

fluxweb: a package to easily estimate energy fluxes in food webs

Benoit Gauzens¹, Andrew Barnes¹, Malte Jochum², Jes Hines^{1,3}, Shaopeng Wang¹, Darren Giling^{1,3,4}, Benjamin Rosebaum¹, and Ulrich Brose¹

¹German Centre for Integrative Biodiversity Research (iDiv)
Halle-Jena-Leipzig, Deutscher Platz 5e, 04103 Leipzig,
Germany

²Institute of Plant Sciences, University of Bern, Altenbergrain
21, 3013 Bern, Switzerland

³Institute of Biology, Leipzig University, Johannisallee 21,
04103 Leipzig, Germany

⁴Institute of Ecology, University of Jena, Dornburger Str. 159,
07743 Jena, Germany

We release here the *fluxweb* package (<https://github.com/gauzens/fluxweb.git>) which merges allometric theory [1] with food web energetics methods [2, 3] to provide a set of functions that estimate energy fluxes in complex food webs based on easily accessible ecological informations. Overall, the *fluxweb* package offers some tools for research on quantitative food webs and ecosystem functions, which facilitates novel research on ecosystem functions in real and dynamic natural landscapes.

1 Introduction

In recent years, there have been multiple calls for the reconciliation of food web structure and ecosystem functioning, to better understand how changes to ecological networks will influence the stability and functioning of ecosystems [4, 5, 6]. Energetic food-web approaches can be used to quantify a key

aspect of ecosystem functioning, energy flux, as a way of characterizing ecological processes that are driven by trophic interactions among nodes in food webs [7, 2, 8]. Furthermore, energy fluxes can be used to quantify interaction strengths in food webs.

Despite interest in using quantitative networks [9, 10, 11] they are still rarely employed for describing natural communities. This is, in part, because quantifying interaction strengths or fluxes in food webs remains a deceptively difficult problem, often requiring intensive experimental and observational efforts. A viable solution is to use mathematical proxies for system, and/or organismal level parameters. For example, at the system level inverse matrix reconstruction (commonly referred to as ‘ecological network analysis’), [12, 13] or the ‘food web energetics approach’ [2, 3] have gained some support. These approaches, which are both based on the same steady state assumption (i.e., populations are at equilibrium densities), require reasonable knowledge of the focal system like network topology. That is, some fluxes must be known for the inverse matrix approach, while the food web energetic assume a top down view (energetic need of predators drive their ingoing fluxes). A major difference relates to the solution provided by the methods. While ecological network analysis produces an infinite number of solutions and requires an a posteriori selection function, the method of the food web energetic approach guarantees a unique solution for each dataset. Previously, however, scientists using the ‘food web energetics’ approach [3, 8, 14], have manually calculated fluxes, which can become exceedingly unmanageable as the complexity of the food web increases. Therefore, there is urgent need for a generalized automation of this important method.

Interaction strengths can also be quantified by focusing on organism-level parameters related to the metabolic theory of ecology [1]. Generalized allometric approaches utilize general patterns of functional responses that depend upon body size ratios [15, 16], opening ways for determining interaction strengths in response to commonly available data such as the abundances and body masses of resources and consumers. Allometric rules were successfully applied to predict fluxes in simplified systems with a few species [17]. However, these results have not yet been generalized for use in complex webs. Here we release the *fluxweb* R package (<https://github.com/gauzens/fluxweb.git>), which merges allometric theory [1] with food web energetics methods to provide a set of functions that estimate energy fluxes in complex food webs. In doing so, we support proposals to create a framework allowing the estimation of energy fluxes in trophic networks using widely available ecological information such as for example, biomass, metabolic demand, efficiencies or network topology [6].

2 The underlying model

The model assumes a steady state hypothesis. It implies that L_i , the total amount of energy lost by a species i , either by consumption or physiological processes, is exactly compensated by the metabolized energy it gains from consumption G_i . It will thus solve the equation

$$G_i = L_i \quad (1)$$

knowing that

$$L_i = X_i + \sum_j F_{ij} \quad (2)$$

and that gains are the part of ingoing fluxes once losses due to feeding efficiency are removed.

