

Ecological Network assembly: how the regional meta web influence local food webs

Abstract

The idea that ecological networks are built in a sequence of colonization events is not new but has been applied mostly to competitive interactions. Similar processes act in trophic networks, i.e. food webs: a regional pool of species act as the source from which species colonize local areas, called the metaweb. Local food webs are realizations of metawebs that result from assembly processes influenced by migration, habitat filtering, stochastic factors, and dynamical constraints imposed by food-web structure. We analyse how the structure of a metaweb influence local food webs with different spatial scales, using an assembly model, a random model and properties at three levels: emergent global properties that take into account the whole network e.g. modularity, sub-structural properties that consider several nodes e.g. motifs, and properties related to one node e.g. topological roles. Three independent data-sets were included: the marine Antarctic metaweb (34.8 million Km²), the Weddell Sea (3.5 million Km²) and Potter Cove (6.8 Km²) food webs. Looking at the global properties, the metaweb presents a structure very different from the random model, while the local food webs follow the same pattern and are very similar to the assembly model. The assembly model only takes into account migration, local extinction and secondary extinctions. For sub-structural properties the metaweb and the local food webs also showed the same pattern against the random model, but we found differences compared to the assembly model that did not increase the local stability of food webs. Topological roles also showed differences between the metaweb and local food webs that were explained by the assembly model. We found that a great portion of the structure of the food webs is determined by evolutionary processes that act on large temporal and spatial scales. On the contrary, dynamical processes that favour stability have a small influence, but habitat filtering or dispersal limitations seem to be important factors that determine food web structure.

Leonardo A. Saravia^{1 2}, **Tomás I. Marina**^{1 2 3}, **Marleen De Troch**⁴, **Fernando R. Momo**^{1 2}

1. Instituto de Ciencias, Universidad Nacional de General Sarmiento, J.M. Gutierrez 1159 (1613), Los Polvorines, Buenos Aires, Argentina.
2. INEDES, Universidad Nacional de Luján, CC 221, 6700 Luján, Argentina.
3. Consejo Nacional de Investigaciones Científicas y Técnicas (CONICET)
4. Marine Biology, Ghent University, Krijgslaan 281/S8, B-9000, Ghent, Belgium.

Introduction

The characterization of ecological systems as networks of interacting elements has a long history (Paine 1966, May 1972, Cohen and Newman 1985). Much of this work has been devoted to investigate network structure and its effects on dynamics and stability (Thebault and Fontaine 2010). In recent years a renewed emphasis on structural stability (Rohr et al. 2014, Grilli et al. 2017) and new developments in random matrix theory has greatly expanded our capability to analyse ecological networks (Allesina et al. 2015). However, the analyses of the effects of ecological dynamical processes on food web structure are not so well developed. One such dynamic process is community assembly: how species from a regional pool colonize a site and build local interactions (Carstensen et al. 2013). Regional species pools are shaped by evolutionary and biogeographical processes that imply large spatial and temporal scales (Carstensen et al. 2013, Kortsch et al. 2018). More specifically, the assembly of local communities is influenced by dispersal, environmental filters, biotic interactions and stochastic events (HilleRisLambers et al. 2012). These processes have been studied by means of metacommunity theory (Leibold et al. 2017), where different spatial assemblages are connected through species dispersal.

Metacommunity theory provides a framework for assessing the roles of local- and regional-scale dynamics (Leibold et al. 2004, Baiser et al. 2013). This approach has been applied mostly to competitive interactions, whereas trophic interactions have received less attention (Baiser et al. 2016). Recently, there has been an increase in food web assembly studies, integrating them with island biogeography (Gravel et al. 2011) and with metacommunity (Pillai et al. 2011, Liao et al. 2016). These were mainly focused on complexity-stability effects (Mougi and Kondoh 2016). Previous attempts to study the food web assembly process have used motifs as building blocks of local communities (Baiser et al. 2016). Motifs are sub-structures in networks composed of species and links whose frequency deviates from the expected in a random network (Milo et al. 2002). By comparing motif representation at different spatial scales—from local to regional—the process of assembly of interactions may be revealed (Baiser et al. 2016), e.g. if the same processes structure the food web across scales, motif representation should be the same. Besides, as local food webs should have tighter links and stronger interaction rates, other structural properties should change as scale changes (Coll et al. 2011).

The objective of the present study is to analyse the process of food web assembly addressing how multilevel network properties change across different spatial scales. For this we considered: the Antarctic metaweb, representing an area of 34.8 million Km², and build from a dietary database (Raymond et al. 2011); the Weddell Sea food web that represents 3.5 million Km² (Jacob et al. 2011); and the Potter Cove network

that represents 6.8 Km², (Marina et al. 2018a).

There are several approaches to characterize the structure of ecological networks: emergent or global properties that focus on averages over the entire network (e.g. modularity or connectance), and more local properties that focus on how one or several species are related to the whole. Global properties related to resilience and stability are fundamental to understand the response of ecological networks to global threats like climate change and biological invasions. One of these properties is the small-world pattern, associated with rapid responses to disturbances and resistance to secondary extinctions (Montoya and Solé 2002). The small-world pattern is related to two global properties: the average of the shortest distance between all species, called the characteristic path length, and the clustering coefficient, the probability that two species linked to the same species are linked. Then, small-world patterns imply a high level of clustering and a short average path length compared to random networks (Watts and Strogatz 1998). This pattern comes from the general network theory, beyond ecological networks, and has been recently applied to marine food webs (Gray et al. 2016, Navia et al. 2016, Bornatowski et al. 2017, Marina et al. 2018b).

Since the early studies of May (1972) stating that larger and more connected ecosystems will be unstable, there has been a search for factors that would stabilize food webs (García-Callejas et al. 2018); one of the potential properties is trophic coherence (Johnson et al. 2014). Trophic coherence is based on the distances between the trophic positions of species and measures how well species fall into discrete trophic levels. Then, more coherence implies a more hierarchical food-web structure, which is directly correlated with linear stability (Johnson et al. 2014). The advantage of coherence as an index of stability is that it does not make any assumptions about interaction strengths. A property related to coherence is the mean trophic level, which historically has been used as an ecosystem health indicator (Pauly et al. 1998), predicting that food webs with higher trophic levels are less stable (Borrelli and Ginzburg 2014).

Food webs have structurally complex and highly non-random patterns that contain internal functional units or sub-modules (Grilli et al. 2016). These are groups of prey and predators that interact more strongly with each other than with species belonging to other modules. These modules (also called compartments) act as a buffer to the propagation of perturbations throughout the network, increasing its persistence (Stouffer and Bascompte 2011). It is interesting to mention that small-world patterns and modularity act in opposite directions. Whereas the small-world structure favours the spread of perturbations through its rapid dissipation (Gray et al. 2016), the presence of a high degree of modularity prevents the dispersal of perturbations (Krause et al. 2003, Stouffer and Bascompte 2011). The detection of these modules is an intermediate approach between the global and local analyses.

