

Movement behaviour responses to environment: fast inference of individual variation with a mixed effects model

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1 **Abstract**

2 Telemetry data provide a rich source of information on animals use of space, habitat pref-
3 erences and movement behaviour. Yet habitat models fit to these data are blind to the
4 underlying behavioural context. Conversely, behavioural models accounting for individual
5 variability are too slow for meaningful analysis of large telemetry datasets. Applying new
6 fast-estimation tools, we show how a model incorporating mixed effects within a flexible
7 random walk movement process rapidly infers among-individual variability in environment-
8 movement behaviour relationships. We demonstrate our approach using southern elephant
9 seal (*Mirounga leonina*) telemetry data. Seals consistently reduced speed and directional-
10 ity (move persistence) with increasing sea ice coverage, had variable responses to chloro-
11 phyll concentration and consistently reduced move persistence in regions where circum-
12 polar deep water shoaled. Our new modelling framework is extensible and substantively
13 advances analysis of telemetry data by allowing fast and flexible mixed effects estimation
14 of potential drivers of movement behaviour processes.

15 **Key Words:** correlated random walk; habitat; individual movement; latent variable;
16 telemetry; Template Model Builder; random effects; southern elephant seal; habitat model

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17 Introduction

18 Understanding animals' use of geographical and environmental space (i.e., where animals
19 are and why they are there) is one of the central aims of ecology (Rosenzweig, 1981). Move-
20 ment is the key process that defines space-use at spatial and temporal scales relevant to
21 individual animals and telemetry is the predominant approach to observe this process
22 (Kays *et al.*, 2015; Hussey *et al.*, 2015). Inferences about the behavioural context of ani-
23 mal movements, such as foraging, resting or predator avoidance, are often made by relat-
24 ing movement behaviour to physical habitat features (e.g., Breed *et al.*, 2017).

25 Various spatial habitat modelling approaches are used to infer animals' space-use and
26 habitat preferences, through combining telemetry and environmental information, e.g.,
27 from remotely sensed data (Aarts *et al.*, 2008; Thurfjell *et al.*, 2014; Raymond *et al.*, 2015).
28 Most habitat models infer animals' habitat preference or selectivity from a combination
29 of observed (presence) and simulated (pseudo-absence) tracking locations (Aarts *et al.*,
30 2008) but are generally blind to the behavioural context (e.g., whether animals are mi-
31 grating, foraging or resting) underlying those inferred preferences. Hidden Markov mod-
32 els (HMMs) and state-space models (SSMs) can provide this context by inferring (un-
33 observed) behavioural states, and relating state-switching probabilities to environmental
34 features (Morales *et al.*, 2004; Patterson *et al.*, 2009; Bestley *et al.*, 2013; Michelot *et al.*,
35 2016).

36 Both HMMs and SSMs offer great flexibility in modelling movement behaviour as a
37 function of extrinsic and/or intrinsic drivers (Bestley *et al.*, 2015; Michelot *et al.*, 2017).
38 Although high individual variation is a commonly reported feature in telemetry analyses,
39 methods to account for individual variability in movement-environment relationships (e.g.,
40 using random effects, Pinheiro & Bates, 2000; Bolker *et al.*, 2009) have been implemented
41 in a limited way to date (Langrock *et al.*, 2012; Bestley *et al.*, 2015). A fully flexible ap-
42 proach where any sensible combination of fixed and random terms can be considered, that
43 allows different environmental responses across individuals, has yet to be implemented.

44 This is mainly because complex mixed effects models (Thorson & Minto, 2015) applied to
45 large time-series can be computationally demanding.

46 Here we present a modelling approach that takes advantage of fast, powerful estimation
47 tools provided by the relatively new R package Template Model Builder (TMB, Kristensen
48 *et al.*, 2016). We illustrate a mixed effects modelling approach for animal tracking data
49 that takes advantage of TMB's fast estimation (Albertsen *et al.*, 2015; Auger-Méthé *et al.*,
50 2017) to parametrize movement behaviour using a time-varying term for movement persis-
51 tence. Our primary aim is to show how the approach can be used to infer relationships be-
52 tween animals' movement behaviour and the environmental features they encounter. These
53 models can be fit flexibly with single or multiple random effects, enabling inference across
54 multiple individuals and assessment of the extent to which relationships differ among in-
55 dividuals. We illustrate our approach using southern elephant seal (*Mirounga leonina*)
56 telemetry data, with cases demonstrating both sea-ice and oceanic foraging trips, to show
57 how seals engaging in different foraging tactics may respond differently to their environ-
58 ment.

59 **Materials and methods**

60 Here we describe our mixed effects modelling approach for inference of covariate relation-
61 ships with movement behaviour. We divide the description of our approach into three sec-
62 tions. First, we focus on a basic move persistence model that can be used to estimate be-
63 havioural change along an animal's observed movement trajectory. Second, we show how
64 this basic model can be expanded to infer how these behavioural changes may be related
65 to environmental features. We focus on relationships with environmental covariates but
66 any combination of extrinsic or intrinsic covariates could be modelled provided they are
67 measured at locations and/or times consistent with the telemetry data. Third, we add ran-
68 dom effects to the model to enable inference about how these behaviour - environmental
69 relationships may differ among individual animals.

