Drone Survey Reveals a Severe Chinstrap Penguin Decline and a Novel Gentoo Colony in an Antarctic Specially Protected Area

Eduardo J. Pizarro*[1,2,3], Miguel Correa[4], Francine Timm[4], Gaspar Mejías[3], Ana Cláudia Franco[4], Juliana A. Vianna[2,3,5], Léa Cabrol[3], Francisco Santa Cruz[3,6], Lucas Krüger*[3,6]

1. Pontificia Universidad Católica de Chile, Facultad de Ciencias Biológicas, Santiago, Chile.
2. Millennium Institute Center for Genome Regulation (CGR), Santiago, Chile.
3. Millennium Institute Biodiversity of Antarctic and Subantarctic Ecosystems (BASE), Santiago, Chile.
4. Department of Microbiology, Immunology and Parasitology, Institute for Basic Health Sciences, Federal University of Rio Grande do Sul, Porto Alegre, RS, Brazil.
5. One Health Institute, Faculty of Life Sciences, Universidad Andres Bello, Santiago, Chile.
6. Departamento Científico, Instituto Antártico Chileno INACH, Punta Arenas, Chile.

*Correspondent: eduardo.pizarro@uc.cl https://orcid.org/0000-0003-3427-1194; lkruger@inach.cl https://orcid.org/0000-0003-4637-5302;

Abstract. Seabird populations in Antarctica are sensitive indicators of environmental change and understanding their dynamics is crucial for assessing the health and future of the Antarctic ecosystems. This study provides an updated population assessment for three key seabird species - the Antarctic Shag (*Leucocarbo bransfieldensis*), Gentoo Penguin (*Pygoscelis papua*), and Chinstrap Penguin (*P. antarcticus*) - within the Antarctic Specially Protected Area (ASPA) 133 on Nelson Island. Using drone-based imagery and photogrammetric analysis, we conducted comprehensive nest counts to evaluate current population size in comparison with historical data. Our results reveal a pronounced 57% decline in the Chinstrap Penguin population at Harmony Point since the 1990s, with 38,080 nests counted compared to historical estimates of nearly 90,000. In contrast, Gentoo Penguin numbers at Harmony Point (3,659 nests) appear stable following a previously documented increase. We also provide the first population estimate for a large Gentoo Penguin colony located on a small peninsula south of Harmony Point, known as the Toe, where 4,641 nests were recorded. Additionally, we report an 84% increase in Antarctic Shags, with 127 nests recorded at Harmony Point. These findings are consistent with regional species-level trends and provide critical baseline data for monitoring the impacts of rapid environmental change on Antarctic seabird communities.

Keywords. breeding colony, drone-based aerial imagery, *Pygoscelis*, Shags

Statements and Declarations

Competing Interests. Authors declare no conflict of interests.

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Data Availability Statement. Data is available in repository (Krüger et al. 2026).

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Introduction

The Maritime Antarctic Peninsula has experienced some of the most rapid warming in recent
decades, leading to significant changes in sea ice dynamics, prey availability, and habitat
conditions that affect seabird species (Morley et al. 2020; Ramos and Pereira 2022). Among
these, the Antarctic shag (*Leucocarbo bransfieldensis*), chinstrap penguin (*Pygoscelis*
antarcticus), and gentoo penguin (*Pygoscelis papua*) occupy important ecological niches,
reflecting differences in foraging strategies, habitat use, and trophic dependence within the
Antarctic marine ecosystem (Lynch et al. 2012; Lee et al. 2021). Gentoo penguins have
demonstrated a notable capacity for range expansion and population growth, with increasing
abundances and a southward shift in their distribution (Clucas et al. 2014; Herman et al.
2020). In contrast, chinstrap penguins have exhibited population declines across multiple
regions, potentially linked to changes in krill availability and sea ice conditions (Trivelpiece et
al. 2011; Strycker et al. 2020). In this context, Antarctic Specially Protected Areas (ASPAs)
provide opportunities to evaluate long-term population trajectories under minimal direct
human disturbance.

