayor, E. O. & Fierer, N. Taxonomic and

environmental distribution of bacterial amino acid auxotrophies. *Nat Commun* **14**, 7608

(2023).

60. Wheatley, R. M. *et al.* Lifestyle adaptations of Rhizobium from rhizosphere to symbiosis.

Proceedings of the National Academy of Sciences **117**, 23823–23834 (2020).

61. Buckel, W. Energy Conservation in Fermentations of Anaerobic Bacteria. *Front. Microbiol.*

12, (2021).

62. Macias-Benitez, S. *et al.* Rhizospheric Organic Acids as Biostimulants: Monitoring

Feedbacks on Soil Microorganisms and Biochemical Properties. *Front. Plant Sci.* **11**, (2020).

63. Vekeman, B. *et al.* New Methyloceanibacter diversity from North Sea sediments includes

methanotroph containing solely the soluble methane monooxygenase. *Environmental*

Microbiology **18**, 4523–4536 (2016).

64. Ding, J. *et al.* Depth heterogeneity of lignin-degrading microbiome and organic carbon

processing in mangrove sediments. *npj Biofilms Microbiomes* **11**, 5 (2025).

65. VanInsberghe, D. *et al.* Non-symbiotic Bradyrhizobium ecotypes dominate North American

forest soils. *ISME J* **9**, 2435–2441 (2015).

66. Willms, I. M. *et al.* Globally Abundant "Candidatus Udaeobacter" Benefits from Release of

Antibiotics in Soil and Potentially Performs Trace Gas Scavenging. *mSphere* **5**,

10.1128/msphere.00186-20 (2020).

67. Button, E. S. *et al.* Deep-C storage: Biological, chemical and physical strategies to enhance

carbon stocks in agricultural subsoils. *Soil Biology and Biochemistry* **170**, 108697 (2022).

68. Jones, D. L. *et al.* Microbial competition for nitrogen and carbon is as intense in the subsoil

as in the topsoil. *Soil Biology and Biochemistry* **117**, 72–82 (2018).

69. Gorka, S., Canarini, A., Schmidt, H. & Kaiser, C. Soil bacterial neutral lipid fatty acids:

Markers for carbon storage or necromass? 2024.12.02.626346 Preprint at

https://doi.org/10.1101/2024.12.02.626346 (2024).

70. Kallenbach, C. M., Grandy, A. S., Frey, S. D. & Diefendorf, A. F. Microbial physiology and

necromass regulate agricultural soil carbon accumulation. *Soil Biology and Biochemistry* **91**,

279–290 (2015).

71. Dinel, H., Schnitzer, M. & Mehuys, G. R. Soil Lipids: Origin, Nature, Content,

Decomposition, and Effect on Soil Physical Properties. in *Soil Biochemistry* (Routledge,

1990).

72. Dai, G. *et al.* Plant-derived lipids play a crucial role in forest soil carbon accumulation. *Soil*

Biology and Biochemistry **168**, 108645 (2022).

73. Zhao, F. *et al.* Microbial functional genes driving the positive priming effect in forest soils

along an elevation gradient. *Soil Biology and Biochemistry* **165**, 108498 (2022).

74. Kalbitz, K. *et al.* Changes in properties of soil-derived dissolved organic matter induced by

biodegradation. *Soil Biology and Biochemistry* **35**, 1129–1142 (2003).

75. Sun, Y., Zhang, T., Lu, B., Li, X. & Jiang, L. Application of cofactors in the regulation of

microbial metabolism: A state of the art review. *Front. Microbiol.* **14**, (2023).

76. Santamaria, G. *et al.* Evolution and regulation of microbial secondary metabolism. *eLife* **11**,

e76119 (2022).

77. Kutschera, A. & Ranf, S. The multifaceted functions of lipopolysaccharide in plant-bacteria

interactions. *Biochimie* **159**, 93–98 (2019).

