

1 Predators or mutualists? Lack of dispersal distance benefits and small caching profits makes  
2 scatterhoarding rodents antagonists of oaks

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15 **Data for the paper will be archived at Dryad repository.**

16 **Abstract**

- 17 1. Numerous interactions between plants and animals vary in their outcome between antagonism  
18 and mutualism, but it proved to be difficult to quantify their final outcome. Interactions  
19 between plants and scatterhoarding animals provide a prime example of this phenomenon.  
20 Scatterhoarders consume large quantities of seeds (potentially reducing plant establishment),  
21 yet also disperse seeds and bury them in shallow caches (potentially improving recruitment).  
22 However, it has been rarely determined which role prevails for particular plant species.
- 23 2. We experimentally evaluated the benefits of rodent seed dispersal in two model oak species  
24 (sessile oak *Quercus petraea*, and red oak *Q. rubra*) and used a heuristic mathematical model  
25 to place the interactions at the antagonism-mutualism continuum.
- 26 3. Our results indicate that during the period of the study, interactions between scatterhoarding  
27 rodents and both focal oaks were antagonistic. Even though acorn burial increased the  
28 likelihood of seedling establishment, this effect was not strong enough to compensate for the  
29 costs of seed predation. Furthermore, we found no evidence that the short-distance  
30 transportation that is usually provided by small mammals benefits early oak recruitment.
- 31 4. Even relatively large improvements in seedling establishment after seed burial do not  
32 necessarily render the plant-scatterhoarder interaction mutualistic. Our study is the first to  
33 separate and directly quantify two most important services provided to plants by their rodent  
34 partners: seed transportation away from parent plants and seed burial in topsoil. It also  
35 demonstrates how readily accessible field data can be used to gauge the outcomes in plant-  
36 granivore conditional mutualisms.

37

38 **Key words:** antagonism, biotic invasion, conditional mutualism, scatterhoarding, seed dispersal,

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## 41 **Introduction**

42 Numerous interactions between plants and animals vary in their outcome between antagonism and  
43 mutualism (Bronstein, 1994; Palmer *et al.*, 2010). Interactions between plants and scatterhoarding  
44 animals, such as rodents or corvids, are a prime example of this phenomenon because scatterhoarders  
45 play a dual role in plant regeneration. On the one hand, they consume large quantities of seeds and  
46 reduce plant establishment (Howe & Brown, 2001; Zwolak *et al.*, 2010; Larios *et al.*, 2017). On the  
47 other hand, however, they disperse seeds and bury them in shallow caches, which for some plant  
48 species provide the only means of successful recruitment (Vander Wall, 1992; Asquith *et al.*, 1999;  
49 Muñoz & Bonal, 2011; Pesendorfer *et al.*, 2016). Determining which role prevails for particular plant  
50 species is crucial for nature conservation, forest management, and control of invasive species because  
51 management strategies will depend on the interaction outcome. Yet, these outcomes have been rarely  
52 quantified.

53 Whether scatterhoarding granivores are beneficial or detrimental for plant population depends  
54 on whether recruitment with granivores is greater or less than recruitment without seed caching  
55 (Jansen & Forget, 2001; Theimer, 2005; Schupp *et al.*, 2010; Zwolak & Crone, 2012). Quantifying  
56 this has proven challenging, but a recent study developed a heuristic model that narrows down the  
57 crucial field measurements to a few essential parameters (Zwolak & Crone, 2012). Intuitively, when  
58 the benefits from seed caching are high, plants can bear higher costs in the form of seed consumption.  
59 Following on this notion, the mathematical model shows that the interaction is mutualistic when the  
60 probability of caching and not retrieving cached seeds exceeds the ratio of seedling emergence from  
61 surface to seedling emergence from caches. To briefly review this approach (Zwolak & Crone, 2012),  
62 it started from the premise that granivores are beneficial when plant recruitment in the presence of  
63 granivores is greater than plant recruitment in the absence of granivores. This inequality is written in  
64 mathematical terms as follows:

$$65 \quad e_S < p_H p_C e_C + (1 - p_H) e_S \quad (0)$$

66 Where  $e_S$  is seedling emergence from surface,  $p_H$  is the proportion of seeds harvested by granivores,  $p_C$   
67 is the probability that seeds will be cached and left uneaten, and  $e_C$  is the seedling emergence from

68 caches. In studies of plant-granivore interactions, it is much easier and more common to measure the  
69 emergence rates ( $e_S$  and  $e_C$ ) than the caching rates ( $p_H$  and  $p_C$ ; see Zwolak and Crone 2012). Therefore,  
70 in order to compare the role of granivores across studies, Zwolak and Crone (2012) rearranged the  
71 equation to calculate the minimum value of  $p_C$  that would be necessary for granivores to increase plant  
72 recruitment:

$$73 \quad p_C < \frac{e_S}{e_C} \quad (0)$$

74 Thus, granivores help plant recruitment when the proportion of cached and uneaten seeds exceeds a  
75 threshold value (hereafter  $p_C$ , after Zwolak and Crone 2012) determined by the seedling emergence  
76 ratio, i.e. the benefits of seed burial.

77         Nonetheless, scatterhoarders not only bury the seeds in the topsoil, but also move them away  
78 from the parent plant. This also modifies recruitment probability, with effects that are usually believed  
79 to be positive, due to colonization of ephemeral microsites or escape from distance- and density  
80 dependent mortality (Jansen *et al.*, 2008; Comita *et al.*, 2010; Johnson *et al.*, 2012; Fricke *et al.*,  
81 2014). However, the effects can also be negative, e.g. when habitat quality is autocorrelated, it often  
82 declines with distance from maternal plants (John *et al.*, 2007; Condit *et al.*, 2013). Furthermore, the  
83 distance to the seed source may alter rodent foraging activity and seed pilferage rates through changes  
84 in local seed availability (Stapanian & Smith, 1984; Gálvez *et al.*, 2009). Nonetheless, even though  
85 factors shaping dispersal distance by scatterhoarders, especially by rodents, are intensively studied  
86 (Jansen *et al.*, 2004; Xiao *et al.*, 2005; Moore *et al.*, 2007; Sunyer *et al.*, 2014; Lichti *et al.*, 2017), the  
87 actual influence of dispersal distance on recruitment probability is seldom quantified. The intertwined  
88 escape-related (transportation distance) and condition-related (burial) benefits constrain our ability to  
89 understand mechanisms that drive the ecological interactions between plants and scatterhoarders.