ASPA No. 133 on Nelson Island, South Shetland Islands, comprises two ice-free areas:
Harmony Point and the Toe (Fig 1a-c). Harmony Point has long been recognized as an
important breeding site for seabird species, especially chinstrap penguins (Silva et al. 1998).
Although historical population data is available for some species, there are inconsistent gaps
in monitoring data (Casaux and Barrera-Oro 2016; Oosthuizen et al. 2020), with the most
recent penguin population estimates dating back to the mid-1990s (Silva et al. 1998).
Against this background, although circumpolar satellite-based assessments have
substantially advanced penguin monitoring, their ability to resolve some locations, including
Nelson Island, has been limited by constraints in satellite imagery (Strycker et al. 2020).
Drone technology and photogrammetric techniques therefore provide an effective
complementary tool for more accurate and comprehensive seabird monitoring in remote and
logistically inaccessible locations (i.e. Borowicz et al. 2018).

This study aims to provide an updated population assessment of Antarctic shags, chinstrap
penguins, and gentoo penguins in ASPA No. 133, a protected area subject to minimal direct
anthropogenic pressure, using drone-based imagery and photogrammetric analysis. By
comparing current population estimates with historical data, we seek to contribute to the
understanding of long-term population dynamics of these key indicator species within a
rapidly changing Antarctic ecosystem.

Methods

Aerial imagery was collected at the study sites Harmony Point and the Toe, using a DJI
Mavic 3 unmanned aerial vehicle (UAV) equipped with a high-resolution camera (Fig 1a,b,c).
Flights were conducted manually during favorable weather conditions at an elevation of 100
m above ground level. The theoretical ground sampling distance (following Stone and Davis
2025) was approximately 2.67 cm per pixel, according to camera specification: sensor width
of 17.3, focal length of 12.29 and image width of 5280. Images were acquired at uniform
resolution every 3s at an average speed of ~5 m/s, resulting in a forward overlap of
approximately 70–80% and side overlap of approximately 50-70%, while minimizing
disturbance to breeding birds (Ratcliffe et al. 2015). At Harmony Point, a total of 7 flights

were conducted over four survey days (6th, the 16th and 19th of December) during the incubation stage of gentoo and chinstrap penguins, yielding 5,932 images. These flights collectively covered the entire breeding area. At the Toe, two flights were conducted, one in 22th of December 2024, and another one on the 10th of January 2025 (to cover a small gap in the area detected after generating the orthomosaic) collecting 284 images. All drone imagery was processed using Agisoft Metashape Professional software v.2.2.1.20502 to generate one high-resolution orthomosaic image per flight. As orthomosaics were generated per flight, we calculated mean ± standard deviation resolution (distance between the center of consecutive pixels), x (longitude) and y (latitude) pixel size (in centimeters).

Flight operations adhered to the Environmental Guidelines for Operation of RPAS in Antarctica, adopted under Antarctic Treaty Consultative Meeting Resolution 4 (2018), and the most up-to-date handbook produced by COMNAP in collaboration with the SCAR committee COMNAP RPAS Operator's Handbook (Version 8/18 December 2023), which provide best-practice recommendations for planning and conducting RPAS operations to minimize environmental impact in the Antarctic Treaty Area. Drone flights complied with all permit requirements for drone use in ASPAs (INACH permits Nº 00623/2024 and Nº 00624/2024).

Nest counts were conducted manually by a single observer placing a geographical point on each identified nest structure within the orthomosaic images using ArcGis Pro (ESRI 2026). The final shapefiles were used to estimate the number of nests. The identification of the Antarctic shags were corroborated by camera images taken on the field. The positions of gentoo penguin breeding groups at Harmony Point were marked on-site using a handheld GPS (Garmin Montana I7) to aid in species identification within the orthomosaic images. As we did not have access to the Toe site, the drone was launched from Harmony Point (~ 3km). Although it is possible to distinguish between the two penguin species based on nest density, additional imagery was captured at an 80 m altitude to confirm species identification.

