1	Title:
2	Weighted trait-abundance early warning signals better predict
3	population collapse
4	
5	<b>Authors:</b> Christopher F. Clements <sup>1,2</sup> *, Martijn van de Pol <sup>3,4</sup> , Arpat Ozgul <sup>1</sup>
6	Affiliations:
7	<sup>1</sup> Department of Evolutionary Biology and Environmental Studies, University
8	of Zurich, Zurich 8057, Switzerland.
9	<sup>2</sup> School of BioSciences, University of Melbourne, Melbourne 3010, Australia.
10	<sup>3</sup> Department of Ecology & Evolution, Research School of Biology,
11	Australian National University, Canberra 2601, Australia.
12	<sup>4</sup> Department of Animal Ecology, Netherlands Institute of Ecology (NIOO-
13	KNAW), Wageningen 6708PB, the Netherlands.
14	Keywords: Body size, extinction, population dynamics, tipping points, trait dynamics
15	
16	Word count: 2489
17	
18	Acknowledgements:
19	This work was made possible by an SNSF Post-doctoral fellowship
20	(P300PA_174359/1) awarded to C.C. and an ERC Starting grant (#337785) to A.O.
21	M.vd.P. was supported by a Future fellowship from the Australian Research Council
22	(FT12010020).

## 23 Abstract:

24 Predicting population collapse in the face of unprecedented anthropogenic 25 pressures is a key challenge in conservation. Abundance-based early warning signals 26 have been suggested as a possible solution to this problem; however, they are known 27 to be susceptible to the spatial and temporal subsampling ubiquitous to abundance 28 estimates of wild population. Recent work has shown that composite early warning 29 methods that take into account changes in fitness-related phenotypic traits - such as 30 body size - alongside traditional abundance-based signals are better able to predict 31 collapse, as trait dynamic estimates are less susceptible to sampling protocols. 32 However, these previously developed composite early warning methods weighted the 33 relative contribution of abundance and trait dynamics evenly. Here we present an 34 extension to this work where the relative importance of different data types can be 35 weighted in line with the quality of available data. Using data from a small-scale 36 experimental system we demonstrate that weighted indicators can improve the 37 accuracy of composite early warning signals by >60%. Our work shows that non-38 uniform weighting can increase the likelihood of correctly detecting a true positive 39 early warning signal in wild populations, with direct relevance for conservation 40 management.

41

42

# 43 Keywords:

44

Body size, extinction, population dynamics, tipping points, trait dynamics

## 46 Introduction

47 Statistical early warning signals (EWSs) calculated from abundance time 48 series data have been suggested as a possible method for predicting approaching 49 population collapses and regime shifts (Drake & Griffen, 2010; Carpenter et al., 2011; 50 Dakos et al., 2012; Kéfi et al., 2013). However, abundance-based early warning 51 signals are known to be susceptible to the spatial and temporal subsampling 52 ubiquitous to wild population abundance estimates (Clements et al., 2015), and have 53 been criticized for not reliably predicting significant declines in natural populations 54 (Burthe et al., 2016). Recent work has sought to resolve these issues by incorporating 55 data on the dynamics of fitness-related phenotypic traits alongside abundance data 56 (Clements & Ozgul, 2018). Traits such as body size are highly responsive to 57 environmental perturbations and changes in the dynamics of these traits often precede 58 demographic responses to deteriorating environments (Anderson et al., 2008; Ozgul et 59 al., 2014; Clements & Ozgul, 2016a). Incorporating information on the shift in the 60 body-size distribution of a population can not only provide an additional measure of 61 stability (Anderson et al., 2008), but has the potential to improve the predictive 62 accuracy of EWS as trait dynamic estimates may be less susceptible to sampling 63 protocols than population abundance estimates are when the distribution of ages and 64 sexes is assumed to be random (spatial partitioning between ages or sexes may affect 65 this) (Clements et al., 2015, 2017; Clements & Ozgul, 2016a). Previous work has 66 shown composite early warning metrics that include data on both abundance and trait 67 dynamics better predict population collapse than those that incorporate abundance-68 only or trait-only data (Clements & Ozgul, 2016a).