90         Here, we address this problem by separating burial- and distance-dependent benefits of rodent  
91 seed dispersal. We used two oak species as model system: sessile oak (*Quercus petraea*), and  
92 Northern red oak (*Q. rubra*). The sessile oak is the dominant native oak in Central European forests.  
93 The Northern red oak was introduced to European forests from North America in the 17th century as  
94 an ornamental species (Woziwoda *et al.*, 2014b). Currently, it is one of the most frequent foreign

95 deciduous species in the region (Woziwoda *et al.*, 2014b), and its occurrence proved to be  
96 troublesome as it suppresses abundance and richness of numerous native species (Chmura, 2013;  
97 Woziwoda *et al.*, 2014a). For both oaks, the primary means of reproduction is thought to be  
98 abandonment of seed caches made by scatter-hoarding rodents and birds (den Ouden *et al.*, 2005;  
99 Steele, 2008; Myczko *et al.*, 2014). Nonetheless, whether caching benefits exceeds the costs imposed  
100 by seed predation has never been experimentally evaluated.

101 Drawing on past studies, we evaluated the following hypotheses: (1) Benefits of seed caching  
102 should be larger than transportation benefits. Seed burial plays a critical role in regeneration of many  
103 plant species because it protects seeds from desiccation and strict seed predators (Haas & Heske, 2005;  
104 Vander Wall, 2010; Zwolak & Crone, 2012; Zwolak *et al.*, 2016). On the other hand, rodents move  
105 seeds over rather short distances (within population: Den Ouden *et al.* 2005), which is unlikely to  
106 result in strong distance-dependent effects. (2) The distance-related benefits should be smaller in red  
107 oak than in sessile oak. This result might occur because invasive species often escape from the  
108 inhibitory effects of soil biota in their alien range (Agrawal *et al.*, 2005; Maron *et al.*, 2014). (3) Seed  
109 removal by rodents should increase with distance from adult plants. This result might occur because  
110 dense seed shadows under the tree may create local satiation effects (Xiao *et al.*, 2015). (4) Pilferage  
111 of cached acorns should be more frequent in red than in sessile oaks. This result might occur because  
112 heavier acorns are more likely to be pilfered (Perea *et al.*, 2016) and red oak acorns are over twice  
113 heavier than sessile oak acorns (Bogdziewicz *et al.*, 2017a). (5) Benefits of seed dispersal and burial  
114 will be large enough to compensate predation, leading to mutualistic relationship in which small  
115 mammals will aid recruitment of focal oaks. This prediction is based on results of a meta-analysis of  
116 previous studies, which suggested that outcomes of plant-rodent interactions tend to be weakly  
117 mutualistic (Zwolak and Crone 2012).

118 We used the estimates of seedling establishment probability obtained in field experiments to  
119 place the focal interactions at the antagonism-mutualism continuum, using the theoretical framework  
120 developed by Zwolak and Crone (2012). However, the original model did not include the potential  
121 changes in caching benefits driven by seed pilferage (Steele *et al.*, 2014; Sunyer *et al.*, 2015; Zwolak

122 *et al.*, 2016). Therefore, as part of this paper, we extend the logic of the  $p_c$  calculation to include  
123 pilferage by naïve foragers.

124

## 125 **Materials and Methods**

126 We established four study sites in sessile – red oak mixed forests in Drawieńska Forest, western  
127 Poland. This area is located in the temperate climate zone, with average annual precipitation of 592  
128 mm and means monthly temperature ranging from 17°C in July to -2°C in January. These mixed  
129 forests comprise almost exclusively of the three oak species (*Q. petraea*, *Q. robur*, and *Q. rubra*), with  
130 single trees of common hornbeam (*Carpinus betulus*) and European beech (*Fagus sylvatica*). The  
131 understory is poorly developed, with some patches of raspberry (*Rubus* sp.) and common nettle  
132 (*Urtica dioica*), and seedlings of oaks and beech. Oak acorns are relatively large (average weight; *Q.*  
133 *petraea* 1.26 g., *Q. rubra*: 2.85 g.), and readily dispersed and eaten by small mammals (Bogdziewicz  
134 *et al.* 2017b, Bogdziewicz, unpublished). As revealed by camera traps, small mammals at our study  
135 sites were *Apodemus* sp., most likely *A. flavicollis*, a seed specialist: (Selva *et al.*, 2012, Gasperini *et*  
136 *al.*, 2017), and *Myodes glareolus*.

137 To quantify the effects of acorn burial on seedling emergence we conducted seed addition  
138 experiments. We randomly chose 16 *Q. petraea* trees (3 per site) and 16 *Q. rubra* trees (3 per site). We  
139 added acorns of the focal species in 20 x 20 x 20 cm wire mesh cages (5 acorns per cage). Cages were  
140 buried 10 cm into the ground in sets of four. In half of the cages we buried acorns 1-2 cm into the  
141 ground and in the other half we placed acorns at the top of the litter layer and covered them with  
142 leaves to mimic autumn leaf fall. This treatment was crossed with rodent exclusion: in half of the  
143 cages we cut 8 x 8 cm holes to allow rodent access, and the other half remained closed to exclude  
144 rodent foraging. A comparison of seedling recruitment from acorns that were buried vs. placed on  
145 surface allowed us to estimate burial-dependent benefits of rodent seed dispersal. Rodent exclusion  
146 allowed us to estimate seed pilferage by comparing recruitment of buried acorns in open vs. closed  
147 cages.