70 **Time-varying move persistence**

71 Our modelling approach focuses on estimation of the persistence (sensu Patlak, 1953) of
72 consecutive pairs of animal relocations (move steps) along an entire movement trajec-
73 tory. Move persistence, which captures autocorrelation in both speed and direction, has
74 been modelled as an average across entire movement trajectories (Jonsen, 2016), indicating
75 whether that trajectory is, on average, uncorrelated (i.e., a simple random walk or Brown-
76 ian motion), correlated (i.e., a correlated random walk), or somewhere in between (Codling
77 *et al.*, 2008). Allowing move persistence to vary along a trajectory means it can be used
78 as an index of behaviour, identifying segments of relatively low or high persistence. This
79 model can be written as:

$$\mathbf{d}_t = \gamma_t \mathbf{d}_{t-1} + N(0, \Sigma) \quad (1)$$

80 where \mathbf{d}_t and \mathbf{d}_{t-1} are the changes in an animal's location at times t and $t - 1$. Σ is a
81 variance-covariance matrix specifying the magnitude of randomness in the 2-dimensional
82 movements. γ_t is the time-varying move persistence (autocorrelation) between displace-
83 ments \mathbf{d}_t and \mathbf{d}_{t-1} . γ_t is continuous-valued between 0 (low move persistence, Fig. 1a,c)
84 and 1 (high move persistence, Fig. 1b,c). To avoid potential parameter identifiability is-
85 sues between γ_t and Σ , we set the covariance term in Σ to 0 but note this constraint could
86 be relaxed. We assume γ_t follows a simple random walk in logit space (to keep γ_t bounded
87 between 0 and 1):

$$\text{logit}(\gamma_t) = \text{logit}(\gamma_{t-1}) + N(0, \sigma_\gamma) \quad (2)$$

88 where σ_γ is a scale parameter describing how much move persistence varies along an ani-
89 mal's observed movement track.

90 This process model (Eqn's 1 and 2) can be fit either directly to location data with min-
91 imal error, such as GPS data, fit to SSM-filtered locations, or coupled with an observation
92 model to fit to error-prone data, such as Argos or light-based geolocation data. We assume

93 the locations occur at regular time intervals, but other implementations can accommodate
94 irregularly observed location data (Auger-Méthé *et al.*, 2017).

95 The time-varying move persistence model can be used to objectively identify changes
96 in movement pattern. The γ_t 's are the behavioural index but unlike switching state-space
97 models (e.g., Jonsen *et al.*, 2005) or hidden Markov models (e.g., Langrock *et al.*, 2012) of
98 animal movement behaviour, these changes are modelled along a continuum (0 - 1) rather
99 than as switches between a pre-specified number of discrete states.

100 **Move persistence in relation to environment**

101 To make inferences about the factors associated with these behaviours, we can model γ_t as
102 a linear function of environmental predictors like proportion of ice cover, or other extrin-
103 sic or intrinsic covariates measured at each location. With this approach, we replace the
104 random walk on $\text{logit}(\gamma_t)$ (Eqn 2) with a linear regression of covariates on $\text{logit}(\gamma_t)$:

$$\text{logit}(\gamma_t) = \beta_0 + \beta_1 m_{t,1} + \dots + \beta_n m_{t,n} \quad (3)$$

105 where $\beta_0, \beta_1 \dots \beta_n$ are the fixed intercept and regression coefficients and $m_{t,1} \dots m_{t,n}$ are
106 the predictor variables. This model can be fit to a single animal track, or multiple tracks
107 could be pooled together. Typically, we wish to make inference across multiple individual
108 tracks and assess the extent to which relationships may differ among individuals.

109 **Incorporating individual variability**

110 To account for variation among individual responses to environment, we can expand Eqn
111 3 to a mixed-effects regression of covariates on $\text{logit}(\gamma_t)$, embedded directly in the be-
112 havioural model:

$$\text{logit}(\gamma_t) = (\beta_0 - b_{0,k}) + (\beta_1 - b_{1,k})m_{t,1,k} + \dots + (\beta_n - b_{n,k})m_{t,n,k} \quad (4)$$

113 where the β 's are the fixed-effect intercept and slope terms as in Eqn 3, $b_{0,k}$ is a random
114 deviation for the intercept of the k -th individual, $b_{1,k}$ through $b_{n,k}$ are random deviations
115 for the slopes of the k -th individual and $m_{t,1,k}$ through $m_{t,n,k}$ are the covariates measured
116 along the k -th individual's track.

117 Estimation

118 In principle, any combination of fixed and random effects can be specified within the move-
119 ment model described in equations 1 and 4. However, estimation of multiple random ef-
120 fects can be extremely computationally demanding and this has limited the use of such
121 models for animal telemetry data. Here we use TMB to fit the move persistence models
122 (Auger-Méthé *et al.*, 2017). The TMB package allows complex latent variable mixed ef-
123 fects models, such as SSMS (Albertsen *et al.*, 2015), to be specified in C++ and fit effi-
124 ciently via maximum likelihood using reverse-mode auto-differentiation and the Laplace
125 approximation (Kristensen *et al.*, 2016). The Laplace approximation avoids the need for
126 high-dimensional integration by using a second-order Taylor expansion that massively
127 speeds the calculation of the marginal likelihood (e.g., Albertsen *et al.*, 2015). Compar-
128 ing Bayesian and TMB versions of the same location-filtering model fit to individual Argos
129 location datasets, Auger-Méthé *et al.* (2017) found a 30-fold decrease in computation time
130 for the TMB fit with no apparent loss of accuracy.