Recently developed trait-abundance composite early warning indicators have
been based upon the method proposed by Drake & Griffen (2010), whereby multiple

71 statistical signals are normalized and then summed to create a single composite signal. 72 Clements & Ozgul (2016a) used this approach to incorporate shifts in mean body size 73 and variance in body size along with concurrent changes in the statistical properties of 74 an abundance time series, and demonstrated that such an approach can significantly 75 improve the reliability of early warning signals in both experimental (Clements & 76 Ozgul, 2016a) and natural (Clements et al., 2017) populations. However, in this 77 method the relative importance of abundance versus trait data in the composite 78 indicators was weighted evenly. Given the known issues with abundance data, a 79 logical extension to this method is to non-evenly weight the relatively importance of 80 abundance and trait data in the composite indicators.

81 Non-uniform weighting of model parameters has a history of use in 82 conservation biology, particularly in determining optimal management strategies to 83 maximize outputs from limited resources (Joseph, Maloney & Possingham, 2009). 84 For example habitat conservation may be prioritized based on the suitability of the 85 habitat for certain species, with such weightings often being determined by expert 86 opinion (Lehtomäki et al., 2009). Such approaches have also been used to assess 87 trade-offs, for example between conservation and carbon sequestration (Thomas et al., 88 2013). As well as expert opinion, weighting may be based on more quantitative 89 measures of data quality; for example by the frequency of sampling of a population to 90 estimate abundances, or the percentage of a habitat sampled when counting 91 individuals, both of which have been shown to affect the reliability of early warning 92 signals (Clements et al., 2015). However, practitioners must first discern if non-93 uniform weightings convey an advantage before implementing such an approach for 94 wild populations.

95 Here we assess whether non-uniform weightings improve the predictive ability 96 of composite EWS of population collapse using data from an experimental protozoa 97 study. We take the most reliable composite early warning metric (as identified by 98 Clements & Ozgul (2016a)), and alter the relative weighting of the abundance and 99 trait data when calculating whether an early warning signal is present or not. We then 100 reanalyze the data from an experimental protist microcosm system, presented in 101 Clements & Ozgul (2016a), and show that alternate weightings can improve the 102 predictive ability of composite EWS by decreasing the frequency of false positive 103 signals, and increasing the frequency of true positive signals.

104

105 Methods

# 106 Experimental Data

107 Data on the population dynamics and body-size (width,  $\mu m - a$  proxy for 108 mass) of individuals of a predatory ciliate protozoa (Didinium nasutum) feeding on a 109 bactiverous ciliate protozoa (Paramecium caudatum) were collected over a 47-day 110 period (Fig. 1). Populations of *D. nasutum* were subjected to four different treatments 111 (15 replicates per treatment), where the number of *P. caudatum* fed to each population 112 per day was manipulated. In one treatment ("Constant") populations of D. nasutum 113 were fed 300 P. caudatum per day for the 47 days of the experiment, whilst in the 114 other three treatments the number of P. caudatum declined through time at three 115 different rates ("Slow", "Medium", "Fast") driving the populations of D. nasutum to 116 extinction at varying points in time, and with varying population dynamics prior to 117 extinction (Fig. 1). None of the populations in the Constant treatment went extinct. 118 For each population the time at which it passed through a tipping point, if at all, was 119 calculated (as in Drake & Griffen, 2010), and early warning signals were then

120 calculated prior to the occurrence of each of these tipping points. Because of the size 121 of the microcosms it was impractical to count every individual of a population, hence 122 a subsample was taken (10% of the habitat, a volume that allowed all individuals to 123 be easily counted with close to zero error) and we assumed that the total number of 124 individuals in each microcosm was reflected by the abundance in the subsample. 125 Whilst this undoubtedly introduced some minor error into the abundance estimates, 126 EWSs were still detectable using this uncorrected subsample data (Clements & Ozgul, 127 2016a). We believe that this uncertainty in abundances is very representative of the 128 ubiquitous spatial subsampling associated with the monitoring of all wild populations, 129 and hence applying such methods to this data is a reliable reflection of the challenges 130 of applying them to real world population dynamics. For full details of the 131 experimental design and protocols see Clements & Ozgul (2016a).