148 To address the distance-related effects of rodent seed dispersal, the above-described cage sets  
149 were buried at transects. Under each tree, we established a transect along one cardinal direction,

150 aiming to maximize the distance of the transect to other conspecifics. This was done assuming that  
151 rodents tend to carry and cache seeds towards areas of lower conspecific seed density (Stapanian &  
152 Smith, 1984; Hirsch *et al.*, 2012; Steele *et al.*, 2014; Yang *et al.*, 2016). Thus, our estimates of  
153 distance-dependent effects may be overly positive, if such directed dispersal does not occur in our  
154 system. We buried five sets of cages at each transect. We used tree crown as a reference point and  
155 buried one set of cages directly underneath the crown border, another set 5 m towards the tree trunk  
156 (underneath the crown), and the remaining 3 sets every 5 m in the opposite direction (i.e. away from  
157 the tree trunk). We used 25 m as the maximum evaluated distance because acorn-tracking experiments  
158 report that vast majority of acorns transported by rodents are cached within that radius (den Ouden *et*  
159 *al.*, 2005; Xiao *et al.*, 2005; Muñoz & Bonal, 2011; Bogdziewicz *et al.*, 2017b). We set up  
160 experimental cages in October 2016 and quantified seedling establishment in August 2017. The  
161 overall sample size equaled 2400 acorns (480 seedling cages).

162

### 163 Statistical analysis

164 To test how acorn burial, distance from the tree, and rodent foraging affect the seedling establishment,  
165 we built a separate generalized linear mixed model (GLMM) for each oak species. We used nested  
166 random effects of cage set, tree, and study site, logit link, and binomial family error distribution, and  
167 implemented the models via lme4 package in R (Pinheiro & Bates, 2015). In each model, we used  
168 seedling establishment (0/1) as the response variable, and burial (surface vs. sowed), rodent access  
169 (excluded vs. allowed), and distance from the tree as fixed effects. We also included all possible 2-  
170 way interaction terms between fixed effects, and the 3-way interaction (which was removed when  
171 non-significant). We calculated marginal (i.e. the proportion of variance explained by fixed effects)  
172 and conditional (i.e. the proportion of variance explained by fixed and random effects)  $R^2$  for GLMMs  
173 using the MuMIn package (Nakagawa & Schielzeth, 2013; Bartoń, 2016).

174

### 175 Calculating the $p_c$ threshold and the effects of seed pilferage

176 We evaluated how the interactions between rodents and oaks are placed along the antagonism  
177 – mutualism continuum (Zwolak & Crone, 2012). The  $p_c$  threshold was calculated as a ratio of

178 emergence from seeds sown on surface vs. emergence from buried seeds, with rodents excluded. Seed  
179 pilferage was gauged as the ratio of seedling recruitment from buried seeds in open vs. closed cages.  
180 Implicitly, the original definition of the proportion of seeds cached and uneaten ( $p_c$ ) combined three  
181 processes (Zwolak & Crone, 2012): the probability that a seed is cached, the probability it is eaten by  
182 the cache owner, and the probability it is pilfered:

$$183 \quad p_c = p_{cached} \times (1 - p_{eaten\ by\ cache\ owner}) \times (1 - p_{pilfered}) \quad (1)$$

184 where  $p_{cached}$  is the probability a seed is cached,  $p_{eaten}$  by cache owner is the probability of  
185 retrieval by individuals responsible for seed burial and  $p_{pilfered}$  is the probability of retrieval by  
186 pilferers. The  $p_c$  threshold is the minimum value of when the benefits for plants balance the costs of  
187 seed consumption. Thus, if the threshold is determined by the proportion of seedling emergence from  
188 surface,  $e_s$  (estimated with data on seedling emergence from seeds sown on surface in closed cages), to  
189 emergence from caches,  $e_c$  (i.e. by benefits of burial, estimated with data on seedling emergence from  
190 seeds buried in closed cages), i.e.:

$$191 \quad \tilde{p}_c = \frac{e_s}{e_c}$$

192 then the equation can be expanded to show the effects of pilferage:

$$193 \quad p_{cached} \times (1 - p_{eaten\ by\ cache\ owner}) \times (1 - p_{pilfered}) = \frac{e_s}{e_c}$$

194 and rearranged to include only the unknown proportion of seeds cached and uneaten by the cache  
195 owner:

$$196 \quad \tilde{p}_{CO} = p_{cached} \times (1 - p_{eaten\ by\ cache\ owner}) = \frac{e_s}{e_c(1 - p_{pilfered})} \quad (2)$$

197 In the above equation,  $\tilde{p}_{CO}$  is the minimum beneficial proportion of seeds cached and uneaten by *cache*  
198 *owner*, and all other parameters are as defined above. In other words,  $\tilde{p}_{CO}$  threshold defines the burial  
199 benefits while taking into account cache pilferage by naïve forages.

200 Confidence intervals for these parameters ( $e_s$ ,  $e_c$ , and  $p_c$ ) were obtained with parametric  
201 bootstrapping, i.e. sampling from the distributions defined by the mean and standard error of each  
202 coefficient to obtain a joint distribution for the derived variables. We repeated the calculations of  $p_c$



203 and  $p_{CO}$  for both near the conspecific probability of germination (i.e. germination rate estimated at the  
204 distance 0 m, and far i.e. germination rate estimated at distance 25m). Probabilities of germination  
205 were derived from the above-described GLMMs.

206 The empirical  $p_{CO}$  value (the ultimate probability that the acorn will be cached and not  
207 retrieved, accounting for retrieval by the whole granivore community) for both oak species was  
208 derived from a parallel study investigating rodent seed dispersal of the focal oaks, i.e. 17% for red  
209 oaks and 2% for sessile oaks (Bogdziewicz et al. unpublished).

