- 1 | Eberhart-Phillips et al. Origins and consequences of adult sex ratios
- 1 Demographic causes of adult sex ratio variation and their consequences for
- 2 parental cooperation

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40 Abstract:

41 The adult sex ratio (ASR) is a fundamental concept in population biology, sexual selection, and social evolution. However, it remains unclear which demographic processes generate 42 43 ASR variation and how biases in ASR in turn affect social behaviour. Here, we evaluate the demographic mechanisms shaping ASR and their consequences for parental cooperation 44 45 using detailed survival, fecundity, and behavioural data on 6,119 individuals from six wild shorebird populations exhibiting flexible parental strategies. We show that these closely 46 47 related populations express strikingly different ASRs, despite having similar ecologies and 48 life histories, and that ASR variation is largely driven by sex differences in the apparent 49 survival of juveniles. Furthermore, families in populations with biased ASRs were 50 predominantly tended by a single parent, suggesting that parental cooperation breaks down 51 with unbalanced sex ratios. Taken together, our results indicate that sex biases emerging 52 during early life have profound consequences for social behaviour.

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53 Main text:

Sex ratio variation is a fundamental component of life-history evolution. At conception, birth, 54 and adulthood, the ratios of males to females have long been hypothesized by evolutionary 55 56 biologists and human demographers as catalysts for social behaviour and population dynamics^{1,2}. In particular, the adult sex ratio (ASR) exhibits remarkable variation throughout 57 58 nature, with birds and mammals tending to have male-biased and female-biased ASRs, 59 respectively³. Recent studies also show extreme shifts in ASR due to climate change in fish⁴, amphibians⁵, and dioecious plants⁶. By influencing mate availability, ASR bias can alter 60 61 social behaviour with divorce, infidelity, and parental antagonism being more frequent in sex-biased populations^{7,8}. Moreover, in human societies, ASR variation is linked to economic 62 decisions, community violence, and disease prevalence⁹⁻¹¹. Yet despite the widespread 63 occurrence of ASR bias and its significance in evolutionary ecology and social science, the 64 demographic source(s) of ASR variation and their ramifications for social behaviour remain 65 unclear¹². 66

67 Sex ratio theory is concerned with the adaptive consequences of sex-biased parental allocation to offspring^{13,14}, with the processes generating sex ratio bias after birth receiving 68 less theoretical and empirical attention¹⁵. Here, we use a demographic pathway model to 69 70 quantify ASR variation among avian populations and to determine whether this variation is 71 predominantly caused by sex biases at birth, during juvenile development, or in adulthood. We parameterized our model with detailed individual-based life history data from Charadrius 72 plovers - small ground-nesting shorebirds that occur worldwide. Plovers exhibit remarkable 73 74 diversity and plasticity in breeding behaviour with sex roles during courtship, mating, and parental care varying appreciably among populations both between and within species^{16,17}. 75 76 This behavioural variation, coupled with their tractability in the field (Supplementary Video

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- 1), allowed us to explore the sources and significance of demographic sex biases among
- 78 closely related wild populations in the light of social evolution.

79 **Results and Discussion**

Over a total of 43 observational years of fieldwork, we monitored the survival, fecundity, and 80 81 breeding behaviour of 6,119 individually marked plovers from six populations of five closely related species worldwide (Fig. 1a). We then employed two-sex stage-structured population 82 matrix models to derive estimates of ASR at equilibrium from stage- and sex-specific 83 demographic rates of annual survival and reproduction (Fig. 1b)¹⁸. For each population, the 84 85 numbers of male and female progeny in our model depended on modal clutch size and 86 hatching sex ratio derived from our field data. Mark-recapture methods were used to estimate 87 the apparent survival of juveniles and adults while accounting for sex differences in detection 88 probability (the term "apparent survival" indicates that mortality cannot be disentangled from permanent emigration)¹⁹. Fecundity was derived from a mating function that depended on the 89 90 extent of polygamy observed in each population and the frequency of available mates (see 91 *Methods* for details).

The hatching sex ratio, based on 1,139 hatchlings from 503 families, did not deviate significantly from parity in any of the populations (Fig. 2a). Conversely, sex biases in apparent survival varied considerably within and among species and, in most populations, juvenile survival was more biased than adult survival – either towards males or females (Fig. 2a). Taken together, these sources of demographic sex bias rendered significant deviations in ASR from parity for three populations (two male-biased and one female-biased; Fig. 2b). Matrix models provide a flexible analytical environment to decompose the feedbacks

between state-dependent vital rates and population response – an important method used in
 conservation biology for understanding life-history contributions to population growth and

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viability²⁰. In our case, we modified this approach to assess the relative contributions of sex 101 allocation and sex-specific survival on ASR bias¹⁸. We found that sex biases in apparent 102 103 survival during the juvenile stage contributed the most to sex ratio bias of the adult 104 population: in populations with significantly skewed ASR, sex biases in juvenile survival 105 contributed on average 7.2 times more than sex biases in adult apparent survival and 24.6 106 times more than sex biases at hatching (Supplementary Fig. 1). Moreover, variation in hatching sex ratio had no effect on ASR – a result that provides empirical support for 107 Fisher's¹³ prediction of unbiased sex allocation regardless of sex-biased survival of 108 109 independent young or adults.