132

## 133 *Early warning signals*

134 Previous work has identified a composite index comprised of the coefficient of 135 variation of the abundance time series (cv), shifts in mean body size of the individuals 136 in the population (size), and shifts in the standard deviation of mean body size 137 (sd.size) as producing the most reliable estimates of whether a population was at risk 138 of collapse in these experimental data (Clements & Ozgul, 2016a). Here we test this 139 composite index by systematically altering the weighting of these three competent 140 parts as a proof of concept of non-uniform weighting increasing the predictive 141 accuracy of the composite metric.

Here we implement the approach developed by Clements & Ozgul (2016a).
Each of the three leading indicators (*cv*, *size*, *sd.size*) was calculated at each day
observations were made, and for each experimental population independently. Each

145 leading indicator was then normalized by subtracting the long-run mean of that 146 indicator from the value of that indicator at each time point, and dividing it by the 147 long run standard deviation (Drake & Griffen, 2010; Clements & Ozgul, 2016a) 148 (Supplementary Information). The value of the composite early warning signal was 149 then calculated by summing the value of each leading indicator (cv, size, sd.size) at 150 each time point. Previous work has suggested an EWS could be considered present 151 when the value of this composite EWS exceeds its running mean by either 1 or  $2\sigma$ 152 (Drake & Griffen, 2010). Recent evidence suggesting a  $2\sigma$  threshold provides more 153 reliable results (Clements & Ozgul, 2016a) and consequently here we consider a 154 signal present at a  $2\sigma$  threshold.

155 The weighting of each of the three leading indicators was altered by 156 multiplying the normalized value of each metric prior to summing them together to 157 calculate the composite EWS. Each leading indicator was weighted from 1 to 10, with 158 every combination of weightings tested (e.g.  $cv_w=1:size_w=2:size.sd_w=5$ , 159  $cv_w = 8:size_w = 4:size_sd_w = 1$ ). The performance of each weighting was assessed by using 160 a "normalized metric score" (Clements & Ozgul, 2016a), calculated by subtracting the 161 proportion of false positives (EWS present in data from the constant treatment) from 162 the proportion of true positives (EWS present in data from the slow, medium, and fast 163 treatments). The highest scoring weighting for each of the slow, medium, and fast 164 treatments was compared to uniform weighting in each of these treatments (Fig. 2a). 165 The best metric when data from all three treatments were grouped together was 166 calculated as the weighting with the highest normalized metric score, and the 167 minimum difference in normalized metric scores between treatments (Fig. 2b). This 168 gave an indication as to the weighting that was most robust to different rates of 169 environmental change, and thus potentially most widely applicable to different170 scenarios.

All analyses were carried out using the statistical software R (R Development
Core Team, 2016), and the code to implement weighted trait-abundance early warning
signals is available as supplementary information.

174

# 175 **Results**

176 Experimental data

177 Non-uniform weighting of the importance of abundance and trait data in early 178 warning indicators can improve the reliability of these methods in predicting 179 population collapses (Fig. 2). The largest improvement (62.5%) was seen when data 180 from the Medium treatment was analyzed, possibly because uniform weighting 181 performed relatively poorly (Fig. 2a). The highest achieved normalized metric score 182 was 0.8 (in the slow treatment), suggesting very high numbers of true positive EWS, 183 and low numbers of false positive EWS (Fig. 2a).

184 When data from three deteriorating treatments was grouped together the 185 weighting that produced the greatest improvement in predictive accuracy weighted 186 the relative importance of cv, size.sd, size as 4:7:4, although the improvement over 187 uniform weighting was not large (Fig. 2b), suggesting that the how fast the pressure 188 on the system changes (known as the rate of forcing (Clements & Ozgul, 2016b)) may 189 be an important factor in determining not only the correct weighting to apply, but 190 also our ability to reliably predict population declines. To highlight this, the 4:7:4 191 weighting performed as well as the best weighting in the fast treatment, average in the 192 medium treatment, and worse than both the uniform and best weighting in the slow 193 treatment (Fig. 2a).

