

1 Large-scale experimental removal of non-native slider turtles has unexpected consequences on
2 basking behavior for both conspecifics and a native, threatened turtle

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39 **Abstract:**

40 The red-eared slider turtle (*Trachemys scripta elegans*; RES) is one of the world's most invasive
41 species. Native to the central United States, RES are now widely established in freshwater
42 habitats across the globe, largely due to release of unwanted pets. Laboratory and mesocosm
43 experiments suggest that introduced RES are competitively dominant to native turtles, but such
44 competition remains untested in the wild. Here, we experimentally removed introduced RES to
45 test whether they compete for critical basking habitat with native, threatened western pond turtles
46 (*Emys marmorata*; WPT), a species being considered for listing under the U.S. Endangered
47 Species Act. Following removal, we found that both the remaining RES as well as WPT altered
48 their basking distribution but in a manner inconsistent with strong interspecific competition.
49 However, these findings suggest strong intraspecific competition for basking sites amongst RES
50 and that interspecific competition between WPT and introduced RES likely occurs at higher RES
51 densities. Our works suggests RES influence the behavior of native species in the wild and
52 indicates that RES removal may be most beneficial at high RES densities. This experiment
53 highlights the importance of considering experimental venue when evaluating competition
54 between native and non-native species and should encourage conservation biologists to treat
55 removal efforts as experiments.

56 **KEYWORDS:** *Actinemys*, *Emys marmorata*, experimental venue, invasive species, *Trachemys*
57 *scripta elegans*, UC Davis Arboretum

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61 **1.0 Background**

62 Invasive species are a major threat to biodiversity (Simberloff et al. 2013) and are an ongoing
63 concern for conservation practitioners (Kuebbing and Simberloff 2015). One species widely
64 considered harmful to native species worldwide is the red-eared slider turtle (*Trachemys scripta*
65 *elegans*; RES). This species is native to the central United States but is now present on every
66 continent except Antarctica, predominantly because of releases of unwanted pet turtles (Arvy
67 1997, Cadi et al. 2008, Kraus 2009, Rhodin et al. 2017). The widespread continued introduction
68 of this species led the International Union for Conservation of Nature (IUCN) to name RES as
69 one of the “worst invasive species” in the world (Lowe et al. 2000). However, despite long-held
70 concerns about the effects of introduced RES on native turtle species (Arvy and Servan 1998,
71 Cadi et al. 2008), few studies have explicitly explored the consequences of RES introductions on
72 wild, native turtle populations (Lambert et al. 2013, Pearson et al. 2013, Costa 2014, H eritier et
73 al. 2017) and there have been no experiments on wild populations.

74 Laboratory and mesocosm experiments suggest that RES can outcompete native turtles
75 for food and basking sites (Cadi and Joly 2003, 2004, Polo-Cavia et al. 2008, 2010, 2011,
76 Pearson et al. 2015). While these simplified, semi-natural experiments allow us to begin isolating
77 causal agents, they also frequently inflate the effects of interspecific competition compared to *in*
78 *situ* manipulations under more natural conditions (Skelly and Kiesecker 2001, Skelly 2002,
79 Winkler and Van Buskirk 2012). Although we recognize that *in situ* experiments come with their
80 own drawbacks, comparing laboratory and mesocosms experiments with field manipulations is
81 critical to understanding the strength of species interactions in wild contexts. To our knowledge,
82 no study has yet experimentally tested whether RES are an important competitor with any native
83 turtle species in the wild.

84 Basking sites are a key resource for evaluating competition between aquatic turtle species
85 because these sites are critical for proper thermoregulation, which directly influences vital
86 physiological parameters like disease control as well as growth and reproductive rates (Ernst and
87 Lovich 2009). Basking sites have repeatedly been identified as a likely axis of competition
88 between introduced RES and native turtle species, with several laboratory and mesocosm
89 experiments suggesting that RES may exhibit dominant aggressive behaviors while basking and
90 may displace native turtles from basking sites (Cadi and Joly 2003, Polo-Cavia et al. 2010,
91 Pearson et al. 2015). In human-modified waterways, competition for basking sites may be
92 especially pronounced because turtles often experience reductions in basking site availability due
93 to the removal of basking objects for flood control and aesthetic reasons (Spinks et al., 2003).

94 One study in the University of California, Davis (UCD) Arboretum waterway found that
95 RES and native western pond turtles (*Emys marmorata*; WPT) are spatially segregated across
96 basking sites (Lambert et al. 2013). Although both RES and WPT sometimes bask at the same
97 sites (Fig. 1), they tended to concentrate in opposite ends of the waterway and at basking sites
98 that differ in slope, water depth adjacent to the site, site substrate, and the degree of human
99 activity (Lambert et al. 2013). It is unclear, however, whether these interspecific differences in
100 basking site use are due to innate preferences or competitive interactions. Because of the
101 biological importance of basking sites in WPT life history (Bury and Germano 2008, Ernst and
102 Lovich 2009), determining whether RES limit WPT use of preferred basking habitat is essential
103 for effective conservation (Thomson et al. 2016), particularly given the widespread occurrence of
104 introduced RES in California (Thomson et al. 2010, Fisher unpubl.)

