Variance partition reveals contrasting random effect contributions on the density

and species composition of malaria-transmitting mosquitoes

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<u>Abstract</u>

Spatial-temporal variations exist in the density and species composition of malaria-

carrying mosquitoes, which will in turn influence the transmission of the disease. While

there has been extensive research on the seasonality and other main drivers of the

vector populations, the heterogeneity induced by random effects is just as important

but did not quite attract the same attention. To investigate the relative contributions of

the between-house, between-village, and between-year variations, as well as other

house-level covariates such as inhabitant number and bed net usage, intensive

Pyrethroid Spray Catches (PSC) sampling was conducted across a 60-month period

between 2012-2019 from four villages in the Sudano-Sahelian region of Burkina Faso.

For mosquito density, measured by counts of female *Anopheles gambiae s.l.*, our

modelling showed that the between-house variation was the largest component,

followed by the between-year then between-village variation, after accounting for

seasonality and other covariates. Density increased with the number of inhabitants

within a household but was uncorrelated with bed net usage. A subset of female

mosquitoes was genotyped for species identification, and the composition of *An.*

coluzzii and *An. gambiae*, the two dominant vectors in the region, varied hugely across

villages without a clear seasonal trend. The between-village variance contributed up to

76% of the total random variation, followed by the between-year variance. The between-

house variation was estimated to be statistically insignificant. Neither household size

nor bed net usage had any impact on species composition.

In short, the relative importance of the random components in mosquito density was in

the reverse order from species composition. The estimates and relative strengths help

parameterise potential field trials for novel vector control programmes and monitoring.

 Introduction

Despite the continued effort and investment in malaria control programmes, regions in

sub-Saharan Africa remain disproportionally affected by this vector-borne disease.

Harrowing figures of 263 million cases and over half a million deaths are still being

reported on the continent, accounting for about 95% of the global malaria burden [1].

Recent modelling predicted that the current intensity and range of interventions (such

as long-lasting insecticidal nets and indoor residual spraying) can at best maintain the

status quo but are unable to drive malaria towards eradication [2]. In fact, the recent

increase in malaria incidence coincides with the vectors' resistance to insecticides [3,

4], highlighting the need for developing complementary cost-effective and sustainable

novel technologies for malaria control. One example is the Sterile Insect Technique

(SIT), which has previously been implemented by the release of radiation-sterilised

individuals across multiple insect species that are deemed harmful [5]. SIT can also be

achieved via Genetic Modification (GM) of the vectors, such as by engineering a

transgene into the males of *Aedes aegypti* for dengue control [6]. Beyond SIT, there has

been extensive theoretical and laboratory-based research on developing gene drive

systems for *Anopheles gambiae s.l.*, a species complex whose members contain the

primary malaria vectors in sub-Saharan Africa. Gene drive allows the transgene to be

passed on to offspring and beyond in a super-Mendelian manner to achieve longer-term

population suppression or replacement [7, 8].

For contagious or vector-borne diseases, the basic reproductive number R_0 is the most

important metric to quantify their transmission, with $R_0 < 1$ being a requirement for

eradication. In malaria, R_0 is generally defined as the expected number of persons to be

infected one generation after the malaria parasite was introduced, or equivalently the

number of secondary cases [9]. As the life cycle of the malaria parasite consists of both

mosquito and human phases, complex parameterisation arises when characterising the

host-parasite-vector interaction [10]. The entomological inoculation rate (EIR), for

example, connects the sporozoite rate (i.e., the prevalence of the parasite in the salivary

glands of female mosquitoes), and human biting rate (HBR) per night, which has the

mosquito density per person embedded [11]. Under the classical models R_0 and

vectorial capacity are functions of these entomological rates at equilibrium. To reduce

the intensity of transmission, most vector control programmes aim to supress factors

concerning the vectors' density or competence [9].

Malaria transmission is never uniform across space and time. Previous estimates of R_0

varied hugely from ~1 to >1000 across Africa [9], and mosquito density fluctuates

across seasons often by orders of magnitude [12]. In fact, heterogeneity exists on

almost all levels: from mosquito density and species distribution across countries,

exposure among households within the same village, down to the variable susceptibility

of individuals. While earlier works have focused on identifying the main drivers (such as

environmental and climatic factors) associated with the trends in malaria vectors and

cases [13], it is equally important to quantify the unexplained portion of the variation.