Species may participate in different ways with respect to modularity, depending on how many trophic links they have within their own module and/or between modules (Guimerà and Nunes Amaral 2005, Kortsch et al. 2015). This participation with respect to modularity is called a species' topological role. Theoretical and empirical results suggest these roles are related to species traits, such as wide niche breadth, environmental tolerance, apex position in local communities and high motility (Rezende et al. 2009, Guimerà et al. 2010, Borthagaray et al. 2014, Kortsch et al. 2015). This is a local property at an individual species level.

If we consider a subset of linked species inside the food web this forms a sub-network, when the abundance of one of these sub-networks deviates significantly from a null model network, this is called a motif (Milo et al. 2002). Besides this definition, in the ecological literature motif has been used as a synonym of sub-network. We analyse here the three-species sub-networks that have been most studied theoretically and empirically in food webs (Prill et al. 2005, Stouffer et al. 2007, Baiser et al. 2016). Specifically, we focused on four of the thirteen possible three-species sub-networks: apparent competition, exploitative competition, tri-trophic chain, and omnivory (Figure 1).

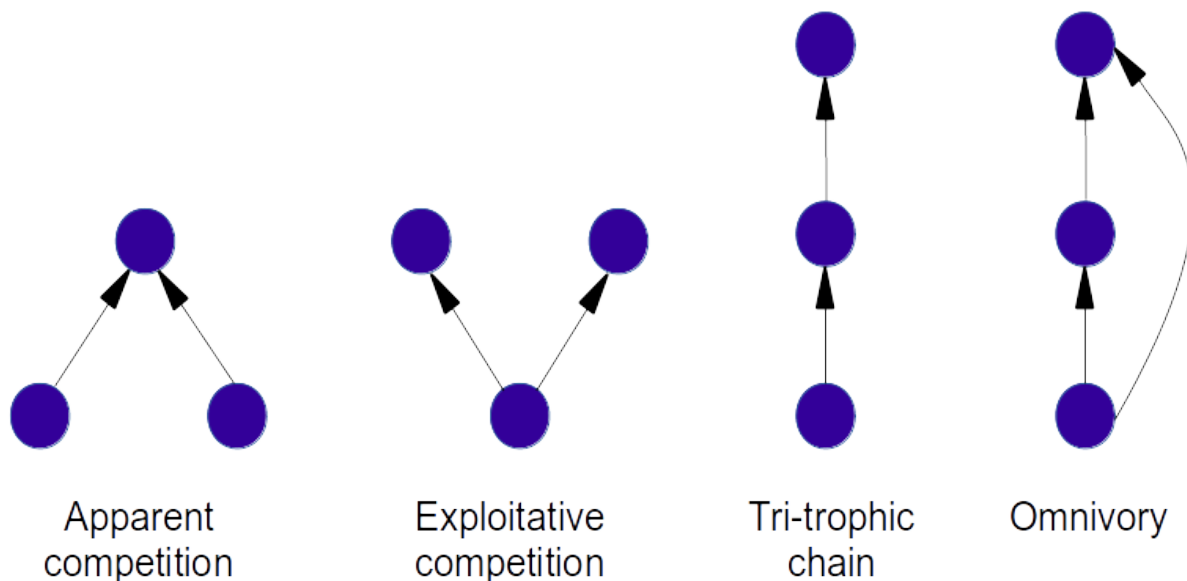


Figure 1: The four three-species sub-networks analysed: apparent competition, exploitative competition, tri-trophic chain, and omnivory. These four sub-networks have been explored both theoretically and empirically in ecological networks and are the most common sub-networks found in food webs

The frequency of motifs at different spatial scales (i.e over-represented, under-represented, or random) may reveal aspects about the process behind the assembly of interactions (Baiser et al. 2016). If the same processes structure the food web across scales, motif representation should be the same. Ecological interactions occur at the local scale so differences may show the importance of local interactions in the assembly of the food web.

During the assembly process those motif structures that are less dynamically stable tend to disappear from the food web (Borrelli 2015), this has been called non-adaptative systemic selection (Borrelli et al. 2015). The process is non-adaptative because the selection process is not adapting the system to local conditions (Borrelli et al. 2015).

In this study, we analyze food web assembly from a metaweb to local networks with different spatial scales. To our knowledge, combining modularity, trophic coherence and motifs at different spatial scales has not been applied in food web studies up to now. First we compared the networks—including the metaweb—against a random network model (i.e. absence of ecological mechanisms), and a metaweb assembly model (i.e. representing an assembly process). We hypothesize that network properties will change from the metaweb to a local scale at which interactions are realized. We particularly expect global properties related to resilience and stability to be close to the random null model at the metaweb scale and significantly different at the local scale; a greater frequency of stable motifs in the local food webs; as well as a change in the frequency of topological roles since habitat filtering or dispersal limitation act at the local food web scale. These last two changes also should be reflected as differences from the metaweb assembly model.

Methods

The three datasets used in this study encompass a wide range of spatial scales and were collected independently. The Southern Ocean database compiled by Raymond et al. (2011) was used to construct the Antarctic metaweb selecting only species located at latitudes higher than 60°S. Raymond et al. (2011) compiled information from direct sampling methods of dietary assessment, including gut, scat, and bolus content analysis, stomach flushing, and observed feeding. We considered that the metaweb is the regional pool of species defined by the biogeographic Antarctic region. Next we analysed two local food webs: the Weddell Sea food web dataset includes species situated between 74°S and 78°S with a West-East extension of approximately 450 km, and comprises all information about trophic interactions available for the zone since 1983 (Jacob et al. 2011). The Potter Cove dataset comes from a 4 km long and 2.5 km wide Antarctic fjord located at 62°14'S, 58°40'W, South Shetland Islands (Marina et al. 2018a). These food web datasets comprise benthic and pelagic habitats of the Antarctic ecosystem, few aggregated low-trophic level groups (e.g. detritus, diatoms, phytoplankton, zooplankton) and a high resolution of the macroalgae community (i.e. 24 biological species of red, brown and green macroalgae). The macroalgae community is responsible for the majority of the primary production and supports a large fraction of secondary production in Antarctic fjords (Quartino and Boraso de Zaixso 2008, Valdivia et al. (2015)). Higher trophic levels comprise:

invertebrate (e.g. ascidians, sponges, isopods, amphipods, bivalves, gastropods, cephalopods, echinoderms) and vertebrate predator groups (e.g. demersal and pelagic fishes, penguins, seals and whales). For more information about these datasets we refer to the original publications. To make datasets compatible, we first checked taxonomic names for synonyms, and second, we added species (either prey or predator) with their interactions to the metaweb when the local food webs contain a greater taxonomic resolution. When the metaweb contained more detailed records, only for predators we added them to the local food webs checking for its geographic range. We removed cannibalistic (self-links) and double arrows (i.d. A eats B and B eats A).

