

1 The Genome of BAM-degrading *Aminobacter* 2 sp. MSH1 with Several Low Copy Plasmids

3 Tue Kjærgaard Nielsen¹, Ole Hylling¹, Lea Ellegaard-Jensen¹, Jens Aamand², Lars Hestbjerg Hansen¹

4 ¹Department of Environmental Science, Aarhus University, Roskilde, Denmark

5 ²Department of Geochemistry, Geological Survey of Denmark and Greenland, Copenhagen, Denmark

6 Abstract

7 As one of the only described degraders of the recalcitrant metabolite 2,6-dichlorobenzamide (BAM) of the
8 pesticide dichlobenil, *Aminobacter* sp. MSH1 has been intensively studied for its characteristics with regards
9 to physiology and its use in bioremediation. Two plasmid sequences from strain MSH1 have previously been
10 published, while the remaining genome sequence has been left uninvestigated. We here present the
11 complete genome sequence of this important strain, which consists of a chromosome, two megaplasmids
12 and five smaller plasmids. Intriguingly, the plasmid copy numbers are mostly below one per bacterial
13 chromosome, indicating that plasmids in strain MSH1 are under very unstable conservation. The results of
14 this report improve our understanding of the genomic dynamics of *Aminobacter* sp. MSH1.

15 Introduction

16 *Aminobacter* sp. MSH1 is a Gram-negative, motile rod isolated from a plant nursery courtyard soil, previously
17 exposed to the pesticide dichlobenil (Sørensen et al. 2007). *Aminobacter* MSH1 and two other *Aminobacter*
18 strains (ASI1 and ASI2) are the only bacteria reported in the literature capable of utilizing the recalcitrant
19 dichlobenil metabolite 2,6-dichlorobenzamide (BAM) as their sole carbon source and mineralize it at
20 nanomolar concentrations (Simonsen et al. 2006; Sørensen et al. 2007). This characteristic is of immense
21 interest since BAM is among the most common micro-pollutant found in groundwater aquifers of several
22 European countries (Porazzi et al. 2005; Törnquist et al. 2007; Schipper et al. 2008; Pukkila and Kontro 2014;
23 Vandermaesen et al. 2016), often exceeding drinking water threshold limits set by the European Union (0.1
24 µg/L) (The Council of the European Union 1998).

25 Since many countries rely partially or solely on groundwater for drinking water production, remediation of
26 pesticide polluted groundwater by bioaugmentation is proposed as a cost-effective and sustainable
27 biotechnological method (Benner et al. 2013). Strain MSH1 has been suggested as a promising candidate to
28 remediate BAM-polluted groundwater, and thus prevent the costly closures of abstraction wells in drinking

29 water production (Björklund et al. 2011). Laboratory and pilot scale studies have applied MSH1 to drinking
30 water treatment sand filter (SF) units for biological removal of BAM contamination (Albers et al. 2015;
31 Ellegaard-Jensen et al. 2016; Horemans, Raes, Vandermaesen, et al. 2017). However, MSH1 is challenged by
32 the oligotrophic environment of SFs that may inhibit degradation and maintenance of the MSH1 population
33 (Ellegaard-Jensen et al. 2017).

34 Over the past decade, studies of MSH1 have uncovered and described several characteristics, important in
35 this context; BAM-catabolic genes, enzyme characterization, metabolic pathways, adhesion properties,
36 surface colonization and invasion of natural SF communities (Simonsen et al. 2012; Albers et al. 2014; T'Syen
37 et al. 2015; Sekhar et al. 2016; Horemans, Raes, Brocatus, et al. 2017; Horemans, Vandermaesen, et al. 2017).
38 However, despite comprehensive attention, whole genome sequencing of MSH1 is still absent from
39 literature. We report for the first time the complicated complete genome of *Aminobacter* sp. MSH1, which
40 consists of a chromosome, two megaplastids, and 5 plasmids in the range of 31-97 kb in size.