Through the extensive mosquito collection survey (see below) this study has the

following aims: 1) To estimate the random or unexplained variance components

concerning mosquito density and species composition at different spatial-temporal

levels, in particular, the between-village, between-house, and between-year variation,

after accounting for seasonality and other fixed covariates. 2) To discuss how these

random components and their relative strengths impact vector density and malaria

transmission. 3) To use these analyses to provide recommendations on the potential

design and sample size considerations for clustered randomised controlled trial (cRCT)

under various intervention scenarios.

<u>Sampling</u>

Mosquitoes were sampled in the houses from four villages in Burkina Faso's Sudano-

Sahelian: Bana Market, Bana Village, Soukroudingan, and Pala (Figure 1), where

morphologically identical *An. gambiae (s.s.)*, *An. coluzzii*, and *An. arabiensis* are the

primary vectors. The first three villages are located to the west of Bobo-Dioulasso,

where our research base situates, while Pala is on the opposite direction from the rest

[14]. Bana's two collection sites, Bana Market and Bana Village, are separated by a

small semi-permanent river with different human activities [15]. Two phases of

collection were conducted: from July 2012 to June 2014 and throughout 2017-2019,

covering a total of 60 months. Monthly collection was scheduled aiming to sample 20

houses per village per month. Houses were chosen randomly from a pool of known

houses based on local information, from which half of the houses were repeatedly

visited over time (fixed houses), while the remaining half were further randomised each

month (random houses). For each village, we attempted to sample all houses in the

same or consecutive days to minimise the effect of nuisance abiotic factors. For Pala

and Souroukoudingan, the sampling interval switched to once every two months in the

second phase (Figure 2). Sampling took place after sunrise via indoor pyrethroid spray

catches (PSC), using a well-known and locally-market available insecticide spray

Kaltox® Paalga (Sphyto company, Burkina Faso). All mosquitoes were morphologically

identified in the field [16]. The targeted organisms *An. gambiae s.l.* were preserved in

80% ethanol for further analysis. During the second phase of collection PCR was

performed on a subset of the preserved samples for species identification, such that

the composition of the species (within *An. gambiae s.l.*) per visit was also known. The

genotyping was based on the detection of the SINE 200x locus [17].

Methods

Each house visit generated a line of record, including the number of female *An. gambiae*

s.l., the date of collection, and a unique house ID. Other potential covariates

concerning the visit were also documented, such as the number of residents, and

whether bed net was in use. Female *An. gambiae s.l.* per visit was the main response to

proxy indoor mosquito density. To model seasonality we used a six-level factor variable,

with two consecutive months being pooled as one (e.g. January and February as one

level, and so on) for a more balanced design. A Poisson generalised linear mixed model

(GLMM) was fitted. Seasonality, household size, and the use of bed nets (yes/no) were

treated as fixed effects. Observational random effects were included as a measure of

overdispersion [18]. Villages, year of sampling, and unique house IDs were considered

as random effects. The interaction between village and seasonality was also included

for additional variability. For species composition, a similar binomial GLMM was fitted

to the proportion between the female counts of the two most dominant species in the

region. Analyses were conducted in R-4.3.1 [19] with packages lme4 and emmeans

[20].

<u>Results</u>

We endeavoured to follow closely the intended sampling plan with 3212 visits made

(Figure 2). After removing entries with missing information, 3133 visits of 383 distinct

houses were retained over the sampling period. Mosquito density as measured by the

number of female *An. gambiae s.l.* is summarised as follows: A total of 28,919 female

An. gambiae s.l. were collected, with an average of 9.2 per visit. The high mean count

came with huge variation across visits ranging from 0 to 333 per visit. The mean

household size was just over 3 persons – this value was subsequently used to classify

household size into categories (none, small with 1-3 persons, large with >3 persons).

Bed nets were installed in about 90% of the visits.

A Poisson GLMM was fitted to the counts (Table 1A). Seasonality was well defined within

a calendar year with a period of low and high counts. Lower counts were observed

between November and February, then increased when transitioning into the next

season towards the peak at the months of September and October. There was about a

30-fold difference between the lowest and highest mean counts. As we grouped

seasonality into six two-monthly factor levels, we performed contrasts to test whether

their mean counts were different. All pairwise comparisons were significant, apart from

the three levels of months between May and October where they shared the same

mean. The presence of bed nets reduced mosquito count by 14%, but the effect was

not statistically significant (p=0.099). Count increased with household size: compared

to an empty house a 3-fold (p=0.013) increase was observed if there were ≤3 occupants,

then a further 53% increase (p<0.001) for a bigger household with >3 occupants. For the

random components (Table 1B), observation-level overdispersion had the highest

contribution, accounting for 65% of the total variance. The second largest term was the

between-house variation (14%). The between-village variance was found to be the

lowest at 4.6%. The village-seasonality interaction allowing for additional variation on

top of the two random components was estimated to be of moderate strength (8%). All

five variance components were statistically larger than zero.