131 All code for fitting these models in R is available at <https://github.com/ianjensen>.
132 This code draws on the `lme4` (Bates *et al.*, 2015) and `glmmTMB` (Brooks *et al.*, 2017) R
133 packages to specify the mixed effects models in a general and flexible manner.

134 Data application

135 We demonstrate our move persistence models with 24 adult female southern elephant seal
136 tracks. The seals were captured at Iles Kerguelen (49.35° S, 70.22° E) between late Jan-
137 uary and mid-March in 2009 and 2013-2015, at the end of their annual moult. Animal

138 handling and instrument attachment details can be found elsewhere (McMahon *et al.*,
139 2000; Field *et al.*, 2012; McMahon *et al.*, 2008). These data were sourced from the Aus-
140 tralian Integrated Marine Observing System (IMOS) deployments at Iles Kerguelen and
141 are publicly available (<http://imos.aodn.org.au>). The tracks comprise a mixture of sea
142 ice foraging trips on or near the Antarctic continental shelf (12 seals; Appendix S1.1a) and
143 entirely pelagic foraging trips in sub-Antarctic waters (12 seals; Appendix S1.1b). Prior
144 to fitting the move persistence models, we used a TMB implementation of a state-space
145 model (Jonsen *et al.*, 2005; Auger-Méthé *et al.*, 2017) to filter the observed locations, ac-
146 counting for error in the Argos telemetry, and to regularize the filtered locations a 12-h
147 time interval (see Appendix S1 for details).

148 We fit the move persistence model (**mpm**; Eqn's 1 and 2) to the SSM-filtered seal tracks.
149 To ascertain whether γ_t adequately captures changes in the seals' movement patterns, we
150 compare the γ_t -based behavioural index from the **mpm** to discrete behavioural states es-
151 timated from a behavioural switching state-space model (SSSM; Jonsen, 2016) fitted us-
152 ing the **bsam** R package. Details on how we fit the **bsam** model are in Appendix S2. We
153 then fit the move persistence mixed effects model (**mpmm**; Eqn's 1 and 4) to the same SSM-
154 filtered seal tracks to infer how the seals' movement behaviour may be influenced by envi-
155 ronmental features encountered during their months-long foraging trips. In both analyses,
156 we fitted separate models to the ice and pelagic foraging trips. For the **mpmm**'s, we specified
157 mixed effects models with random intercept and slopes to account for variability among
158 individual seals. We fit all possible combinations of fixed and random effects and use AIC
159 and likelihood ratios to find the best supported model for each set of tracks.

160 We examined 3 potential environmental correlates of elephant seal movement behaviour:
161 sea ice cover (the proportion of time the ocean is covered by $\geq 85\%$ ice; **ice**), chlorophyll
162 *a* concentration (near-surface summer climatology in mg m^{-3} ; **chl**) and the salinity differ-
163 ence between 600 and 200 m depths (based on winter climatology averaged over 1955-2012
164 in psu, **saldiff**). Sea ice and chl *a* data were obtained from the Australian Antarctic

165 Data Centre (Raymond, 2014). Salinity data were obtained from the World Ocean Atlas
166 (Zweng *et al.*, 2013). All three covariates were spatially interpolated to the same 0.1 x 0.1
167 degree grid covering the spatial domain of the 24 elephant seal tracks (Appendix S3.1).
168 The environmental data values were then extracted at each seal location from the SSM-
169 filtered track data. As `saldiff` could not be calculated in areas where the bathymetry was
170 shallower than 600 m, we did not include this variable in the models fit to the seals mak-
171 ing ice-bound foraging trips as several of them spent considerable time in waters shallower
172 than 600 m (Appendix S2.2). Similarly, `ice` was excluded from the models fit to seals
173 making pelagic foraging trips as they spent relatively little time in regions with sea-ice
174 cover.

175 R code for the model selection exercise is in Appendix S4.

176 Results

177 Time-varying move persistence (`mpm`)

178 The ice-bound seals all exhibited similar movement patterns (Fig. 2a), with high move
179 persistence on their outbound migrations and lower move persistence near the Antarctic
180 continent in areas of higher sea-ice coverage. Return migrations to Iles Kerguelen were
181 more variable, with some individuals travelling back in a persistent fashion and others tak-
182 ing meandering routes, possibly to forage en route. Pelagic foraging seals (Fig. 2b) mi-
183 grated approximately 2000 km either east or west of Iles Kerguelen in relatively persis-
184 tent fashion. Less persistent movements occurred at the distal ends of these migrations,
185 although seals travelling to the west of Iles Kerguelen had markedly less persistent and
186 slower movements, suggestive of more intense search and foraging, compared to those trav-
187 elling to the east (Fig. 2b).