78. Zhang, G., Meredith, T. C. & Kahne, D. On the essentiality of lipopolysaccharide to Gram-negative bacteria. *Current Opinion in Microbiology* **16**, 779–785 (2013).

79. Kästner, M., Miltner, A., Thiele-Bruhn, S. & Liang, C. Microbial Necromass in Soils—Linking Microbes to Soil Processes and Carbon Turnover. *Front. Environ. Sci.* **9**, (2021).

80. Bradley, J. A., Amend, J. P. & LaRowe, D. E. Necromass as a Limited Source of Energy for Microorganisms in Marine Sediments. *Journal of Geophysical Research: Biogeosciences* **123**, 577–590 (2018).

81. Choma, A., Komaniecka, I. & Zebracki, K. Structure, biosynthesis and function of unusual lipids A from nodule-inducing and N2-fixing bacteria. *Biochimica et Biophysica Acta (BBA) - Molecular and Cell Biology of Lipids* **1862**, 196–209 (2017).

82. Cania, B. *et al.* Site-Specific Conditions Change the Response of Bacterial Producers of Soil Structure-Stabilizing Agents Such as Exopolysaccharides and Lipopolysaccharides to Tillage Intensity. *Front. Microbiol.* **11**, (2020).

83. Henao, L. J. & Mazeau, K. Molecular modelling studies of clay–exopolysaccharide complexes: Soil aggregation and water retention phenomena. *Materials Science and Engineering: C* **29**, 2326–2332 (2009).

84. Aroney, S. T. N., Newell, R. J. P., Tyson, G. W. & Woodcroft, B. J. Bin Chicken: targeted metagenomic coassembly for the efficient recovery of novel genomes. *Nat Methods* **22**, 2516–2524 (2025).

85. Woodcroft, B. J. *et al.* Genome-centric view of carbon processing in thawing permafrost. *Nature* **560**, 49–54 (2018).

86. Bickel, S. & Or, D. The chosen few—variations in common and rare soil bacteria across biomes. *The ISME Journal* **15**, 3315–3325 (2021).

87. Kaminsky, R. & Morales, S. E. Conditionally Rare Taxa Contribute but Do Not Account for Changes in Soil Prokaryotic Community Structure. *Front. Microbiol.* **9**, (2018).

88. Shi, C. *et al.* Scaling High-Resolution Soil Organic Matter Composition to Improve Predictions of Potential Soil Respiration Across the Continental United States. *Geophysical Research Letters* **52**, e2024GL113091 (2025).

89. Dittmar, T., Koch, B., Hertkorn, N. & Kattner, G. A simple and efficient method for the solid-phase extraction of dissolved organic matter (SPE-DOM) from seawater. *Limnology*

and Oceanography: Methods **6**, 230–235 (2008).

90. Corilo, Y. E., Kew, W. R., McCue, L. A., Heal, K. & Carr, J. C. EMSL-Computing/CoreMS:

CoreMS 3.0.0. Zenodo https://doi.org/10.5281/zenodo.14009575 (2024).

91. Corilo, Y. E., Kew, W. R., McCue, L. A., Heal, K. & Carr, J. C. EMSL-Computing/CoreMS:

CoreMS 3.0.0. Zenodo https://doi.org/10.5281/zenodo.14009575 (2024).

92. Kim, S., Kramer, R. W. & Hatcher, P. G. Graphical Method for Analysis of

Ultrahigh-Resolution Broadband Mass Spectra of Natural Organic Matter, the Van Krevelen

Diagram. *Anal. Chem.* **75**, 5336–5344 (2003).

93. Degnan, D. J. *et al.* FREDA: A Web Application for the Processing, Analysis, and

Visualization of Fourier-Transform Mass Spectrometry Data. *Rapid Communications in*

Mass Spectrometry **39**, e9980 (2025).

94. G–test of goodness-of-fit - Handbook of Biological Statistics.

https://www.biostathandbook.com/gtestgof.html.