210

## 211 **Results**

212 In accordance with hypothesis (1), benefits of acorn burial were higher than benefits of acorn  
213 transportation away from adult trees. Acorn burial enhanced seedling establishment in both species,  
214 mainly through reducing seed predation. In *Q. petraea*, when rodents had access, acorn burial  
215 increased establishment probability 2-fold (open cages, surface vs. buried acorns: 18% vs. 39%). This  
216 effect was considerably weaker when rodents were excluded (closed cages, 57% for acorns on the  
217 surface vs. 65% for buried acorns; rodent exclusion  $\times$  burial interaction in Table 1a, Fig. 1). Similarly,  
218 burial increased establishment probability almost 3-fold in *Q. rubra* (open cages: 25% vs. 71%, Fig.  
219 1). This effect was weaker when rodents were excluded (50% vs. 73%; rodent exclusion  $\times$  burial  
220 interaction in Table 1b).

221 We did not detect any distance-related benefits of seed dispersal and contrary to hypothesis  
222 (2), the oak species did not differ in this regard. In fact, seedling establishment probability decreased  
223 with the distance from the focal tree (Table 1, Fig. 2 and 3). This phenomenon was caused by an  
224 increase in acorn removal, which supports hypothesis (3). In *Q. petraea* increased removal was  
225 apparent only when acorns were sown on surface (Table 1a, Fig. 2). Pilferage of cached acorns did not  
226 differ with distance from the tree in this species. In *Q. rubra*, this effect occurred both for acorns that  
227 were buried and those that were left on surface.

228 Acorn burial was more beneficial for *Q. rubra* than for *Q. petraea* (see above, and Fig. 1). In  
229 contrast to hypothesis (4), pilferage rates were higher in *Q. petraea* than in *Q. rubra*. In *Q. petraea*  
230 seedling establishment from buried acorns was 1.5 times higher when rodents were excluded (39% in

231 open vs. 65% in closed cages). In *Q. rubra*, burial provided almost complete protection from pilferage  
232 (open vs. closed cages: 71% vs. 73%). Note that the percentages estimates are the model intercepts,  
233 and decrease with distance from the tree in some treatments (see below).

234 Estimated  $p_{CO}$  values (the minimum beneficial proportion of seeds cached and uneaten by  
235 cache owner, i.e. those taking into account cache pilferage) equaled 1.21 (0.86-1.79) for *Q. petraea*  
236 (both near and far from the tree), and 0.69 (0.46-0.97) in *Q. rubra* near, and 1.24 (0.77-1.99) in *Q.*  
237 *rubra* far from the tree (Fig. 4). Note that the  $p_{CO}$  value for *Q. petraea* does not differ with distance  
238 because the pilferage rates were distance-independent (Fig. 2a). These values are either impossible to  
239 reach (when they exceed 1) or would require almost all cached acorns to be never retrieved to  
240 approach the mutualism parameter space of the interaction (Fig. 4). Thus, hypothesis (5) was not  
241 supported: interactions between scatterhoarders and oaks in our study system were clearly  
242 antagonistic.

243 A fraction of removed acorns could be recached rather than consumed. If we assume that  
244 pilfered seeds are as likely to be eaten or re-cached as seeds collected for the first time (Jansen *et al.*,  
245 2004), then the consequences of burial depend on the number of rounds of recaching (Online  
246 Supplement 1). Nonetheless, for the parameters observed in our system, recaching of pilfered acorns  
247 would not affect the conclusion that scatterhoarders acted antagonistically in their interactions with  
248 oaks (Online Supplement 1).

249

## 250 **Discussion**

251 Our study is the first to separate and directly quantify two most important services provided to  
252 plants by their rodent partners: seed transportation away from parent plants and seed burial in topsoil.  
253 Our results suggest that widely-accepted benefits of transportation might in many ecological systems  
254 be smaller than expected. Moreover, they demonstrate that even relatively large improvements in  
255 seedling establishment after seed burial do not necessarily render the plant-scatterhoarder interaction  
256 mutualistic. Finally, and most generally, our study illustrates a straightforward empirical approach that  
257 can be used to evaluate the ambiguous role of scatterhoarding granivores in plant regeneration.

258           Animal dispersers often provide several different services that are not equal in their  
259 importance for their partner plants, but benefits provided by each are rarely separated. Yet, a few case  
260 studies that have done so demonstrated that it may change the way we think about particular  
261 interactions. For example, birds that disperse chili peppers (*Capsicum chacoense*) remove pathogens  
262 from the dispersed seeds (condition-related benefit) and transport seeds far from parent plants (escape-  
263 related benefit). However, only gut passage enhances seed survival (Fricke *et al.*, 2013). Similar  
264 situation was reported for Iberian pears (*Pyrus bourgaeana*), where pulp removal was more important  
265 than transportation distance for plant recruitment (Fedriani *et al.*, 2012). By untwining the role of  
266 caching (condition-related benefit) from transportation distance in oak-scatterhoarders interactions, we  
267 demonstrated that acorn burial was the main benefit of this interaction (hypothesis 1 in Table 2). The  
268 lack of distance-dependent benefits was unsurprising because in locally common plant species, the  
269 benefits should disappear when the species become so common that their predator and pathogen  
270 communities become functionally uniform across the landscape (Janzen, 1971; Schupp, 1992; Fricke  
271 *et al.*, 2013; Garzon-Lopez *et al.*, 2015). This is likely the case in our system because sessile and red  
272 oaks dominate forest stands. This indicates that the generally greater establishment of some species far  
273 from parent plants may be due distance-independent benefits of burial rather than distance-dependence  
274 *per se*. In other words, intertwined caching and transportation may create false positive effect of  
275 transportation, while it is only caching that helps recruitment (as in our system). This calls for  
276 increased attention on condition-dependent benefits of seed dispersal, which have been often  
277 overlooked as researchers focused on dispersal distance and final location of seeds (Fricke *et al.*,  
278 2013).