$$G_i = \sum_j F_{ji} e_a, \quad (3)$$

where X_i defines energetic losses from species i such as through metabolism, and F_{ij} is the flux from species i to species j . e defines a species' feeding efficiency, subscript a is, depending on the ecological assumption, equal to i (when efficiency relates to the identity of the predator), j (when efficiency relates to the identity of the prey) or ij (when efficiency depends on both prey and predator). More details about parameters can be found in section 3. Any flux F_{ij} can be written as $F_{ij} = W_{ij} F_j$, where F_j is the sum of all ingoing fluxes to species j , and W_{ij} defines the proportion of F_j that is obtained from species i ($\sum_i W_{ij} = 1$), and can be obtained by scaling predator preferences w_{ij} to the distribution of prey biomasses

$$W_{ij} = \frac{w_{ij} B_i}{\sum_k w_{kj} B_k}, \quad (4)$$

where B_i sets the biomass of species i . We thus obtain the following model for determining each species' sum of ingoing fluxes:

$$\sum_j W_{ji} F_i e_a = X_i + \sum_j W_{ij} F_j. \quad (5)$$

This equation is solved in two stages: first, the sum of ingoing fluxes for each species is computed. Then, individual fluxes for each pairwise predator-prey interaction are calculated using predator preferences (set in W).

The solution for eq. 5 depends on the chosen definition of feeding efficiency (assigned based on the predator, prey, or link identity) (see SI 1 for demonstrations) and is as follows:

- Efficiencies depending on predator identity

$$F = (\text{diag}(E) - W)^{-1} X. \quad (6)$$

F is the vector such that F_i describes the sum of ingoing fluxes to species i , E is the vector of feeding efficiencies, such that E_i describes the efficiency of a predator i (see related paragraph in section 3 for more precise definitions of feeding efficiencies) with $E_i = 0$ if i is basal, W is the matrix such that W_{ij} sets the proportion of ingoing fluxes to species j from species i and X is the vector defining the sum of energetic losses for each species.

- Efficiencies depending on prey identity

$$F = \left(\text{diag}(W^T E + \vec{b}) - W \right)^{-1} X. \quad (7)$$

In this case, E is the vector such that E_i expresses a prey-related efficiency. \vec{b} is a vector such that b_i equals to 1 if species i is basal, 0 otherwise.

- Efficiencies depending on link identity (both prey and predator)

$$F = \left(\text{diag}(U^T \vec{1}) - W \right)^{-1} X. \quad (8)$$

Here, $\vec{1}$ is a vector of ones, U is the matrix defined by the term by term multiplication of matrices W and E : $U_{ij} = W_{ij} E_{ij}$. E is a matrix such that E_{ij} is the assimilation efficiency of species j feeding on species i .

3 Defining the parameters

A great advantage of this method is that it offers a flexible quantitative framework that can be used to test many different ecological hypotheses related to fluxes in networks [18]. Parameters used to configure the model can be estimated from direct field measurement as well as assessed from general scaling relationships using easily accessible species (e.g., body size) and/or environmental (e.g., temperature) information. Therefore, the *fluxweb* package is a tool that is highly applicable for both experimental/empirical approaches aiming to describe natural systems and for pure theoretical approaches requiring generic solutions.

In the following section, we will describe the different parameters needed and how they can be estimated (see table 1 for example).

Physiological losses: Depending on user assumptions and choices, different ecological processes can be used. Classical choices are often:

- Metabolic rates [1]
- Death rate [19]
- Potentially more complex allometric functions, including time allocated to resting or hunting and associated energy costs [20]

Metabolic rates and death rates can be measured for simplified systems like in microcosms experiments (ref Jon). If the complexity of the network considered prevent such measurements, they can be estimated for each taxonomic group i using the classic allometric equation [1]

$$Y_i = y_0 M_i^b \quad (9)$$

where Y_i represent a parameter related to the physiology of species i . y_0 and b are constants associated with parameter Y_i and M_i is body mass. References for the choice of constant values associated with each model parameter can be found in the descriptions below. Depending on the amount of ecological information available, or precision required, parameters y_0 and b can be quite general (i.e., the same value for all species), or more specific (i.e., applying one parameter value per functional group, taxonomic group, or species). As Y_i is typically estimated per unit biomass, setting the correct value for estimating energy flux is made by a simple multiplication by a species' total biomass.

Efficiencies: *fluxweb* offers the possibility to use a variety of input parameters that define energetic losses, for which different aspects of ecological efficiency must be employed. If using metabolic rate as an input parameter for energetic loss, assimilation efficiency, defined as the proportion of consumed energy that is assimilated (used in respiration and biomass production), needs to be provided. If mass-specific death rates are used in place of metabolism (sensu [18]), users should use the product of assimilation efficiency and production efficiency (percentage of assimilated energy that is used for biomass production). The *fluxweb* package offers three different options for defining ecological efficiencies: consumer-defined, resource-defined, or link-defined (considering both predator and prey identity) efficiencies. These options correspond respectively to the values *prey*, *pred* and *link.specific* for the *ef.level* argument. If, within a single study, each consumer has a relatively homogenous resource pool (i.e., consumers are trophic

specialists such as strict herbivores or strict carnivores), defining efficiencies at the consumer level could be the standard option. However, if a single consumer node draws on a variety of resource nodes (e.g., plants, detritus and animals), efficiencies can be defined at the resource level to account for differences in resource quality ingested by the consumer species.