188 The γ_t -derived behavioural index is comparable but not identical to the discrete be-
189 havioural states estimated from the `bsam` SSSM (Fig. S2.1). The γ_t index captured the

190 same changes in movement behaviour but the magnitudes of those changes generally were
191 smaller. Fitting the move persistence model, including the SSM filtering step, was almost
192 500 times faster than fitting the `bsam` SSSM (Appendix S2.1).

193 **Individual variability in move persistence - environment relationships (mpmm)**

194 *Sea-ice foragers.* The best supported model for elephant seals foraging in the sea-ice zone
195 included fixed and random coefficients for both the proportion of ice cover and chlorophyll
196 *a* concentration (Table 1). On average, seals had movements that became less persistent
197 or directed as sea-ice cover and chlorophyll *a* concentration increased (Fig. 3a,b). Among
198 individuals, the relationship with `ice` was consistently negative but the degree to which
199 move persistence declined differed markedly (Fig. 3a), whereas the relationship with `chl`
200 was highly variable with 4 individuals having strong negative relationships and the rest
201 weak to moderately positive relationships (Fig. 3b). Unsurprisingly, the `chl` fixed-effect
202 was not significant (Z-value = -1.04, p = 0.3). Using the fixed-effects from the best sup-
203 ported model, the spatial prediction of γ_t over the entire spatial domain implies that the
204 best foraging habitat generally lies south of 65° S (south of the black contour line, Fig.
205 3d).

206 *Pelagic foragers.* The best supported model for elephant seals foraging pelagically in-
207 cluded fixed and random coefficients for the salinity difference between 600 and 200 m
208 depths (`saldiff`, Table 2). On average, seals had movements that became strongly less
209 persistent as the salinity difference decreased (Fig. 3c). Among individuals, this relation-
210 ship was moderately variable with two individuals exhibiting relatively small changes in
211 move persistence over the full range of `saldiff` (Fig. 3c). The spatial prediction of γ_t
212 over the entire spatial domain implies that animals generally adopt a movement behaviour
213 indicative of search or forage south of 65° S (south of the black contour line, Fig. 3e) or
214 north in the vicinity of the Subantarctic Front (north of the black contour line, Fig. 3e).

215 Discussion

216 Animal telemetry data obtained at the level of individual animals poses a challenge to
217 scale from individual to population ecology. While correlative statistical analyses using
218 mixed effects models have been widely applied to behavioural datasets (e.g., marine ani-
219 mal diving and bird migration ecology analyses, Hassrick *et al.*, 2010; Mandel *et al.*, 2008),
220 individual variability currently is incorporated into process-based models of movement be-
221 haviour in a relatively limited way. This is partly due to the extra complexity required for
222 building random effects into a process-oriented approach (i.e., the temporal nature of the
223 data are taken into account explicitly) though primarily due to the significant computa-
224 tional overhead entailed. Our method uses TMB estimation for a process model describ-
225 ing animal movement behaviour in direct relation to environmental features. Our results
226 show this enables multiple fixed and random effects in movement-environment relation-
227 ships to be fit simply and efficiently. Taking advantage of TMB's speed and power, this
228 approach provides a feasible solution to analysing increasingly large and detailed telemetry
229 datasets, and for harnessing individual-to-population level information on animal move-
230 ment responses to environment.

231 Environmental responses

232 Our analyses revealed relatively consistent responses by individual animals to environmen-
233 tal variables we tested, however substantial individual variability was also a persistent fea-
234 ture of the telemetry data. Comparisons of model structures allowed these individual-level
235 effects to be directly evaluated. Those animals whose forage migrations went towards the
236 Antarctic continent showed low move persistence once in areas of higher sea ice coverage.
237 Some individuals also showed positive responses to elevated chlorophyll a concentrations,
238 targeting productive coastal polynya areas (Malpress *et al.*, 2017; Labrousse *et al.*, 2018);
239 however this was not a persistent response with many others foraging farther offshore in
240 the marginal ice zone (Labrousse *et al.*, 2015) where chlorophyll a concentrations are lower

241 (Appendix S3.1). For the pelagic foraging animals, our results indicated seals moved per-
242 sistently away from the region in which salty Circumpolar Deep Water was confined to
243 depths (i.e., where the salinity difference was highly positive). The majority then adopted
244 a lower move persistence in areas where the CDW shoaled (salinity difference closer to
245 zero, southern areas) with four animals targeting the vicinity of the Subantarctic Front
246 (salinity difference negative) where cold fresh Antarctic Intermediate Water subducts un-
247 der saline Subantarctic surface waters (northwestern areas, Appendix S3.1).

248 Substantial variability among individuals is a persistent feature reported from animal
249 telemetry data (e.g., Block *et al.*, 2011). Understanding this variability is essential for
250 scaling from data collected on individuals up to inferences of population-level processes
251 (Morales *et al.*, 2010) and for predicting future responses to a changing environment. For
252 example, within the Southern Ocean climatic changes are impacting the sea-ice extent and
253 duration, the location of major oceanic frontal features, and potentially the meridional
254 overturning circulation (whereby water masses sink and rise as governed by density gradi-
255 ents) with large-scale consequences for marine ecosystem structure, function and produc-
256 tivity (Constable *et al.*, 2014).