279           The patterns of seedling establishment suggest that rodent foraging is a strong filter of oaks  
280 spatial recruitment in our system. When rodents were excluded, distance from the tree did not change  
281 the probability of seedling establishment in either species (contrary to hypothesis 2 in Table 2).  
282 However, when rodent foraging was allowed, seedling establishment rate was highest near the trees,  
283 indicating that locally dense seed shadows may allow higher proportion of undispersed seeds to  
284 survive and germinate (hypothesis 3 in Table 2). On the other hand, distance to adult trees influenced  
285 pilferage of buried acorns in a species-specific manner. In *Q. petraea*, the pilferage did not change

286 with distance from the tree, while it decreased with distance in *Q. rubra*. It suggests that ambient seed  
287 density has a stronger effect on cache pilferage rates in *Q. rubra* than in *Q. petraea* (Gálvez *et al.*,  
288 2009). However, we did not measure seed shadows, and thus this pattern might have resulted from  
289 larger crop size of *Q. rubra* at the time of our experiments.

290 *Q. rubra* acorns were generally better protected from seed predation once buried than acorns  
291 of *Q. petraea* (contrary to hypothesis 4 in Table 2). The pilferage rates of buried seeds are related to  
292 their release of odorant molecules (Vander Wall, 1998; Vander Wall & Joyner, 1998; Yi *et al.*, 2016).  
293 *Q. rubra* appears to rely more on rodent dispersal in their native range than does *Q. petraea* in Europe  
294 (den Ouden *et al.*, 2005; Steele, 2008). Thus, the tighter coevolution between small mammals and *Q.*  
295 *rubra* may led to reduced emission of volatile compounds in this species, reducing the overall  
296 pilferage rates. In consequence, lower pilferage rates in *Q. rubra* increased overall burial benefits,  
297 moving the interaction closer to the mutualism.

298 Nonetheless, in contrast to our fifth and final hypothesis (Table 2), results of our experiments  
299 indicate that scatterhoarding rodents reduced recruitment of focal oak species. As anticipated, acorn  
300 burial increased the likelihood of seedling establishment. However, the seedling establishment of  
301 unburied acorns in the absence of small mammal foraging was high for both species. As a final result,  
302 burial benefits were too small to override the costs of seed predation (Zwolak & Crone, 2012). In fact,  
303 even if the probability of seedling establishment from rodent caches would elevate to 100%, the *p*<sub>CO</sub>  
304 value for *Q. petraea* would equal 0.57 while for *Q. rubra* 0.50. This level of survival of cached acorns  
305 appears unlikely, as the reported values range from 1 to 20% (Vander Wall & Joyner, 1998; Vander  
306 Wall, 2002; Gomez *et al.*, 2008; Xiao *et al.*, 2013; Bogdziewicz *et al.*, 2017b; Wróbel & Zwolak,  
307 2017). Therefore, while it is still best for an acorn to be cached and then forgotten, presence of small  
308 mammals did not help recruitment of oaks in our system.

309 We note, however, that our study was conducted during relatively short time frame (1 year). It  
310 is possible that fluctuating environmental conditions (e.g. years with droughts or severe winters) can  
311 increase the benefits of acorn caching and shift the oak-rodent relationship towards mutualism (burial  
312 benefits in our study were remarkably low, e.g. (Kollmann & Schill, 1996; García *et al.*, 2002; Xia *et*  
313 *al.*, 2016) reported higher benefits in oaks). In this situation, the role of small mammals in oak

314 recruitment would fluctuate over time. Moreover, both focal oaks species exhibit mast years (Sork *et*  
315 *al.*, 1993; Bogdziewicz *et al.*, 2017c), which in turn drives fluctuations in small mammal population  
316 abundance and may create satiation effects, both at the seed source and after caching (Kelly, 1994;  
317 Xiao *et al.*, 2013; Zwolak *et al.*, 2016). Nevertheless, the potential increase in cache survival caused  
318 by the phenomenon is unlikely to counterbalance the costs of seed predation – unless it is  
319 accompanied by environmental changes that increase the benefits of seed caching. Thus, an interesting  
320 venue for future studies would be to quantify temporal variation in this interaction.

321         As a final caveat to this study, Janzen-Connell effects are stronger at the seedling than seed-to-  
322 seedling stage (Comita *et al.*, 2014). Therefore, the benefits of transportation may appear at later  
323 stages of plants life cycle. However, several studies that evaluated distance-dependent survival rates at  
324 seedling stage in temperate oaks did not found the effect (Reinhart *et al.*, 2012; Comita *et al.*, 2014).  
325 Furthermore, directed dispersal increases the likelihood of colonization of microhabitats that are  
326 favorable for germination and establishment (Steele *et al.*, 2013; Yi *et al.*, 2013). Although such  
327 effects were never reported for focal oaks, our experimental design could underestimate these effects  
328 and thus, benefits of acorn transportation by scatterhoarding rodents.

329         To conclude, while costs and benefits of plant-scatterhoarder interactions might vary among  
330 particular pairs of engaged species (Zwolak and Crone 2012), our evaluation of the net outcomes in  
331 rodent-oak interactions indicated that for they were antagonistic for both the native and the invasive  
332 tree. Even though acorn burial increased the likelihood of seedling establishment, this effect was not  
333 strong enough to compensate for the costs of seed predation. Furthermore, we found no evidence that  
334 the short-distance transportation that is usually provided by small mammals benefits early oak  
335 recruitment. In all up-to-date studies, scatterhoarder service was treated as a single unit. Yet, seed  
336 burial and transportation are two separate services that differ in their importance. Separating them will  
337 allow better understanding of the mechanisms driving interactions between plants and scatterhoarders.