A subset of female *An. gambiae s.l.* from the second phase of collection (2017-2019)

was genotyped for species identification. After genotyping, 3626 were identified as *An.*

coluzzii and 1102 as *An. gambiae*, yielding an overall ratio of approximately 3:1 from 985

house visits (Figures 5-6). Apart from the two dominant species a small proportion of

An. arabiensis (~5%) was also found but excluded from subsequent statistical analyses.

The species proportions in our four villages were very different (Figure 4). In Bana Market

and Bana Village *An. coluzzii* was the dominant species making up over 90% of the

samples, and similarly 60% in Souroukoudingan, while in Pala it was the minority (24%).

Although species composition varied throughout a year, the four villages did not seem

to follow the same dynamics nor a pre-defined pattern. A binomial GLMM was then

fitted to model the relative proportion of the two species. For the fixed effects, neither

the use of bed nets (p=0.77) nor household size (p=0.89) had an influence on the

proportion (Table 2A). For the random components (Table 2B), the between-village

variation was estimated to be the largest, explaining 77% of the total random variance.

This was followed by overdispersion (11.4%) and the village-seasonality interaction

(3.9%). The between-house variation was not significant (1.5%, p=0.26). In other words,

setting overdispersion aside, the relative importance of the random components were

the exact opposite of those for mosquito density.

<u>Discussion</u>

Several baseline collections have been conducted in Burkina Faso and beyond to

evaluate vector density over space and time [21-23]. Our results agree with the existing

knowledge that mosquito density (as measured by female counts in this study) exhibits

strong seasonal trends. The study region is characterised by a tropical climate marked

by two main seasons, with average annual temperatures between 25-30˚C and rainfall

of 800-1100mm [24]. The rainy season is usually defined as the months from June to

September, with May and October being in transition [15], the same period at which we
consistently observed higher counts. The peak at the late rainy season near October
was also detected by Dao et al. [12] in a Sahelian village in the neighbouring country of
Mali. After reaching the peak a period of low counts immediately followed from
November into February. The clean switch between the two seasons is a signature of
the Sahelian region as the surface water disappears, drastically reducing the availability
of larval sites [25, 26]. In parallel to our census size estimates, genetic estimates of
effective population size also showed similar patterns of seasonal variation [27].

While many existing collections aimed to correlate seasonality and other deterministic
drivers with mosquito dynamics through monthly summaries, curating individual PSC
records gives opportunity to analyse covariates on finer scales such as on observational
and household level. Bed net use, a factor considered as a fixed effect in this study, only
reduced mosquito density by 14% (but not significant) despite a high coverage of over
90%. A separate study discovered the increase in frequency of insecticide resistance
alleles during the same period which potentially explains its relatively low efficacy [3].
We found a strong relationship between density and household size, a factor which was
previously thought to be less important [21]. Another dimension of this work is to
estimate the random effects, which induce uncontrolled heterogeneity that have not
been explained by our main effects. We intentionally fitted a 6-level seasonality variable
with 24 additional random village-seasonality interaction terms to our GLMMs to
saturate the seasonality effect, allowing us to focus on these random components. The
sizes and relative importance of the random components are of interest rather than how
and why they differ. Several random components on various levels were considered:
Overdispersion, assigned to each visit, was also the largest component on mosquito
density. It includes the intrinsic variation of indoor mosquito density, sampling method,
and potentially other abiotic or idiosyncratic factors. By revisiting some households, the
between-house variance was estimated to be the second largest component, hinting
some houses were more susceptible to mosquitoes even though they were exposed to
similar conditions as their neighbours. The difference could be due to building material,
usage, and proximity to larval sites. Further modelling, including the use of GIS data, is

warranted to further dissect this variance component [21]. The between-village and

between-year variances were minor with respect to density, each contributing to only a

few percent of the total random variation.