Preferences: Preferences depict the feeding behavior of predator species and should quantify their foraging choices. Depending on system and user choice, they can be absolute preferences or per capita. The package thereafter offers the possibility to estimate or scale preferences using a linear scaling with prey biomass (like a type one functional response, eq. 4).

Species biomasses: Biomasses are used (depending on user choices) to scale losses (if they are provided per biomass unit) and preferences. It is therefore an optional parameter.

4 Other functionalities

Above we presented the theoretical background used by *fluxweb* to determine fluxes in food webs. However, the package offers several other possibilities. Under the steady state assumption, it is quite straightforward to relate estimated fluxes to the equilibrium state of a set of ordinary differential equations depicting population dynamics (Lotka Volterra like systems of equations). This offers the possibility to gain insight into network stability using the methods established for such equation systems. Thus, the *fluxweb* package offers the *stability.value* and *make.stability* functions using the concept of resilience to quantify the stability of a network with fluxes (see SI 2 for more explanations and the mathematical derivation). The second functionality provided is a sensitivity analysis of outputs regarding the parameters. The *sensitivity* function allows one to assess how functions from the package are sensitive to a specified parameter.

5 Using *fluxweb*

Within the *fluxweb* package, we provide two complete case studies corresponding to different levels of trophic complexity (fig 1). The first example consists of a network of 62 nodes resolved to the species level and 573 edges depicting trophic interactions among mesofauna in forest soil (for details see

Table 1: Description of the different parameters used in our example and their meaning. The units are only examples and might depend on user choices, as long as global coherence is preserved.

Symbol	Parameter name	Description	Unit used in the case study
X_i	Physiological losses	Total amount of energy loss related to physiology for population i	$g.yr^{-1}$
e	Feeding efficiency		Dimensionless (proportion)
B_i	Total biomass of population i		g
F_{ij}	Energy flux from species i to species j		$g.yr^{-1}$
W_{ij}	Absolute preference of species j for species i		Dimensionless
w_{ij}	scaled preference of species j for species i	Sum over j of w_{ij} equals 1. Can be scaled or not with prey abundances	Dimensionless (proportion)

Table 2: Description of the different functions provided by *fluxweb* and their arguments. More details can be found in the help of the package.

Function	Description	Arguments
<i>fluxing</i>	Compute energy fluxes in networks	<ul style="list-style-type: none"> - Interaction Matrix (potentially including preferences) - Physiological losses - Feeding efficiencies - Species biomasses (optional)
<i>stability.value</i>	Return stability of the network of flux (resilience)	<ul style="list-style-type: none"> - Interaction matrix of fluxes - Species biomasses - Physiological losses - Feeding efficiencies - Growth rate
<i>make.stability</i>	Return the smallest multiplicative scalar of losses insuring network stability	<ul style="list-style-type: none"> - Interaction matrix of fluxes - Species biomasses - Physiological losses - Feeding efficiencies - growth rate
<i>sensitivity</i>	Compute the sensitivity of a function to an argument	<ul style="list-style-type: none"> - Function to analyze - Parameter to analyze - Interval of uncertainty for the parameter - Number of replicates to use - Set of parameters needed by the function

[21]). As often for species resolved networks, we only have a binary description of interactions (no weight on trophic links nor feeding preferences are available). The network corresponding to the intermediate level of complexity is a version of the species resolved network where species were aggregated in trophic groups using a group detection method [22]. Reducing complexity by forming aggregated groups can be used to gain basic estimates of predator foraging preferences. Here preferences were estimated by the aggregation process: the foraging preference of a trophic group i on a trophic group j is defined as the number of predation links from species of group i on species of group j . Data used for the species level-food web and the aggregated food web can be accessed using the *species.level* and *groups.level* lists respectively. Each of these lists contains all the necessary information to estimate fluxes. They are automatically loaded with the package.