338

339 **Authors contributions**

340 MB, RZ, and ECC conceived and designed the study, MB collected the data, MB, RZ, and ECC  
341 analyzed the data, MB and RZ wrote the manuscript, all authors contributed substantially to the  
342 revising of the paper.

343

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350

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517 timing of granivore satiation and benefits of seed caching support the predator dispersal  
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- 519

520 Table 1. Effects of burial, distance from the parent tree, and rodent access on the germination  
 521 probability of focal oaks. The marginal  $R^2$  of the model for *Q. petraea* equaled 0.28, while conditional  
 522 0.34. For *Q. rubra*, the marginal  $R^2$  of the model equaled 0.33, while conditional 0.38.

523

524 A) Sessile oak *Quercus petraea*

| Fixed effect                                       | Estimate | SD | z    | p     |        |
|--|----------|----|------|-------|--------|
| Intercept (rodent access allowed, burial: surface) | -1.49    |    | 0.35 | -4.23 | <0.001 |
| Rodents excluded                                   | 1.77     |    | 0.33 | 5.31  | <0.001 |
| Burial: cached                                     | 1.04     |    | 0.33 | 3.10  | 0.001  |
| Distance   | -0.06    |    | 0.02 | -2.16 | 0.03   |
| Rodents excluded × Burial: cached                  | -0.67    |    | 0.31 | -2.19 | 0.03   |
| Rodent excluded × Distance                         | 0.04     |    | 0.02 | 1.82  | 0.07   |
| Burial: cached × Distance                          | 0.04     |    | 0.02 | 2.03  | 0.04   |

525

526 B) Northern red oak *Quercus rubra*

| Fixed effect                                       | Estimate | SD | z    | p     |        |
|--|----------|----|------|-------|--------|
| Intercept (rodent access allowed, burial: surface) | -1.06    |    | 0.39 | -2.71 | 0.006  |
| Rodents excluded                                   | 1.09     |    | 0.37 | 2.95  | 0.003  |
| Burial: cached                                     | 1.97     |    | 0.36 | 5.36  | <0.001 |
| Distance   | -0.10    |    | 0.03 | -3.75 | <0.001 |
| Rodents excluded × Burial: cached                  | -0.97    |    | 0.33 | -2.94 | 0.003  |
| Rodent excluded × Distance                         | 0.06     |    | 0.02 | 2.88  | 0.004  |
| Burial: cached × Distance                          | 0.03     |    | 0.02 | 1.21  | 0.22   |

527 Table 2. Summary of hypotheses and results.

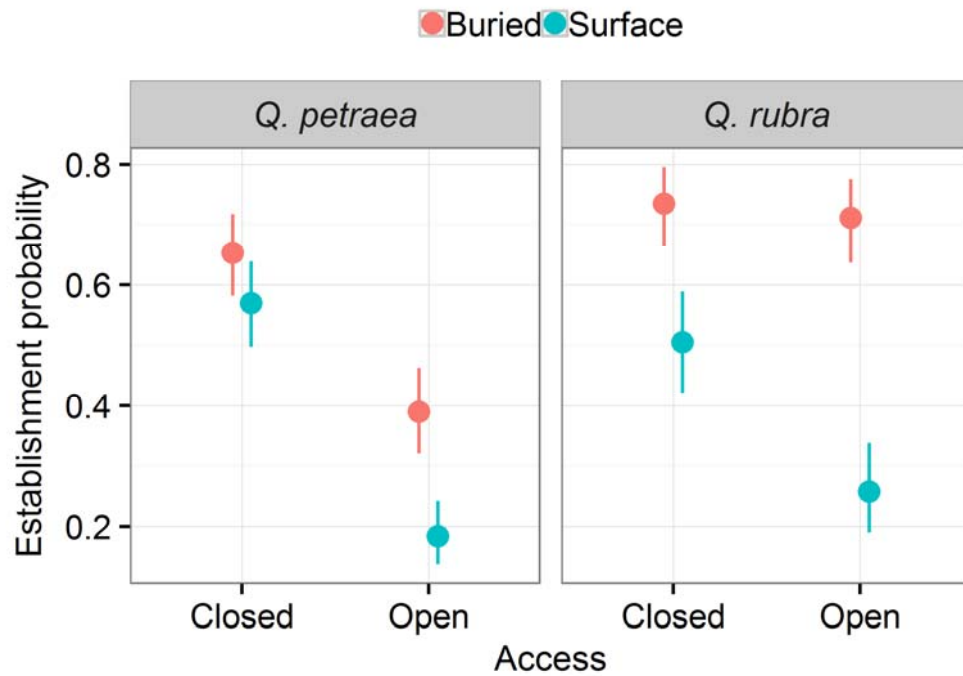
528

| <b>Hypothesis</b>  | <b>Result</b>   |
|--|---|
| (1) Benefits of acorn caching exceed benefits of acorn transportation.                       | Confirmed.  |
| (2) Benefits of acorn transportation is lower in <i>Q. rubra</i> than in <i>Q. petraea</i> . | Not confirmed: in both oak species, seedling establishment did not change with distance of acorn dispersal.                                   |
| (3) Seed removal and pilferage increases with distance from the mother tree.                 | Confirmed.  |
| (4) Pilferage rate is higher in <i>Q. rubra</i> than in <i>Q. petraea</i> .                  | Not confirmed: pilferage rate was lower in <i>Q. rubra</i> , perhaps due to lower emission of volatile compounds.                             |
| (5) Interactions between scatterhoarders and both species of oaks are mutualistic.           | Not confirmed: due to relatively low burial benefits and no distance-related benefits, interaction were antagonistic in both species of oaks. |