105 Here, we present the results of an *in situ* field experiment whereby we dramatically
106 reduced the UCD Arboretum RES population to examine whether WPT subsequently shifted

107 their use of available basking sites in the wild. Our experiment explicitly tests whether invasive
108 species removal, an intensive and commonly-advocated management practice (Simberloff et al.
109 2013, Gaeta et al. 2015), including for RES (Garcia-Díaz et al. 2017), influences the basking
110 behavior of native WPT in the wild. If RES and WPT compete for basking sites and RES are
111 dominant to WPT, then we predict that removing RES would lead to WPT basking activity
112 becoming more concentrated at sites previously dominated by RES. However, if existing
113 basking-site use patterns reflect species-specific habitat preferences, then we predict that
114 removing RES would have minimal impact on WPT basking site use. Results from this
115 experiment provide a useful first test of the impacts of introduced RES on native, wild turtles;
116 these data are immediately relevant to management of WPT across its known range (Thomson et
117 al. 2016), and for the undergoing Status Review for possible listing under the US Endangered
118 Species Act (USFWS 2015)

119 **2.0 Methods**

120 *2.1 Study Site:* The UCD Arboretum waterway runs along the southern border of the university
121 campus in Yolo County, CA, USA and is situated in the former channel of the North Fork of
122 Putah Creek. Various sections of the waterway are bordered by urban, agricultural, and
123 undeveloped natural landscapes (Fig. 1). For more detailed descriptions of the location, see
124 Spinks et al. (2003) and Lambert et al. (2013).

125 *2.2 RES Removal:* In 2011, we captured turtles throughout the UCD Arboretum from 10 July–1
126 August and again from 13–29 September. We primarily used baited traps that can be
127 deployed in water depths of 0.5–2.0 m. Cumulative submersible trap effort was
128 approximately 900 trap-nights. We supplemented our submersible trapping with
129 opportunistic hand captures and dip netting, along with periodic deployment of a fyke net

130 and a basking trap. The submersible traps, fyke net, and basking trap were not biased towards
131 any particular species, but hand captures and dip netting were targeted at RES. We recorded
132 mass and plastron length of each captured RES using digital pan scales and dial calipers. We
133 re-homed several captured RES with responsible pet owners and euthanized all other RES,
134 donating the majority to the UC Davis School of Veterinary Medicine, the Natural History
135 Museum of Los Angeles County, or the Museum of Wildlife and Fish Biology at UC Davis.
136 All turtle handling was authorized under UC Davis IACUC Protocols #15263 and #16227,
137 and California Department of Fish and Wildlife Scientific Collecting Permits #2480, #4307,
138 and #11663.

139 To test whether our RES trapping success plateaued over time, which would suggest that
140 our trapping effort removed the majority of the RES population, we analyzed whether the
141 cumulative number of RES was better modeled by a linear or quadratic relationship across
142 trapping days. We used linear regression and likelihood ratio tests to determine whether our
143 trapping effort had minimal impact on the RES population (a linear fit) or resulted in fewer
144 RES trapped each day (a quadratic fit). We also tested for an interaction between sex and
145 trapping day to estimate whether we reduced the sexes at different rates.

146 *2.3 Turtle Monitoring:* From 18 March–22 April 2012, we conducted visual (with binoculars)
147 surveys of the same set of 24 basking sites studied in spring 2010 prior to the RES removal
148 (Fig. 1). Each basking site is a short stretch of shoreline (1–2 m long) with adjacent sites at
149 least 3 m apart. We conducted surveys 2010 and 2012 surveys within a similar set of dates
150 and times to make them as similar as possible. During each survey, we measured water
151 temperature with a hand-held thermometer; we also obtained maximum daily temperature

152 data from the UC Davis Russel Ranch Weather Station, ca. 4 km northwest of the UCD
153 Arboretum.

154 *2.4 Analysis:* The basking distribution of WPT and RES previously was shown to vary strongly
155 along a west-east gradient, with WPT focused at the west end and RES focused at the east
156 end of the waterway (Lambert et al. 2013). Because of this, we analyzed the relative and
157 absolute basking abundances of both species, and the extent to which these changed after
158 removing RES. We limited our analysis to 24 basking sites that had data available for every
159 survey date within the same date range (March 18 to April 22) in 2010 (pre-RES removal,
160 from Lambert et al. 2013) and in 2012 (post-RES removal, measured here).

161 To test for changes in the relative basking distribution of WPT to RES across the
162 waterway, we used a generalized linear mixed effects model (GLMM) with a binomial
163 family for proportion data using the ‘glmer’ function in the R package lme4. We used the
164 distance of each basking site from the west end of the UCD Arboretum (following Lambert
165 et al. 2013) as well as treatment (pre- or post-RES removal) as fixed effects and used
166 observation date as a random effect to account for repeated measures of basking sites
167 (Lambert et al. 2013). We first tested for a significant interaction between treatment and each
168 basking site’s distance from the west end. If the interaction term was not significant, we
169 removed it from the model. We then assessed the influence of the main effects and tested
170 whether the relative basking distribution of WPT to RES differed pre- and post-RES removal
171 using a Tukey’s post-hoc test with ‘glht’ function in the R package “multcomp”. We also
172 performed binomial GLMMs for each basking site separately to explore whether individual
173 basking sites show changes in the proportion of WPT to RES after the experiment.

174 For the absolute abundance of each species, we applied a similar modeling approach but
175 used Poisson GLMMs for count data. We used the ‘r.squaredGLMM’ function in the
176 package “MuMIn” to calculate R^2 for GLMMs; ‘MuMIn’ calculates both a conditional R^2
177 (cR^2) for the full model including fixed and random effects as well as a marginal R^2 (mR^2)
178 for just the model’s main effects. We conducted all analyses in R (version 3.2.2).