Analysis

We analysed the structure of the networks using properties that focus on different levels: emergent global properties that take into account the whole network, sub-structural properties that consider several nodes, and properties related to one node. To describe food webs as networks each species is represented as a node or vertex and the trophic interactions are represented as edges or links between the nodes. These links are directed, from the prey to the predator, as the flow of energy and matter. Two nodes are neighbours if they are connected by an edge and the degree k_i of node i is the number of neighbours it has. The food web can be represented by an adjacency matrix $A = (a_{ij})$ where $a_{ij} = 1$ if species j predaes species i , else is 0. Then $k_i^{in} = \sum_j a_{ji}$ is the number of preys of species i or its in-degree, and $k_i^{out} = \sum_j a_{ij}$ is the number of predators of i or its out-degree. The total number of edges is $E = \sum_{ij} a_{ij}$.

Null models

We considered two null models, the Erdős-Rényi random graph (Erdős and Rényi 1959), and the metaweb assembly model. An Erdős-Rényi network is constructed fixing the number of edges and nodes and assigning at random the m edges to the n nodes with equal probability (Erdős and Rényi 1959, Baiser et al. 2016). We restricted the random model eliminating double arrows and cannibalistic links. To calculate trophic level and the coherence parameter (see below) we further restricted to random webs with at least one basal node, to make these calculations possible. Since the random model represents the absence of any network assembly mechanism, the comparison against it does not guarantee information on this aspect.

In order to consider network assembly mechanisms we designed a dynamic metaweb assembly model. In this model species migrate from the metaweb to a local web with a uniform probability c , and become extinct from the local web with probability e ; a reminiscence of the theory of island biogeography (MacArthur and Wilson

1967, Gravel et al. 2011), but with the addition of network structure. Species migrate with their potential network links from the metaweb, then in the local web species can only survive if at least one of its preys is present, or if it is a basal species. When a species goes extinct locally it may produce secondary extinctions; we check that the local predators maintain at least one prey if not they become extinct independent of the probability e . We simulated this model in time and it eventually reaches an equilibrium that depends on the migration and extinction probabilities but also on the structure of the metaweb. The ratio of immigration vs. extinction $\alpha = c/e$ is hypothesized to be inversely related to the distance to the mainland (MacArthur and Wilson 1967), and as extinction e should be inversely proportional to population size (Hanski 1999), the ratio α is also hypothesized to be related to the local area.

For the random model we simulated networks with the same number of nodes n and edges m as the empirical networks; for the metaweb model we fitted the parameters g and e to obtain networks with n and m close to the empirical networks. This implies that α should reflect the differences in areas of the two local food webs. For details of the fitting and simulations see Appendix.

Small-world topology

The first global emergent property we used is the small-world pattern, which examines the average of the shortest distance between nodes and the clustering coefficient of the network (Watts and Strogatz 1998). We first calculated the characteristic path length that is the shortest path between any two nodes. Then L is the mean value of the shortest path length across all pairs of nodes. The clustering coefficient of node i was defined as

$$c_i = \frac{2E_i}{k_i(k_i - 1)}$$

where E_i is the number of edges between the neighbours of i . The clustering coefficient of the network is the average of c_i over all nodes. The original definition of small-world networks is conceptual (Watts and Strogatz 1998), a network G is small-world when it has a similar mean shortest path length but greater clustering than an Erdős-Rényi random network with the same number of nodes n and edges m . For the quantitative version of the small-world pattern we followed Humphries and Gurney (2008); we need to define:

$$\gamma_g = \frac{C_g}{C_{null}}$$

and

$$\lambda_g = \frac{L_g}{L_{null}}$$

where C_g and L_g are the clustering coefficient and the mean shortest path length of the network of interest G ; C_g and C_{null} are the same quantities for the corresponding null model. Thus, the quantitative small-world-ness is defined as:

$$S = \frac{\gamma_g}{\lambda_g}$$

and to determine if S is statistically significant Monte Carlo methods are used. We built 1000 null model networks with the same number of nodes n and edges m than the empirical network; then we calculated S for each random network and the lower and higher 99% quantiles of the S distribution are called ql, qh :

$$CI = \frac{qh - ql}{2}$$

the upper 99% confidence limit is then $CL^{0.01} = 1 + CI$. Thus, if a network has $S > CL^{0.01}$ it is considered a small-world network (Humphries and Gurney 2008). We also calculated the small-world-ness and the CI using the metaweb assembly model as a null model.

Coherence

The second global property is called trophic coherence (Johnson et al. 2014), and is related to stability in the sense that small perturbations could get amplified or vanished, which is called local linear stability (May 1972, Rohr et al. 2014). We first needed to estimate the trophic level of a node i , defined as the average trophic level of its preys plus 1. That is:

$$tp_i = 1 + \frac{1}{k_i^{in}} \sum_j a_{ij} tp_j$$

where $k_i^{in} = \sum_j a_{ji}$ is the number of preys of species i , basal species that do not have preys (then $k_i^{in} = 0$) are assigned a $tp = 1$. Then the trophic difference associated to each edge is defined as $x_{ij} = tp_i - tp_j$. The distribution of trophic differences, $p(x)$, has a mean $E^{-1} \sum_{ij} a_{ij} x_{ij} = 1$ by definition. Then the trophic coherence is measured by:

$$q = \sqrt{\frac{1}{E} \sum_{ij} a_{ij} x_{ij}^2 - 1}$$

that is the standard deviation of the distribution of all trophic distances. A food web is more coherent when q is closer to zero, thus the maximal coherence is achieved when $q = 0$, and corresponds to a layered network in which every node has an integer trophic level (Johnson et al. 2014, Johnson and Jones 2017). To compare coherence and trophic level we generated 1000 null model networks with at least one basal species and the same number of species and links—or approximately the same—than the network of interest. Then we calculated the 99% confidence interval using the 0.5% and 99.5% quantiles of the distribution of q ; we also calculated the confidence interval for the mean trophic level tp . We calculated the z-scores as:

$$z_i = \frac{q_{obs} - q_{null}}{\sigma_{qnull}}$$

where q_{obs} is the observed coherence, q_{null} is the mean coherence from the null model networks and σ_{qnull} is the standard deviation. The same formula is used for tp . The z-score thus measures the significance of deviations of the real network from the null hypothesis. If the distribution of the quantity (q , tp) under the null model is normal, a z-score greater than 2 is evidence that the observed quantity is significantly greater than its random counterpart, and a z-score less than 2 means that the quantity is significantly lower. If the distribution under the null model is skewed this is not necessarily true and thus we must rely on confidence intervals.