529

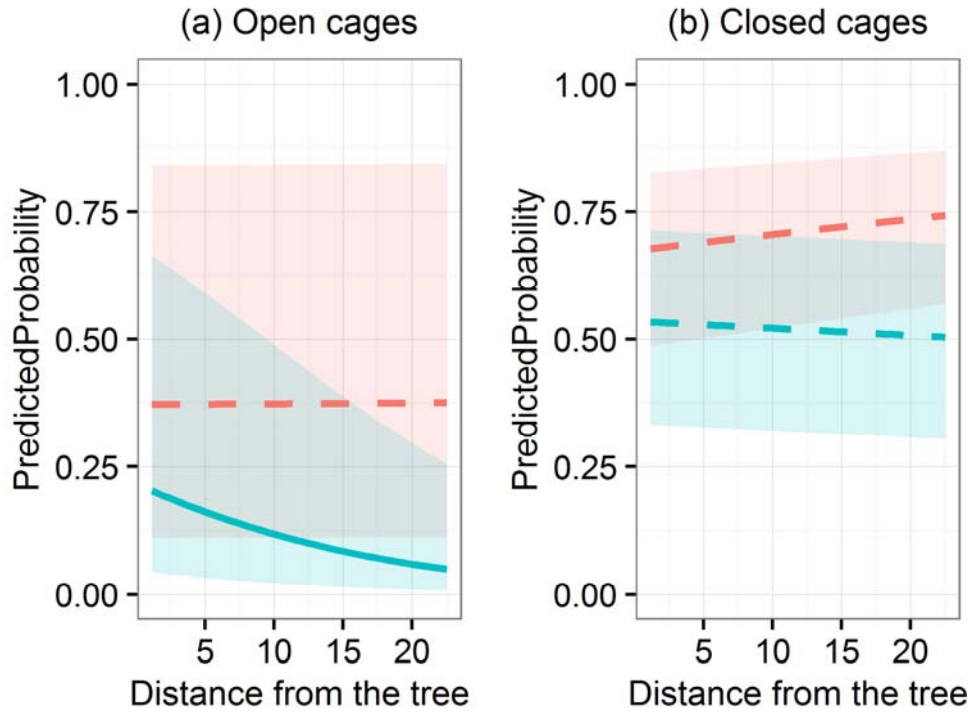
530

531 Figure 1. Probability of seedling establishment of red oaks and sessile oaks in closed (rodent access  
532 excluded), and open (rodent access allowed) cages. Sowed indicate acorns sowed in 1-2 cm in the soil,  
533 while surface indicate acorns sown on soil surface and covered with leaves to mimic autumn leaf fall.  
534 Whiskers indicate standard errors.  
535



536  
537  
538

539 Figure 2. Probability of seedling establishment of sessile oaks (*Quercus petraea*) as a function of  
540 distance from the mother tree. Open cages indicate rodent access, while closed rodent exclusion; red  
541 lines acorns buried in topsoil, blue sowed on soil surface. Solid line indicates relationship significantly  
542 different from 0, while dashed line non-significant relationship.