110 In species where both parents have equal caring capabilities, the desertion of either parent is often influenced by the availability of potential mates²¹ – parental care by the 111 abundant sex is expected to be greater than that of the scarcer sex due to limited future 112 113 reproductive potential²². Detailed behavioural observations of 471 plover families revealed 114 high rates of parental desertion in populations with biased ASRs, whereas desertion was rare 115 in unbiased populations (Fig. 3). We evaluated our *a priori* prediction of a quadratic 116 relationship between parental cooperation and ASR variation using a regression analysis incorporating a bootstrap procedure that acknowledged uncertainty in our estimates of ASR 117 118 and parental care (see Methods for details). We found that families in male- or female-biased populations tended to express higher rates of parental desertion, while unbiased populations 119 120 were more likely to exhibit parental cooperation (Fig. 3a). This is supported by experimental 121 evidence of sex-biased mating opportunities in three of the populations studied here (Supplementary Fig. 2; ref 23). Moreover, the relationship between parental cooperation and 122 local ASR bias was apparent in our within-species contrast of C. alexandrinus: the unbiased 123 124 Cape Verde population exhibited a higher rate of parental cooperation than the male-biased population in Turkey (Fig. 3a). Counterintuitively, we also found a high rate of male-only 125

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care in *C. pecuarius* despite ASR being female-biased (Fig. 3b), although in line with
expectations, *C. pecuarius* also showed the highest proportion of female-only care among our
studied populations (Fig. 3b, Supplementary Fig. 3b). This provides partial support for the
notion that breeding strategies may respond flexibly to local mating opportunities provided
by ASR bias, while also suggesting that other factors may play a role, such as the energetic
costs of egg production imposed on females or because of sex differences in parental
quality²⁴ or the age at maturation²⁵.

There are several ways in which sex-biased juvenile survival could potentially lead to 133 134 a skewed ASR. Natal dispersal rates may differ between the sexes, as is typical of many birds²⁶, which could contribute to our estimates of sex-biased apparent survival. Genetic 135 136 studies of several of the populations presented here are partially consistent with this 137 hypothesis, as island populations of C. alexandrinus and the endemic C. thoracicus have reduced gene flow relative to comparable mainland populations^{27,28}. However, sex-biased 138 juvenile survival in plovers has been reported elsewhere, even after accounting for 139 dispersal^{29,30}, implying that sex-differences in mortality are at least partly driven by intrinsic 140 factors, conceivably for example via genotype-sex interactions 31,32 . An alternative but not 141 necessarily mutually exclusive explanation is that males and females may differ in their 142 143 premature investment into reproductive traits, which could inflict survival costs for the larger or more ornamented sex³³. Although sexual dimorphism among adults is negligible in the 144 species we studied^{34,35}, sex-specific ontogeny does appear to vary in three of these 145 146 populations. In male-biased populations of *C. nivosus* and *C. alexandrinus*, female hatchlings are smaller and grow more slowly than their brothers during the first weeks of life, whereas 147 juveniles of the unbiased C. alexandrinus population exhibit no such sex-specific differences 148 during early development³⁴. 149

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150	The association between sex-specific demography and breeding system evolution
151	represents a causality dilemma because of the feedback that parental strategies impose on
152	ASR bias and vice versa ³⁶ . On the one hand, mating competition and parental care may entail
153	costs via sexual selection that could drive differential survival of males over females and
154	have knock-on effects on ASR ³⁷ . On the other hand, sex-specific survival creates unequal
155	mating opportunities via ASR that may influence mating patterns and parenting strategies ³⁷ .
156	Our study provides empirical support for the latter – sex biases emerge prior to sexual
157	maturity, suggesting that this evolutionary feedback loop is catalysed by intrinsic early-life
158	demographic variation. Moreover, our results add to the growing evidence of unbiased birth
159	sex ratios in nature ¹⁵ and provide a comprehensive empirical test of Fisher's ¹³ original
160	prediction that influential sex biases arise in life-history stages beyond parental control. By
161	unravelling the demographic foundations of adult sex ratio bias and their consequences for
162	parental cooperation, we hope to stimulate future studies to understand the complex
163	relationship between evolutionary demography and behavioural ecology.

164 Methods

165 Field and laboratory methods

166 We studied five *Charadrius* species comprising six populations from four sites worldwide

167 (Fig. 1a, Supplementary Table 1). In Mexico, we monitored the snowy plover (C. nivosus) at

168 Bahía de Ceuta, a subtropical lagoon on the Pacific coast. In Madagascar, we monitored the

169 Kittlitz's plover (C. pecuarius), white-fronted plover (C. marginatus), and the endemic

170 Madagascar plover (*C. thoracicus*), all of which breed sympatrically at a saltwater marsh near

- 171 the fishing village of Andavadoaka. Lastly, we monitored the Kentish plover (C.
- 172 *alexandrinus*) at two independent populations located at Lake Tuzla in southern Turkey and
- 173 at Maio in Cape Verde. The Mexico and Madagascar populations were monitored over a

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seven-year period, whereas the Turkey and Cape Verde populations were monitored over six
and nine years respectively, thus totalling 43 years of data collection (Supplementary Table
1).

177 At each location, we collected mark-recapture and individual reproductive success data during daily surveys over the entire breeding season that typically spanned three to four 178 179 months after a region's rainy season. Funnel traps were used to capture adults on broods or 180 nests³⁸. We assigned individuals to a unique colour combination of darvic rings and an alphanumeric metal ring, allowing the use of both captures and non-invasive resigntings to 181 182 estimate survival. Broods were monitored every five days on average to assess daily survival 183 and identify tending parents. During captures, approximately 25–50 µL of blood was sampled 184 from the metatarsal vein of chicks and the brachial vein of adults for molecular sex typing with two independent markers Z-002B³⁹ and Calex-31, a microsatellite marker on the W 185 186 chromosome⁴⁰. Details of PCR conditions are given in ref. 34.