179 To further explore patterns at individual basking sites, we used contingency table
180 analyses for each species independently to test whether certain basking sites comprised larger
181 or smaller proportions of the total basking observations for either species pre- and post-RES
182 removal. We focused on sites P, O, E, Q, and R which were, respectively, the five most
183 heavily-used turtle basking sites (combined for both species) pre-RES removal. We also
184 examined site X since it was the most heavily-used turtle basking site post-RES removal.

185 **3.0 Results**

186 *3.1 RES Removal:* In summer 2011, we captured and removed 180 RES from the UCD
187 Arboretum. We removed 25 adult males, 71 adult females, and 84 juveniles (individuals with
188 carapace length ≤ 100 mm, Ernst and Lovich 2009), including one individual of 113 mm
189 carapace length that lacked sexually-diagnostic traits. We removed 59 RES that we had
190 captured and marked in previous years and 121 unmarked individuals. All turtle marking
191 previously occurred as part of the UC Davis Herpetology course from 2007 through spring
192 2011. Of these new RES, 70% ($n = 84$) were juveniles, 22% ($n = 27$) were adult females, and
193 8% ($n = 10$) were adult males.

194 A likelihood ratio test supported a model (full model $R^2 = 0.95$, $p < 0.0001$) with a
195 quadratic over a linear fit between cumulative RES trapped and trapping day ($p < 0.0001$)

196 and with an interaction between RES sex and trapping day ($p < 0.001$). During the removal
197 effort, captures of RES declined and leveled off, indicating that we removed a substantial
198 portion of the catchable RES population. Furthermore, the significant interaction term
199 suggests we depleted the male RES population faster than the female RES population (Fig.
200 2). In total, we removed 104.5 kg of RES biomass, of which 79% (82.3 kg) was from adult
201 females, 15% (15.5 kg) was from adult males, and 6% (6.7 kg) was from juveniles. During
202 this same trapping effort, we captured, marked (or re-marked), and released 118 individual
203 WPT, 14 of which were juveniles (≤ 110 mm plastron length; Holland, 1991; Spinks et al.
204 2003). While some aspects of our capture efforts in 2011 specifically targeted RES (e.g., dip
205 netting and hand captures), our data indicate RES outnumbered WPT by about 1.5:1 at the
206 start of the experiment.

207 *3.2 Basking Surveys:* We surveyed for 16 days from 18 March to 22 April 2010 (pre-removal)
208 and 18 days from 18 March to 22 April 2012 (post-removal). Maximum daily air
209 temperatures were not significantly different between years (two-tailed t-test, $p = 0.74$; 2010,
210 $19.2\text{ C} \pm 0.69\text{ SE}$; 2012, $18.8\text{ C} \pm 0.88\text{ SE}$). However, in the two weeks prior to our surveys
211 the maximum daily air temperatures were significantly warmer in 2012 ($18.8\text{ C} \pm 1.08\text{ SE}$)
212 than in 2010 ($15.2\text{ C} \pm 0.65\text{ SE}$); two-tailed t-test, $p < 0.001$). Water temperature was
213 significantly warmer (two-tailed t-test, $p < 0.0001$) in 2010 ($17.0\text{ C} \pm 0.24\text{ SE}$) compared to
214 2012 ($15.4\text{ C} \pm 0.36\text{ SE}$). In 2010, we recorded 283 WPT and 645 RES observations. In
215 2012, we recorded only 43 WPT observations and 61 RES observations.

216 Pre-removal, we recorded WPT basking at 15 of the 24 basking sites, but post-removal
217 we recorded WPT basking at only 8 of the 24 sites (Fig. 3). WPT were absent from eight
218 sites that they used pre-removal, although six of these were used infrequently in 2010. We

219 recorded WPT using one additional site where they were not recorded pre-removal. In
220 general, the basking sites most commonly used by WPT pre-removal were the same sites
221 used post-removal (Fig. 3). Pre-removal, we recorded RES basking at 17 of the 24 basking
222 sites, but post-removal we recorded RES basking at only 8 of the 24 sites (Fig. 3). RES were
223 absent from nine sites that they used pre-removal and were not recorded using any new sites
224 after the removal.

225 *3.3 Relative Abundance:* The interaction between distance from the west end of the waterway
226 and treatment was not significant ($p = 0.18$) and was removed from the model. Both distance
227 from the west end ($p < 0.0001$) and treatment ($p < 0.0001$) were significant and were retained
228 in the model ($cR^2 = 0.31$, $mR^2 = 0.31$). Both pre- and post-RES removal, the relative basking
229 distribution of turtles was WPT-biased in the west end and RES-biased in the east end of the
230 waterway (Fig. 4). Furthermore, the Tukey's post-hoc test indicated that the proportion of
231 basking observations increased from 30.5% WPT pre-RES removal to 41.3% WPT post-RES
232 removal ($p < 0.0001$). The non-significant interaction term indicates that the RES removal
233 did not change the relative basking distribution of the two species across the waterway.

234 Individual binomial GLMMs for each basking site returned a significant treatment effect
235 for site Q ($p = 0.002$, 9% WPT to 55% WPT) and a marginally significant effect for site O (p
236 $= 0.09$, 30% WPT to 75% WPT). All other basking sites showed no significant difference in
237 the proportion of the two species between years (all $p > 0.1$).