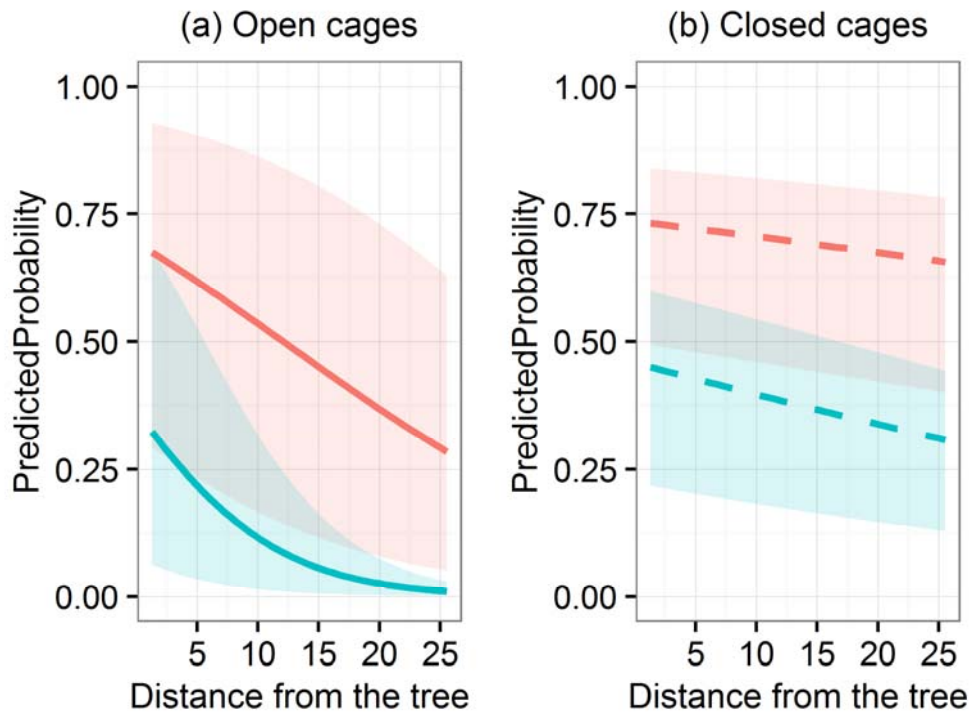


543

544



545 Figure 3. Probability of seedling establishment of red oaks (*Quercus rubra*) as a function of distance  
546 from the mother tree. Open cages indicate rodent access, while closed rodent exclusion; red lines  
547 acorns buried in topsoil, blue sowed on soil surface. Solid line indicates relationship significantly  
548 different from 0, while dashed line non-significant relationship.  
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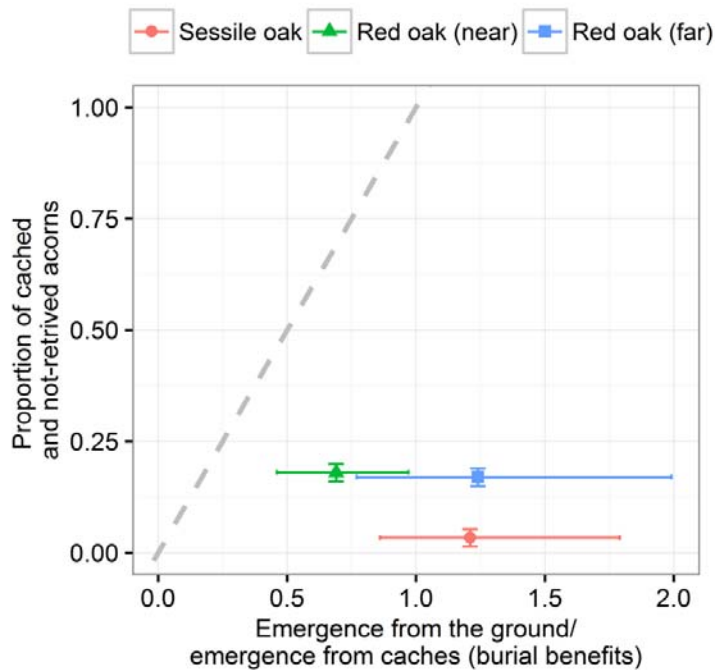
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555 Figure 4. Classification of oak–granivore interactions based on the probability of caching and not  
556 retrieving seeds, and the ratio of seedling emergence from the ground to emergence from caches. The  
557 net effect of granivores is beneficial at any point above the dotted gray line and antagonistic at any  
558 point below. The ‘far’ and ‘near’ categories indicate the establishment ratio calculated based on  
559 germination rate estimated at the distance 0 m (near), and 25 m (far). For sessile oaks, the ratio  
560 components did not differ with the distance from the seed source tree (see Fig. 2). The values on x-  
561 axis (proportion of acorns cached and not retrieved) are derived from parallel study (Bogdziewicz et  
562 al. unpublished).



563

564

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566 Online Supplement 1

567

568 Bogdziewicz M, Crone EE, Zwolak R. Predators or mutualists? Lack of dispersal distance benefits and  
569 small caching profits makes scatterhoarding rodents antagonists of oaks.

570

571 The influence of reaching after pilferage on the outcome of the interaction

572 We took into consideration the possibility of multiple rounds of seed recaching (i.e. some of the  
573 pilfered seeds might be cached again in a different place rather than eaten). We explored possible  
574 effects of recaching, using models based on a pattern from Jansen et al. (2004), who observed nearly  
575 identical rates of seed burial after initial removal (85%) and seed burial after retrieving primary seed  
576 caches (87%). For example, scatterhoarders are mutualists with no recaching if

577 
$$p_{CO}(1 - p_{pilfered}) > \frac{e_s}{e_c} \text{ (eq 2).}$$

578 With one round of recaching, this inequality becomes

579 
$$p_{CO}(1 - p_{pilfered}(p_{CO}(1 - p_{pilfered}))) > \frac{e_s}{e_c},$$

580 with two rounds of recaching, it becomes

581 
$$p_{CO}(1 - p_{pilfered}(p_{CO}(1 - p_{pilfered}(p_{CO}(1 - p_{pilfered})))))) > \frac{e_s}{e_c},$$

582 etc.

583

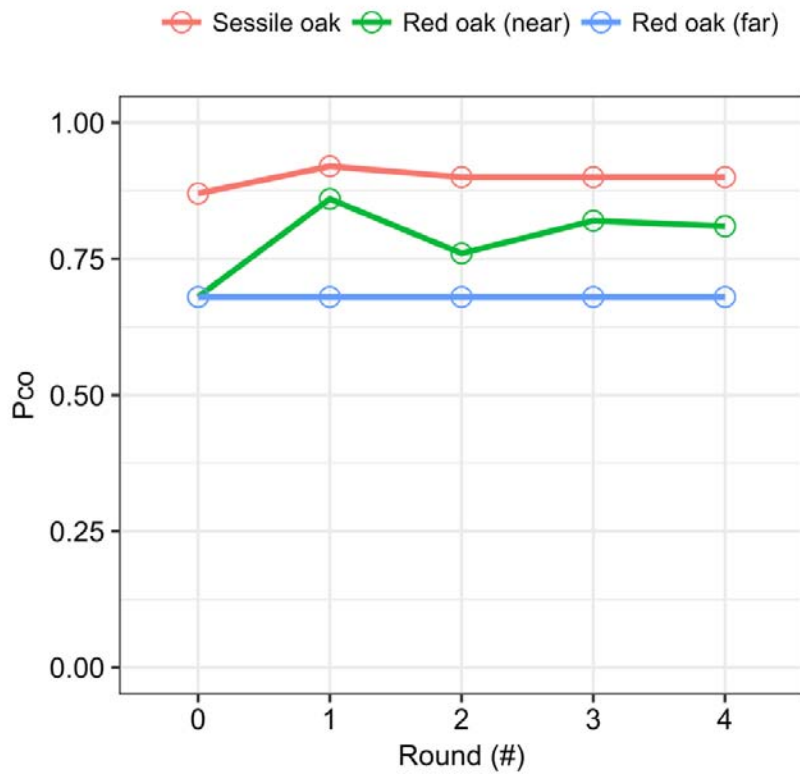
584 If we assume that pilfered seeds are as likely to be eaten or re-cached as seeds collected for the first  
585 time, then the consequences of burial depend on the number of rounds of recaching (Figure 1S).

586 Nonetheless, for the parameters observed in our system, recaching of pilfered seeds would not affect  
587 the conclusion that scatterhoarders are antagonists (see Figure 1S).

588

589

590 **Figure 1S. Possible effects of re-caching on the relative benefits of scatterhoarders for beech**  
591 **emergence.** Following Jansen et al. (2004), these calculations assume that the probabilities of being  
592 eaten, pilfered or recached are the same for all rounds of caching. Lines on the graph are solutions for  
593 the minimum value of  $p_{CO}$  that meets this inequality, for observed values of pilferage and emergence.  
594



595

596

597