Modularity

An intermediate level property (i.e. between the whole network and the individual species) is modularity. It measures how strongly sub-groups of species interact between them compared with the strength of interaction with other sub-groups (Newman and Girvan 2004). These sub-groups are called compartments, and in order to find the best partition we used a stochastic algorithm based on simulated annealing (Reichardt and Bornholdt 2006). Simulated annealing allows to maximize modularity without getting trapped in local maxima configurations (Guimerà and Nunes Amaral 2005). The index of modularity was defined as:

$$M = \sum_s \left(\frac{I_s}{E} - \left(\frac{d_s}{2E} \right)^2 \right)$$

where s is the number of modules or compartments, I_s is the number of links between species in the module

s , d_s is the sum of degrees for all species in module s and E is the total number of links for the network. To assess the significance of our networks we calculate the 99% confidence intervals and z-scores based on 1000 null model networks as previously described.

Motifs

We considered four of the thirteen possible three-species sub-networks: apparent competition, exploitative competition, tri-trophic chain, and omnivory (Figure 1). These are the only motifs present in all networks analysed here. We compared the frequency of these motifs to 1000 null model networks using the 99% confidence interval and the z-score as previously described. To determine if the proportions of motifs change across networks we use the Pearson's Chi-squared test with simulated p-value based on 10000 Monte Carlo replicates.

Topological roles

As a local property that reflect the ecological role of each species we determined topological roles using the method of functional cartography (Guimerà and Nunes Amaral 2005), which is based on module membership (See modularity). The roles are characterized by two parameters: the standardized within-module degree dz and the among-module connectivity participation coefficient PC . The within-module degree is a z-score that measures how well a species is connected to other species within its own module:

$$dz_i = \frac{k_{is} - \bar{k}_s}{\sigma_{k_s}}$$

where k_{is} is the number of links of species i within its own module s , \bar{k}_s and σ_{k_s} are the average and standard deviation of k_{is} over all species in s . The participation coefficient PC estimates the distribution of the links of species i among modules; thus it can be defined as:

$$PC_i = 1 - \sum_s \frac{k_{is}}{k_i}$$

where k_i is the degree of species i (i.e. the number of links), k_{is} is the number of links of species i to species in module s . Due to the stochastic nature of the module detection algorithm we made repeated runs of the algorithm until there were no statistical differences between the distributions of PC_i and dz_i in successive repetitions; to test such statistical difference we used the k-sample Anderson-Darling test (Scholz and Stephens 1987). Then we calculated the mean and 95% confidence interval of dz and PC .

To determine each species' role the $dz - PC$ parameter space was divided into four areas, modified from Guimerà and Nunes Amaral (2005), using the same scheme as Kortsch et al. (2015). Two thresholds were used to define the species' roles: $PC = 0.625$ and $dz = 2.5$. If a species had at least 60% of links within its own module then $PC < 0.625$, and if it also had $dz \geq 2.5$, thus it was classified as a module hub. This parameter space defines species with relatively high number of links, the majority within its own module. If a species had $PC < 0.625$ and $dz < 2.5$, then it was called a peripheral or specialist; this refers to a species with relatively few links, mostly within its module. Species that had $PC \geq 0.625$ and $dz < 2.5$ were considered module connectors, since they have relatively few links, mostly between modules. Finally, if a species had $PC \geq 0.625$ and $dz \geq 2.5$, then it was classified as a super-generalist or hub-connector, because it has high between- and within-module connectivity. To test if the proportion of species' roles changed between networks we performed a Pearson's Chi-squared test with simulated p-value based on 10000 Monte Carlo replicates. Also, we tested if these proportions changed for one realization of the metaweb assembly model fitted for both local networks (i.e. Weddell Sea and Potter Cove food webs).

With the aim of giving an integrated visualization of the sub-structural and local properties of the food webs, we combined in a single plot information about compartments and the topological roles with the trophic level for each species.

All analyses and simulations were made in R version 3.4.3 (R Core Team 2017), using the igraph package version 1.1.2 (Csardi and Nepusz 2006) for motifs and topological roles analyses, and NetIndices (Kones et al. 2009) for trophic level calculations. Source code and data is available at <https://github.com/lisaravia/MetawebsAssembly>.

Results

Global network properties

The number of trophic species (size), links and connectance (Table 1), were in concordance with values found for marine food webs (Marina et al. 2018b). Based on the random null model, all networks presented the small-world topology as their small-world-ness index was larger than the 99% confidence interval (Table 1 & S1). However, we did not find differences between the local food webs and the assembly model (Table 1 & S2). Regarding trophic coherence, all networks presented negative random z-scores and significantly smaller q values (Table 1 & S1). thus they are more locally stable as they are more coherent. Using the metaweb assembly model, the Weddell Sea food web showed negative z-scores lower than 2, and Potter Cove food web exhibited no significant differences (Table 1 & S2). Mean trophic level results were similar among networks

and significantly lower than the random null model (Table 1 & S1), though were not significantly different when compared to the metaweb model trophic levels. Modularity values for the empirical food webs were greater than the random model, but not significantly higher in the Potter Cove web. No differences were found when compared to the metaweb assembly model (Table 1). Overall, networks differed from the random null model though presented similarities with the metaweb assembly model.

Table 1: Network global properties across scales. The *Metaweb* represents the marine predator-prey relationships of Antarctica, the *Weddell Sea* and *Potter Cove* are the local food webs. Z-scores were calculated against 1000 null model networks (random or metaweb assembly models). Quantities marked with '*' are significant at 1% level. A negative z-score means that the quantity is smaller than the expectation from null model simulations; a positive z-score means that is greater.

Network	Potter Cove	Weddell Sea	Metaweb
Size	91	437	859
Links	309	1908	9003
Area (Km ²)	6.8	3.5e6	34.8e6
Connectance	0.037	0.010	0.012
PathLength	1.81	2.20	2.57
Clustering	0.10	0.048	0.22
Small-World-ness random	*2.75	*4.69	*10.87
Small-World-ness Assembly	0.42	0.21	
Coherence	0.53	0.45	0.70
Coherence random z-score	*-0.54	*-2.08	*-3.54
Coherence assembly z-score	-0.20	*-3.41	
Mean Trophic level	2.13	1.98	1.91
Trophic level random z-score	*-0.27	*-0.86	*-1.60
Trophic level assembly z-score	1.29	1.09	
Modularity	0.37	0.48	0.45
Modularity random z-score	0.89	*18.97	*85.75
Modularity assembly z-score	-0.38	0.41	

Sub-structural properties (motifs)

The representation of three-species sub-networks with respect to the random model showed similar patterns in all networks (Figure 2A). While exploitative competition, apparent competition, and omnivory were over-represented, tri-trophic chains were under-represented; all these patterns were significant (Table S3). We found that motifs proportions for the three examined spatial scales were different (Chi-squared = 12612, p-value < 9.9e-05). this means that local and regional networks are not a random sample of the metaweb. With respect to the metaweb assembly model only some of them were significant (Table S4): tri-trophic chains and omnivory were under-represented for Weddell Sea, and apparent competition was over-represented for

Potter Cove (Figure 2B). Contrary to our expectations Potter Cove was more similar to the metaweb than Weddell Sea food web.