41 Materials and Methods

42 *Aminobacter* sp. MSH1 was obtained from the strain collection of the laboratory that originally isolated the
43 bacterium. Strain MSH1 was grown in shaking R2B medium for 72 hours at 22°C. 2 ml culture was used for
44 extraction of high molecular weight (HMW) DNA using the MasterPure™ DNA Purification Kit (Epicentre,
45 Madison, WI, USA), using the protocol for cell samples. The purity and concentration of extracted DNA were
46 recorded with a NanoDrop 2000c and a Qubit® 2.0 fluorometer (Thermo Fisher Scientific, Waltham, MA, USA),
47 respectively. An Illumina Nextera XT library was prepared for paired-end sequencing on an Illumina NextSeq
48 500 with a Mid Output v2 kit (300 cycles) (Illumina Inc., San Diego, CA, USA). For Oxford Nanopore
49 sequencing, a library was prepared using the Rapid Sequencing kit (SQK-RAD004). This was loaded on an R9.4
50 flow cell and sequenced using MinKnow (v1.11.5). Nanopore reads were basecalled with albacore (v2.1.10)
51 without quality filtering of reads. Only reads longer than 5,000 bp were retained and sequencing adapters
52 were trimmed using porechop (v0.2.3). For Illumina sequencing, 2x151 bp paired-end reads were trimmed
53 for contaminating adapter sequence and low quality bases (<Q20) in the ends of the reads were removed
54 using Cutadapt (v1.8.3) (Martin 2011). Paired-end reads that overlapped were merged with AdapterRemoval
55 (v2.1.0) (Martin 2011). A hybrid genome assembly with Nanopore and Illumina reads was performed using
56 Unicycler (v0.4.3) (Wick et al. 2017). An automatic annotation was performed with Prokka (v1.11) (Seemann
57 2014) which was manually curated for plasmids pBAM1 and pBAM2 using BLAST and the associated
58 Conserved Domain Database (Altschul et al. 1990; Marchler-Bauer et al. 2015). Sequence alignment was

59 produced with MUSCLE (v3.0) (Edgar 2008), as implemented in CLC Genomics Workbench (v11) (QIAGEN,
60 Hilden, Germany).

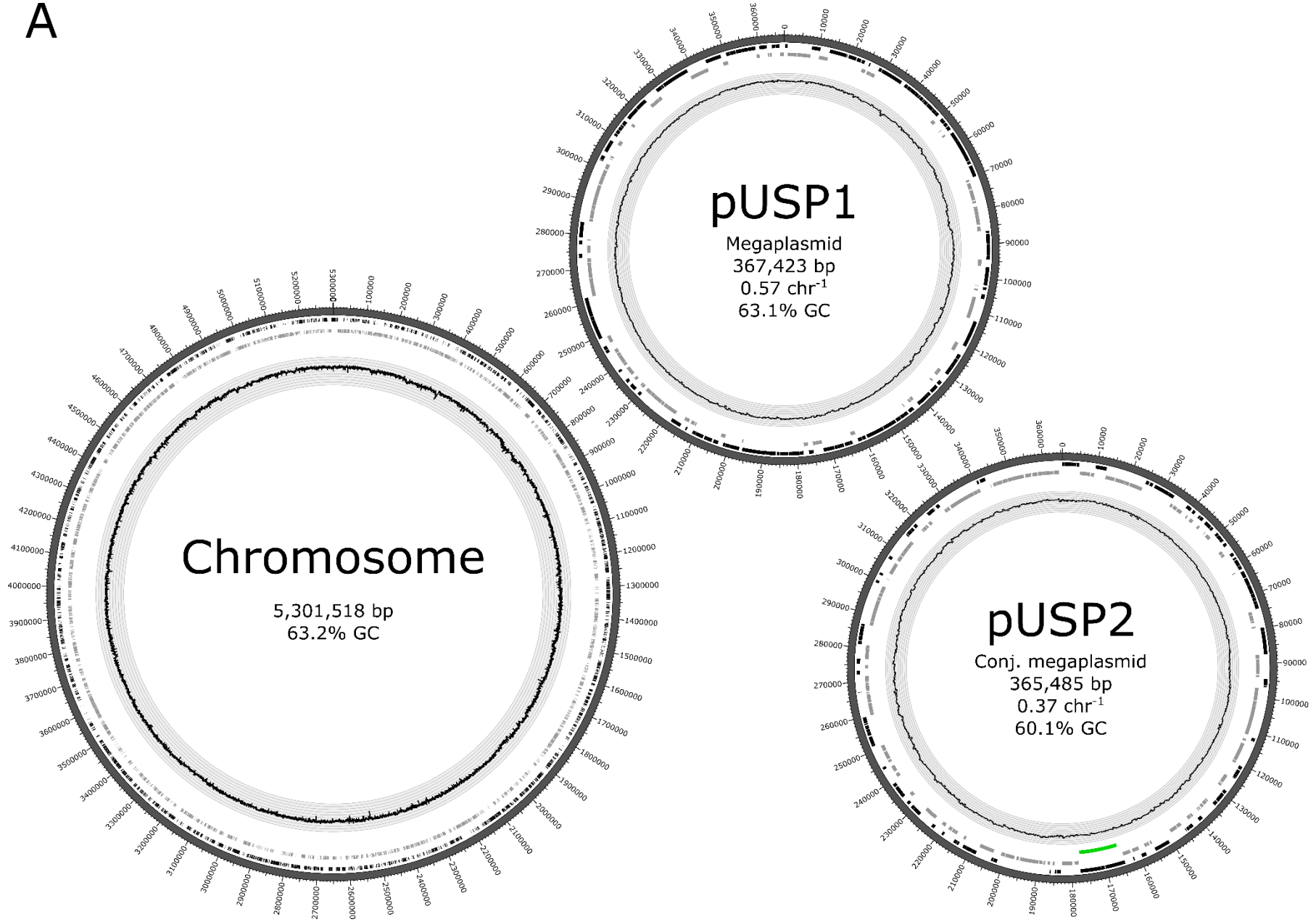
61 The complete genome assembly of *Aminobacter* sp. MSH1 was deposited in GenBank under accession
62 numbers CP028966-CP028973.