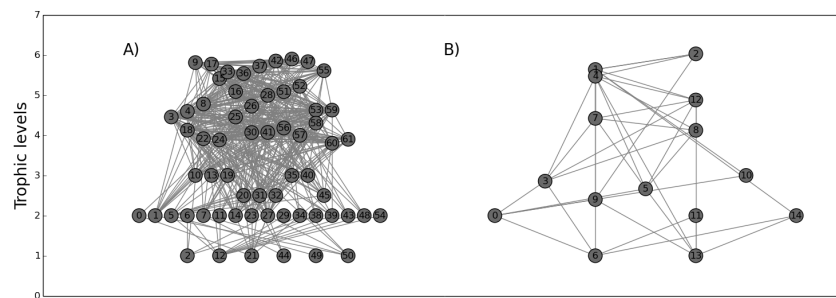


Figure 1: Representation of the *species.level* (A) and *groups.level* (B) food webs

Species-level network

The different parameters of this dataset are:

- The network binary adjacency matrix: value of line i and column j is 1 if species j feeds on species i , 0 otherwise: `species.level$mat`
- The vector of species biomasses (g): `species.level$biomasses`
- The vector of species body masses (g): `species.level$bodymasses`
- The vector of assimilation efficiencies, related to prey identity. Values are 0.45 when prey is a plant, 0.85 when prey is an animal (these

values are commonly used in allometric models of population dynamic; Delmas et al. 2016): `species.level$efficiencies`

We used species metabolic rates to define energetic losses related to physiology (eq. 9).

Network of trophic groups

Data provided here are similar to the ones used for the `species.level` list. Body mass of a group is defined as the average bodymass of species belonging to this group, group biomass as the sum. In addition, the list contains the `species.tgs` data frame indicating the species group composition.

function examples

The different datasets can be loaded using the `load()` function and elements can be thereafter directly accessed after a use of the `attach()` function. In both cases, the matrix of fluxes is simply computed through the call to the *fluxing* function:

```
species.fluxes = fluxing(mat, biomasses, metabolic.rates, efficiencies,  
  bioms.prefs = TRUE, bioms.losses = TRUE, ef.level = 'prey')
```

Here, `bioms.prefs = TRUE` specifies that species preferences depend on prey abundances (eq. 4), `bioms.losses = TRUE` specifies that the `metabolic.rates` input is defined per unit biomass, thus, total physiological losses have to be multiplied by species biomasses. In the same way, the stability of the food web of fluxes is returned by the *stability.value* function:

```
stability.value(val.mat, biomasses, losses, efficiencies, growth.rates,  
  bioms.prefs = TRUE, bioms.losses = TRUE, ef.level = 'pred')
```

with the addition of a vector of growth rates for basal species (parameter `growth.rates`), determined using the classic allometric equation (eq. 9)

6 Conclusion

We presented here the R package *fluxweb*, providing a set of methods allowing the calculation of trophic fluxes in food webs based on the conceptual framework of the ‘food web energetics’ approach [3, 8, 14]. By automating the link between the identity and composition of interacting populations and energy fluxes, these methods facilitate the integration of community and

ecosystem ecology. Fluxes within a system, which have typically been estimated in highly aggregated compartments, can now be quickly calculated at the species level or grouped as appropriate by users to match their objectives. This is a much-needed advance to mechanistically understand how changes in biodiversity may impact ecosystem functioning [5], and is timely given the increasing number and complexity of ecological network data being collected over environmental and disturbance gradients [23, 24].

Under the equilibrium of biomasses assumption, multiple aspects of ecosystem function can be investigated owing to the package's flexibility in the processes represented by parameters, their units, and how the outputs are interpreted. Function parameters can be estimated from general allometric relationships suitable for abstract models or tuned accordingly to precise measurement of specific systems depending on the users' aims. The impact of these estimations on ecological inferences can be assessed with the sensitivity function. The output of the function fluxing is a matrix of the fluxes between nodes. The sum of a single column can be used to assess the fluxes to a particular node of ecological interest, e.g. a top predator. Hypotheses regarding the effect of network structure or community composition on a single function (or multiple single functions; multifunctionality) can also be tested by summing multiple columns, such as secondary production by herbivores or decomposition by detritivores. Assessing such fluxes is important because they are directly linked to ecosystem services but may be mismatched with the standing biomass of these species or trophic groups [8]. Additionally, whole-system flux or throughput, the sum of the entire fluxing matrix, can be used as a single value representing the emergent property of multitrophic functioning [8]. This metric offers promise to assess mechanisms of BEF relationships in multitrophic communities, such as complementarity and selection effects [8].

The functions of *fluxweb* also offer several distinct but related ways to examine network stability that are important in the face of global changes and species loss. First, the biomass fluxes can be interpreted as link weights, and use to assess the distribution of interaction strengths in the network. Second, the stability functions returns the network resilience, its ability to return to its equilibrium state following a small perturbation (see SI 2).