187 *Quantifying sex roles*

We evaluated sex role variation by summarizing for each population the proportion of all 188 189 families that were attended bi-parentally or uni-parentally by a male or female. We restricted 190 our field observations to include only broods that were at least 20 days old. Young chicks are 191 attended by both parents in all populations although, as broods get older, male or female parents may desert the family⁴¹. As chicks typically fledge around 25 days of age, we 192 193 therefore choose broods of between 20 and 25 days of age to quantify parental care given that 194 at this age many parents already deserted the family but some still attend the young. 195 Furthermore, we restricted these data to include only broods that had at least two sightings 196 after 20 days. Given these criteria, our dataset consisted of 471 unique families distributed throughout the six populations and pooled across all years of study (Supplementary Table 2). 197

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198 To account for surveyor oversight while recording tending parents (e.g., observing only one 199 parent when two were present), we took a conservative approach by assigning a bi-parental status to families that had both uni-parental and bi-parental observations after the 20th day. In 200 summary, desertion was most common in C. nivosus and C. pecuarius, whereas bi-parental 201 care was most common in C. alexandrinus (Cape Verde), C. thoracicus, and C. marginatus 202 203 (Fig. 3b, Supplementary Table 2). The C. alexandrinus population in Turkey had 50% bi-204 parental and 50% desertion (Fig. 3b, Supplementary Table 2). C. pecuarius had the highest incidence of male desertion (20%; Fig. 3b, Supplementary Table 2). To acknowledge 205 206 uncertainty in these parental care proportions given variation in sample size, we used a 207 method that estimated simultaneous 95% confidence intervals according to a multinomial distribution⁴² (Supplementary Table 2). 208

209 Estimation of sex- and stage-specific survival

Our structured population model considered sex-specific survival during two key stage classes in life history: juveniles and adults (Fig. 1b). The juvenile stage was defined as the one-year transition period between hatching and recruitment into the adult population. The adult stage represented a stasis stage in which individuals were annually retained in the population.

We used mark-recapture models to account for sex, stage, and temporal variation in encounter (p) and apparent survival (ϕ) probabilities as they allow for imperfect detection of marked individuals during surveys and the inclusion of individuals with unknown fates⁴⁰. We use the term "apparent survival" as true mortality cannot be disentangled from permanent emigration in this framework¹⁹. We used Cormack-Jolly-Seber models to estimate juvenile and adult survival, with one-year encounter intervals. Juvenile and adult survival models were constructed from design matrices that included sex, year, and stage as factors.

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222	Since we were primarily interested in stage- and sex-specific variation in survival, all
223	models included a $\phi \sim sex * stage$ component. Our model selection thus evaluated the best
224	structure explaining variation in detection probability by comparing all interactions between
225	sex, year, and stage (e.g., $p \sim sex * year * stage$). We constructed survival models with the R
226	package "RMark" ⁴³ and estimated demographic parameters via maximum likelihood
227	implemented in program MARK ⁴⁴ . We evaluated whether our data were appropriately
228	dispersed (i.e. $c-hat \le 3$; ref. 19) by employing the "median c-hat" goodness-of-fit bootstrap
229	simulation in program MARK ⁴⁴ .

230 *Estimating hatching sex ratios*

To account for potential sex-biases arising prior to the juvenile stage (i.e. sex allocation or 231 sex-specific embryo mortality), we tested whether the hatching sex ratio deviated 232 significantly from parity. Each population was analysed separately using a general linear 233 mixed effect model with a binomial error distribution and a logit link function (R package 234 235 "lme4"⁴⁵). In this model, the response variable was chick sex, the fixed effect was the 236 intercept, and brood identifier was included as a random factor to control for the non-237 independence of siblings from the same nest. Because plover chicks are precocial, post-hatch 238 brood mixing can occur. Consequently, our dataset for analysing hatching sex ratio included 239 only complete broods (i.e. with no missing chicks) that were captured at the nest on the same 240 day of hatching (503 unique families with 1139 chicks, Supplementary Table 3). The fixedeffect intercepts of all populations were not significantly different from zero, indicating that 241 242 hatching sex ratios did not deviate from parity (Fig. 2a).

243 Matrix model structure

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244 We built two-sex post-breeding matrix models for each plover population that incorporated

two annual transitions denoting juveniles and adults (Fig. 1b). The projection of the matrix

for one annual time step (*t*) is given by:

$$\mathbf{n}_t = \mathbf{M}\mathbf{n}_{t-1} \tag{Eq. 1}$$

248 where *n* is a 4×1 vector of the population distributed across the two life stages and two

sexes:

250
$$\boldsymbol{n} = \begin{bmatrix} \boldsymbol{\varphi} \text{ Juvenile} \\ \boldsymbol{\varphi} \text{ Adult} \\ \boldsymbol{\sigma} \text{ Juvenile} \\ \boldsymbol{\sigma} \text{ Adult} \end{bmatrix}$$
(Eq. 2)

and **M** is expressed as a 4×4 matrix:

252
$$\mathbf{M} = \begin{bmatrix} 0 & R_{\varphi} (1-\rho) & 0 & R_{\sigma} (1-\rho) \\ \phi_{\varphi J} & \phi_{\varphi A} & 0 & 0 \\ 0 & R_{\varphi} \rho & 0 & R_{\sigma} \rho \\ 0 & 0 & \phi_{\sigma J} & \phi_{\sigma A} \end{bmatrix}$$
(Eq. 3)

where transition probabilities (ϕ) between life stages are the apparent survival rates of female 253 254 (\mathcal{Q}) and male (\mathcal{A}) juveniles (\mathcal{J}) and adults (\mathcal{A}) . The hatching sex ratio (ρ) describes the probability of hatchlings being either male (i.e. ρ) or female (i.e. $1 - \rho$), and was estimated for 255 256 each population from our field data (see above). *Per capita* reproduction of females (R_{\circ}) and males (R_{β}) is expressed through sex-specific mating functions used to link the sexes and 257 258 produce progeny for the following time step of the model given the relative frequencies of each sex²⁰. We used the harmonic mean mating function which accounts for sex-specific 259 frequency dependence⁴⁶: 260