238 *3.4 WPT Absolute Abundance:* In 2010, we recorded 283 WPT basking observations and only 43
239 in 2012. The Poisson GLMM indicated a significant interaction between treatment and
240 distance from the west end ($p = 0.012$, $cR^2 = 0.23$, $mR^2 = 0.06$), suggesting a shift in the

241 absolute basking distribution of WPT across the waterway. Individual GLMMs for each year
242 indicate that distance from the west end is significant in the pre-RES removal year ($p <$
243 0.012 , $cR^2 = 0.27$, $mR^2 = 0.03$) but not in the post-RES removal year ($p = 0.55$). These
244 results suggest that, before the RES removal, absolute basking abundance of WPT declined
245 from west-east, and that post-RES removal WPT had a relatively even basking distribution
246 throughout the UCD Arboretum (Fig. 5). Contingency tables indicated that sites Q ($p = 0.01$)
247 and X ($p = 0.001$) comprised larger proportions of total WPT basking observations post-RES
248 removal than pre-RES removal. All other sites analyzed comprised similar proportions of
249 total WPT basking observations before and after the experiment (all $p > 0.1$), although small
250 sample sizes often resulted in relatively little statistical power. Together, these analyses
251 indicate removing RES resulted in a less clustered, more even distribution of WPT across
252 basking sites with two sites towards the center-east and east of the Arboretum comprising
253 more WPT basking activity.

254 *3.5 RES Absolute Abundance:* For RES, the GLMM indicated a significant interaction between
255 distance to the west end and treatment ($p < 0.0001$, $cR^2 = 0.30$, $mR^2 = 0.16$). While the
256 number of RES basking observations was an order of magnitude lower in 2012 ($n = 61$)
257 versus 2010 ($n = 645$), the positive relationship between RES absolute abundance and the
258 west-east gradient in the UCD Arboretum appears to be more pronounced post-RES removal
259 (Figs. 3, 5). Individual GLMMs for each year show that distance to the west end was
260 significant in the pre-RES removal year ($p < 0.0001$, $cR^2 = 0.27$, $mR^2 = 0.02$) and the post-
261 RES removal year ($p < 0.0001$, $cR^2 = 0.14$, $mR^2 = 0.14$). In both years, the absolute
262 abundance of basking RES increased along the west-east gradient pre (Fig. 5). After the
263 RES-removal, RES were relatively sparse through the western and central portions of the

264 waterway and were concentrated in the far eastern end (Fig. 3, 5). Contingency table analyses
265 indicated that sites E ($p = 0.01$), O ($p = 0.0004$), P ($p = 0.0001$), and R ($p = 0.03$) comprised
266 lower proportions of total basking observations after the experiment and site X comprised a
267 higher proportion ($p = 0.001$). Site Q made up similar proportions of total RES observations
268 in both years ($p = 0.14$).

269 **4.0 Discussion**

270 To test whether non-native RES influence WPT basking site use, we removed 180 non-native
271 RES, totaling over 100 kg of turtle biomass. This experiment represents a dramatic alteration to
272 the turtle community inhabiting the UCD Arboretum. Our experiment indicated that removing
273 the majority of the RES population altered the basking distribution of both native WPT and
274 residual RES, and that some form of interspecific interactions is occurring between the two
275 species. However, our results do not necessarily provide evidence for strong interspecific
276 competition between introduced RES and WPT, and suggest that a more nuanced, complex set of
277 interactions may be occurring in wild populations.

278 *4.1 Intraspecific Competition:* One of the clearest effects of our experiment was that the
279 remaining post-removal RES abandoned several basking sites that they previously used heavily
280 (particularly sites O and P) and shifted towards the east end of the transect (e.g., sites V and X).
281 Although it is unclear what drove this shift, this result indicates that RES prefer habitat at this
282 end of the waterway and that, prior to our experiment, RES densities were high enough for
283 intraspecific competition among RES to force some individuals into other areas of the waterway.
284 Our previous work showed that RES basking activity was highest at sites with shallow slopes,
285 deeper water adjacent to the site, a steel mesh (rather than concrete or dirt) substrate, and high

286 human activity (Lambert et al. 2013). Consistent with these observations, the two basking sites
287 that showed the most concentrated RES activity post-removal were comprised of steel mesh and
288 were the sites with some of the flattest slopes and deepest water along the transect as well as the
289 sites with the highest level of human activity (Lambert et al. 2013), indicating that residual RES
290 concentrated their basking activity at the most preferred sites.

291

292 *4.2 Interspecific competition:* Previous work on the UCD Arboretum turtle population found that
293 RES and WPT largely use different basking sites (Lambert et al. 2013). Before and after our
294 experiment, WPT predominantly used the same basking sites but at different frequencies, with a
295 general trend toward a more uniform west-east distribution post-removal. These results indicate
296 that reducing the density of introduced RES allow WPT to spread out in the waterway. Even so,
297 if sites towards the east end of the waterway which were previously dominated by RES (e.g.,
298 sites O, P, Q, and R) are also preferred by WPT, then we would have expected WPT basking
299 behavior to concentrate at these sites post-RES removal. But we did not see this. Rather, our
300 experiment resulted in a shift in WPT basking activity suggesting that WPT basking is
301 contingent on RES densities but we did not observe a dramatic shift that might be indicative of
302 strong interspecific competition for basking sites. Competition is presumably greatest at high
303 densities of RES and perhaps influenced by the relative densities of both species, as has been
304 shown in other biological invasion scenarios (Gurnell et al. 2004). Earlier experiments have
305 concluded that introduced RES outcompete native turtles for resources including basking sites or
306 food. Because these experiments took place in artificial experimental venues and (Cadi and Joly
307 2003, 2004, Polo-Cavia et al. 2008, 2010, 2011, Pearson et al. 2015), it is possible that prior
308 conclusions about the competitive dominance of introduced RES were inflated. Although no

309 prior experiments have focused on WPT, some (Cadi and Joli 2003, 2004) have focused on
310 European *Emys orbicularis*, which is a closely-related congener (Spinks et al. 2016).