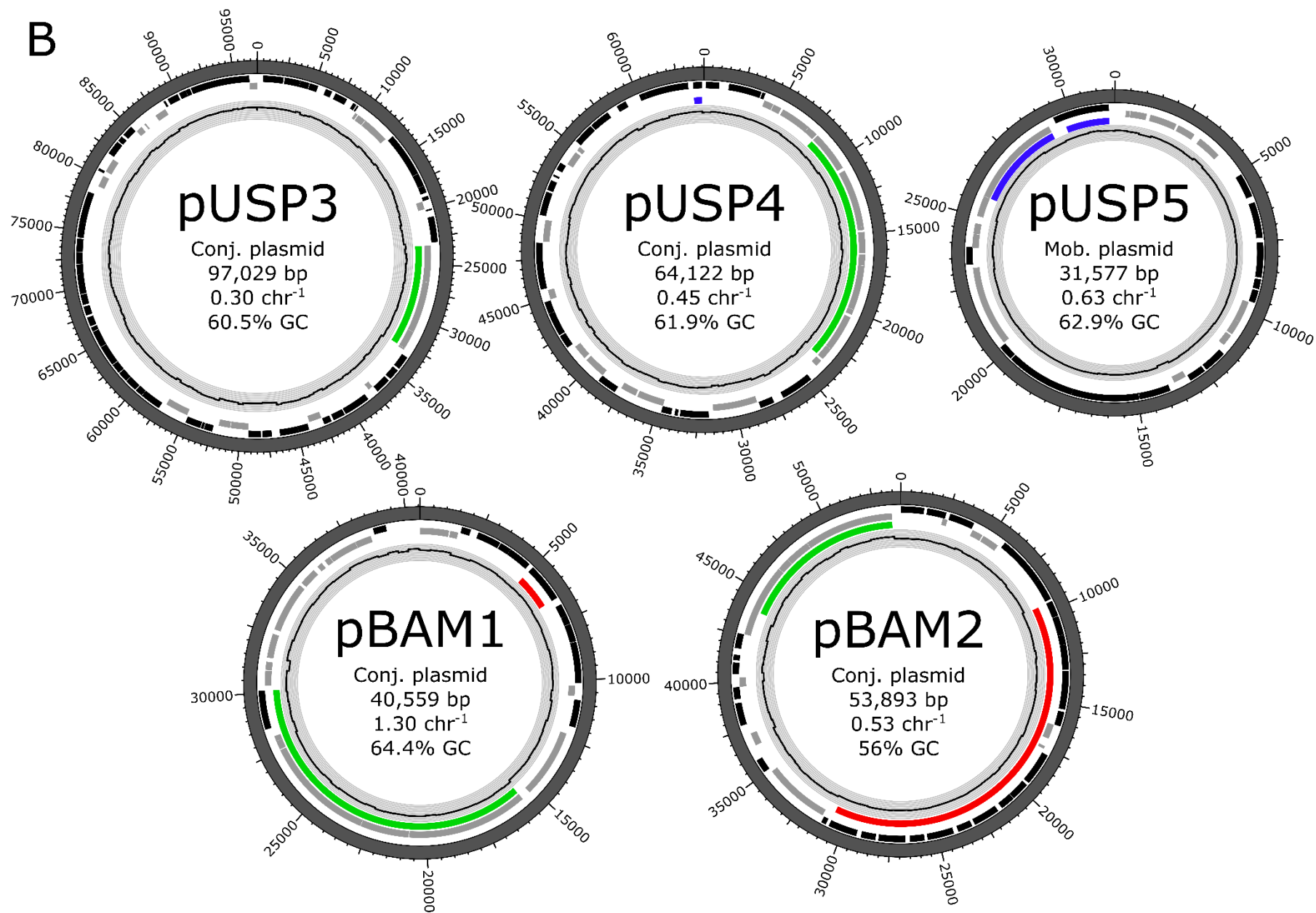
63 Results and Discussion

64 **Complete genome of *Aminobacter* sp. MSH1**

65 Sequencing of *Aminobacter* sp. MSH1 on Illumina NextSeq yielded 8.5M paired-end reads before quality
66 filtering. 2.9M filtered paired-end reads were not overlapping, while 2.7M reads were merged for a total
67 Illumina data size of 661 Mbp. Nanopore sequencing resulted in 287K reads with a total data size of 5.5 Gbp,
68 an average read length of 19 kb, and a read N50 of 33 kb. After trimming of Nanopore adapters and removal
69 of short reads (<5 kb), 213K Nanopore reads with a total of 5.27 Gbp remained for hybrid genome assembly
70 which resulted in eight circular and closed DNA replicons (Figure 1).

A





73 Figure 1 (above). Complete genome assembly of *Aminobacter* sp. MSH1 with eight replicons determined by Unicycler
74 assembler to be complete and circular. The complete genome consists of a chromosome and two megaplasmids (A) and
75 five plasmids ranging in size from 97 kb to 31.6 kb (B). For all replicons, the GC content (window size 500 nt) is shown
76 as a graph with grey background. Genes on forward strands are shown as black bars, while genes on reverse strands are