261
$$R_{\mathcal{P}}(n_{\mathcal{P}}, n_{\mathcal{P}}) = \frac{kn_{\mathcal{P}}}{n_{\mathcal{P}} + n_{\mathcal{P}}h^{-1}}, \quad R_{\mathcal{P}}(n_{\mathcal{P}}, n_{\mathcal{P}}) = \frac{kn_{\mathcal{P}}}{n_{\mathcal{P}} + n_{\mathcal{P}}h^{-1}}$$
(Eq. 4)

where k is the modal clutch size (3 in *C. nivosus*, *C. alexandrinus*, and *C. marginatus*, and 2
in *C. thoracicus* and *C. pecuarius*), h is an index of the annual number of mates acquired per

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female (i.e. mating system, see below), and n_{2} and n_{3} are the densities of females and males,

respectively, in each time step of the model.

266 *Quantifying mating system*

- 267 Demographic mating functions are traditionally expressed from the perspective of males⁴⁶,
- whereby *h* is the average harem size (number of female mates per male). Under this
- definition, h > 1 signifies polygyny, h = 1 monogamy, and h < 1 polyandry; ref. 47).
- 270 Although both sexes can acquire multiple mates in a single breeding season, within-season
- 271 polygamy is typically female biased in plovers. Thus, in accordance with the predominantly
- 272 polyandrous or monogamous mating systems seen across these six populations, h was derived
- 273 from the average annual number of mates per female (μ):

274
$$\mu = \frac{1}{n} \sum_{i=1}^{n} \frac{m_i}{b_i}$$
 (Eq. 5)

where, *n* is the total number of females in a given population, *b* is the total number of years female *i* was seen breeding, and *m* is the total number of mating partners female *i* had over *b* years. Thus, if μ was less than or equal to one, females tended to have only one mating partner annually, and *h* was set to 1. Alternatively, if μ was greater than 1, females were polygamous and *h* was calculated as the inverse of μ :

280
$$h = \begin{cases} 1, \ \mu \le 1\\ \mu^{-1}, \ \mu > 1 \end{cases}$$
 (Eq. 6)

Our dataset to estimate *h* for each population only included females for which we were
confident of the identity of their mates, and had observed them in at least two reproductive
attempts. In summary,
$$\mu$$
 varied among populations (Supplementary Fig. 4), with *C. nivosus*
(*h* = 0.82), *C. alexandrinus* (Turkey; *h* = 0.85), and *C. pecuarius* (*h* = 0.86) having
polyandrous mating systems and *C. alexandrinus* (Cape Verde), *C. thoracicus*, and *C.*
marginatus all having monogamous mating systems (i.e. *h* = 1).

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287 Estimation of the adult sex ratio

310

288 We estimated ASR from the stable stage distribution (w) of the two-sex matrix model:

289
$$ASR = \frac{w_{\heartsuit A}}{w_{\heartsuit A} + w_{\heartsuit A}}$$
(Eq. 7)

where $w_{\oslash A}$ and $w_{\diamondsuit A}$ provide the proportion of the population composed of adult males and 290 291 females, respectively, at equilibrium. To evaluate uncertainty in our estimate of ASR due to 292 sampling and process variation in our apparent survival parameters, we implemented a 293 bootstrapping procedure in which each iteration: (i) randomly sampled our mark-recapture 294 data with replacement, (ii) ran the survival analyses described above, (iii) derived stage- and 295 sex-specific estimates of apparent survival based on the model with the lowest AIC_{C} (i.e. $\Delta AIC_{c} = 0$: Supplementary Fig. 5), (iv) constructed the matrix model (Eq. 3) of these 296 297 estimates, (v) derived the stable stage distribution through simulation of 1,000 time steps, then (vi) derived ASR from the stable stage distribution at equilibrium on the 1,000th time 298 299 step. This approach ensured that parameter correlations within the matrix were retained for 300 each bootstrap and it also accounted for non-linearity in the mating function. We ran 1,000 301 iterations and evaluated the accuracy of our ASR estimate by determining the 95% 302 confidence interval of its bootstrapped distribution. Note that our method estimated ASR as 303 the asymptotic value predicted under the assumption that each population was at equilibrium 304 and thus we could not evaluate inter-annual variation in asymptotic ASR. Nonetheless, our 305 model-derived ASR estimate of the C. nivosus population falls within annual count-based ASR estimates of this population⁴⁸, providing support that our method is robust. Count-based 306 307 estimates of ASR from the remaining populations in our study are unfortunately 308 uninformative due to our limited sample of marked individuals with known sex. 309 Our mark-recapture analysis was based on the encounter histories of 6,119 uniquely

marked and molecularly sexed individuals (Supplementary Table 4). After implementing the

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bootstrap procedure, we found that variation in the encounter probabilities of juveniles and
adults was best explained by sex, year, and age in *C. nivosus*, *C. alexandrinus* (Turkey) and *C. pecuarius* (Supplementary Fig. 5). Encounter probability was best explained by age and
year in *C. alexandrinus* (Cape Verde) and *C. marginatus* (Supplementary Fig. 5). In *C. thoracicus*, encounter probability was best explained by sex and year (Supplementary Fig. 5).