Overall, the *fluxweb* package thus offers some critically important tools for research on quantitative food webs and ecosystem functions, which facilitates novel research on ecosystem functions in real and dynamic natural landscapes [25].

References

- [1] James H. Brown, James F. Gillooly, Andrew P. Allen, Van M. Savage, and Geoffrey B. West. Toward a metabolic theory of ecology. *Ecology*, 85(7):1771–1789, 2004.
- [2] H. W. Hunt, D. C. Coleman, E. R. Ingham, R. E. Ingham, E. T. Elliott, J. C. Moore, S. L. Rose, C. P.P. Reid, and C. R. Morley. The detrital food web in a shortgrass prairie. *Biology and Fertility of Soils*, 3(1-2):57–68, feb 1987.
- [3] P. C. de Ruiter, A.-M. Neutel, and J. C. Moore. Energetics, Patterns of Interaction Strengths, and Stability in Real Ecosystems. *Science*, 269(5228):1257–1260, 1995.
- [4] J. Emmett Duffy, Bradley J. Cardinale, Kristin E. France, Peter B. McIntyre, Elisa Thébault, and Michel Loreau. The functional role of biodiversity in ecosystems: Incorporating trophic complexity. *Ecology Letters*, 10(6):522–538, jun 2007.
- [5] Ross M. Thompson, Ulrich Brose, Jennifer A. Dunne, Robert O. Hall, Sally Hladysz, Roger L. Kitching, Neo D. Martinez, Heidi Rantala, Tamara N. Romanuk, Daniel B. Stouffer, and Jason M. Tylianakis. Food webs: Reconciling the structure and function of biodiversity. *Trends in Ecology and Evolution*, 27(12):689–697, 2012.
- [6] Jes Hines, Wim H. van der Putten, Gerlinde B. De Deyn, Cameron Wagg, Winfried Voigt, Christian Mulder, Wolfgang W. Weisser, Jan Engel, Carlos Melian, Stefan Scheu, Klaus Birkhofer, Anne Ebeling, Christoph Scherber, and Nico Eisenhauer. Towards an integration of biodiversity-ecosystem functioning and food web theory to evaluate relationships between multiple ecosystem services. *Advances in Ecological Research*, 53:161–199, jan 2015.
- [7] Robert V. O’Neill. Indirect estimation of energy fluxes in animal food webs. *Journal of Theoretical Biology*, 22(2):284–290, feb 1969.
- [8] Andrew D. Barnes, Malte Jochum, Steffen Mumme, Noor Farikhah Haneda, Achmad Farajallah, Tri Heru Widarto, and Ulrich Brose. Consequences of tropical land use for multitrophic biodiversity and ecosystem functioning. *Nature Communications*, 5:5351, oct 2014.
- [9] Eric L. Berlow, Anje Margiet Neutel, Joel E. Cohen, Peter C. De Ruiter, Bo Ebenman, Mark Emmerson, Jeremy W. Fox, Vincent A.A. Jansen,

- J. Iwan Jones, Giorgos D. Kokkoris, Dmitrii O. Logofet, Alan J. Mckane, Jose M. Montoya, and Owen Petchey. Interaction strengths in food webs: Issues and opportunities. *Journal of Animal Ecology*, 73(3):585–598, may 2004.
- [10] Carolin Banašek-Richter, Louis Félix Bersier, Marie France Cattin, Richard Baltensperger, Jean Pierre Gabriel, Yves Merz, Robert E. Ulanowicz, Annette F. Tavares, D. Dudley Williams, Peter C. De Ruiter, Kirk O. Winemiller, and Russell E. Naisbit. Complexity in quantitative food webs. *Ecology*, 90(6):1470–1477, jun 2009.
- [11] Dick van Oevelen, Karel van den Meersche, Filip J R Meysman, Karline Soetaert, Jack J. Middelburg, and Alain F. Vézina. Quantifying food web flows using linear inverse models. *Ecosystems*, 13(1):32–45, nov 2010.
- [12] Af F. Vézina and T Piatt. Food web dynamics in the ocean. I. Best-estimates of flow networks using inverse methods. *Marine Ecology Progress Series*, 42(3):269–287, 1988.
- [13] Stuart R. Borrett and Matthew K. Lau. enaR: An R package for Ecosystem Network Analysis. *Methods in Ecology and Evolution*, 5(11):1206–1213, nov 2014.
- [14] Andrew D. Barnes, Patrick Weigelt, Malte Jochum, David Ott, Dorothee Hodapp, Noor Farikhah Haneda, and Ulrich Brose. Species richness and biomass explain spatial turnover in ecosystem functioning across tropical and temperate ecosystems. *Philosophical Transactions of the Royal Society B: Biological Sciences*, 371(1694):20150279, may 2016.
- [15] Olivera Vucic-Pestic, Björn C. Rall, Gregor Kalinkat, and Ulrich Brose. Allometric functional response model: Body masses constrain interaction strengths. *Journal of Animal Ecology*, 79(1):249–256, 2010.
- [16] Gregor Kalinkat, Björn Christian Rall, Olivera Vucic-Pestic, and Ulrich Brose. The allometry of prey preferences. *PLoS ONE*, 6(10):e25937, oct 2011.
- [17] U. Brose, R. B. Ehnes, B. C. Rall, O. Vucic-Pestic, E. L. Berlow, and S. Scheu. Foraging theory predicts predator-prey energy fluxes. *Journal of Animal Ecology*, 77(5):1072–1078, sep 2008.
- [18] John C Moore and Peter C de Ruiter. *Energetic Food Webs: An analysis of real and model ecosystems*. Oxford University Press, 2012.