For Antarctic shags, a nest was defined as a seabird occupying a circular structure within the boundary of a white guano stain. For penguin species, a nest was considered occupied when an adult bird was observed either standing over or sitting on an identifiable nest structure within the boundary of a guano stain, and forming part of a recognizable breeding aggregation (i.e., multiple nests in close proximity, consistent with colony structure). For estimating accuracy of the counts, three different authors (EP, GM and LK) marked nest points on a subset of spatially discrete units of the colonies for chinstrap (n=38) and gentoo (n=24), and all of the colony for the shags (n=10). The difference between the three counts was used to calculate standard deviation, and then the accuracy calculated as the percentage of the standard deviation corresponding to the mean of the three counts (coefficient of variation). We also calculated the coefficient of variation to the total of the counts per observer, in order to compare whether differences in counts at smaller subsets reflects the differences in total count. We tested whether there was correlation between counts using a two-way test intraclass correlation coefficient in the 'irr' R package (Matthias Gamer et al. 2025). The topography of Harmony Point was estimated from previous drone flights (Santa Cruz and Krüger 2023).

Current population estimates were compared with historical data from previous studies
conducted at Harmony Point. For Antarctic shags, population estimates are available from
(Casaux and Barrera-Oro 2016) and (Oosthuizen et al. 2020), while comparable information
for chinstrap and gentoo penguins is provided by (Silva et al. 1998) and (Croxall and
Kirkwood 1979). Early population assessments were largely based on coarse visual surveys
and extrapolations and are generally considered to have high uncertainty, often on the order
of several tens of percent (e.g., Croxall and Kirkwood 1979). In contrast, drone-based
surveys, such as those used in the present study, offer substantially higher spatial resolution
and improved methodological consistency. Although the accuracy of drone-derived
estimates varies depending on image resolution, survey design, and counting protocols, they
are generally expected to provide lower uncertainty than early ground-based assessments
(Humphries et al. 2017; Hilton et al. 2024).

All analyses were conducted in R (R Core Team 2024) and maps generated in ArcGIS Pro
(ESRI 2026). All data and R code is available at Krüger et al. (2026) and (Krüger 2026).
Inter-Quartil Range (IQR) are presented as lower and upper 75%.

Results

The resulting orthomosaics had 10.25 ± 1.16 cm/pix of resolution (min 9 cm, max 12 cm). X
pixel size was 7.30 ± 0.83 cm (min 6.20 cm, max 8.62 cm). Y pixel size was 3.40 ± 0.39 cm
(min 2.89 cm, max 4.01 cm).

Median accuracy of the counts was 5.1% [IQR 2.3%, 8.3%], 4.7% [2.6%, 7.5%] and 4.5%
[0.9%, 6.5%] for chinstrap penguins, gentoo penguins and Antarctic shags, respectively.
Total variation was within the IQR for chinstrap penguins (4.3%) and Antarctic shags (1.2%)
but was below the IQR for gentoo penguins (1.9%). Very high agreement was found
between observers for the three species: chinstrap penguin ICC = 0.994 [IQR 0.988,0.997],
$F_{1,37}$=698.4, p<0.001; gentoo penguin ICC = 0.993 [0.988,0.997], $F_{1,19}$=511.8, p<0.001;
Antarctic shag ICC = 0.949 [0.863,0.986], $F_{1,9}$=53.7, p<0.001.

Chinstrap penguins occupied most of the areas along the shoreline and rocky outcrops (Fig
1d,e). A total of 38,080 and 8,153 nests were recorded in the 2024 season at Harmony Point
and the Toe, respectively. This represents a decrease of 57% and 74% from the 1990s and
1980s estimates respectively, for Harmony Point (Fig 2a). For the Toe colony, it represents a
45% decrease from the 1970s, with a high error around that estimation (±50%) (Fig 2b).
Chinstrap penguins occupied elevations mostly between 9.8 m and 32.9 m, on slopes of 1.8º
to 5.6º inclination, and were typically found within 104 m of the coastline.

Gentoo penguins nests were estimated at 3,659 and 4,641 in the 2024 austral summer
season at Harmony Point and the Toe, respectively (Fig 1d,e). Previous assessments
indicate that the Harmony Point colony has fluctuated between 3,000 and 4,000 pairs since
the late 1980s, with one earlier estimate of 1,642 pairs in 1966 (Fig 2c), representing an
122% increase since the 1960s. Most nests were located at elevations between 10.3 m and
36.0 m, on smooth slopes with inclinations of 1.0º to 5.4º, and at distances of 67.3 m to
151.7 m inland.