194

# 195 Discussion

Predicting population collapse is a key but challenging goal in conservation biology. Because previously developed EWS that take into account both trait and abundance data are non-system specific, and thus widely applicable, they may be of particular interest. Here we analyze data from a small-scale experimental system and show that non-uniform weighting can improve the reliability and strength of traitabundance early warning signals, but that the strength of this improvement is not uniform across different rates of environmental change.

203 Previous work in small-scale experimental systems has identified a composite 204 metric of cv, size, and size.sd as the most reliable predictor of population collapse in 205 experimental microcosm populations. This method provides improved reliability over 206 methods that are based on either abundance-only or size-only data; however, the 207 method still produces false positive and false negative signals in some populations. 208 Because of the known susceptibility of abundance-based early warning signals to poor 209 quality data (Clements et al., 2015), non-uniformly weighting the components of 210 composite metrics provides an obvious extension to this previous work.

211 Here we demonstrate, using the same experimental data with which the 212 original trait-abundance method was developed, that a weighting of 4:7:4 213 (cv:size.sd:size) provides the greatest overall improvement across all three treatments, 214 with the minimum between-treatment variation in this result (Fig. 2b). Resilience to 215 treatment variation in performance in the experiment is important, as it maximizes the 216 reliability of applying such methods to systems where the rate of change remains 217 unknown. However, the among-treatment variation in the potential advantages of 218 non-uniform weighting should not be ignored, as with weightings other than 4:7:4

219 there were significantly higher normalized metric scores in the medium and slow 220 treatments (Fig 3a). Such a result is likely to be driven by the rate of forcing, known 221 to potentially alter the detectability of EWSs (Clements & Ozgul, 2016b), of the 222 system altering the rates of change of the mean and  $\sigma$  body size of individuals. For 223 example, mean body size rapidly declines in the fast treatment (Fig. 1), whilst in the 224 medium treatment body size decline is more gradual and a weighting towards the 225 coefficient of variation of abundance, rather than towards body size, improves 226 predictive accuracy (Fig. 2). These results suggest that the rate of forcing a system 227 can alter the weighting that produces the most reliable predictions of an approaching 228 population collapse.

229 Generalizing such a result to real-world systems may be problematic, as we 230 cannot assume that the population and trait dynamics of the microcosm system 231 analyzed here are truly representative of all real-world population collapses. Ideally 232 one would select the weighting based on an estimate of the reliability of the available 233 abundance or trait data, and possibly based on the rate of forcing of the system, 234 although doing so is likely to be non-trivial. If, for example, available abundance data 235 are known to be estimated from a survey conducted on a small proportion of the 236 known range of a species, or are temporally limited, it may be prudent to calculate the 237 presence of early warning signals with a bias in favor of trait-based data. A less 238 quantitative option would be to weight metrics based on expert opinion of the 239 reliability of the available data. However, whilst criticism has been leveled at the use 240 of expert opinion in conservation management, it has been shown to be useful if 241 approached with caution (Johnson & Gillingham, 2004; Martin et al., 2005, 2012). 242 Similar caution must be applied to the non-uniform weighting of trait and abundance 243 data in the models presented here, however the significant improvements in the

bioRxiv preprint doi: https://doi.org/10.1101/282087; this version posted March 14, 2018. The copyright holder for this preprint (which was not certified by peer review) is the author/funder. All rights reserved. No reuse allowed without permission.

244 predictive accuracy of these approaches when weighting is non-uniform mean that 245 canvasing expert opinion may be a relatively simple and cost-effective solution to 246 improve predictive accuracy.

247 In conclusion, we demonstrate the possible advantages of non-uniform 248 weighting in an early warning signals framework. This work provides a first step to 249 improving the reliability of recently proposed abundance-trait methods (Clements & 250 Ozgul, 2016a), and may be used to negate some of the known issues that affect 251 abundance-based EWSs (Clements et al., 2015). Future work may seek to make more 252 concrete recommendations on weightings based on qualitative measures such as 253 expert opinion, or more quantitative measures such as measures of data quality, the 254 known level of threat to a species or a population, the trophic level of the species, or 255 its connectedness in a network. One option to tackle this is to use complex size-256 structured community models, such as those commonly used in fisheries (Blanchard 257 et al., 2012; Scott, Blanchard & Andersen, 2014), to simulate shifts in the trait 258 dynamics and abundances of multiple interacting species, allowing alternative 259 weightings of data from various trophic levels to be tested on communities where 260 collapse can be invoked by, for example, overfishing or changing climatic variables.