311
312 *4.3 Study Limitations:* Our analyses showed that WPT made up proportionally more of our
313 observations after RES removal (Fig. 4). However, we recorded far fewer turtle basking
314 observations for both species in 2012. For RES, this was expected as it was the goal of our
315 experiment to reduce the RES population. This is not the case for WPT. It is possible that
316 temperature or other environmental variation among years as well as unforeseen consequences of
317 our manipulation resulted in reduced overall turtle basking activity after the RES removal. For
318 instance, aquatic turtles like RES can dramatically influence trophic dynamics and aquatic
319 ecosystem function (Lindsay et al. 2013). By removing a substantial portion of the turtle
320 community, our experiment may have altered the availability and distribution of food resources
321 which may have indirectly impacted where turtles chose to bask and turtle basking behavior
322 generally, resulting in fewer observations. Unfortunately, we cannot distinguish whether the
323 generally lower basking observations of WPT post-RES removal are an effect of our experiment
324 or whether other uncontrolled factors may have resulted in fewer WPT basking observations.
325 Because of logistical constraints, we were only able to collect a single of year of observations
326 post-RES removal. We also recognize that our experiment did not address other putative axes of
327 competition that are important for the continued recruitment and persistence of this WPT
328 population. For example, evidence from experimental mesocosms suggests introduced RES
329 generally eat more and grow faster than native turtles (Cadi and Joly 2003, Pearson et al. 2015),
330 and these effects may have important consequences for native turtles.

331

332 *4.4 Management Implications:* Removing 180 RES from the UCD Arboretum was an intensive
333 effort requiring over 2,000 person-hours of field work across forty days. Although WPT
334 comprised a larger proportion of our basking observations post-removal (Fig. 4), RES still made
335 up the dominant portion of our observations, summing to almost 60% of the total basking
336 observations made after the experiment. In general, removing invasive species is difficult, time
337 and labor intensive, and may still fail to extirpate the entire population, particularly in the face of
338 continued introductions (Gaeta et al. 2015). In Europe, where RES removal is a widely
339 advocated practice, recent work noted the severe challenges of functionally eradicating
340 introduced RES (Garcia-Diaz et al. 2017). As long as RES are readily available in the pet trade,
341 *de novo* introductions are likely to continue, complicating attempts to successfully eradicate
342 introduced RES populations.

343 Our results suggest that a concerted effort at RES reduction in a large, complex water
344 body has the potential to influence native turtle species, but that these influences may be
345 relatively modest in their quantitative effects. Regardless of whether it is known that RES
346 compete with a given native species, both removing non-native RES and stemming the future
347 release of RES are important steps for reducing possible disease and parasite transmission
348 (Héritier et al. 2017; Demkowska-Kutrzepa et al. 2018). Further, removals and reductions in pet
349 releases could help minimize competition if it is occurring, whether that be for food, basking
350 sites, or other resources. Although the commitment of time and energy is large, we encourage
351 conservation biologists to treat RES removal efforts as experiments, as was done here, and test
352 whether removing RES benefits native turtle species along these other ecological axes.

353 Habitat modification due to urban and agricultural land use is a major threat to WPT in
354 California (Thomson et al. 2016). Nonetheless, human-modified habitats can be valuable

355 resources for WPT when appropriate conservation and management efforts are implemented
356 (Spinks et al. 2003, Thomson et al. 2010, 2016). Directly managing urban basking habitat may
357 be a particularly tractable conservation activity for WPT in addition to directly managing non-
358 native RES. Future experiments and management practices can readily manipulate these basking
359 site characteristics to test whether doing so is beneficial for WPT. Emerging research from the
360 UCD Arboretum suggests that experimentally-added floating logs are preferred by WPT over
361 bank-side basking and are more heavily used by WPT than RES especially when placed further
362 from human activities (Cossman et al. unpubl.). In our experiment here, we may have liberated
363 parts of the waterway that were previously dense with turtles generally, thus allowing WPT to
364 spread out across the waterway. However, because WPT did not concentrate their basking
365 activity at sites previously dominated by RES, these two species may not intensely compete for
366 bank-side basking sites in this waterway. WPT may ultimately show little preference for
367 particular bank-side basking site characteristics, although this warrants further study. Providing
368 more basking sites of suitable quality, and particularly further from high levels of human
369 activity, may be a feasible and fruitful management practice in conjunction with removing RES.
370 We encourage additional research into the merits of this strategy.