A total of 127 Antarctic shag nests were recorded at Harmony Point (Table 1, Fig 1f). This
represents an 84% increase compared to the previous assessment in the 2018/19 breeding
season but falls within the expected variability considering counts between 1980 and 2010

(Fig 2b). Compared to the penguin colonies, shags occupied lower elevations (11.2 m to
17.0 m), on relatively steeper slopes (10.4º to 18.4º), and closer to the coastline (within 15 m
inland).

Discussion

We update the population size and spatial distribution of *Pygoscelis* penguins in ASPA No.
133, which had not been revised since the 1990s. The population assessment demonstrates
a marked decline of 57% in one of the largest chinstrap penguin colonies in the South
Shetland Islands, consistent with recent regional assessments (Strycker et al. 2020; Hilton et
al. 2024; Hinke et al. 2025), and report the presence of a previously undescribed, large
gentoo penguin colony at the Toe with over 4,500 pairs, contributing to further evidence of
population expansion in this species (Herman et al. 2020). Together, these findings highlight
the importance of site-based population assessments and the value of drone-derived data
for accurately quantifying breeding populations in remote Antarctic environments.

The number of chinstrap penguins at Harmony Point exhibited a clear decline from the last
available estimate of nearly 90,000 breeding pairs in the mid-1990s (Silva et al. 1998). The
observed 57% decrease over three decades is consistent with regional declines reported
across the Antarctic Peninsula (Talis et al. 2023; Krüger 2023; Hinke et al. 2025), and with
assessments suggesting increasing conservation concern for the species at a local scale
(Krüger 2023). Many chinstrap colonies in the region still lack recent and reliable population
estimates (Harris et al. 2015), highlighting the importance of updated, site-specific
assessments such as those presented here. The concordance between the decline
observed at Harmony Point and broader regional trends supports the robustness of this
result. Although the underlying causes remain poorly understood, several interacting factors
are potentially affecting chinstrap penguin population dynamics, include the combined
effects of climate change on chick survival (Salmerón et al. 2023), shifts in prey availability
combined with increasing krill fisheries (Watters et al. 2020; Krüger et al. 2021), and
competition with recovering whale populations (Savoca et al. 2024). We hypothesize that the
cumulative effects of multiple factors are likely responsible for the observed trends. The
Maritime Antarctic Peninsula holds one-third of the global chinstrap penguin population
(Strycker et al. 2020); even assuming the remaining two-thirds of the population is stable,
the observed decline suggests a substantial global reduction.

In contrast, gentoo penguin numbers at Harmony Point increased between the 1960s and
the 1990s and appear to have stabilized in recent decades (Croxall and Kirkwood 1979;
Silva et al. 1998). As a species with greater dietary flexibility than other pygoscelids, gentoo
penguins have demonstrated a higher capacity to cope with environmental change in the
region (Miller et al. 2009; McMahon et al. 2015), allowing them to maintain or expand
populations while sympatric species decline (Talis et al. 2023). For the Toe colony, where no
previous data was available, our results serve as a baseline for future trend estimations.

Historical data for Antarctic shags from 1964 to the present indicate substantial fluctuations
in colony size over time, with the current estimate of 127 nests representing the highest
value recorded to date, pattern that is consistent with previous studies reporting high spatial

and temporal variability in Antarctic shag populations throughout the Antarctic Peninsula region (Casaux and Barrera-Oro 2006, 2016; Schrimpf et al. 2018).

Our results are consistent with previously reported nesting habitat segregation among penguin species, primarily structured by distance inland and elevation (Gallagher et al. 2025; Korczak Abshire et al. 2026), with gentoo penguins occupying more inland nesting areas than chinstrap penguins. This reflects species-specific nesting preferences and the greater capacity of gentoo penguins to exploit a broader range of ice-free habitats, as previously reported along the West Antarctic Peninsula (Cole et al. 2019; Herman et al. 2020). The preferential use of inland areas by gentoo penguins may facilitate colonization of newly exposed habitats as ice-free land expands under ongoing regional warming (Lee et al. 2022). Such habitat segregation has important implications for future species distributions and potential interspecific interactions in rapidly changing Antarctic coastal ecosystems.