257 We used environmental climatologies to demonstrate our data application, however
258 for many analyses relevant covariates may be extracted from time-varying environmen-
259 tal fields. Many automated options for this exist via websites such as ZoaTrack (<http://www.zoatrack.org/>) managed by the Atlas of Living Australia or Xtractomatic (<http://coastwatch.pfel.noaa.gov/xtracto/>) managed by the US National Oceanic and
262 Atmospheric Administration. We also note here the need to incorporate location uncer-
263 tainty when sampling environmental covariates from spatially gridded remote-sensing data.
264 This can be done using multiple imputation methods as implemented in momentuHMM R
265 package (McClintock & Michelot, 2018), i.e., drawing realizations of the locations from the
266 uncertainty of the location-filtering SSM estimates.

267 **Individual variation**

268 Although the ultimate source of observed individual differences in behaviour - environ-
269 ment relationships is often unclear, two non-exclusive explanations seem likely. First, we
270 often use relatively few predictors and these may represent the proximate influences which
271 predators are actually responding to (i.e., prey density and/or distribution) indirectly or
272 imperfectly. This may inflate apparent individual differences in predator movement be-
273 haviour. Modelling more direct indices of prey availability, and/or reducing error within
274 covariates by accounting for location uncertainty as discussed above, may help to reduce
275 apparent variation among individuals.

276 Second, individual variation is likely a real feature of foraging ecology (Magurran, 1993),
277 where individual quality and personality (Dall *et al.*, 2004; Stamps, 2007) may confer real
278 differences in foraging behaviour with relatively little difference in fitness (Mangel & Stamps,
279 2001). For example, consistent boldness in foraging can generate important ecological
280 trade-offs, effecting increases in both growth and mortality rates (Stamps, 2007; Bergvall
281 *et al.*, 2011; Chapman *et al.*, 2011). Research into behavioural syndromes along axes, such
282 as boldness-shyness or proactiveness-reactiveness (Sih *et al.*, 2004), may provide insight
283 into the functional connection between individual behavioural traits and physiological con-
284 sequences (e.g. via metabolic rates, reproductive success or mortality rates), and hence the
285 evolutionary significance for ecological patterns and processes. Individual differences likely
286 represent yet another characteristic contributing to survival and resilience in a complex
287 and variable environment.

288 **Modelling approach and extensions**

289 Our model is composed of a linear mixed effects regression embedded within a correlated
290 random walk process model for animal movement behaviour. While the linear mixed ef-
291 fects approach allows flexible combinations of fixed and random effects, there is scope for
292 further enhancement. In many cases parametric, linear fixed effects may not adequately

293 capture the complexity of movement behaviour - environment relationships and a nonpara-
294 metric approach using penalised splines may yield improved inference (Langrock *et al.*,
295 2017). Our random effects currently use an unstructured covariance matrix that may be
296 less appropriate given the serial dependence structure typical of telemetry data. A first-
297 order autoregressive covariance structure may better account for this dependence (Pinheiro
298 & Bates, 2000). Finally, diagnosing lack of fit in latent variable models can be problematic
299 as there is no "observed response" variable. One-step-ahead prediction residuals provide a
300 useful model validation tool and can be estimated when fitting the model (Thygesen *et al.*,
301 2017).

302 This work addresses a key improvement in the quantitative integration of animal move-
303 ment behaviour and environment. Habitat models are presently the dominant method for
304 inference of environmental drivers of species' habitat preferences and space-use but largely
305 ignore the behavioural context underlying observed animal locations. By modelling ani-
306 mal movement behaviours as a mixed effects function of environmental variables, we gain
307 deeper insight into how individuals and populations actually use habitat. Additional ef-
308 fort is required to converge movement behaviour and habitat modelling approaches. For
309 example, our behavioural models do not account for availability/accessibility of habitat in
310 any way but this clearly must be considered when inferring habitat preferences (Wakefield
311 *et al.*, 2011). A reasonable approach for this might be to use the movement process param-
312 eters to simulate animal tracks and examine implications of including/excluding environ-
313 mental covariates. These pseudo-absence tracks may be used as the basis for developing a
314 habitat accessibility surface and generating spatial predictions of animal behaviour condi-
315 tional on this (e.g., Raymond *et al.*, 2015).

316 Our results show that TMB facilitates the fast estimation of multiple random effects
317 by using the Laplace approximation to calculate the marginal likelihood of a movement
318 behaviour process model. The model selection we conducted on the 24 southern elephant
319 seal tracks took a total of 8 minutes to complete. This includes the time required to SSM

320 filter the original Argos tracks and to fit the `mppmm`'s and is approximately 1500 times faster
321 than a more limited hierarchical Bayesian model selection exercise, using Markov chain
322 Monte Carlo simulation (Bestley *et al.*, 2013). The dramatically faster computation times
323 achieved by our TMB-enabled approach means that similar analyses of movement be-
324 haviour - environmental relationships can be scaled up to very large telemetry datasets.
325 This computation speed also opens up possibilities for far more realistic models of animal
326 movement, incorporating the third dimension for diving or flying animals and/or high-
327 volume accelerometry data.