316 Our mark-recapture data were not over-dispersed (Supplementary Table 4).

317 Life table response experiment of ASR contributions

318 Perturbation analyses provide information about the relative effect that each component of a 319 matrix model has on the population-level response, in our case ASR. To assess how 320 influential sex biases in parameters associated with each of the three life stages were on ASR 321 dynamics, we employed a life-table response experiment (LTRE). A LTRE decomposes the difference in response between two or more "treatments" by weighting the difference in 322 parameter values by the parameter's contribution to the response (i.e. its sensitivity), and 323 summing over all parameters²⁰. We compared the observed scenario (**M**), to a hypothetical 324 scenario (M_0) whereby all female survival rates were set equal to the male rates (or vice 325 *versa*) and the hatching sex ratio was unbiased (i.e. $\rho = 0.5$). Thus, our LTRE identifies the 326 drivers of ASR bias by decomposing the difference between the ASR predicted by our model 327 and an unbiased ASR¹⁸. 328

329 The contributions (C) of lower-level demographic parameters (θ) were calculated 330 following Veran and Beissinger¹⁸:

331 $C(\theta) = \left(\theta_{\heartsuit} - \theta_{\heartsuit}\right) \times \frac{\partial ASR}{\partial \theta}$ (Eq. 8)

332 where $\frac{\partial ASR}{\partial \theta}$ is the sensitivity of ASR to perturbations in the demographic rate θ in matrix **M**', 333 which is a reference matrix "midway" between the two scenarios¹⁸:

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334

$$\mathbf{M}' = \frac{\mathbf{M} + \mathbf{M}_0}{2} \tag{Eq. 9}$$

The two-sex mating function makes our model non-linear in the sense that the 335 336 projection matrix, and specifically the fertility elements (Eq. 4), depends on sex-specific 337 population structure. Perturbation analyses must therefore accommodate the indirect effects of parameter perturbations on population response via their effects on population structure. 338 339 such as the relative abundance of males and females which can affect mating dynamics and 340 fecundity. To estimate the sensitivities of the ASR to vital rate parameters, we employed 341 numerical methods that independently perturbed each parameter of the matrix, simulated the model through 1,000 time steps, and calculated ASR at equilibrium. This produced 342 parameter-specific splines from which $\frac{\partial ASR}{\partial \theta}$ could be derived. This approach appropriately 343 344 accounts for the non-linear feedbacks between vital rates and population structure, though it does not isolate the contribution of this feedback^{47,49}. 345

Our LTRE revealed that across all populations, sex differences in juvenile apparent
survival made the largest overall contribution to ASR bias (Supplementary Fig. 1). Likewise,
for all populations, sex biases at hatching and in mating system had negligible effects on
ASR variation (Supplementary Fig. 1).

350 Evaluating the association between ASR bias and parental cooperation

To test the relationship between ASR bias and parental cooperation, we conducted aregression analysis of the following quadratic model:

353
$$P_{a^{\gamma}} = \beta_0 + \beta_1 A + \beta_2 A^2 + \varepsilon$$
 (Eq. 11)

where $P_{\sigma} \neq \beta$ is the proportion of families exhibiting parental cooperation, β_i are the regression parameters (i.e. intercept and coefficient), *A* is the ASR, and ε is random error. We chose a quadratic model *a priori* as we expected maximum parental cooperation at unbiased ASR but minimum cooperation at both male- and female-biased ASRs (see inset in Fig. 3a).

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358 This relationship was assessed with a bootstrap procedure that incorporated uncertainty in our 359 estimates of ASR and parental care. Each iteration of the bootstrap (i) randomly sampled an ASR value from the 95% confidence interval of each population shown in Fig. 2b, (ii) 360 361 randomly sampled a parental care value from the truncated 95% confidence interval of each population shown in Supplementary Table 2, then (iii) fitted the regression model. We ran 362 363 1,000 iterations of the bootstrap and evaluated overall relationships by visualizing the central 364 tendency of the regressions. We also evaluated the relationship between ASR variation and 365 male-only or female-only care using a similar bootstrap procedure of the following models:

366
$$P_{\heartsuit} = \beta_0 + \beta_1^A + \varepsilon \quad , \qquad P_{\updownarrow} = \beta_0 - \beta_1^A + \varepsilon \quad (Eq. 12)$$

where P_{\triangleleft} and P_{\downarrow} are the proportions of families exhibiting male-only or female-only care, 367 368 respectively. In this case, we chose exponential models *a priori* as we expected a non-linear 369 increase in uni-parental care by the abundant sex under biased ASR (Supplementary Fig. 3a). 370 This analysis demonstrated that male-only care tended to be more common in populations with male-biased ASR (mean $\beta_1 = 0.682$ [-0.366, 1.555 95% CI]) and female-only care 371 tended to be more common in female-biased populations (mean $\beta_1 = -0.205$ [-0.502, 0.037 372 95% CI]; Supplementary Fig. 3b). However, the overall magnitude of the effect of ASR 373 374 variation on female-only care was less than that of male-only care.

375 *Data and code availability*

376 All of our modelling and statistical analyses were conducted using R version *Kite-eating*

377 *Tree*⁵⁰ with significance testing evaluated at $\alpha = 0.05$. We provide all computer code and

378 documentation as a PDF file written in RMarkdown (Supplementary Material A) together

379 with all the raw datasets needed to reproduce our modelling and analyses (Supplementary

380 Material B). These can be downloaded from our GitHub repository:

381 <u>https://github.com/leberhartphillips/Plover_ASR_Matrix_Modeling.</u>

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- 522 M.C.L, I.C.C., and T.S. planned and collected the field data. L.E.-P., C.K., J.I.H., and T.B.
- 523 performed or supervised the molecular sexing. L.E.-P., T.M., and O.K. implemented the
- 524 demographic modelling. L.E.-P. wrote the manuscript and the RMarkdown file. All authors
- 525 contributed substantially to revisions of the paper.