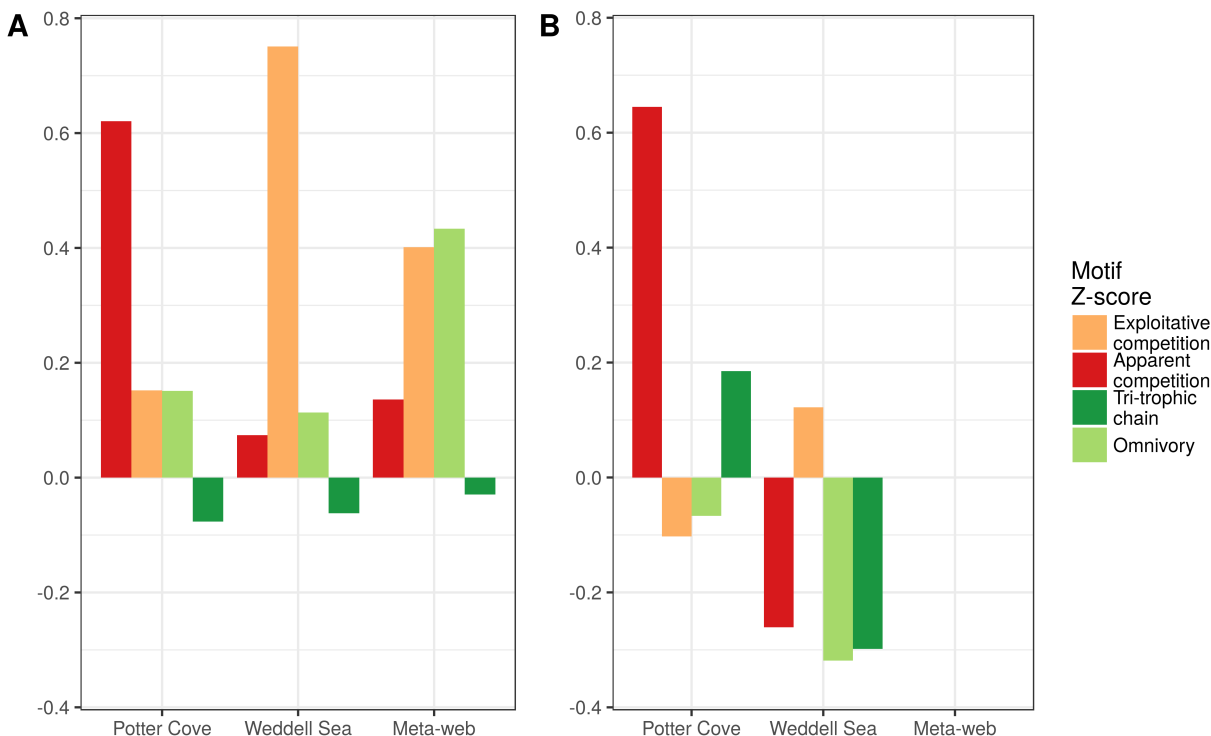


Figure 2: Network motifs z-scores across scales. Motifs are three-node sub-networks counted on each of the networks.: the *Metaweb* represents the marine predator-prey relationships of Antarctica (34.8 million Km²); the *Weddell Sea* (3.5 million Km²) and *Potter Cove* (6.8 Km²) are local food webs. A. Z-scores estimated with the random null model. B. Z-scores estimated with the metaweb assembly model. Z-scores were normalized by the square root of the sum of the squared z-scores for that food web.

Node-level properties (Topological roles)

The proportion of species displaying the four topological roles was different among food webs (Chi-squared = 79.31, p-value = 9.9e-05). A higher presence of module connectors (few links, mostly between modules) was observed in Weddell Sea, while a lack of module hubs (high number of links inside its module) was found in Potter Cove (Figure 3 A), which can be related to its low modularity value (Table 1). The proportions obtained with the metaweb assembly model were not different from the food webs (Chi-squared = 5.95, p-value = 0.41)(Figura 3 B).

The plot of topological roles combined with trophic levels and modularity revealed important details of the food webs (Figure 4): the metaweb has densely connected compartments but some of them have few low-connected species (module connectors or module specialists). Additionally, we observed in the Weddell Sea

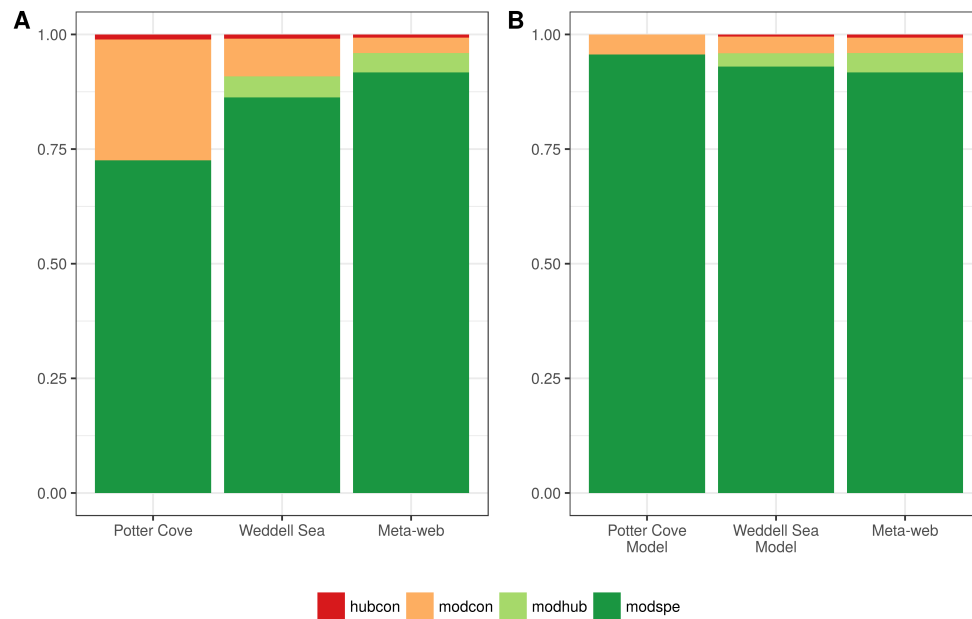


Figure 3: Proportion of topological roles across scales., the *Metaweb* represents the marine predator-prey relationships of Antarctica (34.8 million Km²); the *Weddell Sea* (3.5 million Km²) and *Potter Cove* (6.8 Km²) are local food webs. The topological roles are: *Hub connectors*, high number of between-module links; *Module connectors*, low number of links mostly between modules; *Module hubs*, high number of links within its module; *Module specialists*, low number of links within its module. A. Observed proportions for each food web; significant differences were found between them (Chi-squared = 79.31, p-value = 9.9e-05). B. Proportions for local networks obtained from the metaweb assembly model; no differences were found (Chi-squared = 5.95, p-value = 0.41)

food web hub connectors with a basal trophic level (Table S5). These are aggregated nodes that represent generic preys, e.g. fish or zooplankton, they only have incoming links or predators and they cannot have outgoing links or preys because they comprise several species. Different fish species are present in the Weddell Sea food web with detailed information about preys and predators, but for some predators there is insufficient knowledge of its preys and aggregated nodes must be added. Thus the existence of these basal hub connectors is a spurious result of aggregating prey species. The other non-aggregated hub connectors are highly mobile species with an intermediate trophic level like krill (Table S5). The variation of maximum trophic levels is evidenced in Figure 4, where both Potter Cove and Metaweb networks had similar values and Weddell Sea food web exhibited a lower maximum trophic level.