There is no surprise that the two primary vectors *An. coluzzii* and *An. gambiae* were

abundant in the region. Our binomial GLMM concluded that the proportion of the two

species was unaffected by bed net usage or household size. The four sampled villages

tended to have their own underlying proportions and seasonal variations:

Souroukoudingan and the two sites from Bana were predominantly *An. coluzzii*, while

Pala was the opposite. There was no observable switch in the proportions across

seasons. The large between-village and village-seasonality interaction illustrated how

diverse species composition can be across landscape. It is known that the two species

coexist but with rather contrasting preferences. *An. coluzzii* prefers sites with

permanent freshwater while *An. gambiae* can tolerate intermittent access to water [25].

Bana and Souroukoudingan have access to water year-round due to agricultural

activities that explains the strong presence of *An. coluzzii*. Their respective persistence

mechanisms may also affect the relative abundance in the dry season and transition

months. It is generally thought that *An. coluzzii* aestivates locally during the dry season

thus an immediate population expansion is expected when the aestivators reemerge,

whereas *An. gambiae* takes days to respond to the arrival of the rainy season as

immigrants from further afield recolonise the focal population [28]. In short, the large

between-village and lack of between-house variance were the key signatures of species

composition. In fact, the relative importance of the random components in species

composition is the exact opposite of that of density (Table 3), which itself is a very

interesting finding. The design of population monitoring programmes and control

strategies will therefore be influenced by the choice of response variable (see below).

Future work should include the monitoring of *An. arabiensis*, which was also present in

our study, especially in urban settings where it appears to be colonising [29].

Both mosquito density and malaria transmission are subject to strong spatial-temporal

variation. Besides our findings on density, a previous survey in the same region

estimated a 25-fold increase in EIR from the dry to rainy season, mainly driven by the

sharp rise in human landing rate [22]. Both malaria incidence and mosquito density

found their peaks towards the late rainy season in Burkina Faso [30], meaning seasonal

strategies may be required [31], further straining the healthcare system. Seasonality

may also affect the overall efficacy of GM vector control and the optimal schedule of

releases [32, 33]. While the heterogeneity in transmission across seasons or climatic

zones have been anticipated, the unexplained variance components impose additional

challenges on the road to the elimination of malaria. Persistent transmission can be

maintained even under continued and effective malaria control [34]. It only requires a

minority of human individuals with dry season exposure to constitute an infectious

reservoir [35]. The higher the random variances in mosquito density the higher chance

these small pockets of space are created for permanent transmission. Classical

epidemiological models showed that with heterogeneous biting, a factor of $(1 + \alpha)$ is

applied to R_0 where α is the squared coefficient of variation (CV) in biting rate. In other

words, R_0 increases linearly with the variance in biting rate if all else is equal. While

biting rates are best estimated from human landing catches, our PSC collection and

between-house variance estimate in density could be carefully extrapolated to inform

the degree of heterogeneity linking to indoor exposure and biting. With the many layers

of random variations, the compounding effect on R_0 could potentially be significant

even if the mean estimate suggests otherwise. Biting can also occur outdoor, thus

measuring the outdoor mosquito density and species composition with other trapping

methods would complement this study [36]. In terms of reporting, increased spatial and

temporal heterogeneity in malaria cases in low transmission settings reduces the

usefulness of national or regional level trends in incidence or prevalence [37].

Any new intervention needs to undergo phases of rigorous testing and trial. A cRCT is

the gold-standard study design for evaluating public health programmes, where

interventions are addressed to some clusters then compared to those without [38, 39].

Our entomological dataset and parameter estimates can contribute towards baseline

data collection for a potential upcoming field trial, and more importantly, inform some

of the key considerations concerning trial design. In many mosquito intervention

programmes treatments must be addressed to a cluster of households (e.g. a village) as

a whole, hence it is the basic unit for randomisation. If the proposed intervention targets

the entire species complex to suppress mosquito density by a certain percentage, then

its effect can be evaluated from the difference in counts between the control and

treatment arms via a Poisson GLMM like the one we have fitted but with an additional

variable indicating the treatment arm. We suggest sampling should prioritise months

with high abundance to maximise the chance of detecting the intervention effect. There

are two reasons for this: The large baseline counts in the rainy seasons provide

headroom for population suppression, while in the drier months the monthly

aggregated counts are low with numerous zero readings from individual houses. The

second reason is that the intervention effect is likely to be estimated with greater

precision (i.e., a lower coefficient of variation). Observation-level covariates such as

household size were shown to be correlated with the baseline counts. For the same

reason, the recruitment of households should target those with habitants. These

covariates are essential when the cRCT is analysed from observation-level data [40],

and remain useful under the use of cluster-level summaries, where they can inform the

pairing of clusters in restricted randomisation, or even a matched-pair cRCT design to

further improve balance and statistical power [41].