371 *4.5 Conclusions:* Evidence from laboratory and mesocosm studies indicates that introduced RES
372 are competitively dominant to native turtles. Here, we offer the first experimental test for
373 competition between native turtles and non-native RES in the wild; our work provides insight
374 into the seemingly complex nature of competition between introduced RES and native turtles.
375 Our population manipulation suggests that reducing the density of RES may alter the basking
376 activity of threatened WPT but that RES and WPT may not compete intensely for basking sites.
377 We found strong evidence for strong intraspecific competition for basking sites at high RES

378 densities, and that reducing that competition may have had additional effects on the distribution
379 of WPT basking. We hope that our study will encourage further field-based experiments to better
380 understand the extent to which RES are competing with native turtles for basking sites and/or
381 other resources, and to explore which management practices are both reasonable and effective.

382

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393

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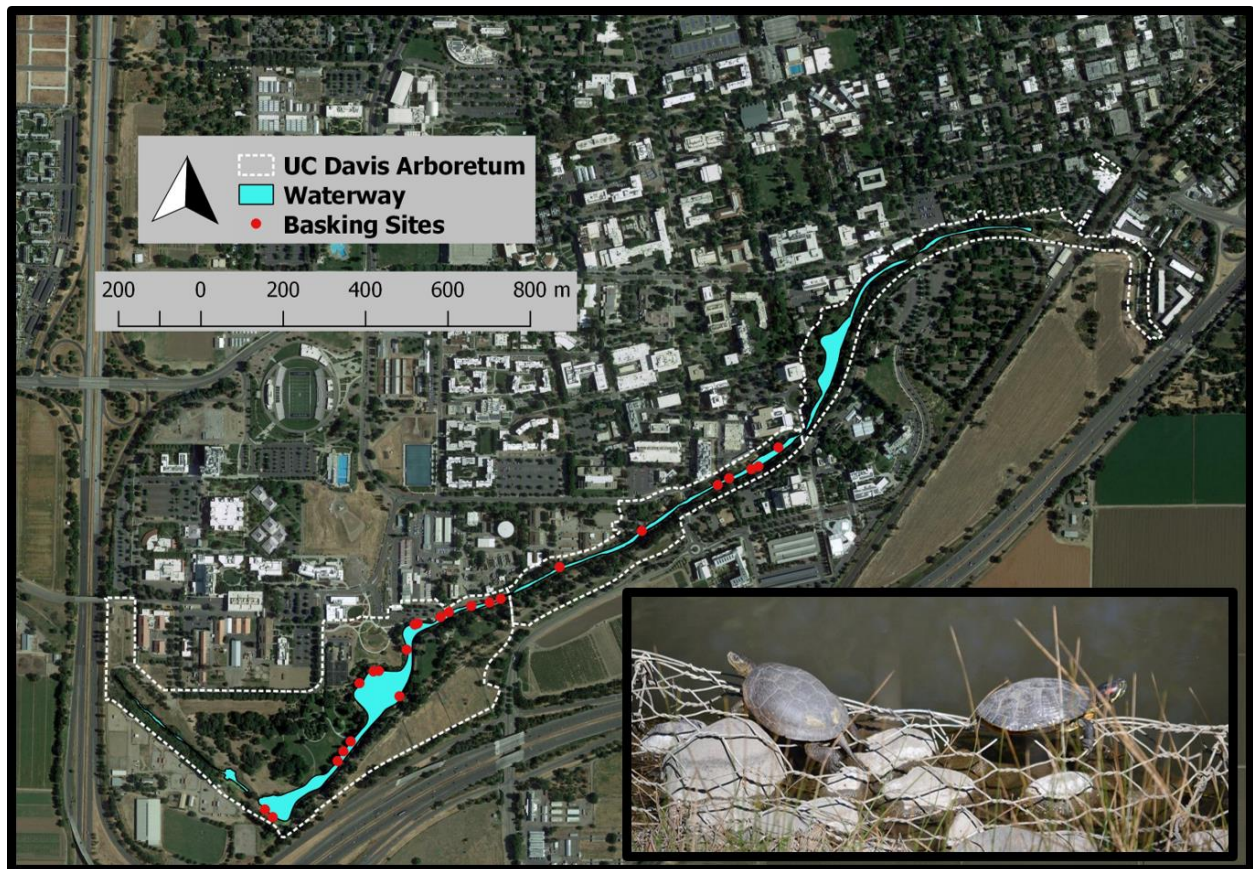
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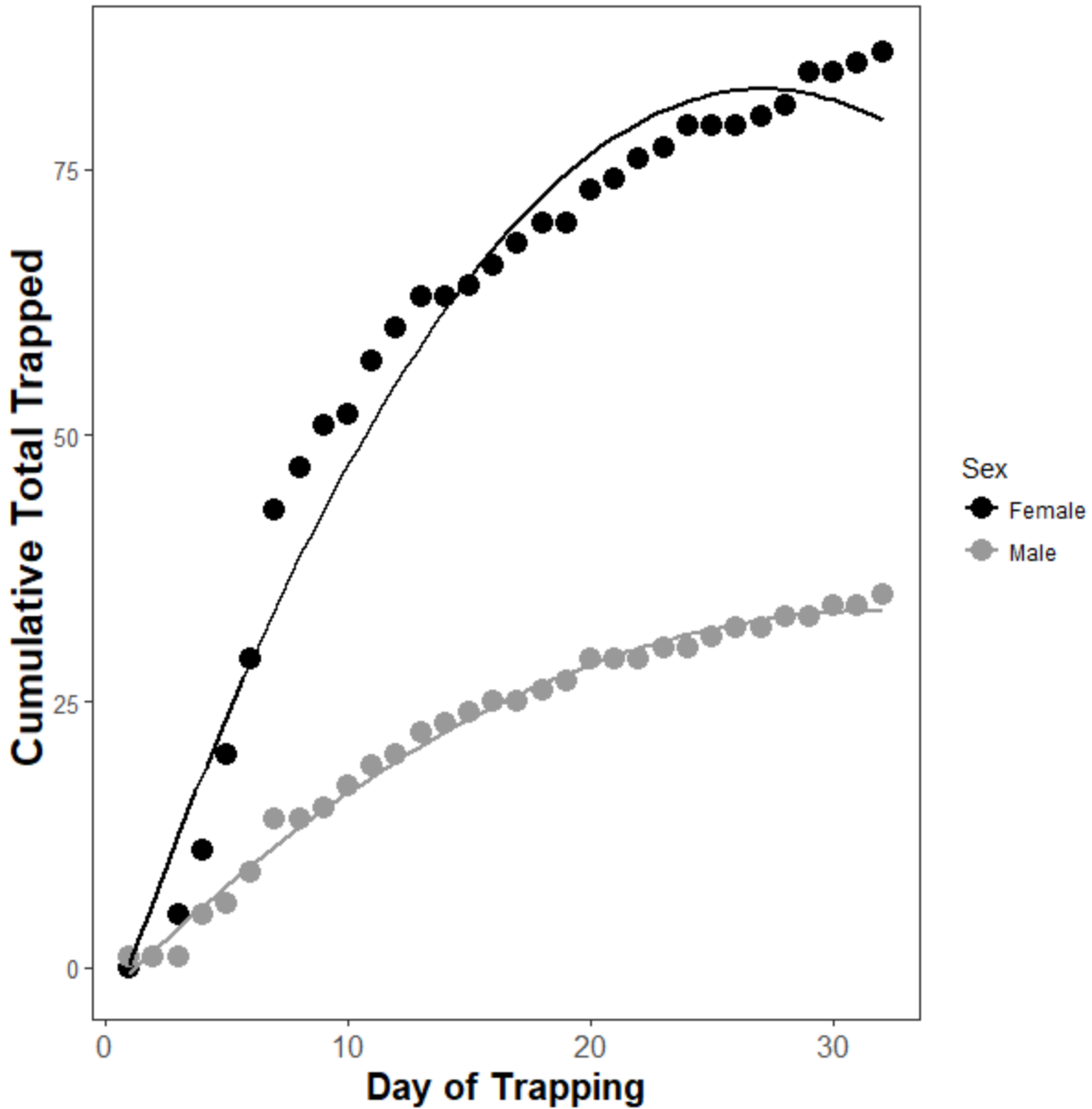
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500 **Figures and Figure Legends**