- [19] Marisa Rossetto, Giulio A. de Leo, Daniele Bevacqua, and Fiorenza Micheli. Allometric scaling of mortality rates with body mass in abalones. *Oecologia*, 168(4):989–996, 2012.
- [20] Chris Carbone, Amber Teacher, and J. Marcus Rowcliffe. The costs of carnivory. *PLoS Biology*, 5(2):0363–0368, 2007.
- [21] Christoph Digel, Alva Curtsdotter, Jens Riede, Bernhard Klarner, and Ulrich Brose. Unravelling the complex structure of forest soil food webs: Higher omnivory and more trophic levels. *Oikos*, 123(10):1157–1172, oct 2014.
- [22] Benoit Gauzens, Elisa Thébault, Gérard Lacroix, and Stéphane Legendre. Trophic groups and modules: two levels of group detection in food webs. *J. R. Soc. Interface*, 12:20141176, 2015.
- [23] Benoit Gauzens, Stéphane Legendre, Xavier Lazzaro, and Gérard Lacroix. Intermediate predation pressure leads to maximal complexity in food webs. *Oikos*, 125(4):595–603, aug 2015.
- [24] Jason M. Tylianakis and Rebecca J. Morris. Ecological Networks Across Environmental Gradients. *Annual Review of Ecology, Evolution, and Systematics*, 48(1):25–48, nov 2017.
- [25] Ulrich Brose and Helmut Hillebrand. Biodiversity and ecosystem functioning in dynamic landscapes. *Philosophical transactions of the Royal Society of London. Series B, Biological sciences*, 371(1694):20150267, may 2016.

7 Supplementary information I

7.1 Efficiency depending on predator identity

We will consider in the following that feeding efficiencies depend on predator identity. We define E as the vector of efficiencies and W as the matrix such that W_{ij} is the proportion of energy entering j that is obtained from i ($\sum_j W_{ij} = 1$). F_{ij} is the flux between species i and j . L_i , the energy loss of species i is defined by:

$$L_i = X_i + \sum_{j=1}^s F_{ij}, \quad (10)$$

where s is the number of species and X_i are the physiological losses of species i . Thus, for satisfying the equilibrium criteria, F_i , the sum of fluxes entering i is:

$$F_i = \frac{1}{E_i} \left(X_i + \sum_{j=1}^s F_{ij} \right). \quad (11)$$

As W_{ij} set the proportion of energy entering j obtained from species i , using $F_{ij} = W_{ij}F_j$, we can write

$$F_i = \frac{1}{E_i} \left(X_i + \sum_{j=1}^s W_{ij}F_j \right), \quad (12)$$

where values W_{ij} are estimated accordingly to species preferences (w_{ij}) and prey abundances:

$$W_{ij} = \frac{w_{ij}B_i}{\sum_{k \in p_i} w_{kj}B_k}. \quad (13)$$

We then have:

$$E_i F_i = X_i + \sum_{j=1}^s W_{ij} F_j, \quad (14)$$

which can be rewritten as:

$$\text{diag}(E)F = X + WF, \quad (15)$$

where $\text{diag}(E)$ is the diagonal matrix such as $\text{diag}(E)_{ii} = E_i$. Provided that $(\text{diag}(E) - W)$ is invertible, the system solve as:

$$F = (\text{diag}(E) - W)^{-1} X. \quad (16)$$

Then, all fluxes $F_{ij} = W_{ij}F_j$ are derived from F_j using W .

7.2 defining efficiencies accordingly to prey identity

Another common method is to define feeding efficiencies accordingly to prey identity. This section proposes a method to adapt the previous framework to this case.