The random effect estimates will also influence the design of the cRCT. The current

sampling of 20 houses per month per village aimed to illustrate seasonality by averaging

out the overly dispersed count data. One can find the optimal sampling strategy by

altering the number of clusters and houses per month per cluster while keeping the

total sampling effort constant. Given the relatively high importance of the between-

house variation, sampling the same set of houses repeatedly may risk inducing

systematic bias. Small but non-zero between-year variation in density was detected

across the 6-year period. Yearly variation is less of a problem for the classical parallel

cRCT design as comparisons are made mostly within the same time period (vertical)

between the two arms. A stepped-wedged design, however, contains both vertical and

within-cluster (horizontal) comparisons, may be prone to this effect. In the stepped-

wedged design, clusters are randomised into several streams. All clusters are in control

condition initially and will have received the intervention at the end, with intervention

being gradually rolled out at different time points of the trial. As all clusters now include

pre- and post-intervention phases, a good year (with respect to mosquitoes' survival)

alone may result in a sudden increase in mosquito density even in clusters receiving the

intervention. On a positive note, the between-year variance was estimated to be small

with limited impact, contributing to only a few percent of the total variation in both

mosquito counts and species proportions. While our four villages were chosen based

on logistical reasons, they were considered as a random effect in the GLMMs as this

would be the case in a cRCT. A non-zero between-village variance was estimated

despite them being in the close vicinity (~30km radius) of Bobo-Dioulasso. A full-scale

cRCT would likely be conducted on a regional or even national basis with many more

clusters, thus this variance is expected to increase with geographical coverage. Note

that these parameter estimates (Tables 1A, 1B) are reported in the log scale, which

means they should be compared to the intervention strength of $\log(1-G)$ if the

intervention suppresses the density by G. In trial designs, the between-village variance

informs the intra-cluster correlation coefficient (ICC), which measures relative

similarity of the observations from the same cluster compared to those from the others.

Since individual variances (i.e. overdispersion and between-house variation) were

estimated to be much larger than between-village variance, the ICC is expected to be

reasonably small, which helps reduce the number of clusters required.

Some novel vector control strategies can be tailored to target one or a subset of species

while keeping the rest intact. The results from such a cRCT can be analysed with a

similar Poisson GLMM on the counts of the targeted species, potentially with the non-

targeted species as covariates. Another way is to analyse the relative proportion of the

two as a pair of joint response with a binomial GLMM, whose intervention effect is $\log(1-$

$G)$ on the logit scale. Between *An. coluzzii* and *An. gambiae*, the ICC in proportion is

inferred to be much higher than in their combined density. In other words, most

variation comes from clusters rather than the individual houses, thus the effective

sample size approaches the number of clusters. Under high ICC the stepped-wedged

design may be preferred as the efficiency is improved by the inclusion of horizontal

comparisons [42]. The lack of between-house variation in species composition has the

potential to simplify the recruitment process, as one or a few households could

represent the entire cluster. To conclude, our data and analyses provide crucial

parameter estimates to be used in trial design, power analyses and sample size

determination. With the many combinations of variables and scenarios, quantitative

studies are necessary to fully explore their complex interplay, often aided by computer

simulations.

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Author contributions

P.S.E., A.A.M., F.A.Y. led the field collection. P.S.E., D.K. conducted the genotyping.

P.S.E., A.A.M., F.N., T-Y.J.H curated and processed the data. T-Y.J.H. conducted the

statistical analyses. T-Y.J.H wrote the manuscript with inputs from the other authors.

Conflicts of interest

Authors declare no conflicts of interest.

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Data availability

The dataset will be made available upon acceptance.

Fixed effect	df	Comparison(s)	Estimate (s.e.)	p-value
Seasonality	5			0.000***
		(See Table 1C)		
Bed net	1			
		No vs Yes	-0.15072 (0.091)	0.099
Household size	2			0.000***
		None vs Small	1.110 (0.446)	0.013*
		Small vs Large	0.427 (0.066)	<0.001***
		None vs Large	1.537 (0.449)	0.000***

515 Table 1A. Fixed effect estimates from the Poisson GLMM on the counts of female *An.*

516 *gambiae s.l.*. The significance of the fixed effects and p-values were based on

517 Likelihood Ratio Tests (LRTs), while Wald p-values were computed for pairwise

518 comparisons. Effect estimates are on log-scale. Asterisks are given to indicate

519 statistical significance: (***) p<0.001, (**) p<0.01, (*) p<0.05.