77 shown as grey bars. Type IV secretion systems for conjugative plasmid transfer are highlighted with green bars, while
78 genes putatively involved in plasmid mobilization are marked with blue bars. The *bbdA* gene on pBAM1 and the *bbd*
79 gene cluster on pBAM2 are highlighted with red bars. The copy number per chromosome of the plasmids are reported
80 by the Unicycler assembler and are based solely on Illumina data.

81 Megaplasmid pUSP1 has no apparent genes encoding conjugative machinery or plasmid mobilization,
82 suggesting that this replicon might be a possible secondary chromosome. However, the low copy number of
83 0.57 per chromosome indicates that this replicon is not essential for survival of strain MSH1. The second
84 megaplasmid, pUSP2, does harbor a *vir* operon encoding a type IV secretion system (T4SS) for plasmid
85 conjugation. Furthermore, plasmids pUSP3-4 and pBAM1-2 each have genes encoding putative T4SS
86 proteins, whereas pUSP5 seems to possibly be a mobilizable plasmid due to presence of genes encoding a
87 TraG coupling protein and TraA conjugative transfer relaxase but no T4SS. Plasmids pUSP1-3 and pBAM2 are
88 all *repABC* family plasmids. pBAM1 was previously found to be an IncP1 plasmid through analysis of the
89 plasmid replication initiator gene *trfA* (T'Syen et al. 2015).