502 **Figure 1** (1.5 columns, color online only): Map of the UC Davis Arboretum waterway and turtle
503 basking sites monitored before and after the red-eared slider population reduction. Inset are a
504 native western pond turtle (left) and a non-native red-eared slider (right) basking in the UC Davis
505 Arboretum. Photo by M. Lambert.

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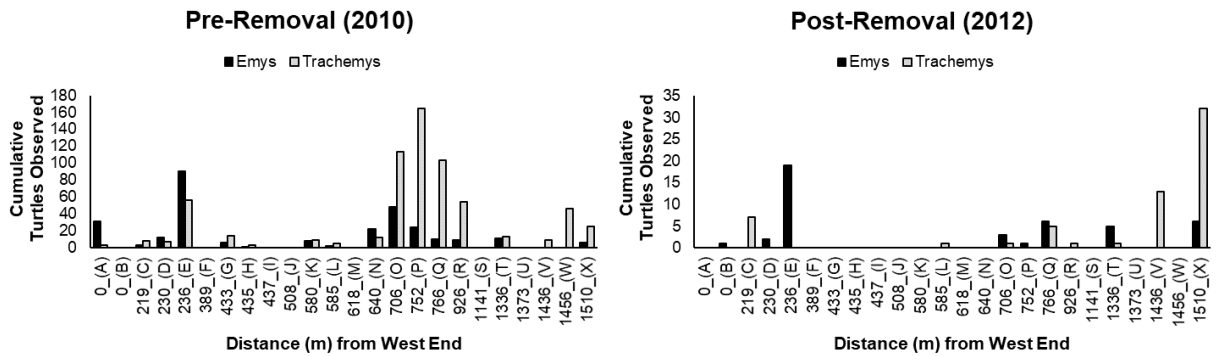


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508 **Figure 2** (single column). Cumulative total number of adult female and male red-eared sliders
509 (RES; *Trachemys scripta elegans*) removed from the UC Davis Arboretum in 2011. Trap Day 1–
510 18 are in July and 19–32 are in September.