Discussion

We expected the metaweb structure to reflect the evolutionary constraints of the species interactions, and the local networks to be influenced and determined by the assembly processes and the local environment. Our results showed that the structure of the metaweb does not change significantly in many properties as the spatial scale changes. We did not find a clear pattern in the properties expected to be maximized by dynamical assembly and local stability, but we found clear differences in the properties influenced by habitat filtering and dispersal limitation. As a consequence, food webs would be mainly shaped by evolutionary forces and local environment drivers and less constrained by dynamical assembly processes.

Global level network properties showed a similar pattern across scales; most of them were significantly different from the random null model but not from the assembly model. Modularity for Potter Cove food web was the only property that is similar to the random model. Studies suggest that modularity enhances local stability (Stouffer and Bascompte 2011), even though stability strongly depends on the interaction strength configuration (Grilli et al. 2016) and on the existence of external perturbations (Gilarranz et al. 2017). We did not find modularity is greater than the metaweb assembly model, the existence of a modular structure could be related to different habitats—in marine environments could be benthic and pelagic (Zhao et al. 2017). Thus we observed that the modular structure is present in the metaweb and not produced by a maximization of local stability. Dynamical stability is expected to be maximized at the local level but we only found that Weddell Sea food web exhibited a greater trophic coherence than the assembly model. Thus, although this evidence is not conclusive with regard to the importance of dynamical processes in the assembly of food webs, the structure of the local food webs examined here seem to be a consequence of the metaweb structure.

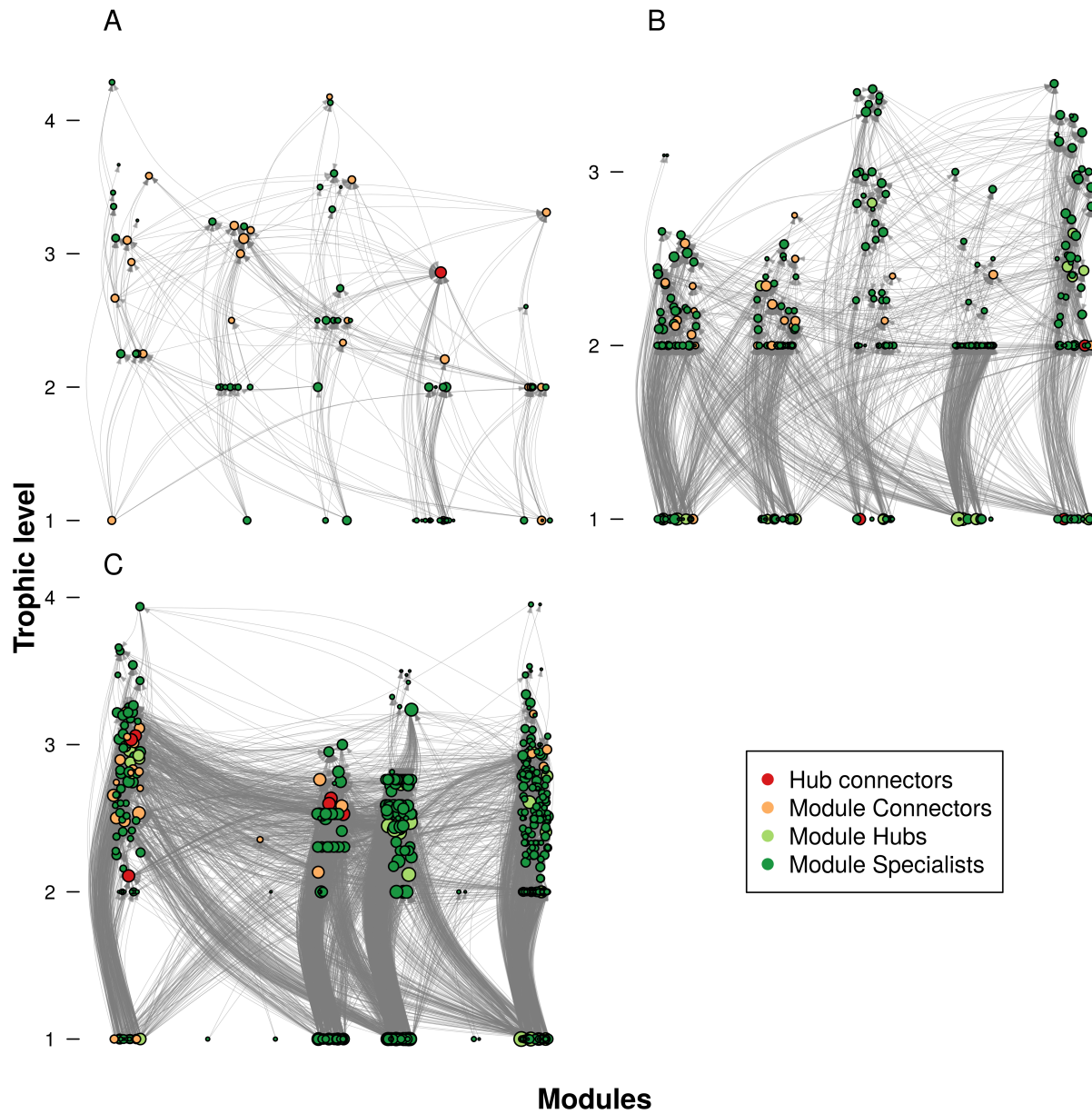


Figure 4: Plot of topological roles combined with trophic levels and modularity for each food web. The topological are: *Hub connectors* have a high number of between module links, *Module connectors* have a low number of links mostly between modules, *Module hubs* have a high number of links inside its module. *Module specialists* have a low number of links inside its module. Size of the nodes is proportional to the log of the species degree.

All networks have a significant value of small-world-ness compared with the random model but the local food webs are not different from the assembly model. In general, food webs do not show the small-world topology (Dunne et al. 2002, Marina et al. 2018b), which suggests that the small-world property is inherited from the metaweb and is less influenced by the greater percentage of realized interactions. If small-world-ness was determinant for an increased resilience and robustness to secondary extinctions (Bornatowski et al. 2017), local food webs should reflect significantly higher values than those obtained from the metaweb assembly models.