90 Surprisingly, only plasmid pBAM1 had a copy number per chromosome higher than one. It was previously
91 reported that, under non-selective BAM-free conditions, a phenotype of MSH1 incapable of BAM
92 mineralization rapidly increased in abundance. It was speculated to be caused by a loss of plasmid pBAM2,
93 harboring genes responsible for degradation of 2,6-DCBA – the first metabolite of BAM. The stability of BAM
94 mineralization was monitored with qPCR targeting the single copy *bbdA* and *bbdD* genes on pBAM1 and
95 pBAM2, respectively, and calculating their abundance relative to 16S rRNA gene copy numbers. It was found
96 that a culture grown in mineral salt medium with 200 mg/liter BAM as sole carbon source sustained gene
97 copy ratios of *bbdA*/16S and *bbdD*/16S of 2.4 and 2.2, respectively. While the *bbdA*/16S rRNA ratio was
98 sustained under varying incubation settings, the *bbdD*/16S ratio rapidly diminished when other C sources
99 than BAM were available (Horemans, Raes, Brocatus, et al. 2017). The complete genome of strain MSH1
100 includes two 16S rRNA gene copies on the chromosome. This results in a plasmid copy number of pBAM1
101 and pBAM2 under BAM selective conditions of 1.2 and 1.1, respectively, based on the previously determined
102 *bbdA*- and *bbdD* to 16S rRNA gene ratios. For pBAM1, this approximately agrees with the observed plasmid
103 copy number based on Illumina data coverage. However, for the remaining six plasmids, the copy number
104 per chromosome is curiously below one, indicating that these plasmids may be unstable in strain MSH1 or
105 may be maintained at unusually low copy numbers compared to the chromosome. Future studies should
106 investigate the unusually low copy numbers per chromosome of the plasmids in MSH1, perhaps owing to a
107 possible chromosomal polyploidy as seen in multiple bacterial lineages (Oliverio and Katz 2014).

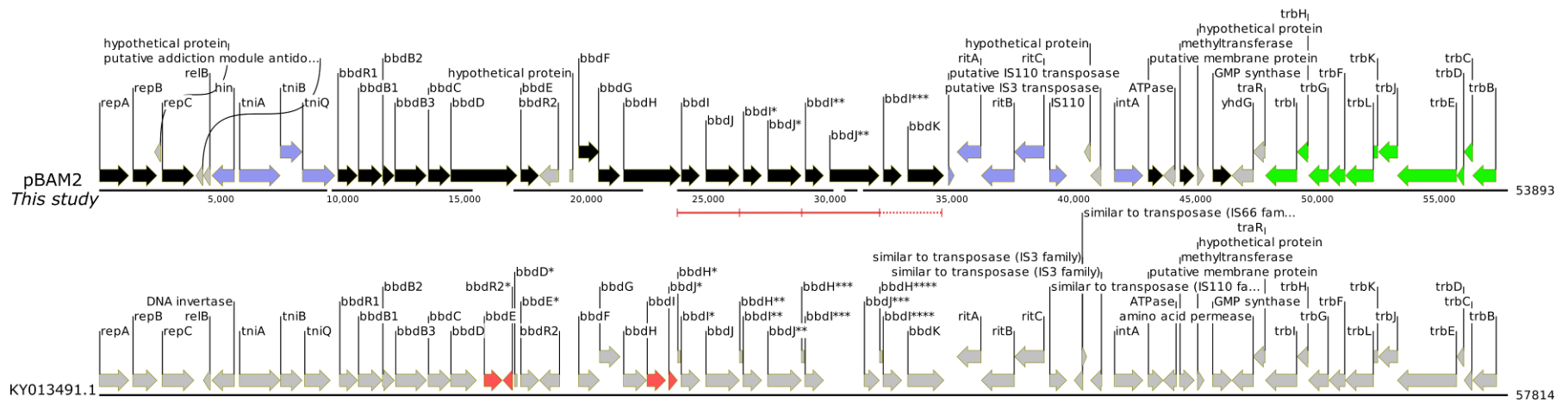
108 **Characteristics of pBAM1**

109 The first step in degradation of BAM is performed by the BbdA amidase which converts it to 2,6-
110 dichlorobenzoic acid (DCBA) (T'Syen et al. 2015). The *bbdA* gene is located on pBAM1 (Figure 1), while two
111 *bbd* gene clusters located on pBAM2 encodes enzymes converting DCBA to intermediate compounds in the
112 Krebs cycle (Horemans, Raes, Brocatus, et al. 2017). Only pBAM1 is maintained in the population at a copy
113 number of more than one per chromosome, suggesting that it is either more stable than the other plasmids
114 or that there are multiple copies of the chromosome per cell.

115 The gene encoding the pivotal BbdA amidase on pBAM1, is located immediately upstream of two genes
116 encoding a putative IS5 transposase. While no closely related homologs of BbdA currently exists in the NCBI
117 nr database, the putative IS5 transposase has 90% nucleotide similarity to sequences from *Comamonadaceae*
118 bacteria, indicating that this transposase has a broad phylogenetic dispersal. The proximity of *bbdA* to a
119 transposase furthermore indicates that this gene has been transferred from an unknown origin to plasmid
120 pBAM1.