511

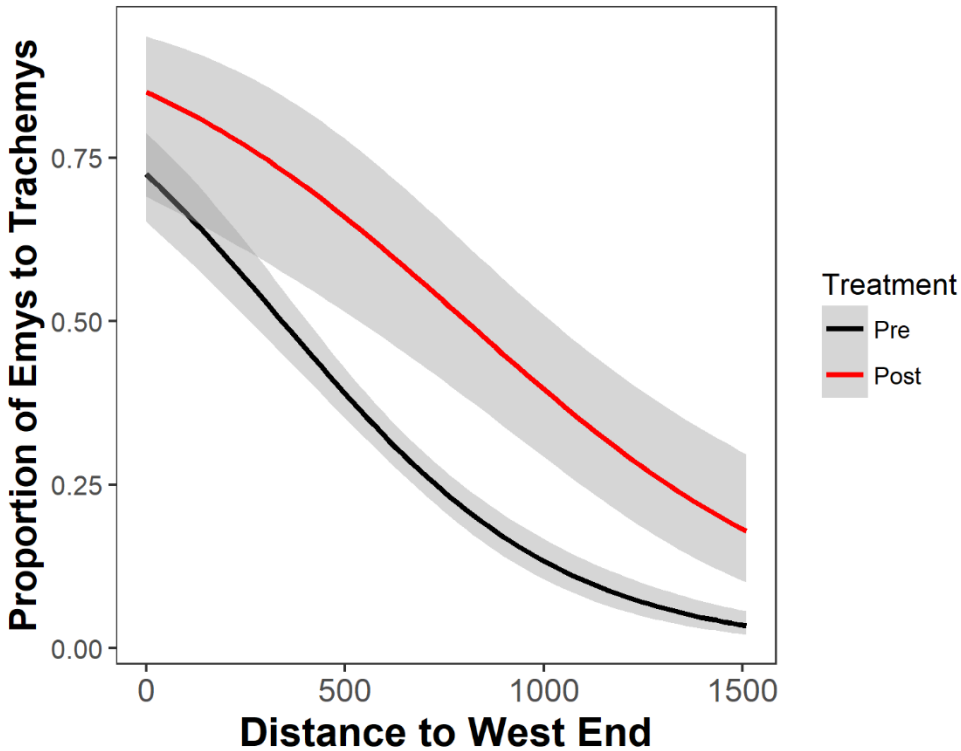
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514 **Figure 3** (double column). The cumulative number of western pond turtle (WPT; *Emys*
515 *marmorata*) and red-eared slider (RES; *Trachemys scripta elegans*) basking observations across
516 sampling dates in the pre- and post-RES removal years. Letters in parentheses under the x-axis
517 are basking site identifiers. Note that y-axes of the two panels have different scales.

518



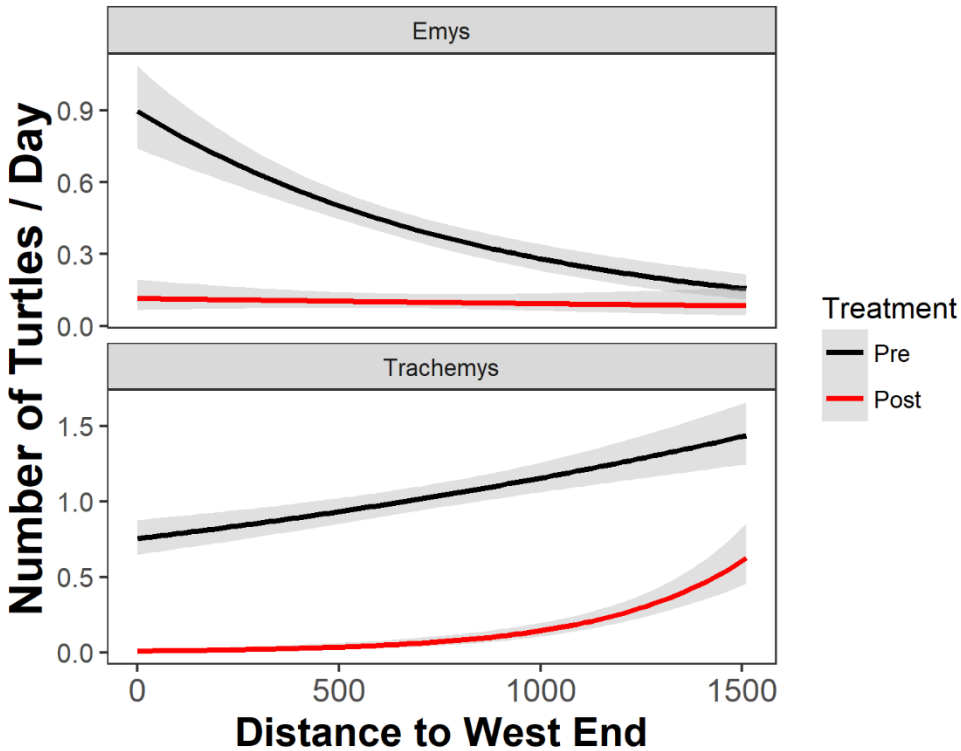
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520 **Figure 4** (single column, color in print). The modeled proportion of western pond turtles (WPT;
521 *Emys marmorata*) to red-eared sliders (RES; *Trachemys scripta elegans*) basking along a west-
522 east gradient in the UC Davis Arboretum. While WPT made up a greater proportion of
523 observations in 2012 than in 2010, the relative basking distribution of the two species along the
524 Arboretum did not change between pre- and post-RES reduction years. Curves are binomial fits
525 and gray shading represents 95% confidence intervals.

526

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529

530 **Figure 5** (single column, color in print). The modeled number of western pond turtles (WPT;
531 *Emys marmorata*) and red-eared sliders (RES; *Trachemys scripta elegans*) observed basking
532 along a west-east gradient at the UC Davis Arboretum pre- and post-RES removal. Note that
533 fewer turtles were observed basking in the post-RES removal survey. Curves are Poisson fits and
534 gray shading represents 95% confidence intervals.