95. Zorrilla, F., Buric, F., Patil, K. R. & Zelezniak, A. metaGEM: reconstruction of genome scale

metabolic models directly from metagenomes. *Nucleic Acids Research* **49**, e126 (2021).

96. Chen, S., Zhou, Y., Chen, Y. & Gu, J. fastp: an ultra-fast all-in-one FASTQ preprocessor.

Bioinformatics **34**, i884–i890 (2018).

97. Li, D., Liu, C.-M., Luo, R., Sadakane, K. & Lam, T.-W. MEGAHIT: an ultra-fast single-node

solution for large and complex metagenomics assembly via succinct de Bruijn graph.

Bioinformatics **31**, 1674–1676 (2015).

98. Li, H. Toward better understanding of artifacts in variant calling from high-coverage

samples. *Bioinformatics* **30**, 2843–2851 (2014).

99. Kang, D. D. *et al.* MetaBAT 2: an adaptive binning algorithm for robust and efficient

genome reconstruction from metagenome assemblies. *PeerJ* **7**, e7359 (2019).

100. Wu, Y.-W., Simmons, B. A. & Singer, S. W. MaxBin 2.0: an automated binning algorithm to

recover genomes from multiple metagenomic datasets. *Bioinformatics* **32**, 605–607 (2016).

101. Alneberg, J. *et al.* Binning metagenomic contigs by coverage and composition. *Nat Methods*

11, 1144–1146 (2014).

102. Uritskiy, G. V., DiRuggiero, J. & Taylor, J. MetaWRAP—a flexible pipeline for

genome-resolved metagenomic data analysis. *Microbiome* **6**, 158 (2018).

103. Chklovski, A., Parks, D. H., Woodcroft, B. J. & Tyson, G. W. CheckM2: a rapid, scalable

and accurate tool for assessing microbial genome quality using machine learning. *Nat*
Methods **20**, 1203–1212 (2023).

104. Woodcroft, B. J. *et al.* Comprehensive taxonomic identification of microbial species in
metagenomic data using SingleM and Sandpiper. *Nat Biotechnol* 1–6 (2025)
doi:10.1038/s41587-025-02738-1.

105. Ayala-Ortiz, C., Freire-Zapata, V. & Tfaily, M. M. Stochastic assembly and metabolic
network reorganization drive microbial resilience in arid soils. *Commun Earth Environ* **6**,
647 (2025).

106. Aroney, S. T. N., Camargo, A. P., Tyson, G. W. & Woodcroft, B. J. Galah: More scalable
dereplication for metagenome assembled genomes. https://doi.org/10.5281/zenodo.13637856
doi:10.5281/zenodo.13637856.

107. Jain, C., Rodriguez-R, L. M., Phillippy, A. M., Konstantinidis, K. T. & Aluru, S. High
throughput ANI analysis of 90K prokaryotic genomes reveals clear species boundaries. *Nat*
Commun **9**, 5114 (2018).

108. Aroney, S. T. N. *et al.* CoverM: Read coverage calculator for metagenomics. Zenodo
https://doi.org/10.5281/zenodo.10531254 (2024).

109. Chaumeil, P.-A., Mussig, A. J., Hugenholtz, P. & Parks, D. H. GTDB-Tk: a toolkit to
classify genomes with the Genome Taxonomy Database. *Bioinformatics* **36**, 1925–1927
(2020).

110. Matsen, F. A., Kodner, R. B. & Armbrust, E. V. pplacer: linear time maximum-likelihood
and Bayesian phylogenetic placement of sequences onto a fixed reference tree. *BMC*
Bioinformatics **11**, 538 (2010).

111. Goldfarb, T. *et al.* NCBI RefSeq: reference sequence standards through 25 years of curation
and annotation. *Nucleic Acids Res* **53**, D243–D257 (2025).