Motifs also have the same representation patterns against the random model. Local food webs should have motif patterns that increase its persistence. The expected pattern, based in dynamical models, is an over-representation of omnivory and tri-trophic chains and an under-representation of apparent and exploitative competition (Stouffer and Bascompte 2010). Other studies showed that tri-trophic chains, exploitative and apparent competition should be displayed more frequently (Borrelli 2015), but this was based on the stability of isolated three-species sub-networks and the assumption that during the assembly process these 3 species sub-networks were selected because of their intrinsic stability (Borrelli et al. 2015). The fact that tri-trophic (or n-trophic) interactions cannot account fully for the properties of the food webs (Cohen et al. 2009) make the results based on isolated modules less convincing. Our results showed an over-representation of omnivory, exploitative competition and apparent competition, this configuration is not the most stable one based on either criteria. A first hypothetical explanation is that the structure observed in the metaweb is maintained in local food webs. Despite the fact that almost all motifs were not different from the assembly model, neither tri-trophic chains and omnivory (lower in Weddell Sea) nor apparent competition (higher in Potter Cove) followed the pattern expected from theoretical studies. Thus, the assembly process is not random, there are differences in the frequencies of motifs as the scale change, but the selection of motifs due to its dynamical stability is not the main driver. This implies that other dynamical processes that influence the presence or absence of species like habitat filtering or dispersal limitation are acting and probably modifying motif frequencies in empirical food webs. This kind of structures that are a sub-product of process that happen at a different level have been called spandrels of assembly (Solé and Valverde 2006).

As expected all the networks have a short mean trophic level (Williams et al. 2002, Borrelli and Ginzburg 2014) compared with the random model. Different hypothesis were posed to explain this pattern: the low efficiency of energy transfer between trophic levels, predator size, predator behaviour, and consumer diversity (Young et al. 2013). These have contradictory support, reviewed by Ward and McCann (2017). Recently, it has been proposed that maximum trophic level could be related to productivity and ecosystem size depending on the context but related to energy fluxes that promote omnivory (Ward and McCann 2017). A different

mechanism based on dynamic stability of the whole web was proposed: food webs with shorter trophic levels (between 2 and 4) that have more omnivore chains are more likely to be stable, which increase the probability of being observed in nature [Borrelli2014]. We found that mean trophic level of the local food webs was not different from the assembly model, and omnivory was under-represented. This combination suggests that the trophic level could also be a spandrel of assembly, inherited from the metaweb structure.

Topological roles are useful to detect the existence of functional roles of species, like super-generalists (or hub connectors). These roles may change as the scale changes. A simple explanation is that modules also change. It was demonstrated in Arctic and Caribbean marine food webs that modules are usually associated with habitats (Rezende et al. 2009, Kortsch et al. 2015). For example, the Antarctic cod (*Notothenia coriiceps*) is a super-generalist for Potter Cove, and a module hub—a species with most of their links within its module—for the metaweb. This means that the same species can have different influences on the food web depending on the type or extension of the habitat considered. Although the metaweb assembly model showed no change in the frequency of topological roles, we found a change in topological roles with the scale. That means that in smaller areas there will be a smaller amount of different habitats, thus habitat filtering should be an important factor and will change the frequency of species that represent a particular topological role.

The spatial scales involved in our study do not represent a continuity; the metaweb and regional web have a 10 to 1 ratio but the local web is 10^6 smaller, besides that most of the global network properties and the motif structure showed no changes relative to the null model. Thus, we found evidence that local processes that limit species from the metaweb to local scales are influencing the assembly process but the structure of the food web originated through an emergent process from evolutionary or co-evolutionary forces seems to be more important. This could guide us to find a unified theory of ecological interactions that could describe the diversity of patterns observed in the natural world.

Acknowledgements

We are grateful to the National University of General Sarmiento for financial support. LAS would like to thank Susanne Kortsch that shared with us her source code for topological analysis and also made comments on an earlier version of the manuscript. This work was partially supported by a grant from CONICET (PIO 144-20140100035-CO).

References

- Allesina, S. et al. 2015. Predicting the stability of large structured food webs. - *Nature communications* 6: 7842.
- Baiser, B. et al. 2013. Predicting food-web structure with metacommunity models. - *Oikos* 122: 492–506.
- Baiser, B. et al. 2016. Motifs in the assembly of food web networks. - *Oikos* 125: 480–491.
- Bornatowski, H. et al. 2017. Topological redundancy and ‘small-world’ patterns in a food web in a subtropical ecosystem of Brazil. - *Marine Ecology* 38: e12407.
- Borrelli, J. J. 2015. Selection against instability: stable subgraphs are most frequent in empirical food webs. - *Oikos* 124: 1583–1588.
- Borrelli, J. J. and Ginzburg, L. R. 2014. Why there are so few trophic levels: Selection against instability explains the pattern. - *Food Webs* 1: 10–17.
- Borrelli, J. J. et al. 2015. Selection on stability across ecological scales. - *Trends in Ecology & Evolution* 30: 417–425.
- Borthagaray, A. I. et al. 2014. Inferring species roles in metacommunity structure from species co-occurrence networks. - *Proceedings of the Royal Society B: Biological Sciences* in press.
- Carstensen, D. W. et al. 2013. Introducing the biogeographic species pool. - *Ecography* 36: 1310–1318.
- Cohen, J. E. and Newman, C. M. 1985. When will a large complex system be stable? - *Journal of Theoretical Biology* 113: 153–156.
- Cohen, J. E. et al. 2009. Food webs are more than the sum of their tritrophic parts. - *Proceedings of the National Academy of Sciences* 106: 22335 LP–22340.
- Coll, M. et al. 2011. Food-Web Structure of Seagrass Communities across Different Spatial Scales and Human Impacts. - *PLOS ONE* 6: e22591.
- Csardi, G. and Nepusz, T. 2006. The igraph software package for complex network research. - *InterJournal Complex Sy*: 1695.
- Dunne, J. A. et al. 2002. Food-web structure and network theory: The role of connectance and size. -

Proceedings of the National Academy of Sciences 99: 12917–12922.

Erdős, P. and Rényi, A. 1959. On random graphs. - *Publicationes Mathematicae Debrecen* 6: 290–297.

García-Callejas, D. et al. 2018. Multiple interactions networks: towards more realistic descriptions of the web of life. - *Oikos* 127: 5–22.

Gilarranz, L. J. et al. 2017. Effects of network modularity on the spread of perturbation impact in experimental metapopulations. - *Science* 357: 199 LP–201.

Gravel, D. et al. 2011. Trophic theory of island biogeography. - *Ecology Letters* 14: 1010–1016.

Gray, C. et al. 2016. Chapter Ten - Recovery and Nonrecovery of Freshwater Food Webs from the Effects of Acidification. - In: Dumbrell, A. J. et al. (eds), *Large-scale ecology: Model systems to global perspectives*. Academic Press, ppp. 475–534.

Grilli, J. et al. 2016. Modularity and stability in ecological communities. - *Nature Communications* 7: 12031.

Grilli, J. et al. 2017. Feasibility and coexistence of large ecological communities. - *Nature Communications* 8: 14389.