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526 Figure legends

527 Figure 1. Modelling the demographic pathway of adult sex ratio bias among plovers

worldwide. (a) Location of the six study populations. C. pecuarius, C. marginatus, and C. 528 529 thoracicus breed sympatrically in south-western Madagascar, whereas the two populations of C. alexandrinus are geographically disparate, inhabiting southern Turkey and the Cape Verde 530 531 archipelago. The studied C. nivosus population is located on the Pacific coast of Mexico. All populations inhabit saltmarsh or seashore habitats characterized by open and flat substrates. 532 533 (b) Schematic of the stage- and sex-specific demographic transitions of individuals from 534 hatching until adulthood and their contributions to the adult sex ratio (depicted here is C. 535 *nivosus*). The hatching sex ratio (ρ , proportion of male hatchlings) serves as a proxy for the 536 primary sex ratio and allocates progeny to the male or female juvenile stage. During the 537 juvenile ('juv') stage, a subset of this progeny will survive (ϕ) to recruit and remain as adults ('ad'). Dotted clusters illustrate how a cohort is shaped through these sex-specific 538 539 demographic transitions to derive the adult sex ratio (mortality indicated by grey dots). The 540 reproduction function, $R(n_{\beta}, n_{\circ})$, is dependent on mating system and the frequency of available mates (see Methods for details). 541

Figure 2. Inter- and intra-specific variation in sex-biased demography. (a) Hatching sex 542 ratios of successful clutches (proportion of chicks that are male) are shown as point estimates 543 $(\rho \pm 95\% \text{ CI}; \text{ left y-axis})$, and sex bias (i.e. difference between males and females) in annual 544 545 apparent survival rates of juveniles (ϕ_{iuv}) and adults (ϕ_{ad}) are shown as violin plots (right y-546 axis). Horizontal lines within violin plots indicate the median and interguartile ranges of the 547 bootstrapped estimates (see *Methods* for details). (b) Bootstrap distributions of the derived 548 ASRs based on the sex- and stage-specific rates shown in panel a. Vertical bars on the right side of histograms indicate the 95% CI of ASRs based on 1,000 iterations of the bootstrap 549

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550 (mean ASR [95% CI]: *C. nivosus* = 0.644 [0.499, 0.778], *C. alexandrinus* [Turkey] = 0.586

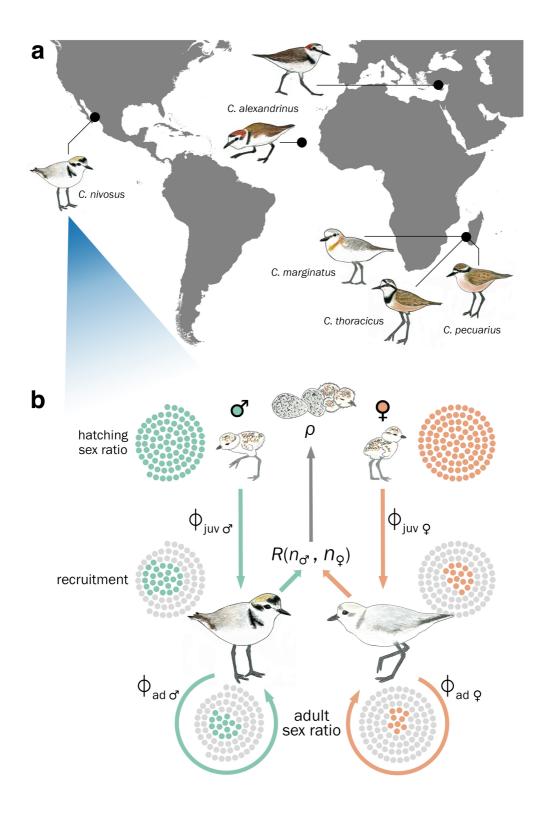
551 [0.510, 0.652], *C. alexandrinus* [Cape Verde] = 0.469 [0.396, 0.543], *C. thoracicus* = 0.421

552 [0.171, 0.681], *C. marginatus* = 0.430 [0.340, 0.522], *C. pecuarius* = 0.386 [0.289, 0.498].

- 553 **Figure 3. Relationship between parental cooperation and the adult sex ratio. (a)** Faint
- white lines illustrate each iteration of the bootstrap, which randomly sampled an adult sex
- 555 ratio and parental care estimate from each population's uncertainty distribution and fitted
- them to the *a priori* quadratic model (shown in inset, Eq. 11). (b) Proportion of monitored
- 557 plover families that exhibit parental cooperation (white) or single-parent care by males
- 558 (green) or females (orange). Sample sizes reflect number of families monitored per
- 559 population.

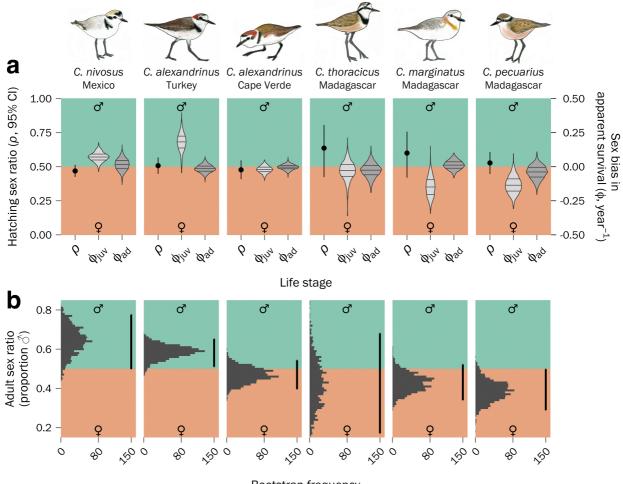
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Figure 1



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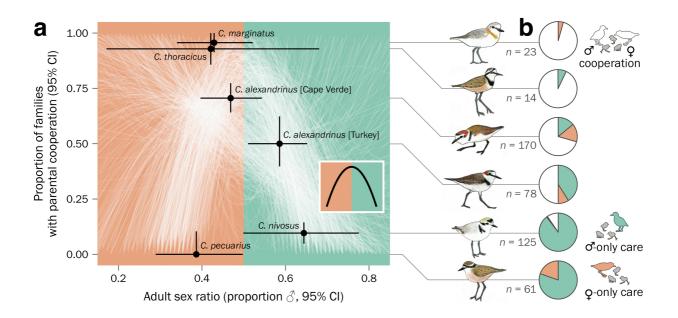
Figure 2



Bootstrap frequency

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Figure 3



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563 Supplementary Information

564 Supplementary Tables

565 Supplementary Table 1. Study site metadata: geographic location (see Fig. 1) and

566 duration of monitoring effort.