261

#### 262 **References**

Anderson CNK., Hsieh C., Sandin S a., Hewitt R., Hollowed A., Beddington J., May
RM., Sugihara G. 2008. Why fishing magnifies fluctuations in fish abundance. *Nature* 452:835–9.

Blanchard JL., Jennings S., Holmes R., Harle J., Merino G., Allen JI., Holt J., Dulvy
NK., Barange M. 2012. Potential consequences of climate change for primary
production and fish production in large marine ecosystems. *Philosophical*

# 269 transactions of the Royal Society of London. Series B, Biological sciences

- 270 367:2979–89. DOI: 10.1098/rstb.2012.0231.
- 271 Burthe SJ., Henrys PA., Mackay EB., Spears BM., Campbell R., Carvalho L., Dudley
- 272 B., Gunn IDM., Johns DG., Maberly SC., May L., Newell MA., Wanless S.,
- 273 Winfield IJ., Thackeray SJ., Daunt F. 2016. Do early warning indicators 274 consistently predict nonlinear change in long-term ecological data? *Journal of*
- 275 *Applied Ecology* 53:666–676. DOI: 10.1111/1365-2664.12519.
- 276 Carpenter SR., Cole JJ., Pace ML., Batt R., Brock W a., Cline T., Coloso J., Hodgson
- 277 JR., Kitchell JF., Seekell D a., Smith L., Weidel B. 2011. Early Warnings of
- 278 Regime Shifts: A Whole-Ecosystem Experiment. Science (New York, N.Y.)
  279 332:1079–1082.
- 280 Clements CF., Blanchard JL., Nash K., Hindell M., Ozgul A. 2017. Body size shifts
- and early warning signals precede the historic collapse of whale stocks. *Nature Ecology & Evolution* 1:188.
- Clements CF., Drake JM., Griffiths JI., Ozgul A. 2015. Factors influencing the
  detectability of early warning signals of population collapse. *Am Nat* 186:50–58.
  DOI: in press.
- 286 Clements CF., Ozgul A. 2016a. Including trait-based early warning signals helps
- 287 predict population collapse. *Nature Communications* 7:10984. DOI:
  288 10.1038/NCOMMS10984.
- Clements CF., Ozgul A. 2016b. Rate of forcing and the forecastability of critical
  transitions. *Ecology and Evolution* 6:7787–7793.
- Clements CF., Ozgul A. 2018. Indicators of transitions in biological systems. *Ecology Letters* In press.
- 293 Dakos V., Carpenter SR., Brock WA., Ellison AM., Guttal V., Ives AR., Kéfi S.,

- Livina V., Seekell DA., van Nes EH., Scheffer M. 2012. Methods for detecting
- 295 early warnings of critical transitions in time series illustrated using simulated
- ecological data. *PloS one* 7:e41010. DOI: 10.1371/journal.pone.0041010.
- 297 Drake J., Griffen B. 2010. Early warning signals of extinction in deteriorating
  298 environments. *Nature* 467:456–459.
- Johnson CJ., Gillingham MP. 2004. Mapping uncertainty: Sensitivity of wildlife
  habitat ratings to expert opinion. *Journal of Applied Ecology* 41:1032–1041.
- 301 Joseph LN., Maloney RF., Possingham HP. 2009. Optimal allocation of resources
- among threatened species: a project prioritization protocol. *Conservation Biology*23:328–338. DOI: 10.1111/j.1523-1739.2008.01124.x.
- 304 Kéfi S., Dakos V., Scheffer M., Van Nes EH., Rietkerk M. 2013. Early warning
- 305 signals also precede non-catastrophic transitions. *Oikos* 122:641–648. DOI:
   306 10.1111/j.1600-0706.2012.20838.x.
- Lehtomäki J., Tomppo E., Kuokkanen P., Hanski I., Moilanen A. 2009. Applying
  spatial conservation prioritization software and high-resolution GIS data to a
  national-scale study in forest conservation. *Forest Ecology and Management*258:2439–2449. DOI: 10.1016/j.foreco.2009.08.026.
- 311 Martin TG., Burgman MA., Fidler F., Kuhnert PM., Low-Choy S., Mcbride M.,
- Mengersen K. 2012. Eliciting Expert Knowledge in Conservation Science. *Conservation Biology* 26:29–38.
- 314 Martin TG., Kuhnert PM., Mengersen K., Possingham HP. 2005. The power of expert
- 315 opinion in ecological models using Bayesian methods: Impact of grazing on
- 316 birds. *Ecological Applications* 15:266–280. DOI: 10.1890/03-5400.
- 317 Ozgul A., Bateman AW., English S., Coulson T., Clutton-Brock TH. 2014. Linking
- body mass and group dynamics in an obligate cooperative breeder. *Journal of*