112. Quast, C. *et al.* The SILVA ribosomal RNA gene database project: improved data
processing and web-based tools. *Nucleic Acids Res* **41**, D590–D596 (2013).

113. Zhang, H. *et al.* dbCAN2: a meta server for automated carbohydrate-active enzyme
annotation. *Nucleic Acids Research* **46**, W95–W101 (2018).

114. Honeker, L. K. *et al.* Drought re-routes soil microbial carbon metabolism towards emission
of volatile metabolites in an artificial tropical rainforest. *Nat Microbiol* **8**, 1480–1494 (2023).

115. Seemann, T. Prokka: rapid prokaryotic genome annotation. *Bioinformatics* **30**, 2068–2069

(2014).

116. Kanehisa, M., Sato, Y., Kawashima, M., Furumichi, M. & Tanabe, M. KEGG as a reference

resource for gene and protein annotation. *Nucleic Acids Research* **44**, D457–D462 (2016).

117. Boyd, J. geronimp/enrichM. (2025).

118. Graham, E. B. *et al.* Deterministic influences exceed dispersal effects on

hydrologically-connected microbiomes. *Environmental Microbiology* **19**, 1552–1567 (2017).

119. Shannon, P. *et al.* Cytoscape: A Software Environment for Integrated Models of

Biomolecular Interaction Networks. *Genome Res.* **13**, 2498–2504 (2003).

120. Banerjee, S. *et al.* Network analysis reveals functional redundancy and keystone taxa

amongst bacterial and fungal communities during organic matter decomposition in an arable

soil. *Soil Biology and Biochemistry* **97**, 188–198 (2016).

121. Arkin, A. P. *et al.* KBase: The United States Department of Energy Systems Biology

Knowledgebase. *Nat Biotechnol* **36**, 566–569 (2018).

122. Brettin, T. *et al.* RASTtk: A modular and extensible implementation of the RAST algorithm

for building custom annotation pipelines and annotating batches of genomes. *Sci Rep* **5**, 8365

(2015).

123. Seaver, S. M. D. *et al.* The ModelSEED Biochemistry Database for the integration of

metabolic annotations and the reconstruction, comparison and analysis of metabolic models

for plants, fungi and microbes. *Nucleic Acids Research* **49**, D575–D588 (2021).

124. Orth, J. D., Thiele, I. & Palsson, B. Ø. What is flux balance analysis? *Nat Biotechnol* **28**,

245–248 (2010).

Figure 1. **Distribution of bacterial and archaeal MAGs across soil layers and rates of potential microbial respiration**. The bar chart shows the number of the dereplicated MAGs that was assigned to each genus-level lineage, while the intensity of the heatmap represents the average numbers of the reads from each soil types that were mapped to the genomes.

SOM Classes

- Condensed HC-like
- Protein-like
- Amino sugar-like
- Lignin-like
- Lipid-like
- Carbohydrate-like
- Tannin-like
- Unsaturated HC-like

Sample Type Comparisons

- Unique to Subsoil/High
- Unique to Subsoil/Low
- Present in both samples under comparison

Figure 2. **Molecular compositions of the SOMs across soil types according to the pairwise G-test of uniqueness.** For each comparison, compounds unique to a soil type ($p < 0.005$) are represented as dots with colors other than gray, as specified in the legend. The boxes with colored edges on each van Krevelen plot denote the SOM classifications determined by thresholds of H:C and O:C ratios. The numbers located beside the stacked bar charts indicate the total number of compounds unique to each soil type being compared. The sizes of the colored segments in each bar represent the distribution (as percentages) of SOM classes among the total unique compounds in each soil type. The comparisons between the soil samples types are as follows: A) Surface/High vs. Surface/Low; B) Surface/High vs. Subsoil/High; C) Surface/High vs. Subsoil/Low; D) Surface/Low vs. Subsoil/High; E) Surface/Low vs. Subsoil/Low; and F) Subsoil/High vs. Subsoil/Low

Figure 3. **Modules with significant correlations between SOMs and KEGG Orthology (KO)-annotated genes across the soil layers and respiration rates, according to the MCODE network analysis**. The shapes of the nodes within each module signify either KOs or SOMs, with their colors representing KEGG pathway categories for KOs or SOM classes, as detailed in the legend. Positive correlations between SOMs and KOs are depicted as edges situated above the dashed "Correlation" line, whereas edges indicating negative correlations are positioned below the dashed line. These correlations meet the criteria of rho > 0.6 and FDR-corrected P < 0.01.