Satellite imagery provides a valuable broad-scale framework for monitoring seabird populations; however, its application at local scales is constrained by image resolution, atmospheric conditions, and classification uncertainty (Román et al. 2022; Hilton et al. 2024). Using medium-resolution satellite imagery from 2014, (Strycker et al. 2020) estimated 3,000 and 10,000 nests at Harmony Point and the Toe, respectively, whereas our drone-based survey recorded substantially higher nest numbers at both sites. The differences found are likely due to methodological limitations of satellite data rather than true population change, since Strycker et al. (2020) could not retrieve good quality images for those sites. Together, satellite imagery and UAV surveys provide complementary information especially in remote areas, with satellite data supporting broad-scale monitoring and drone-based assessments delivering the fine-scale accuracy required for reliable population models and future seabird assessments.

ASPAs such as Harmony Point play a crucial role in conserving seabird hotspots by providing protected refuges where direct human disturbances, including tourism and research station infrastructure, are minimized. These areas contribute to local ecosystem stability and the maintenance of species interactions within sensitive Antarctic habitats, as shown for benthic communities (Moya et al. 2025). However, ASPAs alone cannot mitigate large-scale threats such as the ongoing warming , pollution, climate change and fisheries, which affect seabirds at regional scales. Therefore, maintenance of seabird populations requires conservation strategies that complement local protection with broader, ecosystem-based management approaches to safeguard Antarctic biodiversity.

References

Borowicz A, McDowall P, Youngflesh C, et al (2018) Multi-modal survey of Adélie penguin mega-colonies reveals the Danger Islands as a seabird hotspot. Sci Rep 8:3926. https://doi.org/10.1038/s41598-018-22313-w

Casaux R, Barrera-Oro E (2016) Linking population trends of Antarctic shag (Phalacrocorax bransfieldensis) and fish at Nelson Island, South Shetland Islands (Antarctica). Polar Biol 39:1491–1497. https://doi.org/10.1007/s00300-015-1850-5

Casaux R, Barrera-Oro E (2006) Shags in Antarctica: their feeding behaviour and ecological role in the marine food web. Antarct Sci 18:3–14. https://doi.org/10.1017/S0954102006000010

Clucas GV, Dunn MJ, Dyke G, et al (2014) A reversal of fortunes: Climate change "winners"
and "losers" in Antarctic Peninsula penguins. Sci Rep 4:1–7.
https://doi.org/10.1038/srep05024

Cole TL, Dutoit L, Dussex N, et al (2019) Receding ice drove parallel expansions in
Southern Ocean penguins. Proc Natl Acad Sci U S A 116:26690–26696.
https://doi.org/10.1073/pnas.1904048116

Croxall JP, Kirkwood ED (1979) The distribution of penguins on the Antarctic Peninsula and
Islands of the Scotia Sea. British Antarctic Survey

ESRI (2026) ArcGIS Pro Version 3.3.2

Gallagher KL, Herman R, Walton K, et al (2025) *Pygoscelis* penguins prefer different
oceanographic and terrestrial habitats during the austral summer. Ecosphere 16:e70091.
https://doi.org/10.1002/ecs2.70091

Harris CM, Carr R, Lorenz K, Jones S (2015) Important Bird Areas in Antarctica. BirdLife
International and Environmental Research & Assessment Ltd., Cambridge

Herman R, Borowicz A, Lynch M, et al (2020) Update on the global abundance and
distribution of breeding Gentoo Penguins (Pygoscelis papua). Polar Biol 43:1947–1956.
https://doi.org/10.1007/s00300-020-02759-3

Hilton M, Humphries GR, Nichol C, et al (2024) A Comparison of Uncrewed Aerial Vehicle
and On-Ground Surveys of Penguin Populations on the Antarctic Peninsula

Hinke JT, Alvstad KF, Feinberg SC, et al (2025) First full census in 45 years of a large
colony of breeding penguins at False Round Point, King George Island. Polar Biol 48:103.
https://doi.org/10.1007/s00300-025-03428-z

Humphries GRW, Naveen R, Schwaller M, et al (2017) Mapping Application for Penguin
Populations and Projected Dynamics (MAPPPD): data and tools for dynamic management
and decision support. Polar Rec 53:160–166. https://doi.org/10.1017/S0032247417000055