As preferences are defined at the prey level, we need to adapt the previous framework by simulating the addition of a nutrient node on which feed all

basal species with an efficiency of one. Then, eq.5 becomes:

$$F_i \left(\sum_{j=1}^s W_{ji} e_j + b_i \right) = X_i + \sum_{j=1}^s W_{ij} F_j \quad (17)$$

$$\Leftrightarrow F_i \left(\sum_{j=1}^s W_{ji}^T e_j + b_i \right) = X_i + \sum_{j=1}^s W_{ij} F_j \quad (18)$$

where b_i is 1 if i is a basal species, 0 otherwise. This can be rewritten as:

$$\text{diag}(W^T E + \vec{b}) F = X + W F, \quad (19)$$

and, provided that $(\text{diag}(W^T E + \vec{b}) - W)$ is invertible, solved by:

$$F = (\text{diag}(W^T E + \vec{b}) - W)^{-1} X \quad (20)$$

7.3 defining efficiencies accordingly to the link level

It is possible to generalise this approach to efficiencies defined for each prey-predator couple. The solution needs the definition of matrix U such as $U_{ij} = W_{ij} e_{ij}$. The system then reads:

$$\text{diag}(U^T \vec{1}) F = X + W F \quad (21)$$

where $\vec{1}$ is the vector of ones. System is solved as:

$$F = (\text{diag}(U^T \vec{1}) - W)^{-1} X \quad (22)$$

8 Supplementary information II

This document presents how the fluxes calculated under the steady state hypothesis can easily be used to assess system stability, following the framework of Moore and De Ruiter. Here we use resilience as a definition of stability. Resilience is determined from the Jacobian matrix. The system is in a stable equilibrium only if the real parts of eigenvalues from the Jacobian are all negative. In this case, resilience is the absolute value of the real part of the larger eigenvalue. It is this value which is returned by the *stability* function from the fluxweb package

Another measure of stability, provided by the function *make.stability*, is to find the minimal value of a scalar s defining the proportion of physiological losses related to species density. In such case, physiological losses terms in the diagonal of the Jacobian matrix are now defined as sX_i and directly

affect resilience value, s being the measure of stability. We will show in the following section how fluxes at equilibrium can relate to a Lotka Volterra system in an equilibrium state, and how to compute the Jacobian matrix, first assuming that feeding efficiencies relates to predator identity and then assuming them to depend on prey identity.

8.1 Derivation of the Jacobian matrix

8.1.1 preferences defined at predator level

$$\frac{dB_i}{dt} = r_i B_i - \sum_{j \in P_j} c_{ij} B_i B_j \quad (\text{for producers}) \quad (23a)$$

$$\frac{dB_i}{dt} = -X_i B_i + \sum_{j \in p_j} e_i c_{ji} B_i B_j - \sum_{j \in P_j} c_{ij} B_i B_j \quad (\text{for consumers}) \quad (23b)$$

c_{ij} is the coefficient of interaction between prey i and predator j and r_i is the relative growth rate of producer i . P_i and p_i define respectively the sets of prey and predator of species i . This model assumes a type I functional response f_{ij} defined as:

$$f_{ij} = c_{ij} B_i, \quad (24)$$

As the whole method assumes that fluxes and biomasses are at an equilibrium state, we have:

$$F_{ij} = c_{ij} B_i^* B_j^*, \quad (25)$$

B_i^* denoting biomass of species i at equilibrium. Then, off diagonal elements α_{ij} from the Jacobian matrix correspond to the per capita effects (effect of one unit of species biomass). Considering the possible presence of cycles of length 1 (species i is at the same time a prey and a predator of species j), off diagonal elements are

$$\alpha_{ij} = \frac{\delta \frac{dB_i}{dt}}{\delta B_j} = e_i c_{ji} B_i - c_{ij} B_i \quad i \neq j \quad (26)$$

and at equilibrium, using $B_i^* = \frac{F_{ij}}{c_{ij} B_j^*}$ and $B_i^* = \frac{F_{ji}}{c_{ji} B_j^*}$ we obtain:

$$\alpha_{ij} = e_i \frac{F_{ji}}{B_j^*} - \frac{F_{ij}}{B_j^*} \quad i \neq j \quad (27)$$