121 **Characteristics of pBAM2**

122 Curiously, a 2,557 bp region on pBAM2, encoding the genes for glutathione S-transferase (*bbdI*) and
123 glutathione-disulfide reductase (*bbdJ*), has apparently undergone duplication and occurs in three perfect
124 consecutive repeats followed by one imperfect repeat (Figure 2). These genes are part of the cluster that is
125 responsible for the conversion of 2,6-DCBA to Krebs cycle intermediates, as previously established
126 (Horemans, Raes, Brocatus, et al. 2017). The fourth, imperfect, repeat encompasses the gene previously
127 annotated as the 1,500 bp gene *bbdK* encoding a glutathione-disulfide reductase. This gene is likely derived
128 from the 1,392 bp *bbdJ* genes occurring in the upstream three perfect repeats, as the first 1,377 bp are
129 identical in *bbdJ* and *bbdK* (results not shown). It is not known whether *bbdK* is a non-functional pseudogene
130 derived from *bbdJ*, but considering the redundancy of genes encoding glutathione-disulfide reductases on
131 pBAM2, it seems likely that *bbdK* has been subjected to spurious mutations.



133

134 Figure 2. Plasmid pBAM2 from this study aligned with MUSCLE to the already published variant. The alignment is shown as black lines under the genes for both
 135 sequences. Notice that gaps are introduced by MUSCLE in the 53,893 bp pBAM2 sequence to align to the previously published 57,814 bp KY013491.1 sequence. For
 136 pBAM2 (this study), genes on forward strands are in black, while genes on reverse strands are in grey. Genes encoding putative mobile genetic elements are highlighted
 137 in blue, while the T4SS genes are highlighted in green. The four repeats of *bbdII* are underscored in red brackets with the fourth, imperfect, repeat shown with a dotted
 138 bracket.

139 The *bbdB-KR* gene region in the presented complete genome assembly shows some discrepancies with the
140 sequence of pBAM2 that was previously deposited in GenBank (accession number KY013491.1). Specifically,
141 the previous pBAM2 assembly is 3,921 bp longer and contains additional copies of the *bbdE*, *bbdR2*, *bbdI*,
142 and *bbdJ* genes (Figure 2). The structure of pBAM2, as shown in this study, was verified by mapping of 250
143 Nanopore reads (minimum length 50,000 bp) to the plasmid sequence, confirming that the pBAM1 assembly
144 is complete. While the MSH1 strain used for sequencing in this study was obtained from the laboratory from
145 which it was originally isolated, it cannot be excluded that genetic rearrangements may have occurred
146 between laboratory strains which have led to the observed discrepancies.

147 Except for the plasmid backbone (*trb* and *rep* genes), there are no significantly similar sequences in the NCBI
148 nr/nt database (BLASTN (Altschul et al. 1990) search on April 11th 2018), which obstructs any attempts to
149 derive how pBAM2 and the *bbd* gene cluster has evolved. Future studies will hopefully lead to the discovery
150 of sequences related to pBAM2. However, several putative mobile genetic elements are located near the *bbd*
151 gene cluster, including the *ritABC* genes that were recently shown to form circular DNA intermediate
152 molecules as a likely intermediate in transposition (Nielsen et al. 2017). This makes it probable that the *bbd*
153 genes have been transferred from another replicon to pBAM2. Furthermore, pBAM2 has a lower GC content
154 than the remaining genome, indicating that it has been transferred to MSH1 from another bacterium.
155 Whereas BAM mineralization occur only rarely in environmental samples, 2,6-DCBA mineralization is much
156 more common (Vandermaesen et al. 2016), supporting that the *bbdB-KR* genes are more widely spread than
157 the *bbdA* gene.