Korczak☐Abshire M, Jaworska KE, Pabjanek P, et al (2026) UAV ☐based terrain metrics
reveal species☐specific nest☐site preferences in *Pygoscelis* penguins. Ibis ibi.70030.
https://doi.org/10.1111/ibi.70030

Krüger L (2026) Seabird Colony Nest Number Error Estimation.
https://github.com/drlucaskruger/Seabird_Colony_Nest_Numbers.git

Krüger L (2023) Decreasing Trends of Chinstrap Penguin Breeding Colonies in a Region of
Major and Ongoing Rapid Environmental Changes Suggest Population Level Vulnerability.
Diversity 15:327. https://doi.org/10.3390/d15030327

Krüger L, Huerta MF, Santa Cruz F, Cárdenas CA (2021) Antarctic krill fishery effects over
penguin populations under adverse climate conditions: Implications for the management of
fishing practices. Ambio 50:560–571. https://doi.org/10.1007/s13280-020-01386-w

Krüger L, Pizarro EJ, Santa Cruz F, et al (2026) Aerial Survey of Harmony Point and The
Toe (ASPA 133, Nelson Island , Antarctica) for Pygoscelis penguins and Antarctic Shag
counts. Zenodo. https://doi.org/10.5281/zenodo.19470477

Lee JR, Waterman MJ, Shaw JD, et al (2022) Islands in the ice: Potential impacts of habitat
transformation on Antarctic biodiversity. Glob Change Biol 28:5865–5880.
https://doi.org/10.1111/gcb.16331

Lee WY, Park S, Kim KW, et al (2021) Inter-Specific and Intra-Specific Competition of Two
Sympatrically Breeding Seabirds, Chinstrap and Gentoo Penguins, at Two Neighboring
Colonies. Anim Open Access J MDPI 11:482. https://doi.org/10.3390/ani11020482

Lynch HJ, Naveen R, Trathan PN, Fagan WF (2012) Spatially integrated assessment
reveals widespread changes in penguin populations on the Antarctic Peninsula. Ecology
93:1367–1377. https://doi.org/10.1890/11-1588.1

Matthias Gamer, Jim Lemon, Puspendra Singh (2025) irr: Various Coefficients of Interrater
Reliability and Agreement. 0.84.1

McMahon KW, Polito MJ, Abel S, et al (2015) Carbon and nitrogen isotope fractionation of
amino acids in an avian marine predator, the gentoo penguin (*Pygoscelis papua*). Ecol
Evol 5:1278–1290. https://doi.org/10.1002/ece3.1437

Miller AK, Karnovsky NJ, Trivelpiece WZ (2009) Flexible foraging strategies of gentoo
penguins Pygoscelis papua over 5 years in the South Shetland Islands, Antarctica. Mar
Biol 156:2527–2537. https://doi.org/10.1007/s00227-009-1277-z

Morley SA, Abele D, Barnes DKA, et al (2020) Global Drivers on Southern Ocean
Ecosystems: Changing Physical Environments and Anthropogenic Pressures in an Earth
System. Front Mar Sci 7:547188. https://doi.org/10.3389/fmars.2020.547188

Moya F, Schwob G, Ugas-Bravo N, et al (2025) Human footprint alters morphological traits
and gut microbiome assembly of Antarctic sea urchins. Mar Pollut Bull 220:118448.
https://doi.org/10.1016/j.marpolbul.2025.118448

Oosthuizen WC, Krüger L, Jouanneau W, Lowther AD (2020) Unmanned aerial vehicle
(UAV) survey of the Antarctic shag (Leucocarbo bransfieldensis) breeding colony at
Harmony Point, Nelson Island, South Shetland Islands. Polar Biol 43:187–191.
https://doi.org/10.1007/s00300-019-02616-y

R Core Team (2024) R: A Language and Environment for Statistical Computing

Ramos JA, Pereira L (2022) Seabird Biodiversity and Human Activities, 1st edn. CRC Press,
Boca Raton

Ratcliffe N, Guihen D, Robst J, et al (2015) A protocol for the aerial survey of penguin
colonies using UAVs. J Unmanned Veh Syst 3:95–101. https://doi.org/10.1139/juvs-2015-
0006