Random effect	Number of levels	Variance estimate	Percentage	p-value
Overdispersion	3133	1.4413	64.88	0.000***
Between-house	383	0.3127	14.07	0.000***
Between-year	6	0.1881	8.47	0.000***
Between-village x Seasonality	24	0.1768	7.96	0.000***
Between-village	4	0.1025	4.61	0.048*

521 Table 1B. Random effect estimates and percentage contributions from the same

522 Poisson GLMM, ranking from the largest. LRTs were performed to obtain the p-values.

	0102	0304	0506	0708	0910	1112
0102		0.000***	0.000***	0.000***	0.000***	0.000***
0304	2.151		<0.001***	0.003**	0.000***	0.000***
0506	3.212	1.061		0.638	0.285	0.000***
0708	3.066	0.915	-0.146		0.130	0.000***
0910	3.544	1.394	0.332	0.479		0.000***
1112	0.798	-1.353	-2.414	-2.268	-2.746	

527 Table 1C. Pairwise contrast for the seasonality effect in the Poisson GLMM (Table 1A).

528 Seasonality was categorised into 6 factor levels. For example, "0102" refers to the

529 counts of Jan and Feb combined. Contrasts (lower-triangular) and the unadjusted p-

530 values (upper-triangular, shaded in grey) were computed via the R package "emmeans".

Fixed effect	df	p-value
Seasonality	5	0.028*
Bed net	1	0.771
Household size	2	0.886

532 Table 2A. Fixed effect estimates from the binomial GLMM on the proportion between

533 female *An. coluzzii* and *An. gambiae*, with p-values obtained from LRTs. Effects are

534 reported on logit scale.

Random effect	Number of levels	Variance estimate	Percentage	p-value
Between-village	4	3.476	76.47	0.000***
Overdispersion	985	0.520	11.43	0.000***
Between-village x Seasonality	24	0.304	6.68	0.000***
Between-year	3	0.179	3.92	0.000***
Between-house	200	0.0678	1.49	0.263

542 Table 2B. Random effect estimates and percentage contributions from the binomial

543 GLMM on species proportion, ranking from the largest. LRTs were performed to obtain

544 the p-values.

Factor	Density	Species composition
Seasonality	Clear seasonal pattern with a rainy and dry season per year	No clear rainy and dry season patterns. Villages tend to have their own seasonality
Bed net installation	Not significant	Not significant
Household size	Density increased with household size	Not significant
Between-house variance	Large	Not significant
Between-year variance	Small but non-zero	Small but non-zero
Village-seasonality interaction variance	Moderate	Moderate
Between-village variance	Small but non-zero	Large

546 Table 3. Descriptive summary of the fixed and random factors affecting mosquito

547 density and species composition, based on the numerical results from the two GLMMs

548 reported in Tables 1-2. For the random effects, the relative importance is reported.

552 Figure 1. Location of our sampling locations in the Houet Province of western Burkina

553 Faso. The coordinates of the sampling locations are: Bana [11.233, -4.472],

554 Souroukoudingan [11.235, -4.537], Pala [11.151, -4.234]. In the centre of the map is

555 Bobo-Dioulasso, the second largest city in Burkina Faso, where our research base

556 locates. There are two sampling locations from Bana: Bana Village and Bana Market,

557 which are about 23km from Bobo-Dioulasso. Souroukoudingan is about 7km further

558 away in the same direction. Pala is on the opposite direction approximately 6km from

559 Bobo-Dioulasso.

563   Figure 2. Sampling effort per village over time. Each bar represents the number of

564   houses visited for PSC collection per two months. Sampling was paused from July 2015

565   to Dec 2016.

Figure 3. Mean female *An. gambiae s.l.* collected per visit by village over time. Note that

sampling was paused after June 2014 then resumed at the beginning of 2017.

572 Figure 4. The relative proportion of *An. coluzzii* and *An. gambiae* by village over time.

573 Species identification was performed to a subset of mosquitoes collected during 2017-

574 2019.

577   Figure 5. Mean number of female *An. coluzzii* per house visit by village over time.

581   Figure 6. Mean number of female *An. gambiae* per house visit by village over time.