328 The process model used here differs markedly from SSM used by Bestley *et al.* (2013).
329 They used discrete behavioural state Markov-switching (Patterson *et al.*, 2009; Langrock
330 *et al.*, 2012) embedded in a correlated random walk process model (Jonsen, 2016). Here,
331 we used a time-varying move persistence parameter γ_t as a behavioural index that varied
332 continuously between 0 and 1. This continuous behavioural index provides another tool for
333 characterising animal movement patterns and for making inferences about the possible en-
334 vironmental drivers of animal movement behaviour. In some cases, a continuous index may
335 offer more nuanced insight into variable behavioural sequences (Gurarie *et al.*, 2009; Breed
336 *et al.*, 2012), whereas a discrete state approach may offer more flexibility in capturing the
337 known structure of animal movement patterns (e.g., Michelot *et al.*, 2017).

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Table 1: Model rankings by Δ AIC and likelihood ratios (LR) for the MPMM’s fit to the 12 ice foraging seals. Absolute AIC and deviance values for the best ranked model are displayed on the first row, under the Δ AIC and LR headings. All other Δ AIC and LR values are relative to the best ranked model. Computation time to convergence is also reported. Random effects are included in parentheses in the model formulas, following the `lme4` convention (Bates *et al.*, 2015).

Model formula	df	Δ AIC	LR	Time (s)
\sim ice + chl + (ice + chl id)	12	-9954.21	-9978.21	4.76
\sim ice + chl + (chl id)	9	0.78	6.78	3.61
\sim ice + chl + (1 id)	7	21.06	31.06	4.17
\sim ice + (1 id)	6	21.08	33.08	2.63
\sim ice + chl + (ice id)	9	23.59	29.59	5.76
\sim ice + (ice id)	8	24.14	32.14	4.55
\sim chl + (chl id)	8	219.74	227.74	4.09
\sim chl + (1 id)	6	245.16	257.16	3.48
\sim 1 + (1 id)	5	339.28	353.28	2.79

Table 2: Model rankings by Δ AIC and likelihood ratios (LR) for the MPMM’s fit to the 12 ice foraging seals. Absolute AIC and deviance values for the best ranked model are displayed on the first row, under the Δ AIC and LR headings. All other Δ AIC and LR values are relative to the best ranked model. Computation time to convergence is also reported. Random effects are included in parentheses in the model formulas, following the lme4 convention (Bates *et al.*, 2015).

Model formula	df	Δ AIC	LR	Time (s)
\sim saldiff + (saldiff id)	8	-13897.26	-13913.26	3.87
\sim saldiff + chl + (saldiff id)	9	1.68	-0.32	4.96
\sim saldiff + chl + (chl id)	9	3.25	1.25	3.97
\sim saldiff + chl + (1 id)	7	29.81	31.81	4.04
\sim saldiff + (1 id)	6	36.35	40.35	3.21
\sim chl + (chl id)	8	51.37	51.37	4.54
\sim chl + (1 id)	6	107.41	111.41	4.19
\sim 1 + (1 id)	5	129.93	135.93	2.34
\sim saldiff + chl + (saldiff + chl id)	12	NA*	NA*	6.02