Román A, Navarro G, Caballero I, Tovar-Sánchez A (2022) High-spatial resolution UAV
multispectral data complementing satellite imagery to characterize a chinstrap penguin
colony ecosystem on deception island (Antarctica). GIScience Remote Sens 59:1159–
1176. https://doi.org/10.1080/15481603.2022.2101702

Salmerón N, Belle S, Santa Cruz F, et al (2023) Contrasting environmental conditions
precluded lower availability of Antarctic krill affecting breeding chinstrap penguins in the
Antarctic Peninsula. Sci Rep 13:5265. https://doi.org/10.1038/s41598-023-32352-7

Santa Cruz F, Krüger L (2023) Breeding Population and Nesting Habitat of Skuas in the
Harmony Point Antarctic Specially Protected Area. Diversity 15:.
https://doi.org/10.3390/d15050638

Savoca MS, Kumar M, Sylvester Z, et al (2024) Whale recovery and the emerging human-
wildlife conflict over Antarctic krill. Nat Commun 15:7708. https://doi.org/10.1038/s41467-
024-51954-x

Schrimpf M, Naveen RON, Lynch HJ (2018) Population status of the Antarctic shag
Phalacrocorax (atriceps) brans fi eldensis. Antarct Sci 9:151–159.
https://doi.org/10.1017/S0954102017000530

Silva MP, Favero M, Casaux R, Baroni A (1998) The status of breeding birds at Harmony
Point, Nelson Island, Antarctica in summer 1995/96. Mar Ornithol 26:75–78

Stone TC, Davis KJ (2025) Using unmanned aerial vehicles to estimate body volume at
scale for ecological monitoring. Methods Ecol Evol 16:317–331.
https://doi.org/10.1111/2041-210X.14457

Strycker N, Wethington M, Borowicz A, et al (2020) A global population assessment of the
Chinstrap penguin (Pygoscelis antarctica). Sci Rep 10:1–11.
https://doi.org/10.1038/s41598-020-76479-3

Talis EJ, Che-Castaldo C, Hart T, et al (2023) Penguindex: a Living Planet Index for
Pygoscelis species penguins identifies key eras of population change. Polar Biol.
https://doi.org/10.1007/s00300-023-03148-2

Trivelpiece WZ, Hinke JT, Miller AK, et al (2011) Variability in krill biomass links harvesting
and climate warming to penguin population changes in Antarctica. Proc Natl Acad Sci
108:7625–7628. https://doi.org/10.1073/pnas.1016560108

Watters GM, Hinke JT, Reiss CS (2020) Long-term observations from Antarctica
demonstrate that mismatched scales of fisheries management and predator-prey
interaction lead to erroneous conclusions about precaution. Sci Rep 10:2314.
https://doi.org/10.1038/s41598-020-59223-9

Fig 1. Location of the study area and distribution of seabird colonies. (a) Global context of the South
Shetland Islands. (b) Position of Nelson Island within the Antarctic Peninsula. (c) Detail of Antarctic
Specially Protected Area (ASPA) No. 133, showing the locations of Harmony Point (HP) and the Toe
(TT) within the dashed squares. (d, e) Distribution of gentoo penguin (*Pygoscelis papua*) (red colour)
and chinstrap penguin (*Pygoscelis antarcticus*) (blue colour) colonies at Harmony Point (d) and the
Toe (e). (f) Distribution of Antarctic shag (*Leucocarbo bransfieldensis*) colonies at Harmony Point
(orange colour). All distributions are mapped on high-resolution orthomosaics derived from drone
imagery.

Figure 2. Colony size ± 10% or ± 50% (for penguins counts before 1985) error for chinstrap penguin (a,b), *Pygoscelis antarcticus* (blue filled shapes), gentoo penguin (c), *Pygoscelis papua* (red empty shapes), and Antarctic shag (*Leucocarbo bransfieldensis*) (d), measured as number of nests, in Harmony Point (HP, circle) and the Toe (TT, triangles), ASPA 133, Nelson Island. Year 2024 corresponds to data from this study, as previous years are from previous publications (please see methods section).

a. Chinstrap penguins HP

b. Chinstrap penguins TT

c. Gentoo Penguins

d. Antarctic shag