158

159 References

- 160 Albers CN, Feld L, Ellegaard-Jensen L, Aamand J. 2015. Degradation of trace concentrations of the persistent
161 groundwater pollutant 2,6-dichlorobenzamide (BAM) in bioaugmented rapid sand filters. *Water Res.* 83:61–
162 70. doi:10.1016/j.watres.2015.06.023.
- 163 Albers CN, Jacobsen OS, Aamand J. 2014. Using 2,6-dichlorobenzamide (BAM) degrading *Aminobacter* sp.
164 MSH1 in flow through biofilters - Initial adhesion and BAM degradation potentials. *Appl. Microbiol.*
165 *Biotechnol.* 98:957–967. doi:10.1007/s00253-013-4942-6.
- 166 Altschul SF, Gish W, Miller W, Myers EW, Lipman DJ. 1990. Basic local alignment search tool. *J. Mol. Biol.*
167 215:403–10. doi:10.1016/S0022-2836(05)80360-2.
- 168 Benner J, Helbling DE, Kohler H-PE, Wittebol J, Kaiser E, Prasse C, Ternes TA, Albers CN, Aamand J,
169 Horemans B, et al. 2013. Is biological treatment a viable alternative for micropollutant removal in drinking
170 water treatment processes? *Water Res.* 47:5955–5976. doi:10.1016/j.watres.2013.07.015.
- 171 Björklund E, Styrishave B, Anskjær GG, Hansen M, Halling-Sørensen B. 2011. Dichlobenil and 2,6-
172 dichlorobenzamide (BAM) in the environment: What are the risks to humans and biota? *Sci. Total Environ.*
173 409:3732–3739. doi:10.1016/j.scitotenv.2011.06.004.
- 174 Edgar RC. 2008. Muscle. *BMC Bioinformatics.*
- 175 Ellegaard-Jensen L, Horemans B, Raes B, Aamand J, Hansen LH. 2017. Groundwater contamination with 2,6-
176 dichlorobenzamide (BAM) and perspectives for its microbial removal. *Appl. Microbiol. Biotechnol.*
177 101:5235–5245. doi:10.1007/s00253-017-8362-x.
- 178 Ellegaard-Jensen L, Albers CN, Aamand J. 2016. Protozoa graze on the 2,6-dichlorobenzamide (BAM)-
179 degrading bacterium *Aminobacter* sp. MSH1 introduced into waterworks sand filters. *Appl. Microbiol.*
180 *Biotechnol.* 100:8965–8973. doi:10.1007/s00253-016-7710-6.
- 181 Horemans B, Raes B, Brocatus H, T'Syen J, Rombouts C, Vanhaecke L, Hofkens J, Springael D. 2017. Genetic
182 (in)stability of 2,6-dichlorobenzamide catabolism in *Aminobacter* sp. strain MSH1 biofilms under carbon
183 starvation conditions. *Appl. Environ. Microbiol.* 83. doi:10.1128/AEM.00137-17.
- 184 Horemans B, Raes B, Vandermaesen J, Simanjuntak Y, Brocatus H, T'Syen J, Degryse J, Boonen J, Wittebol J,
185 Lapanje A, et al. 2017. Biocarriers Improve Bioaugmentation Efficiency of a Rapid Sand Filter for the
186 Treatment of 2,6-Dichlorobenzamide-Contaminated Drinking Water. *Environ. Sci. Technol.* 51:1616–1625.
187 doi:10.1021/acs.est.6b05027.
- 188 Horemans B, Vandermaesen J, Sekhar A, Rombouts C, Hofkens J, Vanhaecke L, Springael D. 2017.