*model failed to converge

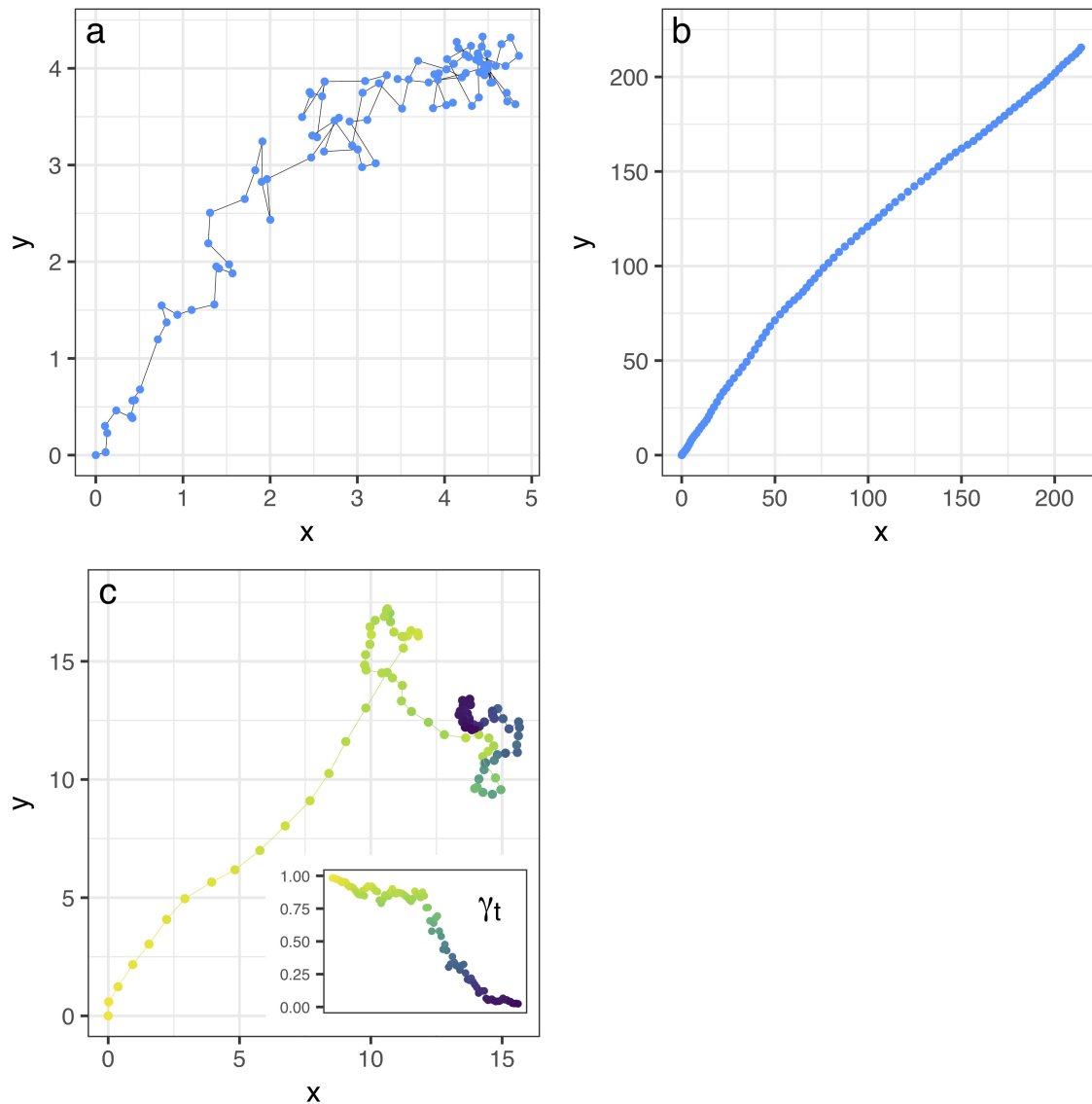


Figure 1: Example tracks simulated from the move persistence model with γ_t set to a constant 0.01 (low persistence) (a), γ_t set to a constant 0.99 (high persistence) and a time-varying γ_t (c). Locations in c are coloured by γ_t values with the random walk time-evolution of γ_t displayed inset in c. Note the substantially different scales of movement across panels a - c, despite sharing the same process covariance matrix (Σ). See Appendix S1 for simulation code.

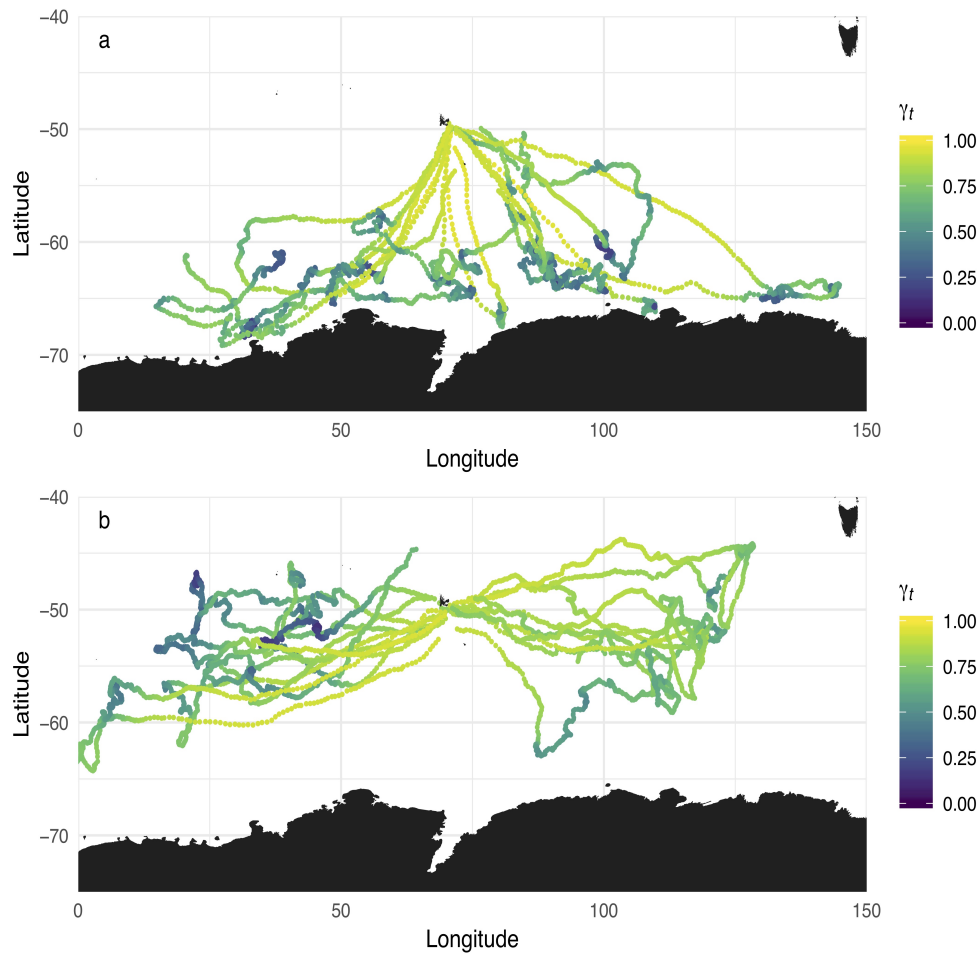


Figure 2: Maps of SSM-filtered southern elephant seal tracks originating from Iles Kerguelen. Ice-bound foraging trips (a) were predominantly directed to locations south of 60°S, whereas pelagic foraging trips (b) are predominantly north of 60°S. Each location is coloured according to its associated move persistence (see γ_t scale bar) estimated from the move persistence model.