				Years
Species	Population	Latitude	Longitude	monitored
C. nivosus	Mexico	23°54'N	106°57'W	2006-2012
C. alexandrinus	Turkey	36°43'N	35°03'E	1996-2001
	Cape Verde	15°8'N	23°13'W	2007-2015
C. thoracicus	Madagascar	22°6'S	43°15'E	2009-2015
C. marginatus	Madagascar			2009-2015
C. pecuarius	Madagascar			2009-2015
				43 years

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567 Supplementary Table 2. Summary of parental care sex roles. Percentages reflect the

568 within-population proportion of families with a given parental care system.

~ •		Bi-parental	Female-only care	Male-only care	
Species	Population	[95% CI]	[95% CI]	[95% CI]	n
C. nivosus	Mexico	9% [5, 14]	1% [0, 6]	90% [85, 94]	125
C. alexandrinus	Turkey	50% [40, 62]	9% [0, 21]	41% [31, 53]	78
	Cape Verde	71% [64, 77]	15% [9, 22]	14% [8, 21]	170
C. thoracicus	Madagascar	93% [86, 100]	0% [0, 12]	7% [0, 36]	14
C. marginatus	Madagascar	96% [91, 100]	4% [0, 12]	0% [0, 7]	23
C. pecuarius	Madagascar	0% [0, 10]	20% [11, 30]	80% [72, 91]	61
				-	471

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569 Supplementary Table 3. Summary of hatching sex ratio data, where ρ is the average

- 570 hatching sex ratio (expressed as the proportion of hatchlings in a brood that are male) and
- 571 95% CIs are calculated using a binomial distribution.

Species	Population	N_{Families}	$N_{ m Hatchlings}$	ρ	95% CI
C. nivosus	Mexico	198	484	0.469	[0.425, 0.514]
C. alexandrinus	Turkey	102	262	0.508	[0.447, 0.568]
	Cape Verde	107	197	0.477	[0.408, 0.547]
C. thoracicus	Madagascar	11	22	0.636	[0.423, 0.807]
C. marginatus	Madagascar	13	30	0.600	[0.419, 0.757]
C. pecuarius	Madagascar	72	144	0.528	[0.446, 0.757]

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572 Supplementary Table 4. Sample size and over-dispersion summary of mark-recapture

573 dataset used to estimate apparent survival.

	Population	Juveniles ¹		Adults ²		Total		
Species		9	Q Q		8	individuals	Median ĉ	
C. nivosus	Mexico	438	388	221	212	1358	1.70	
C. alexandrinus	Turkey	310	293	557	504	1664	1.49	
	Cape Verde	377	383	254	213	1227	1.37	
C. thoracicus	Madagascar	38	56	83	68	245	2.72	
C. marginatus	Madagascar	76	96	99	95	366	1.31	
C. pecuarius	Madagascar	274	286	382	416	1358	1.77	
					_	6119		

574 ¹Individuals first marked as hatchlings (i.e. known age).

575 ²Individuals first marked as breeding adults (i.e. 1+ years old).

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576 Supplementary Figure legends

577 Supplementary Figure 1. Contributions of sex-specific parameters to adult sex ratio

578 **bias.** These results are based on a life-table response experiment (LTRE) that compared the

579 empirically-derived sex-specific model to hypothetical scenarios with no sex differences in

- 580 demographic rates (top panel: female-only rates, bottom panel: male-only rates). ASR is the
- 581 proportion of the adult population that is male, thus changes in female-biased parameters
- 582 have a negative effect on ASR and consequently their LTRE statistics are negative. Notation:
- 583 h = mating system index (Eq. 6), $\rho =$ hatching sex ratio, Juvenile = sex-biased apparent
- 584 survival of juveniles, Adult = sex-biased apparent survival of adults.

585 Supplementary Figure 2. Interspecific variation in sex-specific mating opportunities

among three plover species. Experimental assessment of sex-differences in remating times

587 in three of the six populations analysed indicate that *C. alexandrinus* males in Tuzla, Turkey

588 (n = 19) take longer to find a mate than females (n = 15) after induced divorce. This trend is

reversed in *C. pecuarius* ($n_{\circ} = 10$, $n_{\circ} = 6$) whereas there are no differences in the *C*.

590 *marginatus* ($n_{c} = 6$, $n_{c} = 6$). Significant sex-differences are indicated by asterisks (***: P <

591 0.001, *: P < 0.05, n.s.: P > 0.05). Figure adapted from Parra et al. (ref. 23).

592 Supplementary Figure 3. Relationship between uni-parental care and the adult sex

ratio. (a) Predicted prevalence of male-only care (left panel) or female-only care (right panel) in response to adult sex ratio variation. (b) Observed relationship between parental care strategies and adult sex ratio estimates among the six studied populations. Faint white lines illustrate each iteration of the bootstrap, which randomly sampled an adult sex ratio and parental care estimate from each population's uncertainty distribution and fitted them to the *a priori* exponential model (Eq. 12). (c) Proportion of monitored plover families that exhibit

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- 599 parental cooperation (white) or uni-parental care by males (green) or females (orange).
- 600 Sample sizes reflect the number of families monitored per population, circled numbers
- 601 correspond to the data point labels shown in panel b.