- 189 Aminobacter sp. MSH1 invades sand filter community biofilms while retaining 2,6-dichlorobenzamide
190 degradation functionality under C- and N-limiting conditions. FEMS Microbiol. Ecol. 93.
191 doi:10.1093/femsec/fix064.
- 192 Marchler-Bauer A, Derbyshire MK, Gonzales NR, Lu S, Chitsaz F, Geer LY, Geer RC, He J, Gwadz M, Hurwitz
193 DI, et al. 2015. CDD: NCBI's conserved domain database. Nucleic Acids Res. 43:D222–D226.
194 doi:10.1093/nar/gku1221.
- 195 Martin M. 2011. Cutadapt removes adapter sequences from high-throughput sequencing reads.
196 EMBnet.journal 17:10. doi:10.14806/ej.17.1.200.
- 197 Nielsen TK, Rasmussen M, Demanèche S, Cecillon S, Vogel TM, Hansen LH. 2017. Evolution of
198 Sphingomonad Gene Clusters Related to Pesticide Catabolism Revealed by Genome Sequence and
199 Mobilomics of *Sphingobium herbicidovorans* MH. Genome Biol. Evol. 9:2477–2490.
200 doi:10.1093/gbe/evx185.
- 201 Oliverio AM, Katz LA. 2014. The dynamic nature of genomes across the tree of life. Genome Biol. Evol.
202 6:482–488. doi:10.1093/gbe/evu024.
- 203 Porazzi E, Martinez MP, Fanelli R, Benfenati E. 2005. GC-MS analysis of dichlobenil and its metabolites in
204 groundwater. Talanta 68:146–154. doi:10.1016/j.talanta.2005.04.044.
- 205 Pukkila V, Kontro MH. 2014. Dichlobenil and 2,6-dichlorobenzamide (BAM) dissipation in topsoil and
206 deposits from groundwater environment within the boreal region in southern Finland. Environ. Sci. Pollut.
207 Res. 21:2289–2297. doi:10.1007/s11356-013-2164-1.
- 208 Schipper PNM, Vissers MJM, van der Linden AMA. 2008. Pesticides in groundwater and drinking water
209 wells: Overview of the situation in the Netherlands. Water Sci. Technol. 57:1277–1286.
210 doi:10.2166/wst.2008.255.
- 211 Seemann T. 2014. Prokka: Rapid prokaryotic genome annotation. Bioinformatics 30:2068–2069.
212 doi:10.1093/bioinformatics/btu153.
- 213 Sekhar A, Horemans B, Aamand J, Sørensen SR, Vanhaecke L, Bussche J Vanden, Hofkens J, Springael D.
214 2016. Surface Colonization and Activity of the 2,6-Dichlorobenzamide (BAM) Degrading *Aminobacter* sp.
215 Strain MSH1 at Macro- and Micropollutant BAM Concentrations. Environ. Sci. Technol. 50:10123–10133.
216 doi:10.1021/acs.est.6b01978.
- 217 Simonsen A, Badawi N, Anskjær GG, Albers CN, Sørensen SR, Sørensen J, Aamand J. 2012. Intermediate
218 accumulation of metabolites results in a bottleneck for mineralisation of the herbicide metabolite 2,6-
219 dichlorobenzamide (BAM) by *Aminobacter* spp. Appl. Microbiol. Biotechnol. 94:237–245.

- 220 doi:10.1007/s00253-011-3591-x.
- 221 Simonsen A, Holtze MS, Sørensen SR, Sørensen SJ, Aamand J. 2006. Mineralisation of 2,6-
222 dichlorobenzamide (BAM) in dichlobenil-exposed soils and isolation of a BAM-mineralising *Aminobacter* sp.
223 *Environ. Pollut.* 144:289–295. doi:10.1016/j.envpol.2005.11.047.
- 224 Sørensen SR, Holtze MS, Simonsen A, Aamand J. 2007. Degradation and mineralization of nanomolar
225 concentrations of the herbicide dichlobenil and its persistent metabolite 2,6-dichlorobenzamide by
226 *Aminobacter* spp. isolated from dichlobenil-treated soils. *Appl. Environ. Microbiol.* 73:399–406.
227 doi:10.1128/AEM.01498-06.
- 228 T'Syen J, Tassoni R, Hansen L, Sorensen SJ, Leroy B, Sekhar A, Wattiez R, De Mot R, Springael D. 2015.
229 Identification of the Amidase BbdA That Initiates Biodegradation of the Groundwater Micropollutant 2,6-
230 dichlorobenzamide (BAM) in *Aminobacter* sp. MSH1. *Environ. Sci. Technol.* 49:11703–11713.
231 doi:10.1021/acs.est.5b02309.
- 232 The Council of the European Union. 1998. Council Directive 98/93/EC of November 1998 on the equality of
233 water intended for human consumption. *Off. J. Eur. Communities* 330:32–54.
- 234 Törnquist M, Kreuger J, Adielsson S. 2007. Occurrence of pesticides in Swedish water resources against a
235 background of national risk-reduction programmes - results from 20years of monitoring.
- 236 Vandermaesen J, Horemans B, Degryse J, Boonen J, Walravens E, Springael D. 2016. Mineralization of the
237 Common Groundwater Pollutant 2,6-Dichlorobenzamide (BAM) and its Metabolite 2,6-Dichlorobenzoic
238 Acid (2,6-DCBA) in Sand Filter Units of Drinking Water Treatment Plants. *Environ. Sci. Technol.* 50:10114–
239 10122. doi:10.1021/acs.est.6b01352.
- 240 Wick RR, Judd LM, Gorrie CL, Holt KE. 2017. Unicycler: Resolving bacterial genome assemblies from short
241 and long sequencing reads. *PLoS Comput. Biol.* 13. doi:10.1371/journal.pcbi.1005595.
- 242