Figure 4. **Distribution of high-level BRITE pathway categories based on recruitment of KO-annotated genes mapped to the FTICR-MS derived compounds.** The histogram depicts the number of genes detected in the soil samples. The sizes of the bubbles indicate the relative distribution of genes associated with the designated BRITE pathway category, expressed as a proportion of the total gene counts in each sample. The adjusted p-values, represented by the colors of the bubbles, were calculated using the ClusterProfiler R package.

Figure 5. Carbohydrate-active enzymes (CAZymes) detected in MAGs of the selected genus-level lineages affiliated with the orders *Rhizobiales*, *Chthoniobacterales* and *Nitrososphaerales*. The bar chart on the left illustrates the total number of CAZyme families identified across the selected genera. Meanwhile, the combined dot-and-line plot along with the bar chart at the top displays the count of CAZyme families found in the specified genera or genus. Additionally, the stacked bar chart highlights the proportional distribution of CAZyme categories among these families. The CAZyme categories, as defined in the legend, include GH (glycoside hydrolase), GT (glycosyl transferase), CE (carbohydrate esterase), AA (auxiliary activity), CBM (carbohydrate-binding module), and PL (polysaccharide lyase).

Figure 6. Metabolic potentials of amino acid catabolism in the selected bacterial and archaeal genera, based on the flux-balance-analyses (FBA) and the KO annotation of the protein coding genes detected in MAGs. The box plot in the top-left shows the average log-corrected uptake rates of amino acids for each genus, as determined through FBA, while the bubble plot in the top-right highlights the average copy numbers of each KO associated with the corresponding genes. The color-coding of the dots/bubbles in both plots represents one of the chosen bacterial or archaeal genera. In addition, the shapes within the bubble plot in the leftmost column denote specific reactions, which are depicted in the cell diagram at the bottom.

Supplementary Figure 1. **Geographic coordinates of 47 soil samples spanning 37 sites across the continental United States (CONUS)**. Sample types, defined by both soil depths and respiration levels, are represented by color codes as specified in the legend. The map and geographic coordinates were generated using the rnaturalearth, ggplot2, and dplyr R packages.

Supplementary Figure 2. Modules with significant correlations between SOMs and KEGG Orthology (KO)-annotated genes detected in high-respiration soils, with the MCODE cluster score of < 5. The shapes of the nodes within each module signify either KOs or SOMs, with their colors representing KEGG pathway categories for KOs or SOM classes, as detailed in the legend. Positive correlations between SOMs and KOs are depicted as edges situated above the dashed "Correlation" line, whereas edges indicating negative correlations are positioned below the dashed line. These correlations meet the criteria of rho > 0.6 and FDR-corrected P < 0.01.

Supplementary Figure 3. Carbohydrate-active enzymes (CAZymes) detected in fungal contigs. The bar chart on the top illustrates the total number of CAZyme families identified from the contigs derived from the surface or subsoils under high or low respiration. Meanwhile, the combined dot-and-line plot along with the bar chart on the left displays the count of CAZyme families found in one or more of the soil sample types. Additionally, the stacked bar chart highlights the proportional distribution of CAZyme categories among these families. The CAZyme categories, as defined in the legend, include GH (glycoside hydrolase), GT (glycosyl transferase), CE (carbohydrate esterase), AA (auxiliary activity), CBM (carbohydrate-binding module), and PL (polysaccharide lyase).