Guimerà, R. and Nunes Amaral, L. A. 2005. Functional cartography of complex metabolic networks. - *Nature* 433: 895–900.

Guimerà, R. et al. 2010. Origin of compartmentalization in food webs. - *Ecology* 91: 2941–2951.

Hanski, I. 1999. *Metapopulation Ecology*. - OUP Oxford.

HilleRisLambers, J. et al. 2012. Rethinking Community Assembly through the Lens of Coexistence Theory. - *Annual Review of Ecology, Evolution, and Systematics* 43: 227–248.

Humphries, M. D. and Gurney, K. 2008. Network ‘Small-World-Ness’: A Quantitative Method for Determining Canonical Network Equivalence. - *PLOS ONE* 3: e0002051.

Jacob, U. et al. 2011. The role of body size in complex food webs: A cold case. - In: Research, A. B. B. T. A. in E. (ed), *Advances in ecological research*. Elsevier B. V., ppp. 181–223.

Johnson, S. and Jones, N. S. 2017. Looplessness in networks is linked to trophic coherence. - *Proceedings of the National Academy of Sciences* 114: 5618–5623.

Johnson, S. et al. 2014. Trophic coherence determines food-web stability. - *Proceedings of the National*

Academy of Sciences 111: 17923–17928.

Kones, J. K. et al. 2009. Are network indices robust indicators of food web functioning? a Monte Carlo approach. - *Ecological Modelling* 220: 370–382.

Kortsch, S. et al. 2015. Climate change alters the structure of arctic marine food webs due to poleward shifts of boreal generalists. - *Proceedings of the Royal Society B: Biological Sciences* in press.

Kortsch, S. et al. 2018. Food-web structure varies along environmental gradients in a high-latitude marine ecosystem. - *Ecography* in press.

Krause, A. E. et al. 2003. Compartments revealed in food-web structure. - *Nature* 426: 282.

Leibold, M. A. et al. 2004. The metacommunity concept: a framework for multi-scale community ecology. - *Ecology Letters* 7: 601–613.

Leibold, M. A. et al. 2017. Community assembly and the functioning of ecosystems: how metacommunity processes alter ecosystems attributes. - *Ecology* 98: 909–919.

Liao, J. et al. 2016. An extended patch-dynamic framework for food chains in fragmented landscapes. 6: 33100.

MacArthur, R. H. and Wilson, E. O. 1967. *The theory of island biogeography*. - Princeton University Press.

Marina, T. I. et al. 2018a. The Food Web of Potter Cove (Antarctica): complexity, structure and function. - *Estuarine, Coastal and Shelf Science* 200: 141–151.

Marina, T. I. et al. 2018b. Architecture of marine food webs: To be or not be a ‘small-world’. - *PLoS ONE* 13: 1–13.

May, R. M. 1972. Will a Large Complex System be Stable? - *Nature* 238: 413–414.

Milo, R. et al. 2002. Network Motifs: Simple Building Blocks of Complex Networks. - *Science* 298: 824 LP–827.

Montoya, J. M. and Solé, R. V. 2002. Small world patterns in food webs. - *Journal of Theoretical Biology* 214: 405–12.

Mougi, A. and Kondoh, M. 2016. Food-web complexity, meta-community complexity and community stability. - *Scientific Reports* 6: 24478.

Navia, A. F. et al. 2016. The structure of a marine tropical food web, and its implications for ecosystem-

based fisheries management. - *Ecological Modelling* 328: 23–33.

Newman, M. E. J. and Girvan, M. 2004. Finding and evaluating community structure in networks. - *Phys. Rev. E* 69: 26113.

Paine, R. T. 1966. Food Web Complexity and Species Diversity. - *The American Naturalist* 100: 65–75.

Pauly, D. et al. 1998. Fishing Down Marine Food Webs. - *Science* 279: 860 LP–863.

Pillai, P. et al. 2011. Metacommunity theory explains the emergence of food web complexity. - *Proceedings of the National Academy of Sciences* 108: 19293–19298.

Prill, R. J. et al. 2005. Dynamic Properties of Network Motifs Contribute to Biological Network Organization. - *PLOS Biology* 3: e343.

Quartino, M. L. and Boraso de Zaixso, A. L. 2008. Summer macroalgal biomass in Potter Cove, South Shetland Islands, Antarctica: its production and flux to the ecosystem. - *Polar Biology* 31: 281–294.

R Core Team 2017. R: A Language and Environment for Statistical Computing.

Raymond, B. et al. 2011. A Southern Ocean dietary database. - *Ecology* 92: 1188.

Reichardt, J. and Bornholdt, S. 2006. Statistical mechanics of community detection. - *Phys. Rev. E* 74: 16110.

Rezende, E. L. et al. 2009. Compartments in a marine food web associated with phylogeny, body mass, and habitat structure. - *Ecology Letters* 12: 779–788.

Rohr, R. P. et al. 2014. On the structural stability of mutualistic systems. - *Science* 345: 1253497.

Scholz, F. W. and Stephens, M. A. 1987. K-Sample Anderson–Darling Tests. - *Journal of the American Statistical Association* 82: 918–924.

Solé, R. V. and Valverde, S. 2006. Are network motifs the spandrels of cellular complexity? - *Trends in Ecology & Evolution* 21: 419–422.

Stouffer, D. B. and Bascompte, J. 2010. Understanding food-web persistence from local to global scales. - *Ecology Letters* 13: 154–161.

Stouffer, D. B. and Bascompte, J. 2011. Compartmentalization increases food-web persistence. - *Proceedings of the National Academy of Sciences of the United States of America* 108: 3648–52.

Stouffer, D. B. et al. 2007. Evidence for the existence of a robust pattern of prey selection in food webs. -

Proceedings of the Royal Society B: Biological Sciences 274: 1931 LP–1940.

Thebault, E. and Fontaine, C. 2010. Stability of Ecological Communities and the Architecture of Mutualistic and Trophic Networks. - *Science* 329: 853–856.

Valdivia, N. et al. 2015. Consistent richness-biomass relationship across environmental gradients in a marine macroalgal-dominated subtidal community on the western antarctic peninsula. - *PLoS ONE* 10: 1–20.

Ward, C. L. and McCann, K. S. 2017. A mechanistic theory for aquatic food chain length. - *Nature Communications* 8: 2028.

Watts, D. J. and Strogatz, S. H. 1998. Collective dynamics of 'small-world' networks. - *Nature* 393: 440–442.

Williams, R. J. et al. 2002. Two degrees of separation in complex food webs. - *Proceedings of the National Academy of Sciences* 99: 12913–12916.

Young, H. S. et al. 2013. The roles of productivity and ecosystem size in determining food chain length in tropical terrestrial ecosystems. - *Ecology* 94: 692–701.

Zhao, L. et al. 2017. Identifying compartments in ecological networks based on energy channels. - *Ecology and Evolution* 8: 309–318.