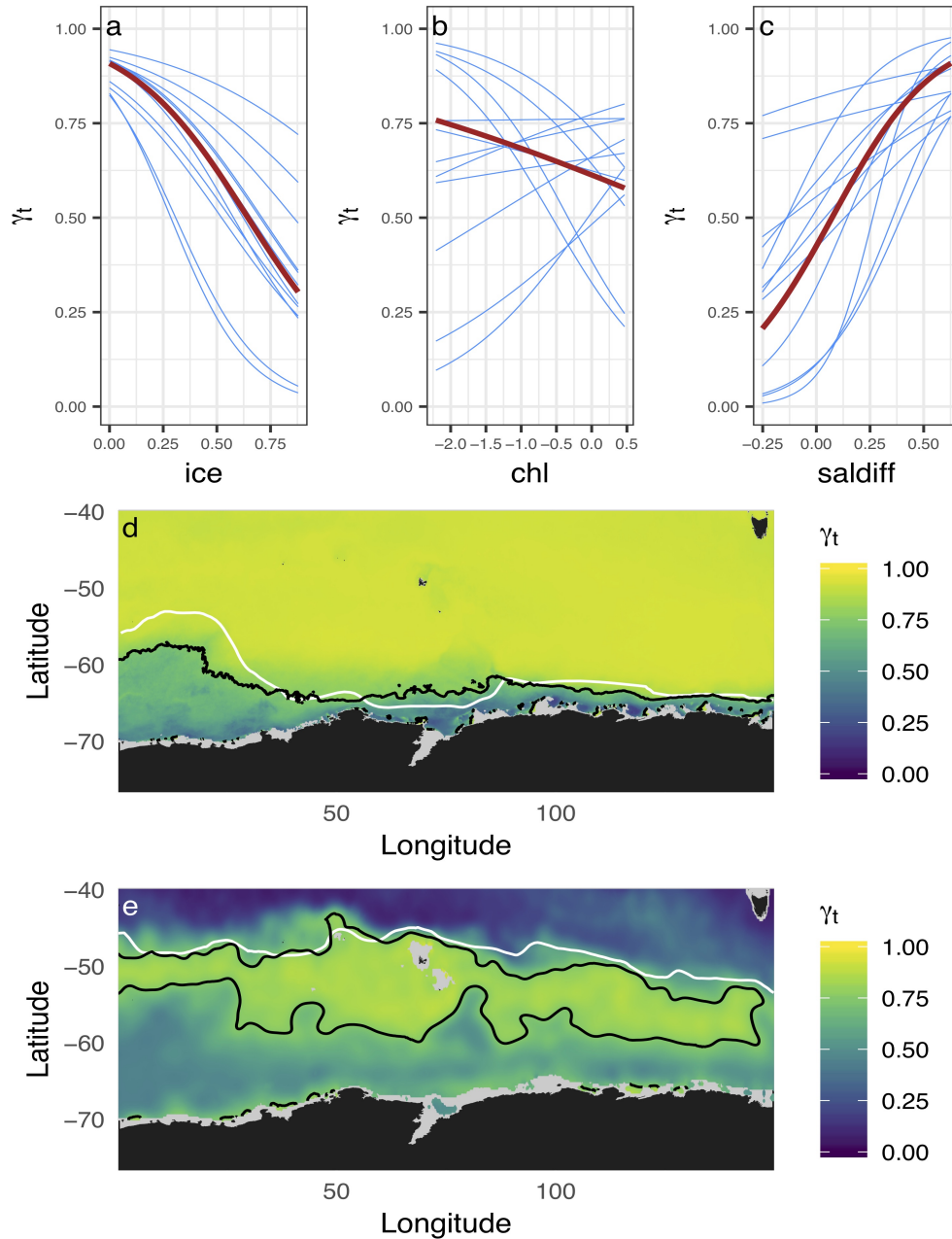


Figure 3: Fixed (red) and random (blue) effects relationships between move persistence γ_t and the proportion of ice cover (a) and chlorophyll *a* concentration (b) for ice foraging seals, and between γ_t and the salinity difference between 600 and 200m (c) for pelagic foraging seals. All three panels display both random intercept and slopes, as per the best ranked models in Tables 1 and 2. Spatial predictions of γ_t based on the fixed effect coefficients for the best fitting models for ice foraging seals (d) and pelagic foraging seals (e). The $\gamma_t = 0.75$ contour (black line) is displayed to aid delineation of predicted high move persistence ($\gamma_t > 0.75$; green - yellow) and low move persistence regions ($\gamma_t \leq 0.75$; green - blue). The southern boundary of the Antarctic Circumpolar Current (d) and the Subantarctic Front (e) are displayed for reference (white lines).