319	Animal Ecology.
-----	-----------------

- R Development Core Team. 2016. R: A language and environment for statisticalcomputing.
- 322 Scott F., Blanchard JL., Andersen KH. 2014. *mizer* : an R package for multispecies,
- trait-based and community size spectrum ecological modelling. *Methods in Ecology and Evolution* 5:1121–1125. DOI: 10.1111/2041-210X.12256.
- 325 Thomas CD., Anderson BJ., Moilanen A., Eigenbrod F., Heinemeyer A., Quaife T.,
- 326 Roy DB., Gillings S., Armsworth PR., Gaston KJ. 2013. Reconciling
- 327 biodiversity and carbon conservation. *Ecology Letters* 16:39–47.

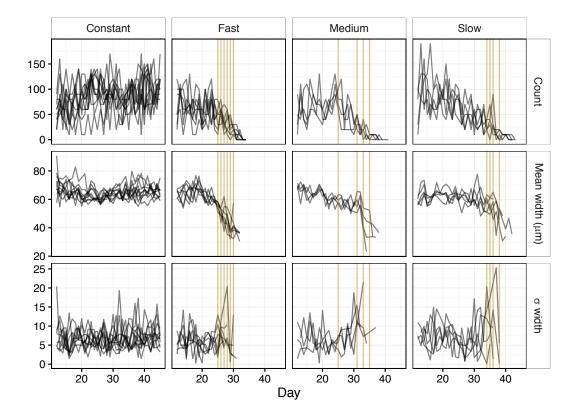


Figure 1. Black lines describe the population and body size dynamics of individual populations of *Didinium nasutum* subjected to four different experimental treatments (constant, fast, medium, and slow rates of decline in prey availability). Data from day 0 to 12 were removed to minimize the effects of transitory dynamics. Each vertical gold line indicates an inferred tipping points for a collapsing population.

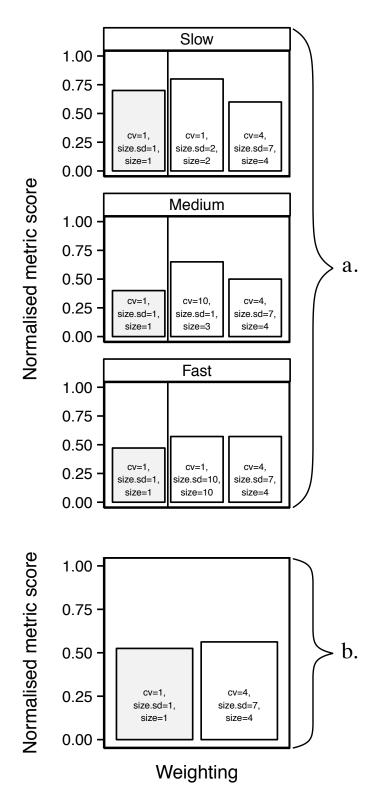


Figure 2. (a) The highest scoring weighting across each of the experimental treatments compared to even weighting and the best weighting when data from all treatments were combined, and (b) the weighting with the highest normalized metric score across all three treatments, and the lowest difference in normalized metric score amongst treatments.