Diagonal elements, considering possible cannibalistic loops, for producers (p) and consumers (c) are:

$$\alpha_{pp} = r_p - \sum_{j \in P_p} c_{pj} B_j \quad (28a)$$

$$\alpha_{cc} = -X_c + 2e_c c_{cc} B_c + \sum_{\substack{j \in P_c \\ j \neq c}} e_c c_{jc} B_j - 2c_{cc} B_c - \sum_{\substack{j \in P_c \\ j \neq c}} c_{cj} B_j \quad (28b)$$

$$= -X_c + e_c c_{cc} B_c + \sum_{j \in P_c} e_c c_{jc} B_j - c_{cc} B_c - \sum_{j \in P_c} c_{cj} B_j \quad (28c)$$

with $c_{ii} \neq 0$ only if species i is cannibalistic. Again, using $B_i^* = \frac{F_{ij}}{c_{ij} B_j^*}$ and $B_i^* = \frac{F_{ji}}{c_{ji} B_j^*}$ we obtain at equilibrium:

$$\alpha_{pp} = r_p - \frac{1}{B_p^*} \sum_{j \in P_p} F_{pj} \quad (29a)$$

$$\alpha_{cc} = -X_c + \frac{F_{cc}}{B_c^*} (e_c - 1) + e_c \sum_{j \in P_c} \frac{F_{jc}}{B_c^*} - \sum_{j \in P_c} \frac{F_{cj}}{B_c^*} \quad (29b)$$

We can observe here that α_{cc} can be rewritten as

$$\alpha_{cc} = -X_c + \frac{F_{cc}}{B_c^*} (e_c - 1) + \frac{1}{B_c^*} (e_c F_c - L_c) \quad (30)$$

where F_i sets the sum of fluxing ingoing species i and L_i its losses due to predation. As we assume a steady state hypothesis, ingoing fluxes compensate outgoing fluxes plus physiological losses: $e_c F_i = L_i + X_i$. From that, we obtain:

$$\alpha_{cc} = -X_c + \frac{F_{cc}}{B_c^*} (e_c - 1) - \frac{X_c}{B_c^*} \quad (31)$$

8.1.2 efficiencies defined at prey level

The Lotka Voltera system is now written as:

$$\frac{dB_i}{dt} = r_i B_i - \sum_{j \in P_j} c_{ij} B_i B_j \quad (\text{for producers}) \quad (32a)$$

$$\frac{dB_i}{dt} = -X_i B_i + \sum_{j \in P_j} e_j c_{ji} B_i B_j - \sum_{j \in P_j} c_{ij} B_i B_j \quad (\text{for consumers}) \quad (32b)$$

Here e_j defines efficiency of prey species j . At equilibrium, off diagonal elements of the jacobian are like above:

$$\alpha_{ij} = e_j \frac{F_{ji}}{B_j^*} - \frac{F_{ij}}{B_j^*} \quad i \neq j \quad (33)$$

Diagonal elements, considering possible cannibalistic loops, for producers (p) and consumers (c) are:

$$\alpha_{pp} = r_p - \sum_{j \in P_p} c_{pj} B_j \quad (34a)$$

$$\alpha_{cc} = -X_c + 2e_c c_{cc} B_c + \sum_{\substack{j \in P_c \\ j \neq c}} e_j c_{jc} B_j - 2c_{cc} B_c - \sum_{\substack{j \in P_c \\ j \neq c}} c_{cj} B_j \quad (34b)$$

$$= -X_c + e_c c_{cc} B_c + \sum_{j \in P_c} e_j c_{jc} B_j - c_{cc} B_c - \sum_{j \in P_c} c_{cj} B_j \quad (34c)$$

which, at equilibrium leads, like above, to:

$$\alpha_{pp} = r_p - \frac{1}{B_p^*} \sum_{j \in P_p} F_{pj} \quad (35a)$$

$$\alpha_{cc} = -X_c + \frac{F_{cc}}{B_c^*} (e_c - 1) + \sum_{j \in P_c} e_j \frac{F_{jc}}{B_c^*} - \sum_{j \in P_c} \frac{F_{cj}}{B_c^*} \quad (35b)$$

8.1.3 preferences defined at link level

following the same mathematical derivation as before, we obtain:

$$\alpha_{ij} = e_{ij} \frac{F_{ji}}{B_j^*} - \frac{F_{ij}}{B_j^*} \quad i \neq j \quad (36a)$$

$$\alpha_{pp} = r_p - \frac{1}{B_p^*} \sum_{j \in P_p} F_{pj} \quad (36b)$$

$$\alpha_{cc} = -X_c + \frac{F_{cc}}{B_c^*} (e_{cc} - 1) + \sum_{j \in P_c} e_{jc} \frac{F_{jc}}{B_c^*} - \sum_{j \in P_c} \frac{F_{cj}}{B_c^*} \quad (36c)$$