602 Supplementary Figure 4. Variation in annual female mating rates (μ) among the six

- 603 plover populations. Sample sizes indicate the number of individual females in each
- 604 population that had at least two recorded breeding attempts with identified male(s) during the
- study. Values below one represent females that bred over multiple years with the same mate
- 606 (i.e. between season monogamy), whereas values greater that one represent females that have
- had more than one mate per year (i.e. within season polyandry). Values equal to one
- 608 represent individuals that have had one mate per year, but have switched mates between years
- 609 (i.e. between season polyandry but within season monogamy). White data points illustrate
- 610 individual females, and black points are population averages ($\mu \pm 1$ SD).

611 Supplementary Figure 5. Summary statistics of bootstrapped mark-recapture modelling

- 612 of juvenile and adult encounter probability. Left panels illustrate variation in AIC_C w_i.
- 613 Right panels illustrate variation in ΔAIC_c . Model structure of encounter probability (*p*) is
- 614 shown as labels on the y-axes. See *Methods* for further details.

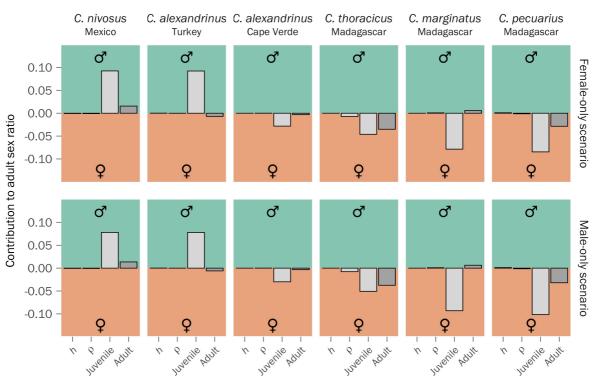
615 Supplementary Movie 1

- 616 Field methods for capturing and monitoring plovers (11min 46sec):
- 617 <u>https://drive.google.com/file/d/0B4RO-u-356yiRjNSQ0RUMWpxYlU/view?usp=sharing</u>

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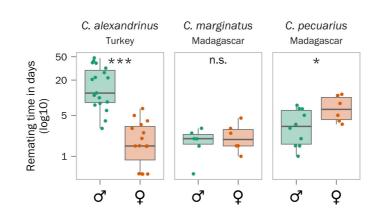
Supplementary Figure 1



Sex bias in parameter

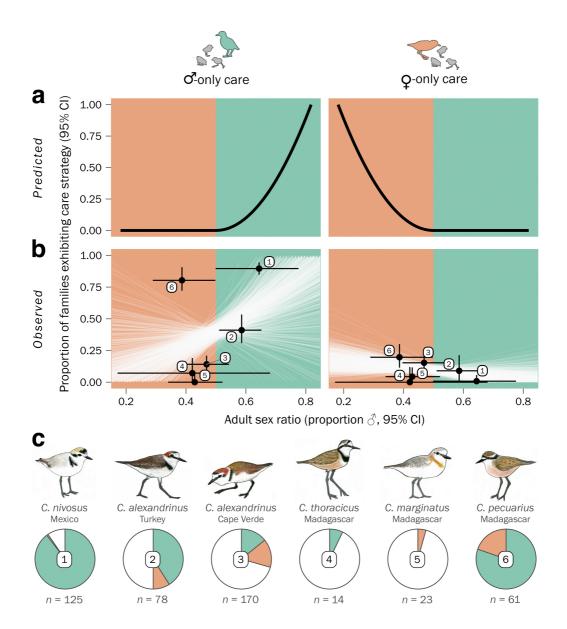
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Supplementary Figure 2



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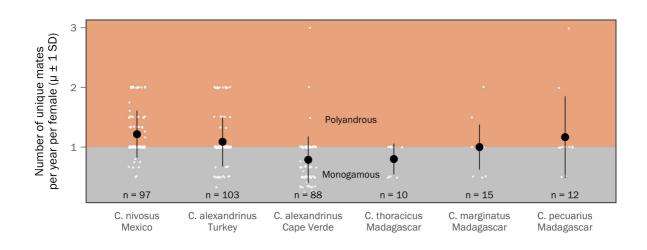
Supplementary Figure 3



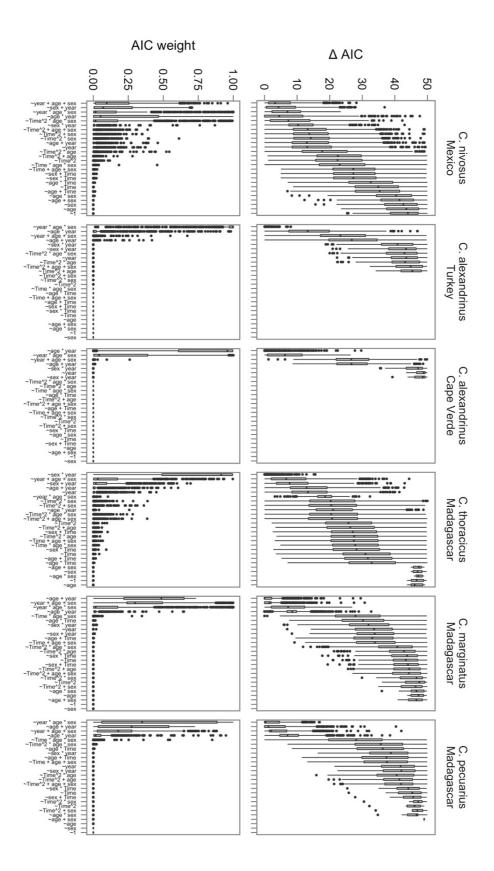
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Supplementary Figure 4



Supplementary Figure 5



622